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# Structure, food and shade attract juvenile coral reef fish to mangrove and seagrass habitats: a field experiment

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ABSTRACT: Mangroves and seagrass beds are considered nurseries for juvenile fish, but little experimental evidence exists to elucidate which factors make them attractive habitats. A multifactorial field experiment on the use of these habitats by juvenile reef fish and their behaviour was performed during daytime with experimental units (EUs:  $1 \times 1 \times 0.8$  m), each representing a unique combination of the factors structure, shade, and food, using artificial seagrass leaves (AS) and artificial mangrove roots (AM). Diurnally active herbivores were most abundant in EUs containing food, and grazed on algae growing on the structures, but were not attracted to structures in the absence of food. The most abundant diurnally active zoobenthivores (Eucinostomus spp.) were present in highest numbers in any EU with food, where they fed on zoobenthos or rested on the bottom. The nocturnally active zoobenthivore/zooplanktivore Ocyurus chrysurus and the diurnally active piscivore Sphyraena barracuda were primarily attracted to structure, in which they rested and were not observed to feed. Haemulon flavolineatum was mainly attracted to AS, Lutjanus mahogoni was attracted to AS or shade, whereas L. apodus, O. chrysurus and S. barracuda were found in AM as well as in AS. The data suggest that during daytime, herbivores and diurnally active zoobenthivores are probably attracted to mangroves and seagrass beds primarily by food, and nocturnally active zoobenthivores by structure (in interaction with shade) that offers shelter from predation. S. barracuda is also attracted primarily to structure, but the larger individuals probably use this for ambush predation rather than for protection. In conclusion, our experiment clarifies that presence of structure, food and shade significantly contribute to the attractiveness of mangroves and seagrass beds to juvenile reef fish.

KEY WORDS: Artificial mangrove roots  $\cdot$  Artificial seagrass  $\cdot$  Coral reef fish  $\cdot$  Shelter habitat  $\cdot$  Feeding habitat  $\cdot$  Habitat utilisation  $\cdot$  Behaviour  $\cdot$  Caribbean

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# INTRODUCTION

Mangroves and seagrass beds are assumed to function as nursery habitats for a variety of reef fish species (Parrish 1989, Jackson et al. 2001, Heck et al. 2003), though direct evidence of connectivity between juvenile and adult habitats has seldom been provided (Beck et al. 2001, Gillanders et al. 2003, Chittaro et al. 2004). Although mangroves and seagrass beds harbour high numbers of various juvenile reef fish, little is known about the factors determining this. It has been suggested that the main reasons why mangroves and seagrass beds may be attractive to juvenile fish are low predator abundance, presence of structure, high food availability, low predation efficiency of predators, and a high interception of planktonic fish larvae due to the

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extensive surface area of mangroves and seagrass beds (Orth et al. 1984, Pollard 1984, Schulman 1985, Parrish 1989, Blaber 2000, Laegdsgaard & Johnson 2001).

The frequency of encounter rates with piscivores is higher on the coral reef than at 20 m away from the reef (Schulman 1985); hence shallow water habitats, such as mangroves and seagrasses, are believed to contain less piscivores than the reef (Blaber & Blaber 1980, Parrish 1989, Ronnback et al. 1999). However, Sheaves (2001) argued that there is little unequivocal evidence for this theory. Structure provides fish with shelter, but also with food (Laegdsgaard & Johnson 2001). Food is abundant in mangroves and seagrass beds and is present in the form of invertebrates attracted to structure or hiding in the sediment, seagrass leaves, algae, and epibionts associated with seagrass leaves and mangrove roots (e.g. Orth et al. 1984, Orth & van Montfrans 1984, Pollard 1984, Howard et al. 1989, Parrish 1989, Ley et al. 1994, Tomascik et al. 1997, Valentine & Heck 1999, Nagelkerken et al. 2000). Fish are also attracted to mangroves by the presence of decomposing mangrove leaves, which possibly function as a food source to detritivores (Odum & Heald 1975, Rajendran & Kathiresan 1999). Predation efficiency on fishes and shrimps is negatively affected by the structural complexity of seagrass beds (e.g. Heck & Thoman 1981, Orth et al. 1984, Rooker et al. 1998, Hindell et al. 2000, Stunz & Minello 2001) and mangroves (Primavera 1997, Laegdsgaard & Johnson 2001). The mere presence of structure does not necessarily provide protection; the density of structural units and species-specific behaviour of both predator and prey should be taken into account (Main 1987). For example, presence of mangrove pneumatophores did not decrease predation efficiency of an actively hunting fish predator on shrimps, whereas it did reduce predation on shrimps by a more passive fish predator (Primavera 1997). Water turbidity can also negatively affect predator efficiency due to scattering and reduction of light by suspended particles (Blaber & Blaber 1980, Benfield & Minello 1996).

A factor that has seldom been investigated, and may be partly responsible for the attraction of fish to mangroves and seagrass beds, is shade. Shade is provided by long seagrass leaves and, more importantly, by an extensive roof of branches and leaves of mangroves above the water surface. Shade is believed to reduce predation risk (Helfman 1981, Cocheret de la Morinière et al. 2004). For example, predation by the ambush predator *Hippocampus erectus* on shrimps in seagrass beds decreased under shaded conditions (James & Heck 1994), and visual prey detection by piscivorous salmonids of their prey decreased with decreasing light levels (Mazur & Beauchamp 2003). In experimental pools, bluegills *Lepomis macrochirus* used shaded areas when a predator was in open unshaded waters, but avoided shade when the predator was in the shade (McCartt et al. 1997). In artificial mangrove units, juvenile reef fish preferred shaded over unshaded units (Cocheret de la Morinière et al. 2004).

Few experiments have studied which factors attract fish to mangroves and seagrass beds, and a multifactorial design has seldom been applied. Such an experimental study on several combined factors can unravel effects of single factors, their interactions, and a possible hierarchy in their attractiveness. So far, only 2 field studies have partially attempted multifactorial investigation by using artificial mangrove roots: Laegdsgaard & Johnson (2001) studied the interactive effects between food and shelter, while Cocheret de la Morinière et al. (2004) focused on structure and shade. A third multifactorial study used living mangrove roots in the laboratory and tested the factors of turbidity, prey density, substrate type and pneumatophore density on the predation of thorn fish Terapon jarbua on shrimp (Macia et al. 2003). Multifactorial experiments in (artificial) seagrass beds on factors attracting fish have been conducted more often (e.g. Bell & Westoby 1986, Edgar 1999, Levin et al. 1997, Levin & Hay 2003).

The majority of field studies mentioned above have investigated only 1 or 2 factors in 1 experimental design; the attractiveness of mangroves and seagrass beds has not yet been studied simultaneously. In the present study, we applied a multifactorial design to study the interactive effects of 4 different factors in a single experiment: structure, shade, food and habitat type (artificial mangrove roots and artificial seagrass leaves). Because little is known of the behaviour of fishes within such artificial habitats, we studied this as well. Here, we tested the hypothesis that fish species of different feeding guilds probably make use of mangroves and seagrass beds during daytime for different reasons. We expected that nocturnally active fish species use mangroves and seagrass beds as shelter habitats during daytime, whereas diurnally active fish species use these habitats as feeding grounds.

# MATERIALS AND METHODS

**Study area.** The study was carried out in December 2003, in the Spanish Water Bay (total surface area approx. 3 km<sup>2</sup>) in south-western Curaçao, Netherlands Antilles (Fig. 1). The bay has a narrow entrance (70 m wide) and a 1.1 km long and 11 to 18 m deep channel that connects the bay to the sea and a fringing coral reef. The main part of the bay is relatively shallow (<6 m deep). The mean daily tidal range is about 30 cm (de Haan & Zaneveld 1959). Besides rainwater, the bay has no freshwater input. Seagrass beds and (macro-) algal



Fig. 1. Map of the Spanish Water Bay at Curaçao ('C') indicating the location of the study site (encircled)

flats cover the bottom of the bay. The shoreline consists partly of fossil coral reef and is fringed by mangroves, dominated by the red mangrove *Rhizophora mangle*. The experiment was carried out near an isolated mangrove stand adjacent to a seagrass bed (Fig. 1). The mangrove stand was approximately 36 m long and 3 m wide and consisted solely of *R. mangle*. The mangroves were separated from the adjacent seagrass bed *Thalassia testudinum* by a strip of muddy substratum approximately 3 m wide. Mean (±SD) seagrass density of this seagrass bed was 289.0 ± 222.1 m<sup>-2</sup>, and mean leaf length was 19.7 ± 1.0 cm. Mean water clarity at the study site was 7.0 ± 1.1 m, as measured by horizontal Secchi disk distance. Mean water depth at the outer mangrove border was 97.6 ± 6.6 cm.

**Experimental design.** Experimental units (EUs) were constructed with iron rods (1 m long, 6 mm in diameter, see Cocheret de la Morinière et al. 2004). The lower vertical 20 cm of the iron rods protruded from the bottom of the EU and could be pushed into the sediment (Fig. 2a). Above the sediment, the open EUs were

1.0 m wide, 1.0 m long and 0.8 m high. In total, 16 EUs were used (Fig. 2b). Each EU represented a unique combination of the factors structural complexity (artificial mangrove roots = AM, artificial seagrass leaves = AS, both AM and AS, or neither), shade (present or absent), and food in the form of fouling algae and access to zoobenthos (present or absent).

PVC pipes (diameter 1.5 cm, length 40 cm) mimicked mangrove proproots. A density of pipes of  $64 \text{ m}^{-2}$  was chosen because in a previous study, highest fish abundances were found in EUs with this level of structural complexity (Cocheret de la Morinière et al. 2004). The PVC pipes were evenly distributed in an EU by suspending them from nylon twine lines which were tied to the upper horizontal iron ribs (Fig. 2a). In this way, they mimicked aerial mangrove prop-roots hanging in the water column, as is the case for mangroves in Spanish Water Bay.

Artificial plastic *Thalassia testudinum* leaves of width 9 mm and length 19 cm (Pangea Rocks APS) were evenly distributed and attached to an iron grid (density 200 leaves  $m^{-2}$ ) on the bottom of an EU by using copper staples. Leaf lengths and densities of artificial seagrass were similar to those in the natural seagrass bed (see above). Up-

ston & Booth (2003) showed that fish assemblages in natural and artificial seagrass beds are comparable.

Shade was created by covering the top of the EU with black shading fabric (polyethylene monofilament). The shading cloth was always submerged. Underwater light intensities were measured at 10:00, 12:30 and 14:00 h by a snorkelling observer holding a Li-Cor light meter at arm length depth. This way, light intensities were measured at approximately 30 cm above the substratum under natural conditions (above the muddy substratum adjacent to the mangroves in full sunlight, at the mangrove rim, and 0.5 m inside the mangroves) and in the centre of EUs with different types of structure and/or shade in the top part (empty EUs, shade only, AM only, AM and shade). Because algae growing on the shading cloth probably decreased light intensity, the shading cloth of all EUs with the treatment 'shade' was cleaned every day.

Presence of food for herbivores was created in the EUs by natural growth of algae on the PVC pipes, the ribs of the EUs, and the artificial seagrass leaves. The EUs 'with

Fig. 2. (a) Schematic drawing of an experimental unit (EU) with artificial mangrove roots, artificial seagrass leaves, and shade. (b) Schematic representation of the 16 treatments. Vertical bars at top of EU represent artificial mangrove roots and vertical bars at bottom of EU represent artificial seagrass leaves. Horizontal grey bars above EUs represent presence of shade. Food was either absent (EUs 1 to 8) or present in the form of fouling algae and access to zoobenthos in the sediment (EUs 9 to 16). (c) Placement of EUs in the field

food' were placed in the field for a minimum of 20 d and a maximum of 32 d before the actual start of the experiment, during which algae had time to grow. The presence of the fouling algae on the EUs attracted small crustaceans. The crustaceans on the EU structures and in the sediment on the bottom of the EUs (which was freely accessible in the EUs 'with food', see next paragraph) served as a food source for zoobenthivores. For all EUs 'without food', algal growth was removed from the EUs every 2 d during the experiment. EUs were always cleaned at the end of the day after the last observation. Moreover, the bottom of all EUs 'without food' was covered by a cotton sheet, so that zoobenthivores were not able to forage in the sediment. It should be noted that in all EUs, presence/absence of food only refers to presence/absence of algae and macrozoobenthos, and not to that of prey fish in the case of piscivores.

In the field, the EUs were lined up in a randomised order, on a muddy substratum bordered by the mangrove stand on one side and the seagrass bed on the other side (Fig. 2c). The muddy substratum did not contain any natural submerged aquatic vegetation which could add additional structure to the bottom of the EUs. A distance of 1 m was maintained between an EU and the mangrove and seagrass border, and between EUs. Before observations started, EUs were left undisturbed for 3 d after placement in the field.

Observations. Observations took place at 3 time intervals during daytime: from 10:00 to 10:45 h (T1), from 11:30 to 12:15 h (T2), and from 13:00 to 13:45 h (T3). Observations were carried out by 2 observers using snorkelling gear. The observers each started at one end of the line-up of EUs and then swam in the opposite direction to each other. One observer swam between the EUs and the mangrove border, while the other swam between the EUs and the seagrass border. After surveying all EUs, observers waited for a period of 5 min, after which a second observation round started in which the 16 EUs were surveyed again. Each observer carried out 4 observation rounds per time interval, resulting in a total of 8 observations per time interval, per day, per EU. All fishes inside the EU at the moment of observation were identified, their numbers were counted, and individual sizes were estimated in 2.5 cm length classes. The behaviour of each fish (resting, feeding, swimming, aggressive interactions, and hunting) and its location inside the EU (bottom, mid-water, top) were recorded by one of the trained observers. Each EU was observed daily for 3 consecutive days, during which time they remained in the same randomised order in the field. EUs were put in a new randomised order after the last observation on every 3rd day, in order to diminish possible effects of location in the line of EUs. Every 4th d (i.e. day after randomly reordering EUs), EUs were left to rest and no observations took place. In total, EUs were placed in 8 different random orders (configurations) and were observed for 3 d per configuration. Water depth was measured at the beginning of each observation round. EUs were completely submerged during all observations; at all times water levels were higher than 80 cm.



Data analysis. Because Eucinostomus jonesii and E. gula were difficult to discriminate under water, these 2 species were recorded as 'Eucinostomus spp.'. All other fishes could be determined to species-level. For the data analysis, all encountered fish species were subdivided into 5 feeding guilds that were either diurnally or nocturnally active (following Randall 1967, Carr & Adams 1973). Mean fish numbers were calculated per EU (treatment) for each feeding guild and for the most abundant species within each quild. The 8 configurations were regarded as replicates through time per treatment. Fish numbers were calculated by (1) taking the mean of the 8 observation rounds of the 2 observers per time interval for each day, which resulted in 9 values (3 time intervals of 3 consecutive days) for 8 configurations (replicates); (2) these 9 values were then averaged for each of the 8 replicates. Mean and SE shown in all figures were calculated from these 8 replicates.

**Statistical analysis.** Differences in light intensities under natural and experimental conditions were tested with a Mann-Whitney *U*-test, for the combinations 'full sunlight' measured above the muddy substratum versus 'empty, unshaded EU', and 'mangrove rim' versus 'EU with AM and shade', using the statistical programme SPSS 11.0.

Prior to statistical analysis, fish abundances were calculated by taking the sum of the 72 observations (8 observation rounds for 3 time intervals for 3 consecutive days) per treatment and per replicate, in order to maintain discrete values to enable a Poisson regression (McCullagh & Nelder 1989). We pooled data from the 8 observation round because of their apparent dependency. The 3 time intervals and 3 d within 1 configuration were pooled for 2 reasons: (1) we were not interested in time or day effects, and (2) by pooling these values we did not have to make the (questionable) assumption of independent data between time intervals within 1 d, or between days at the same time interval. The pooling process was the same as applied by Cocheret de la Morinière et al. (2004). Pooling resulted in 128 observations: 8 replicates (through time) for 16 treatments.

The data were counts and did not have a normal distribution due to the many 0 counts. EUs were relatively small compared to the vast expanses of natural mangroves and seagrass beds surrounding the EUs. As a result, the chance of encountering a fish inside an EU was small (explaining the 0 counts).

To test the main and interaction effects of structure, food and shade on fish abundances, a Poisson regression with a log-link function was performed for each feeding guild and the most abundant species, using the statistical programme Genstat 7.2. Since the observed counts tended to be overdispersed compared to Poisson variance, a correction was applied using the scaled deviance as an estimate for the dispersion parameter (McCullagh & Nelder 1989). The model contained 1 dependent factor (fish abundance), 3 treatment factors (structure, food, and shade) with their interactions, and 2 blocks (replication through time and EU location). The EU location, although randomized in the experiment, was included because location effects were still possible. In order to test the main effects of the factor 'food' and 'shade' for the pooled herbivore species between 2 EUs with the same type of structure (with/without food, with/without shade), *t*-tests for pairwise differences (127 degrees of freedom) were performed for such a pair of EUs within the context of the Poisson regression.

### RESULTS

## **Light intensity**

Underwater light intensities measured in the field under natural and experimental conditions are shown in Table 1. Light intensities in full sunlight and in an empty EU were not significantly different (p = 0.749, Z = -0.320, Mann-Whitney *U*-test) whereas the light intensity was significantly lower at the mangrove rim than in an EU with AM and shade (p = 0.004, Z = -2.882, Mann-Whitney *U*-test).

#### Fish abundance

Table 2 shows the results of the Poisson regression for all feeding guilds and their most abundant species. In total, 15 fish species were encountered in the experimental EUs (see Table 3). Pooled abundances of diurnally active herbivores were significantly affected

Table 1. Light intensities ( $\mu E m^{-2} s^{-1}$ ; mean  $\pm SE$ ) measured under natural and experimental conditions. Light intensities in 'full sunlight' were measured above the muddy substratum adjacent to the mangroves. –: without; +: with. EU + AM: EU with artificial mangrove roots only. EU – AM: empty EUs and EUs with artificial seagrass leaves only

Condition	Light intensity
Full sunlight	$1112 \pm 61$
Mangrove rim	$109 \pm 25$
Inside mangroves	$17 \pm 1$
EU – AM – shade	$1149 \pm 70$
EU – AM + shade	$386 \pm 27$
EU + AM – shade	$1060 \pm 48$
EU + AM + shade	$369 \pm 38$

Table 2. Results of the Poisson regression for treatments 'structure' (struct), 'food' and 'shade', their interactions, and the blocks 'replicate through time' (rep) and 'location' (loc). Test results are given for the different feeding guilds (in bold) and for the most abundant species within each guild. D: diurnally active, N: nocturnally active. ns: not significant (p > 0.05), \*0.01 < p < 0.05, \*\*0.001 < p < 0.01, \*\*\*p < 0.001

	Tre	eatme	nts		the standard states and stat	Blo	cks		
	struct	1000	snade	struct ×100d	struct×snade	100d ×shade	struct ×100d ×shade	rep	10C
Herbivores (D)	* * *	* * *	ns	***	ns	**	ns	ns	* * *
Acanthurus chirurgus	***	* * *	ns	***	ns	**	ns	ns	* * *
Zoobenthivores (D)	**	**	ns	***	ns	ns	*	***	***
Eucinostomus spp.	ns	***	ns	ns	ns	ns	ns	***	***
Zoobenthivores (N)	***	* * *	ns	* * *	***	*	ns	**	* * *
Haemulon flavolineatum	***	* * *	***	*	ns	***	ns	***	* * *
Lutjanus mahogoni	***	ns	ns	ns	***	ns	ns	ns	* * *
Lutjanus apodus	***	**	***	**	**	ns	ns	**	* * *
Zoobenthivores/ Zooplanktivores (N)									
Ocyurus chrysurus	***	*	ns	* *	* *	ns	ns	***	**
<b>Piscivores (D)</b> Sphyraena barracuda	***	ns	ns	ns	ns	ns	ns	**	***

by the factors structure and food (Table 2), and significant interactions were present between structure and food and between food and shade (Table 2). When food was present (in the form of algae), numbers of herbivores increased with increasing structure surface in the order: AS only, AM only, and both AS and AM (Fig. 3a). In all EUs without food, fish numbers were similarly low for all levels of structure (Fig. 3a). In EUs containing AM only, herbivores were more abundant in EUs with food than in EUs without food (Fig. 3a, p < breve0.001 for EUs 'without shade' and 'with shade', paired *t*-test). The same was true for EUs with both AM and AS when shade was absent (p = 0.004, paired *t*-test), but not true when shade was present (p = 0.854, paired t-test) (Fig. 3a). Shade seemed to have a negative effect on herbivore abundances in EUs with food and AM structures, compared to the same EUs when unshaded (Fig. 3a); however, this effect was not significant (EU with AM only: p = 0.105, and EU with AM and AS: p = 0.097; paired *t*-test). At species level, Acanthurus bahianus (Fig. 3b, Table 2), A. chirurgus and Scarus guacamaia showed the same trends as those of pooled herbivores.

Pooled abundances of diurnally active zoobenthivores were significantly affected by the interaction between structure and food, and by the 3-way interaction between structure, food and shade (Table 2). However, pooled diurnally active zoobenthivores seemed equally abundant for all levels of structure, food and shade (Fig. 4a). An exception was the EU without shade, without food and with AM only, where fish numbers were higher compared to the same EU with food because of high numbers of *Gerres cinereus* in 1 of the 8 replicates. As a result of this outlier the SE is very large (Fig. 4a). The most abundant diurnally active zoobenthivore species, *Eucinostomus* spp., showed significantly higher numbers in EUs with food than in EUs without food (Fig. 4b, Table 2). Their abundances were not significantly affected by structure or shade (Fig. 4b, Table 2).





Fig. 3. Mean fish abundance for each treatment (left part) and for each level of the separate factors structure, food and shade (right part) of (a) pooled diurnally active herbivores and (b) *Acanthurus bahianus*. N: total sum of fishes of all observation rounds, time intervals and days. Left part shows all 16 EUs with and without food. Right part shows mean fish abundance in EUs for each structure type and in EUs with and without food. Empty = EU without structure, AS = artificial seagrass leaves only, AM = artificial mangrove roots only, AS + AM = artificial seagrass leaves and artificial mangrove roots. If the SE of the mean is not visible it is very small

Pooled abundances of nocturnally active zoobenthivore species were significantly affected by structure and food, and significant interactions were present between structure and food, between structure and shade, and between food and shade (Table 2). Their abundances were highest in any type of EU containing AS (Fig. 5a). This was also observed at species level for Haemulon flavolineatum, Lutjanus mahogoni and Lut*janus apodus* (Fig. 5b–d), but in the presence of shade this pattern disappeared for the latter 2 species that correspondingly showed significant interactions between structure and shade (Table 2). In presence of shade *L. apodus* showed higher numbers in EUs with any type of structure than in EUs without structure (Fig. 5d, Table 2); in contrast, in presence of shade, numbers of *L. mahogoni* were equally high in all EUs including the EU without structure (Fig. 5c, Table 2). H. flavolineatum and L. apodus were significantly affected by food, in interaction with shade and/or structure (Table 2). Numbers of *H. flavolineatum* and *L. apodus* were higher in AS EUs without food than in AS EUs with food (Fig. 5b,d), particularly in the presence of shade, which is attributable to reduced structural complexity of fouled AS (see 'Discussion'). These 2 highly abundant species were responsible for the significant effect of the factor food, and the significant interaction between food and structure and between food and shade, for the pooled nocturnally active zoobenthivores (Fig. 5a, Table 2).



Fig. 4. Mean fish abundance for each treatment (left part) and for each level of the separate factors structure, food and shade (right part) of (a) pooled diurnally active zoobenthivores, and (b) *Eucinostomus* spp. See Fig. 3 legend for other details

Nocturnally active zoobenthivores/zooplanktivores were only represented by *Ocyurus chrysurus*. Abundances were significantly affected by the factors structure and food and by interactions between structure and food and between structure and shade (Table 2). Similar to the nocturnally active zoobenthivores, abundances of *O. chrysurus* were highest in any EU with AS when shade was absent (Fig. 6a). Just like *H. flavolin*-



Fig. 5. Mean fish abundance for each treatment (left part) and for each level of the separate factors structure, food and shade (right part) of (a) pooled nocturnally active zoobenthivores,
(b) Haemulon flavolineatum, (c) Lutjanus mahogoni, and (d) Lutjanus apodus. See Fig. 3 legend for other details

*eatum* and *L. apodus*, numbers of *O. chrysurus* were higher in AS EUs without food than in AS EUs with food (Fig. 6a). As in the case of *Lutjanus apodus*, abundances of *O. chrysurus in* the presence of shade were higher in EUs with any type of structure compared to empty EUs (Fig. 6a).

Abundances of diurnally active piscivores, represented by *Sphyraena barracuda*, were only significantly affected by structure (Table 2). Abundances were higher in all EUs with structure than in empty EUs (Fig. 6b).

# **Fish behaviour**

Herbivore fish species were mainly foraging in the EUs (69 to 92%, Table 3). On average, herbivores grazed mostly on algae growing on AM (44 to 85%), and less on algae growing on AS (6 to 23%) and least on the bottom substratum (1 to 12%). When comparing EUs with and without food, the percentage of herbivores foraging on AM was much higher in EUs with food (67.4%) than in EUs without food (1.5%), as would be expected. The percentage of herbivores foraging on AS was not higher in EUs with food (17.0%) than in EUs without food (26.3%), but the number of herbivores present in EUs with AS without food, which should be kept in mind when interpreting the percentages.

Nocturnally and diurnally active zoobenthivores mostly rested in the EUs (94.9 and 72.2%, respectively, Table 3). An exception was the diurnally active *Mulloidichthys martinicus*, which was searching for food on the substratum in 54.1% of the cases (Table 3), and did so in EUs both with and without food (40.5 and 67.4%, respectively). Furthermore, *Haemulon bonariense* was actively swimming in and out of the EUs in 53.7% of the cases. All zoobenthivorous species rested on the bottom of the EUs, except for *Lutjanus apodus* that also rested (i.e. hovered) in the upper part of the EU (Table 3).

Most specimens of the nocturnally active zoobenthivore/zooplanktivore *Ocyurus chrysurus* and the piscivorous *Sphyraena barracuda* rested in the EUs (80.7 and 98.2%, respectively). In contrast to the zoobenthivores, they rested on the bottom, in the middle, and also in the upper part of the EUs (Table 3). When this effect was separated for type of structure, the following pattern was observed: when only AS was present, they rested mostly in between the artificial leaves (bottom of EU, 70.0 to 90.2%); when only AM was present, they rested mostly in between the artificial roots (middle and top of the EU, 70.0 to 85.8%); when both structures were present, they rested in between the leaves (32.6



Fig. 6. Mean fish abundance for each treatment (left part) and for each level of the separate factors structure, food and shade (right part) of (a) nocturnally active zoobenthivores/zooplanktivores (*Ocyurus chrysurus*), and (b) diurnally active piscivores (*Sphyraena barracuda*). See Fig. 3 legend for other details

to 40.8%) as well as in between the roots (45.6 to 65.6%). The same pattern was observed for *Lutjanus apodus* (97.3, 91.3, and 49.6 vs. 50.5%, respectively).

#### DISCUSSION

The experiment imparted 2 artefacts that should be considered when interpreting some data points in the figures. The artificial seagrasses could not be completely cleaned from calcareous algae, resulting in a few herbivores that also foraged on artificial seagrass leaves 'without food'. Secondly, some artificial seagrass leaves 'with food' did not stand completely upright due to the weight of the fouling algae and sediment, which is often also the case in a natural seagrass bed. This resulted in artificial seagrass leaves 'with food' sometimes harbouring lower fish abundances than seagrasses 'without food' (mainly Haemulon flavolineatum). When discussing the results, we have taken these artefacts into account and drawn conclusions accordingly. These differences had no effect on the general patterns and the conclusions of this study, because a comparison of fish abundances between artificial seagrass leaves with and without food is not made in the following paragraphs.

Herbivores were more abundant in EUs with food than in EUs without food (i.e. algae), dependent on the level of structure and presence of shade. The majority Table 3. Behaviour and location in the EUs per species of different feeding guilds. For each behaviour and location in the EU, percentages of cases are given, as counted by one trained observer. Numbers in bold represent values for the different feeding guilds, whereas values in plain text are for separate species within each guild. Grey areas indicate most frequent behaviour. D: diurnally active; N: nocturnally active; Total N: sum of fishes counted in all observation rounds by a single observer; Resting: at the bottom, in the middle, or on the top part of an EU; Foraging: from substratum (Sub), algae on artificial seagrass leaves (AS), or algae on artificial mangrove roots (AM); S: swimming through an EU; Other: other types of behaviour (for *M. martinicus* it indicates searching for food)

Scientific name	Common name	Total N	Size range (cm)	All I	Res Bottom	ting Middle	e Top	All	Fora Sub	aging AS	AM	S	Other
Herbivores (D)		506		11.3	8.4	1.6	1.4	80.4	10.3	19.5	50.6	8.2	0.2
Acanthurus chirurgus	Doctorfish	36	7.5-22.5	13.9	8.3	2.8	2.8	69.4	8.3	16.7	44.4	16.7	-
Acanthurus bahianus	Ocean surgeonfish	398	5.0 - 22.5	10.8	8.3	1.3	1.3	80.9	12.3	22.6	46.0	8.0	0.3
Scarus guacamaia	Rainbow parrotfish	72	7.5-15.0	2.8	-	1.4	1.4	91.7	1.4	5.6	84.7	5.6	-
Zoobenthivores (D)		381		72.2	70.6	1.3	0.3	3.2	0.8	1.8	0.5	12.6	12.0
Eucinostomus spp.	Slender mojarra & Silver jenny	242	5.0-15.0	87.6	86.0	1.2	0.4	2.1	1.2	0.8	-	10.3	-
Gerres cinereus	Yellowfin mojarra	54	5.0 - 40.0	87.0	85.2	1.9	-	_	-	-	_	13.0	-
Mulloidichthys martinicus	Yellow goatfish	85	7.5-35.0	18.8	17.6	1.2	-	8.3	-	5.9	2.4	18.8	54.1
Zoobenthivores (N)		1467		94.9	85.8	0.8	8.3	1.0	0.5	0.2	0.2	3.0	1.2
Haemulon bonariense	Black grunt	54	10.0 - 27.5	35.2	35.2	-	-	9.3	9.3	_	_	53.7	1.9
Haemulon flavolineatum	French grunt	808	2.5 - 12.5	98.0	96.5	0.1	1.4	0.9	0.1	0.4	0.4	-	1.1
Haemulon sciurus	Bluestriped grunt	9	10.0 - 22.5	100.0	100.0	-	_	_	-	-	_	-	-
Lutjanus mahogoni	Mahogany snappe	r 263	2.5 - 12.5	94.3	93.5	0.4	0.4	0.8	0.8	-	_	4.2	0.8
Lutjanus analis	Mutton snapper	74	2.5 - 10.0	95.9	95.9	-	-	-	-	-	-	1.4	2.7
Lutjanus apodus	Schoolmaster	259	5.0 - 25.0	97.7	51.4	3.9	42.5	-	-	-	-	1.2	1.2
Zoobenthivores/ Zooplanktivores (N) Ocyurus chrysurus	Yellowtail snapper	244	2.5-22.5	80.7	38.1	29.1	13.5	0.8	0.4	0.4	_	18.0	0.4
Piscivores (D)													
Sphyraena barracuda	Great barracuda	660	2.5-47.5	98.2	37.0	15.9	45.3	-	-	-	-	1.4	0.5

of herbivores were feeding inside the EUs, and did so mostly in EUs with highest structural complexity (i.e. with artificial mangrove roots). These harboured higher herbivore abundances than empty EUs and EUs with artificial seagrass leaves, probably because the higher surface area of artificial mangrove roots provided more substrate for algae and thus more food. The attraction to algae was supported by the fact that unshaded EUs with AM appeared to harbour higher herbivore abundances than shaded EUs: the former possibly contained more algae due to the higher light intensities in the EUs, and therefore also more food. The experimental data thus suggest that juvenile herbivores utilise mangrove and seagrass habitats primarily as feeding habitats. In absence of algae, herbivores did not show any preference for presence or absence of structure. Although several studies have suggested that juvenile fish use structure provided by mangroves and seagrass beds as shelter (Orth et al. 1984, Jenkins & Sutherland 1997, Cocheret de la Morinière et al. 2004), the present study indicates that is not the case for herbivores, which are diurnally active foragers.

Pooled diurnally active zoobenthivores showed no clear preference for any of the factors or their combinations, despite significant (interaction) effects. *Eucinosto*- *mus* spp., on the other hand, were more attracted to EUs with macrozoobenthos than without. Their abundances were comparable in all EUs with access to macrozoobenthos, irrespective of structure and shade. The experimental data thus suggest that these diurnally active species utilise mangroves and seagrass beds primarily as feeding habitats and not as shelter habitats. Nevertheless, these species were mainly observed resting in the EUs. However, because these species feed by taking sporadic bites, and because the behavioural observations were only snap-shots, the importance of feeding was underestimated in terms of duration.

Our results on herbivores and diurnally active zoobenthivores are in agreement with those of Laegdsgaard & Johnson (2001), who found that artificial mangrove roots with fouling algae and associated invertebrates attracted significantly higher abundances of some juvenile fish species than structures cleaned of algal growth. Hence, it can be concluded that for several diurnally active herbivore and zoobenthivore species, Caribbean mangroves and seagrass beds function as foraging habitats, but are not used continously for shelter during daytime. This shelter may become temporarily important, however, when fishes are attacked by a predator (Laegdsgaard & Johnson 2001).

Nocturnally active zoobenthivores and zooplanktivores were attracted to the EUs during daytime by structure (occasionally in interaction with shade) and not by food. These species all shelter in structurally complex habitats during daytime and feed in the seagrass beds at night (e.g. Ogden & Zieman 1977, Rooker & Dennis 1991, Nagelkerken et al. 2000). Accordingly, in the present study they did not show elevated abundances in EUs with food (i.e. macrozoobenthos) and were observed resting instead of foraging during daytime. The significant interaction between food and structure found for most nocturnally active zoobenthivorous and zooplanktivorous species was due to the reduced structural complexity of artificial seagrass leaves fouled by algae (as explained in the first paragraph of 'Discussion'), which were not attractive to these species that were mostly observed to rest in between structurally complex areas. In the absence of shade, Haemulon flavolineatum, Lutjanus mahogoni, L. apodus and Ocyurus chrysurus showed highest abundances in any EU containing artificial seagrass leaves, and none favoured EUs with artificial mangrove roots only. This indicates that unshaded mangrove roots probably provide an unnatural shelter habitat that is not recognised by the fishes. In the presence of shade, EUs with artificial mangrove roots alone were indeed used by the fishes (except H. flavolineatum) to a similar extent as EUs with artificial seagrass leaves. When the unnatural unshaded artificial mangrove roots were not considered, species-specific patterns for habitat preference and shade could be distinguished. H. lavolineatum was attracted mainly by artificial seagrass leaves: its abundance was highest in any EU with artificial seagrass, and it was observed resting only in between the artificial leaves. The same was true for L. mahogoni in absence of shade, which also showed highest abundances in unshaded artificial seagrass (which is a natural situation for seagrass beds). However, in presence of shade, L. mahogoni occurred in equally high abundances in all habitat types. L. apodus and O. chrysurus did not show a different preference for either artificial mangrove roots or artificial seagrass leaves; when the former habitat type was present they were found in the top or middle of the EU (i.e. near or in between the artificial mangrove roots), when the latter habitat type was present they were found on the bottom of the EU (i.e. in between the artificial seagrass leaves), and when both were present they were found in either part of the EUs.

Many studies have shown that structural complexity reduces predation efficiency and offers shelter to both fish and shrimps (Heck & Thoman 1981, Orth et al. 1984, Primavera 1997, Laegdsgaard & Johnson 2001). In the latter study, for example, laboratory experiments were performed showing that juvenile fish avoided artificial mangrove pneumatophores when predators were absent, whereas they actively sought shelter inside the structure when predators were introduced. Furthermore, risk of predation was higher on bare mudflats than in structurally complex seagrasses and mangroves (Laegdsgaard & Johnson 2001). So, in the present study, nocturnally active fish species that rested during daytime were very likely to use mangrove-like and seagrass-like structures as shelter, since the experiment took place in an environment where predators were not excluded.

The piscivore Sphyraena barracuda was only attracted to the EUs by structure. Regardless of presence of food (algae and macrozoobenthos, which are no food sources for this species) or shade. Abundances were higher in EUs with any type of structure than in empty EUs, suggesting that this species may not have a preference for seagrass beds over mangroves or vice versa. Indeed, in EUs with both seagrass leaves and mangrove roots, S. barracuda of equal percentages rested in between structures of both habitat types. Although smaller individuals of this species may have used the structure for shelter, larger individuals most likely used the structure for ambush to forage on sardines, anchovies, and silversides that were schooling around the EUs. This is supported by field observations showing that S. barracuda often roams around at the mangrove/seagrass interface in search of schools of these silverfish (Nagelkerken & van der Velde 2004). High levels of structural complexity have been suggested to aid ambush predators by camouflaging them and decreasing their visibility to their prey (Heck & Orth 1980, Coen et al. 1981, Howard & Koehn 1985).

In contrast to other studies (Hair et al. 1994, Cocheret de la Morinière et al. 2004), our study suggests a relatively small effect of shade as a single factor. This may be due to the fact that light levels inside the shaded EUs were still 20 times higher than inside natural mangroves. Indeed, for visual predators, reductions in light intensities may influence prey detection and reduce predation rates only at very low light intensities (Benfield & Minello 1996). However, the present study used similar light levels as the study of Cocheret de la Morinière et al. (2004), which was also conducted in the Spanish Water Bay. The fact that they found an effect of shade is perhaps caused by the placement of the EUs on an isolated seagrass bed far away from mangroves that offer high levels of shade, whereas our EUs were placed directly adjacent to mangroves.

# CONCLUSIONS

Our experiments provided evidence that herbivorous fish probably use mangroves and seagrass beds during daytime primarily to feed on fouling algae and epiphytes. Diurnally active zoobenthivoric *Eucinosto*- *mus* spp. use mangroves and seagrass beds during daytime to feed on zoobenthos. Nocturnally active zoobenthivores/zooplanktivores primarily use seagrass and mangrove structures (in interaction with shade) during daytime to rest and shelter. The piscivore *Sphyraena barracuda* is probably attracted to mangroves and seagrass beds primarily by structure: larger-sized specimens use this to ambush their prey, while smaller-sized specimens may also use the structure for protection. The hierarchy in importance of the various factors clearly differed between feeding guilds and species. Overall, it can be concluded that during daytime the presence of structure, food and shade significantly contribute to the attractiveness of mangroves and seagrass beds to juvenile reef fish.

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