

1 **Testing the efficiency of extraction of incurred residues from soil with optimized multi-**
2 **residue method**

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Abstract

The efficient extraction of pesticide residues from various matrices is of primary importance for obtaining unbiased results. The reproducibility of extraction of residues from spiked soil samples and from soils containing incurred residues was tested with ¹⁴C-labeled test compounds of different physical-chemical properties. Nearly 100% of the compounds added to the sample before extraction could be recovered with an average reproducibility relative standard deviation (CV) of 5.4%. The additional steps of the determination process (cleanup, evaporation, etc.) contributed to the major part of the variability of the results (CV=10-20%). The incurred residues were most efficiently extracted with acetone for 30 min followed by the mixture of acetone/ethyl acetate 1:1 for additional 30 min. However, they could only be recovered at various extent (64-90% of total residues), underlying the importance of testing the efficiency of extraction. The residues were identified and quantified by gas chromatography applying thermionic detector. The performance parameters of the method complied with the international method validation guidelines, and they proved to be robust and suitable for determination of pesticide residues in soils of widely different physical-chemical properties.

Keywords: residue analysis, pesticide residues in soils, efficiency of extraction, incurred residues.

30 **Introduction**

31

32 There are several extraction methods for determining pesticide residues in soil. Traditionally
33 the Soxhlet ^[1, 2] and the solid phase (SPE) extractions ^[3-6] are used as official methods in
34 many countries. Their main drawbacks are requiring large volume of solvents, and the lengthy
35 extraction time. Among the new techniques, the supercritical fluid extraction (SFE), ^[2, 7, 8]
36 pressurized solvent extraction (PLE), ^[9, 10] accelerated solvent extraction (ASE), ^[11, 12]
37 microwave-assisted solvent extraction (MASE, MAE) ^[12-17] are the most widely used methods
38 for extraction of pesticides from environmental samples. These methods produce high
39 recovery of residues applying specific expensive instruments and large solvent volumes in
40 some cases. Ultrasonic solvent extractions (USE) is one of the preferred techniques ^[12, 18-22] as
41 it can be performed with less solvents and shorter time. The solid phase micro extraction
42 (SPME) ^[6, 17, 23] is mainly used for determining volatile compounds.

43

44 Wide range of solvents have been used depending on the purpose of the analysis. Acetone or
45 acetone – water mixture is frequently used ^[18, 20, 23-25] in which the soil clods fall apart
46 facilitating the complete partition of compounds between the soil and solvent phase.

47 Disintegration of soil particles is further assisted by adding ammonium chloride (NH₄Cl),
48 ammonium phosphate (NH₄)₃PO₄. ^[26-28] The mixtures of acetone with hexane, ethyl acetate
49 (EtAc) or toluene improve the recovery of compounds of wide polarity range. ^[1, 9, 16]

50 Previously dichloromethane was also used, ^[24, 25] but currently its use is restricted for
51 protecting the environment. For its replacement EtAc and cyclohexane are applied. ^[12, 19, 20, 29]

52 The presently applied methods are the variants of those used for residue analysis in food
53 matrices, such as the QuEChERS method ^[12, 28, 30, 31] involving acetonitril for extraction. The

54 EN 12393-2:2014 Standard, [32] based on acetone or EtAc/cyclohexane 1:1 solvent
 55 extraction, was also applied for soil matrices. [12]
 56
 57 The methods should be validated before use to provide evidence that they fit for the intended
 58 purposes. [33, 34] The generally acceptable main performance parameters are: specificity: signal
 59 resulted from untreated control sample is less than 30% of limit of quantification (LOQ)
 60 which is the lowest concentration that can be quantified reproducibly with known uncertainty;
 61 [35] sensitivity (LOD): typically 0.2 [36] - 0.3 LOQ set generally at 10 times the noise level; [37]
 62 matrix effect <±20% compared to response of pure standard solution; the mean recovery:
 63 within 70-120%. [34] The linearity and goodness of calibration should be tested with minimum
 64 5-point calibration covering the analytical range. Its measure is the standard deviation of
 65 relative residuals [38] (S_{rr}) instead of the usually applied coefficient of regression (R^2).

66

$$67 \quad S_{rr} = \sqrt{\frac{\sum(Y_{rel,i} - \bar{Y}_{rel})^2}{n-2}} \quad (1)$$

68

$$69 \quad Y_{rel}\% = 100 \times \frac{Y_i - \hat{Y}}{\hat{Y}} \quad (2)$$

70 Where Y_i is the response and \hat{Y}_i is read from calibration line for x_i calibration concentration, n
 71 is the number of ≥ 5 calibration points. Since the standard deviation of the residuals is usually
 72 proportional to the injected analyte, the standard deviation of the relative residuals reflects the
 73 average variability of the calibration points. Applying weighted linear regression S_{rr} should be
 74 $\leq 20\%$. [34]

75 The uncertainty of the measured residue values should be $\leq 20\%$. It is usually expressed as the
 76 relative standard deviation obtained from repeatability and within laboratory reproducibility
 77 determined from minimum 5 recovery studies.

78 Though the Codex Method Validation GLs ^[33] and the FAO/WHO Joint Meeting on Pesticide
79 Residues (JMPR) ^[39] list the efficiency of extraction and homogeneity of analytical sample
80 obtained from the laboratory sample as a basic performance characteristics to be tested, the
81 published validation reports rarely include these important parameters. It should be
82 emphasized that neither the analyses of proficiency tests and collaborative study samples nor
83 recovery studies performed with spiked test portions removed from the analytical sample
84 provide information on the homogeneity of analytical sample. The efficiency of extraction can
85 only be determined from these studies, if the samples contain incurred residues.

86

87 The objectives of our study are to test the applicability of widely used solvents (acetone, ethyl
88 acetate and hexane) that can be used with GC-NPD (nitrogen phosphor selective detector) and
89 ECD if GC-MS/MS or LC-MS/MS systems are not available for determination of pesticide
90 residues, optimize the extraction procedure and assess its efficiency for extraction of ¹⁴C-
91 labeled incurred residues from soil.

92

93 **Materials and methods**

94

95 *Equipment*

96

97 In addition to the usual laboratory glassware and devices, the following major equipment was
98 used: Beckman 6000 TA liquid scintillation counter (LSC) with automatic quenching
99 compensation; OX400 Biological Oxidizer; Stephan UM 5 Universal and Tecator 2096
100 laboratory homogenizers; Sigma 4K15 centrifuge; Mettler top load (0.01 g) and analytical
101 (0.00001 g) balances; Edmund Bühler SM 25 and Certomat SII sieve shakers; TurboVap
102 (Zymark) solvent evaporator; Varian 3800 gas chromatograph equipped with thermoionic

103 (TSD) detector and PTV injector (1079); CP-Sil-8CB Low Bleed MS column (25m ×
104 0.32mm, df = 0.25µm); and 2.5m × 0.32mm methyl deactivated retention gap. Aglient GC
105 with split/splitless injector and nitrogen, phosphor sensitive detector.

106

107 *Materials*

108

109 Ultima Gold™ liquid scintillation cocktail (Perkin Elmer) for LSC;

110 Calibration Standard for LSC: normal activity of ¹⁴Carbon standard (code CRF 101,

111 Amersham International plc, UK) is 5000 disintegrations per minute (dpm). ¹⁴C radionuclide

112 purity >99.9%. Half-life 5730 ± 40 years.

113 Absorption Solution for ¹⁴CO₂ for biological oxidizer: 10 mL of ethanolamine/methanol

114 (12.5/87.5, v/v).

115 Filter paper to determine blank background activity of biological oxidizer.

116 Anhydrous calcium chloride (Merck reag. grade).

117 Analytical reference standards >98% purity (Dr. Ehrenstorfer GmbH): azinphos-ethyl (S-

118 (3,4-dihydro-4-oxobenzo[d]-[1,2,3]-triazin-3-ylmethyl) O,O-diethylphosphorodithioate),

119 chlorfenvinphos (2-chloro-1-(2,4-dichlorophenyl)ethenyl phosphate), chlorpyrifos (O,O-

120 diethyl O-3,5,6-trichloro-2-pyridyl phosphorothioate), dimethenamid ((RS)-2-chloro-N-(2,4-

121 dimethyl-3thienyl)-N-(2-methoxy-1-methylethyl)acetamide), oxyfluorfen (2-chloro-α,α,α-

122 trifluoro-p-tolyl 3-ethoxy-4-nitrophenyl ether), pendimethalin (N-(1-ethylpropyl)-2,6-

123 dinitro-3,4-xylidine), promertyn (N²,N⁴-diisopropyl-6-methylthio-1,3,5-triazine-2,4-diamine),

124 propazine (6-chloro- N²,N⁴-diisopropyl-1,3,5- triazine-2,4-diamine), terbuthylazine (N²-tert-

125 butyl-6-chloro-N⁴-ethyl-1,3,5-triazine-2,4-diamine), terbutryn (N²-tert-butyl-N⁴-ethyl-6-

126 methylthio-1,3,5-2,4-diamine).

127 ¹⁴C-labeled reference standards: triazol-ring- ¹⁴C-atrazine (6-chloro-N²-ethyl-N⁴-isopropyl-
128 1,3,5-triazine-2,4-diamine) (specific activity 1.6 Mbq/mg radioactive purity 96.5%, provided
129 by Syngenta, 96,5%), (2,2-dimethyl, 3)-¹⁴C-carbofuran (CA) (2,3-dihydro-2,2-
130 dimethylbenzofuran-7-yl methylcarbamate), ethyl-1-¹⁴C-chlorfenvinphos (CF), Ethyl-1- ¹⁴C-
131 chlorpyrifos (CP) and ethyl-1-¹⁴C-p,p'-DDT (DT) (1,1,1-trichloro-2,2-bis(4-
132 chlorophenyl)ethane) >95% radioactive purity provided by the International Atomic Energy
133 Agency (IAEA).

134 The characteristic physical properties and chemical structural formula indicating the label
135 position(s) are summarized in Table 1 and shown in Figure 1, respectively.

136

137 *Soils used in the experiments*

138

139 About 20 kg soil was collected from the top 15 cm layer at six sites having different physical
140 characteristics (Table 2.) which could affect the recovery of the pesticide residues.

141 The samples were prepared following the ISO 11464:2006 Standard^[40] and processed as
142 described by Suszter et al. ^[41] The two terms are synonyms, but in pesticide residue analysis
143 they indicate different operations.

144

145 Sample preparation: the procedure used, if required, to convert the laboratory sample into
146 the analytical sample, by removal of parts (soil, stones, bones, etc.) not to be included in the
147 analysis. ^[33]

148 Sample processing: the procedure(s) (e.g. cutting, grinding, mixing) used to make the
149 analytical sample acceptably homogeneous with respect to the analyte distribution, prior to
150 removal of the analytical portion. The processing element of preparation must be designed to
151 avoid inducing changes in the concentration of the analyte. ^[33]

152

153 **Methods**

154

155 *Determination of dry matter content of soil samples*

156

157 Clean porcelain dishes were pre-heated at 105 °C until constant weight (c [g] and stored over
158 activated anhydrous CaCl₂ in desiccator until use. Ten g of processed and homogenized
159 sample were weighted to the porcelain dishes (a [g]) and heated at 105 °C until constant
160 weight, cooled to room temperature in desiccator and weighted again (b [g]). The dry matter
161 content (dm [%]) of the soils was calculates as:

162
$$dm = 100 - 100 \times \frac{a-b}{a-c} \quad (3)$$

163 All residue values were expressed on dry matter basis in this study.

164

165 *Determination of ¹⁴C activity of samples.*

166

167 Before the series of radioactivity measurements were started, the efficiency of the biological
168 oxidizer, used for determining the ¹⁴C activity in soil samples, was tested. Ten mL of
169 absorption solution and 5 mL of scintillation cocktail were pipetted into a scintillation vial to
170 absorb the evolved ¹⁴CO₂. First the background activity was measured by placing about 500
171 mg filter paper into the combustion boot followed by the measurement of the activity of a
172 complete strip of ¹⁴C standard paper for calibration. The absorbed ¹⁴C activity was measured.
173 The efficiency of the oxidizer was calculated as:

174

175
$$Efficiency = \frac{dpm\ standard\ after\ combustion - dpm\ blank\ after\ combustion}{dpm\ standard} \quad (4)$$

176

177 A recovery of 98% or greater indicated that the oxidizer worked efficiently.

178 The ^{14}C activities of the reference standard, the background activity of soil, as well as the

179 treated soils (500 mg) were determined following the same procedure.

180 For the determination of the ^{14}C activity of the extracts, 12 mL scintillation cocktail and 5 mL

181 extract were transferred into 20 mL LSC vial. The vial was tightly closed, shaken and placed

182 in the Beckman LSC counter.

183 Each sample was measured three times for 5 minutes, after running the self-calibration

184 program, and their average activity was used for further calculations. The average relative

185 standard deviation of the replicate LSC measurements of 29 test portions was 0.0073.

186

187 *Determination of the reproducibility of extraction of spiked samples*

188

189 Six soils of different physical-chemical properties (U129, V01, V02, W33, X65, and Y97)

190 shown in Table 3 were used for studying the reproducibility of extraction.

191 For the treatment of soil test portions, an acetone stock solution containing ^{14}C -labeled

192 atrazine and cold atrazine analytical standard at 0.05 mg/mL concentration with 50,000

193 Bq/mL (3,000,000 dpm/mL) target specific activity was prepared.

194 Test portions of 20 g of processed soil samples were weighed into Petri dish. One thousand

195 μL of 0.001mg/mL atrazine standard solution, prepared from the stock solution, was spread

196 over the soil surface with Hamilton syringe (spike level 0.05 mg/kg). The spiked sample was

197 kept in fume hood for 30 min to evaporate the acetone, then transferred into a 250 mL

198 centrifuge tube and 2.8 mL of 0.2 mol NH_4Cl solution and 40 mL acetone were added. The

199 tightly closed tube was shaken for 30 min at 200-250 rpm with Certomat SII shaker. The

200 shaking frequency was selected to keep the whole amount of soil continuously moving. The
201 soil and the extract was separated with centrifuging (Sigma 4K15) at 3000 rpm.

202

203 Based on the accurate weights of soil and spiking solutions and the latter one's measured
204 activities, the expected activities (A_{spike}) were calculated. The recovery of the residue (Q) was
205 calculated from the average of the three replicate ^{14}C activity measurements of the extract
206 (A_E) taking into account the background activity of the soil sample (A_0).

$$207 \quad Q = \frac{A_E - A_0}{A_{spike}} \quad (5)$$

208 Test portions of each soil sample were spiked and extracted by different analysts 4 or 5 times.

209 The standard deviation of the recovery (S_Q) values, obtained from the repeated tests, was
210 calculated and divided by the average of recoveries to obtain the reproducibility relative
211 standard deviation ($CV_{QR} = S_Q / \bar{Q}$).

212 The combined uncertainty of extraction based on all results (n) of testing the five different soil
213 samples was calculated from the pooled variances (S_Q^2) and the grand average of recoveries
214 (\bar{Q}):

$$215 \quad CV_e = \frac{\sqrt{\frac{1}{n} \sum_{i=1}^n S_Q^2}}{\bar{Q}} \quad (6)$$

216

217

218 *Determination of the efficiency of extraction*

219

220 The efficiency of extraction was tested with four ^{14}C -labeled pesticide (carbofuran,
221 chlorfenvinphos, chlorpyriphos and p-p-DDT) and 3 different soils (V01, V02 and X65). The
222 test compounds were prepared separately in acetone containing the ^{14}C -labelled (target
223 specific activity 400 Bq/mL (24000 dpm/mL) and unlabeled standards in 0.3 $\mu\text{g/mL}$

224 concentration. The exact activities of the standard solutions were determined with LSC in
225 three replicates.

226 Twenty grams of processed soils were weighed in 250 mL round bottom flask and treated
227 with 30 mL acetone containing the standard solution at 0.05 mg/kg dry soil equivalent. Each
228 standard solution was applied to different portions of soil. The flask was fixed on rotary
229 evaporator and rotated for 15 minutes at ambient temperature to thoroughly mix the soil and
230 the solvent, then the solvent was evaporated under gentle vacuum immersing the flask into
231 water bath kept at 35°C. The dry, free flowing soil powder was transferred to 100 mL
232 centrifuge tube with screw cap. Distilled water was added until water holding capacity of the
233 soil and the container was stored in the greenhouse of the IAEA at about 25°C for 6 months.
234 The evaporated water was replaced regularly. Twelve replicates were prepared from each of
235 the soil-pesticide combinations. Untreated soils were processed similarly to fortified ones and
236 they were used to determine the background activity. They also served as blank sample for
237 validation of the optimized method.

238 The exact initial ¹⁴C activities of the fortified soils containing the incurred residues and the
239 blank soils were determined just before their extraction, as described above.

240

241 After 6 months of storage, the soil samples were extracted with either of hexane:acetone (1:1
242 v/v), acetone and ethyl acetate (EtAc). The extracting solvents were selected from those
243 which have been most frequently used for determination pesticide residues in soil and plant
244 materials. Dichloromethane was not considered in view of protection of the environment. The
245 three solvents have high and medium polarity and non-polar character. They were primarily
246 suitable for extraction of pesticides of similar polarity.

247

248 The extraction procedure, performed in 3 replicates, consisted of 4 steps:

249 1. Before extraction, 2.8 mL 0.2 mol NH₄Cl was thoroughly mixed with the 20 g soil
250 sample and let to stand for 15 minutes, then 40 mL of one of the extraction solvents was
251 added, the container was tightly closed and agitated with horizontal shaker at 200-250 rpm for
252 30 min.

253 2. The tube was centrifuged at 3000 rpm, 1-1 mL of clean extract were withdrawn and
254 mixed with 12-12 mL scintillation cocktail in LSC vials. The radioactivity was determined for
255 3 × 5 minutes with Beckman LSC counter.

256 3. The tube was agitated again with horizontal shaker, for another 30 mins (total
257 extraction time 1 hr)

258 4. Step 2 was repeated and the extraction was continued for another hour (total extraction
259 time 2 hours).

260 The radioactivity of the extract was measured after 30, 60 and 120 minutes.
261

262 Based on the results of the first series of tests, an additional extraction procedure was tested:
263 the 20 g soil was first extracted with 20 mL acetone, then 20 mL ethyl-acetate was added and
264 the agitation of the soil was continued for 30 mins (total extraction time 1 hour). The use of
265 combination of solvents was necessary, because acetone completely disintegrated the soil
266 particles which increased the efficiency of extraction and the ethyl acetate extracted non-polar
267 residues as well.

268 The extracting solvent was decanted after the end of the extraction. The soil was rinsed with
269 20 mL extracting solvent, centrifuged, the supernant solvent was decanted and the soil was
270 kept under fume hood until constant weight was reached. The radioactivity of the extracted
271 soil was determined from 500 mg portions.

272 All measured residue concentrations were expressed on dry soil basis.

273

274 *Description of optimized analytical procedure*

275

276 Azinphos-ethyl, dimethenamid, chlorfenvinphos, chlorpyrifos, oxyfluorfen, pendimethalin,

277 promertyn, propazine, terbuthylazine and terbutryn were selected as test compounds

278 representing wide range of water solubility, volatility and octanol – water partition coefficient

279 (supplementary information Table S1) like those of ¹⁴C-labeled test compounds used for

280 studying the efficiency of extraction (Table 1).

281 The untreated soils were spiked with the mixtures of standard solutions at concentration levels

282 equivalent to LOQ, 20LOQ and 100LOQ. Three different types of soil samples (X65, V01

283 and V02) were processed with adding sufficient water as described by Suszter et al. ^[41]

284 Twenty grams of processed soil was weighed into centrifuge tube, 2.8 mL 0.2 mol NH₄Cl was

285 added and mixed with the soil with a glass rod. Twenty mL acetone containing 120 ng

286 chlorpyrifos/mL internal standard (ISTD) was added, the tube was tightly closed and agitated

287 on a horizontal shaker at 200 rpm for 30 minutes. Twenty mL ethyl acetate was added to the

288 extract and the shaking was continued for 30 minutes. The soil was let to settle and the tube

289 was centrifuged at 3000 rpm. Ten mL extract (equivalent to 5 g soil) was pipetted into 20 mL

290 test tube, it was dried by shaking with 6 ± 0.1 g anhydrous sodium sulfate for 20 seconds, and

291 transferred into a 20-mL calibrated glass test tube through filter paper inserted in a glass

292 funnel. The centrifuge tube and the filter funnel was rinsed with 3×2 mL EtAc. The solvent

293 was evaporated with nitrogen to about 0.5 mL with TurboVap®VL evaporator at maximum

294 30 °C and 1 psi pressure. The final volume was adjusted exactly to 2 mL. No further cleanup

295 was employed.

296

297 The qualitative and quantitative determination of the residues was carried out with Varian

298 3800 GC equipped with nitrogen and phosphorus selective thermionic detector (TSD) and

299 PTV injector. Aglient 7890A GC with NPD was used for confirmation of the identity of the
300 analytes. The chromatographic conditions are summarized in supplementary information
301 (Table S2). The condition of the chromatographic system (resolution, phosphorus-carbon and
302 nitrogen-carbon selectivity, peak asymmetry, stability of retention times) was checked by
303 injecting the system suitability test mixture ^[38] at the beginning and at the end of each batch of
304 chromatographic analyses of sample extracts. An example chromatogram of the SST mixture
305 is shown in Figure 2. If the system suitability test indicated malfunction the appropriate
306 maintenance actions were taken.

307 The matrix effect was compensated by preparing the calibration standard solutions from the
308 blank soil extracts. The concentration of the test compounds in the calibrating standard
309 solutions ranged from 0.5 LOQ to 150 LOQ. The weighted linear regression lines, based on 5
310 concentration points, and their confidence intervals, as well as the S_{IT} values were calculated
311 with a self-made Excel template. Examples for typical calibration charts are given in Figure 3.
312 The LOD, LOQ, RT and RRT are summarized in Table 3.

313
314 The specificity of the detection was checked with injecting the standard mixture, extracts of
315 the soil and reagent blanks. The specificity was acceptable if no interfering peak was larger
316 than 0.3LOQ. Examples for the three chromatographic runs are shown in Figures 4 and 5.

317
318 The compounds were identified based on their retention times relative to chlorpyrifos ISTD
319 (RRT) (Table 3). The ratio of peak areas of analytes and ISTD were evaluated with Star 6.2
320 software for the quantitative determination.

321

322 **Results and discussion**

323

324 ***Reproducibility of extraction***

325

326 It was tested on five different soils with ^{14}C -atrazine at spike level of 0.05 mg/kg. Four or five
327 test portions from each soil was spiked and extracted with acetone on different days by 3
328 analysts. The radioactivity of each extract was measured 3 times for 5 min with Beckman
329 LSC. The recovery of ^{14}C -atrazine was calculated with equation 5 from the average recovered
330 activity and the activity of the spiking solution. The potential outliers were tested with
331 Grubb's and the homogeneity of variances with Cochran tests. ^[42] The average recoveries
332 obtained from 28 independent recovery tests performed with the 6 different soils were
333 compared with analysis of variances (ANOVA) ^[42], however none of the neighboring ones, in
334 the rank ordered values, differed more than the least significance difference (0.249) indicating
335 that there was no difference in the recoveries from different soils. Consequently, the grand
336 average of recoveries could be calculated from all recovery data. The calculation of the
337 reproducibility of extraction (CV_{eQ}) is shown with an example in Table 4. The reproducibility
338 of extraction of 5 types of soil samples was calculated by pooling the CV_e values obtained
339 with different soils. The results, indicating complete recovery (100.9%) of all tests with a
340 pooled CV_{extr} (0.0054), are summarized in Table 5.

341

342

343 ***Efficiency of extraction of incurred residues***

344

345 The ^{14}C -activities derived from incurred residues of carbofuran, chlorfenvinphos,
346 chlorpyrifos and p-p-DDT were determined in three different soils (V01, V02 and X65)
347 after 6-month storage at about 25 °C.

348 In some cases, the recovered ^{14}C activity has remained practically the same after 30 minutes.
349 However, in other cases the recovery improved (DDT–EtAc, chlorpyrifos–acetone,
350 chlorfenvinphos–EtAc) significantly if the shaking was continued for 60 minutes, and
351 remained practically constant afterwards. Acetone completely disintegrated the soil particles,
352 while in case of hexane and EtAc some clods were formed or remained in the extracted soil.
353 During the extraction with acetone-hexane mixture two phases were formed. The upper
354 hexane-acetone phase contained mainly the non-polar compounds, while the polar compounds
355 partitioned into the lower (acetone–water) phase, which is not desirable for quantitative
356 determination of residues.

357 Based on the experience gained with various solvents and extraction time, we concluded that
358 starting the extraction with acetone for 30 mins, adding EtAc and continuing the shaking for
359 another 30 minutes would give the highest recoveries for pesticide residues having wide range
360 of polarity. The optimal proportion of soil extracting solvent was not tested, but taken from
361 many publications applying the soil/solvent ratio of 1:2. Taking into account the vast
362 experience with the application of QuEChERS method ^[43] acetonitrile would be a generally
363 applicable solvent for extracting residues from soil ^[28, 30, 31] if MS detection would be used,
364 however acetonitrile cannot be directly used with N-P selective thermoionic and electron
365 capture detectors, therefore its applicability was not tested.

366

367 The remaining activities in the extracted soil was measured after the combined acetone–EtAc
368 extraction procedure. The results, summarized in Table 6, show that the proportion of ^{14}C
369 activity in the soil varied in different pesticide–soil combinations.

370 As the adsorption of pesticide residues to soil particles and their partition between
371 soil–extracting solvent depend on the combination of several physical-chemical properties of
372 soil, the number of tests and combinations did not allow detailed analyses of their

373 relationship. Nevertheless, our experiments clearly indicate the importance of testing the
374 efficiency of extraction with incurred residues as part of the validation or extension of the
375 scope of a method. The most convenient way of testing the efficiency of extraction is to use
376 ^{14}C -labeled test compounds, but they are not readily applicable in routine pesticide residue
377 laboratory. Therefore, the Codex GLs on method validation ^[33] and the FAO JMPR Manual
378 ^[39] provide some generally applicable alternative procedures.

379 It is emphasized that the measured activities include the parent compound and its metabolites
380 which contain ^{14}C . The proportion of parent compound and metabolites depends on several
381 factors such as the time between pesticide application and sampling, microbiological activity,
382 pH and physical properties of soil, therefore the concentration of the parent compound would
383 be lower than that indicated by ^{14}C measurement.

384

385 *Applicability of optimized procedure for analysis of pesticide residues in soil*

386

387 The test mixture of 10 pesticide active substances having wide range of physical-chemical
388 properties (Table S1) were used to spike 3 different types of soils at 3 concentration levels of
389 100-fold difference. The linearity of the response of components of the standard mixture was
390 established in the range of 0.5 LOQ and 150LOQ. The goodness of calibration, was
391 characterized by the coefficient of regression (R^2) and the standard deviation of relative
392 residuals (S_{rr}). Both parameters were well within the acceptable range specified by the
393 European Union Quality control guidance document. ^[34]

394 The reproducibility of determination of residues from spiked samples was tested with 5
395 replicates in each soil and spike level. The results revealed that there was no difference among
396 the reproducibility of analyses depending on the type of soil, which is in line with the findings
397 of reproducibility of extraction. The average recoveries (\bar{Q}_{L1}) and reproducibility relative

398 standard deviations (CV_Q), summarized in Table 7, are within the acceptable limits of the
399 corresponding quality control guidelines.

400

401 **CONCLUSIONS**

402

403 Use of ^{14}C -labeled compounds enabled quantifying the analytes present in the LSC cocktail
404 with an average 0.0073 relative uncertainty. Our results proved that the residues can be
405 extracted from spiked soil samples with an average CV_e of 0.54%. The major part of the
406 variability of results of residue analysis derived from the further steps (evaporation, cleanup
407 and instrumental analyses), which may require special attention if the combined relative
408 reproducibility uncertainty of the results is getting close to the upper acceptable limit of 25%.

409 ^[34] The efficiency of extraction depends on several factors and up to about 35% of total
410 residue might remain unextracted which can lead to biased results. The recovery tests
411 performed with spike samples do not reveal the required information. Therefore, the
412 efficiency of extraction should be tested with incurred residues in every case when a new
413 extraction procedure is validated or an established method is extended to a new matrix. For
414 this purpose, alternative methods ^[39] are available if the application of ^{14}C -labelled
415 compounds is not feasible.

416

417 In the lack of GS-MS/MS, LC-MS/MS instruments, the GC with specific detectors and
418 appropriate cleanup procedures can be reliably used for determination of pesticide residues
419 especially in samples of know pesticide treatment history, or in selective field surveys
420 targeted for specific pesticide residues.

421

422 **ACKNOWLEDGEMENT**

423

424 The authors are grateful for the valuable contribution of Mariana Schweikert Turcu, Philip
425 Klaus and Marivil D. Venida to the performance of laboratory measurements.

426

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FIGURE CAPTIONS

Figure 1. Structural formula indicating labeled positions of test compounds

Figure 2. Typical chromatogram of the system suitability test mixture containing EPTC, propoxur, tributyl-phosphate, dimethoate, pirimicarb, chlorpyrifos-methyl, parathion-methyl, carbaryl, chlorpyrifos, chinalphos, methidathion and phosalone.

Figure 3. Calibration charts of terbutryn on different days. The blue and red lines indicate the confidence and tolerance limits around the weighted regression line. Note the difference in R^2 and S_{TR} .

596 **Figure 4.** Example for specificity of detection of test compounds in extracts of Y97 soil (red
597 colour), spiked at 20 LOQ level (Blue colour), and reagent blank green colour)

598 **Figure 5.** Example for specificity of detection of test compounds in extracts of X65 soil
599 spiked at 20 LOQ level (red colour), bank extract (green colour), blank soil extract
600 blue colour

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TABLE CAPTIONS

Table 1. Physical properties of labeled compounds

Table 2. Summary of soil parameters

Table 3. Performance characteristics of GC determination of test compounds

Table 4. Example for the calculation of reproducibility of extraction

Table 5. Reproducibility of extraction of soil samples

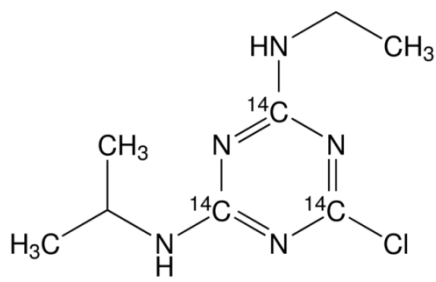
Table 6. Efficiency of extraction of ¹⁴C-labeled test compounds from soil

Table 7. Reproducibility of

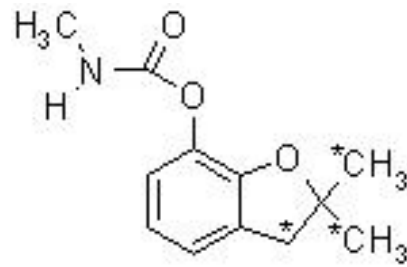
SUPPLEMENTARY MATERIALS

Table S1. Test compounds used for method validation

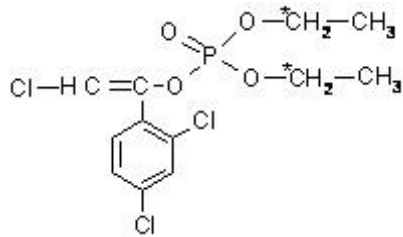
Table S2. Operation conditions of gas chromatographs



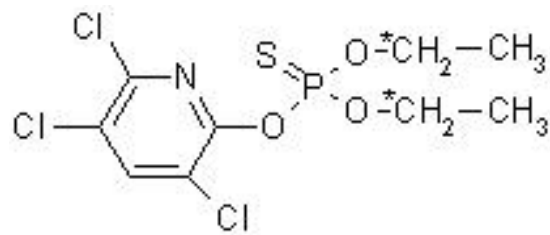
Atrazine [triazol ring- ^{14}C]



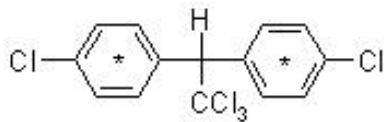
Carbofuran [(2,2-dimethyl,3)- ^{14}C]



Chlorfenvinphos [ethyl-1- ^{14}C]



Chlorpyrifos [ethyl-1- ^{14}C]

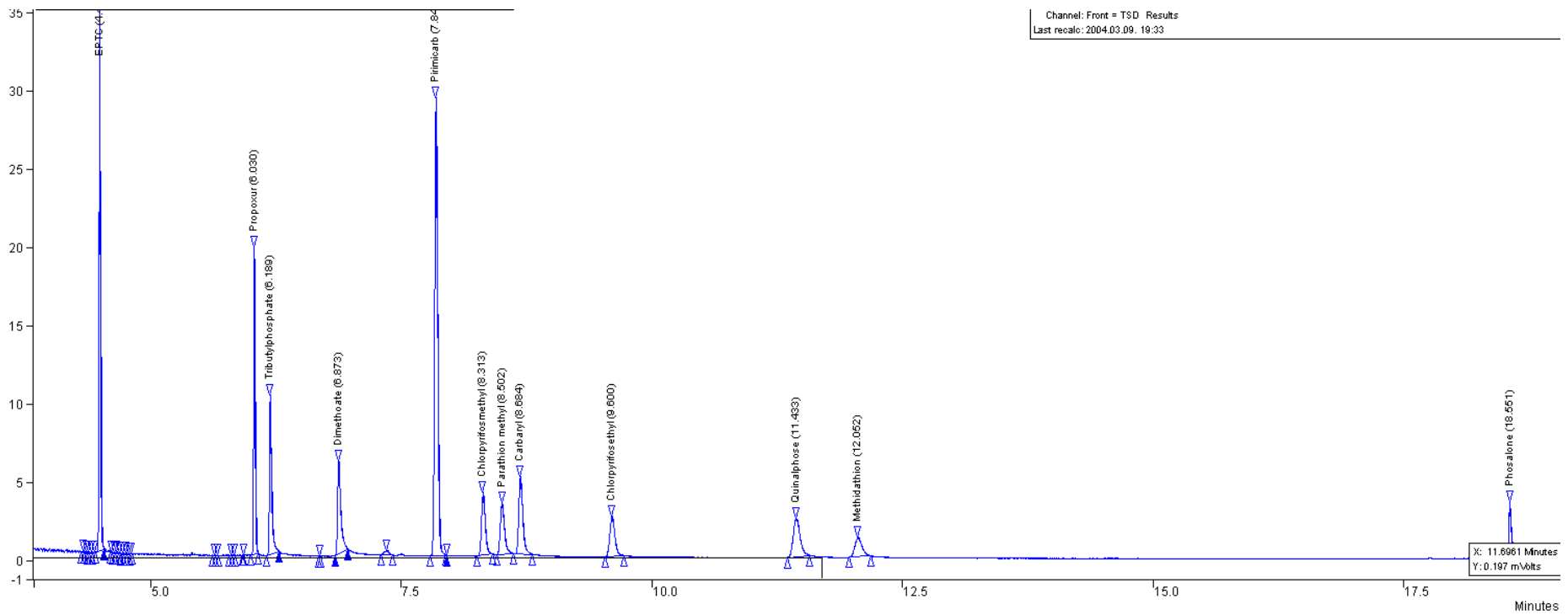


p,p' DDT, [ring-U- ^{14}C]

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630 **Figure 1.** Structural formula indicating labeled positions of test compounds

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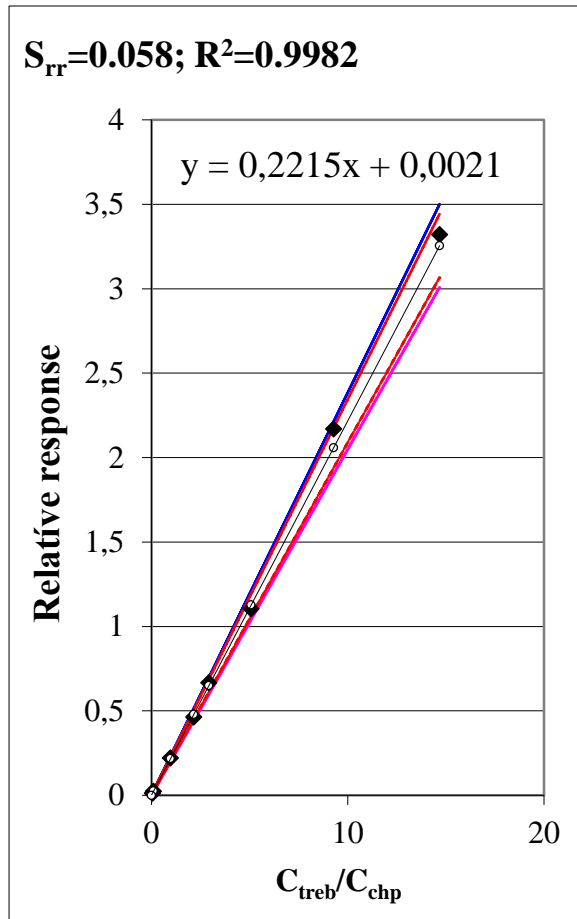
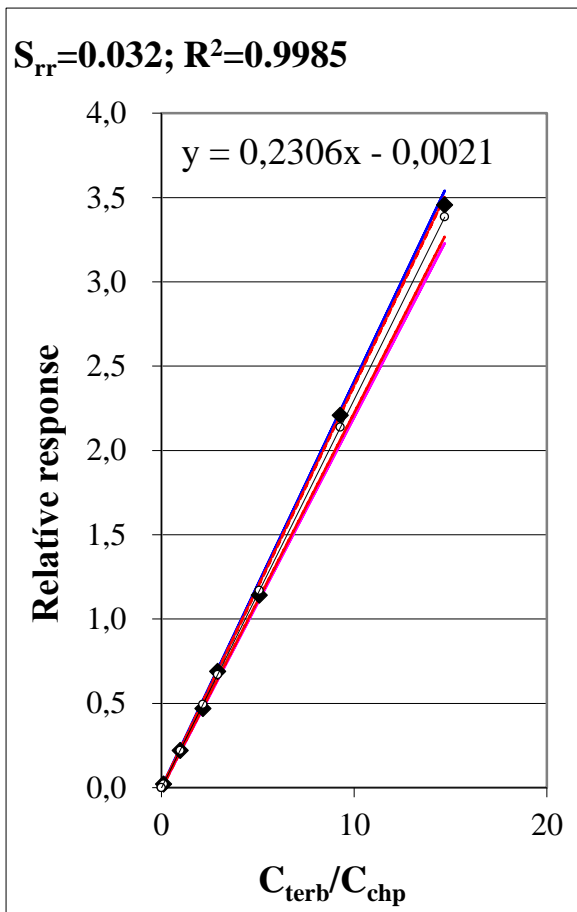
634 **Figure 2** typical chromatogram of the system suitability test mixture containing EPTC, propoxur, tributyl-phosphate, dimethoate, pirimicarb,

635 chlorpyrifos-methyl, parathion-methyl, carbaryl, chlorpyrifos, chinalphos, methidathion and phosalone.

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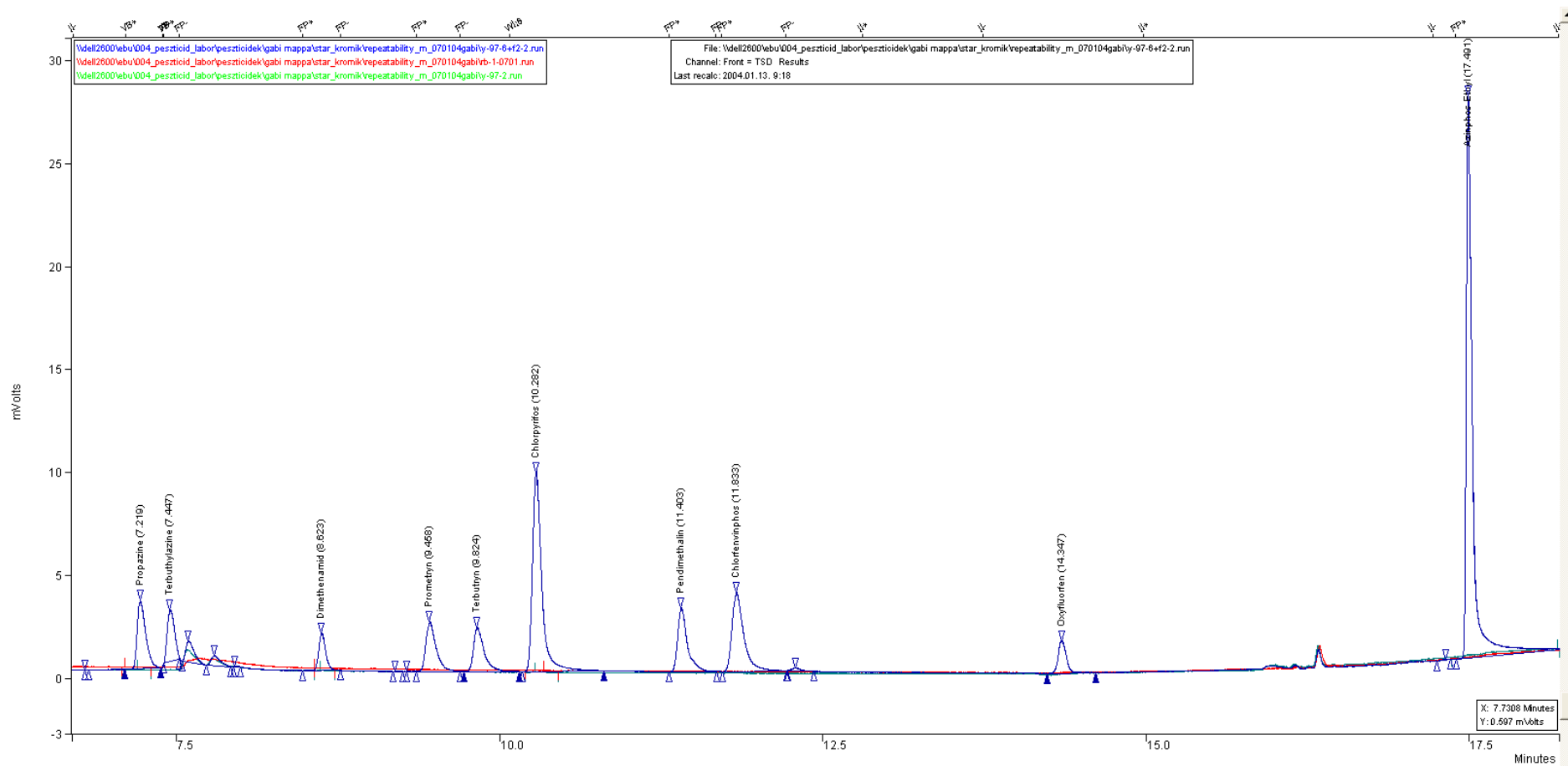


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640 **Figure 3.** Calibration charts of terbutryn on different days. The blue and red lines indicate the
 641 confidence and tolerance limits around the weighted regression line. Note the difference in R^2
 642 and S_{rr} .

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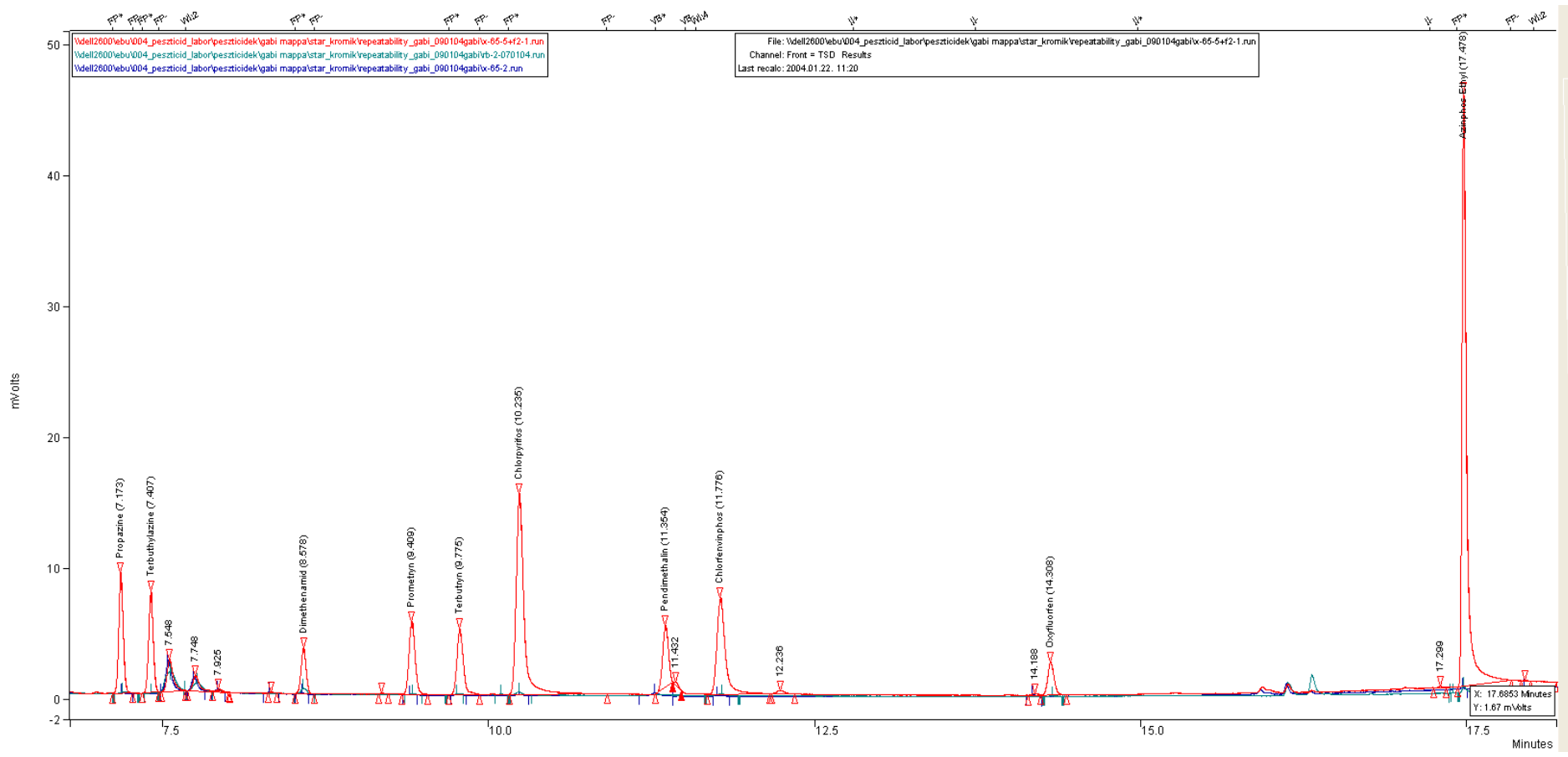
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646 **Figure 4.** Example for specificity of detection of test compounds in extracts of Y97 soil (red colour), spiked at 20 LOQ level (Blue colour), and
 647 reagent blank (green colour).

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650 **Figure 5.** Example for specificity of detection of test compounds in extracts of X65 soil spiked at 20 LOQ level (red colour), bank extract (green
 651 colour), blank soil extract (blue colour).

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653 **Table 1.** Physical properties of labeled compounds

| Name | Water solubility mg/L (20- 25 °C) | Vapour pressure mPa (25 °C) | Henry constant Pa m ³ mol ⁻¹ | log K _{ow} |
|---|--|-----------------------------------|--|---------------------|
| Atrazine, (riazol ring ¹⁴ C) | 33 | 3.85×10 ⁻² | 1.5×10 ⁻⁴ | 2.5 |
| Carbofuran, [(2,2-dimetil ,3)- ¹⁴ C] | 320 | 0.031 ×10 ⁻² | 2.4×10 ⁻⁵ | 1.52 |
| Chlorfenvinphos, [etil-1- ¹⁴ C] | 121 | 1.0 | NA | 3.85 |
| Chlorpyrifos [etil-1- ¹⁴ C] | 1.4 | 2.7 | 0.6761 | 4.7 |
| p,p'DDT, [ring-U- ¹⁴ C] | 0.0055 | 0.025* | NA | 6.91 |

654 * measured at 20 °C; NA: not available

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656 **Table 2.** Summary of soil parameters

| Site, code | dm [%] | Organic matter% | pH | Sand % | Silt % | Clay % |
|--------------------|--------|--------------------|------|--------|--------|--------|
| Hercegkút, Y97 | 86.2 | 3.14 | 6.41 | 33.8 | 41.6 | 24.6 |
| Mezőkövesd X65 | 88.4 | 2.4 | 6.8 | 36.0 | 26.5 | 37.5 |
| Olaszliszka, V02 | 94.0 | 1.89 | 6.34 | 26.3 | 26.7 | 46.9 |
| Olaszliszka, U129 | 92.9 | 2.09 | 6.37 | 37.1 | 36 | 28 |
| Hejőkeresztúr, V01 | 95.7 | 3.5 | 6.74 | 58.2 | 23.1 | 18.8 |
| Velm, W33 | 85.0 | 3.6 | 7.69 | 43 | 27.5 | 29.4 |

657 The measurements were carried at the Soil Testing Laboratory of Agricultural Service
 658 Institute of Fejér County, Hungary

659 dm: dry matter content

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663 **Table 3 .** Performance characteristics of GC determination of test compounds

| Compound | Varian GC | | Aglient 7890 | | Varian GC | |
|-----------------|-----------|------|--------------|-------|-----------|-------------|
| | RT [min] | RRT | RT | RRT | LOD pg | LOQ [mg/kg] |
| Azinphos-ethyl | 17.85 | 1.69 | 9.913 | 0.928 | 5 | 0.01 |
| Dimethenamid | 8.87 | 0.84 | 9.154 | 0.857 | 20 | 0.02 |
| Chlorfenvinphos | 12.12 | 1.15 | 10.52 | 0.985 | 10 | 0.01 |
| Chlorpyrifos | 10.56 | 1.00 | 10.68 | 1.000 | 5 | 0.01 |
| Oxyfluorfen | 14.66 | 1.39 | 8.858 | 0.829 | 50 | 0.05 |
| Pendimethalin | 11.68 | 1.11 | 9.966 | 0.933 | 20 | 0.02 |
| Promertyn | 9.71 | 0.92 | 9.25 | 0.886 | 10 | 0.01 |
| Propazine | 7.43 | 0.70 | 10.21 | 0.955 | 10 | 0.01 |
| Terbuthylazine | 7.67 | 0.73 | 10.14 | 0.950 | 10 | 0.01 |
| Terbutryn | 10.09 | 0.96 | 11.15 | 1.044 | 10 | 0.01 |

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667 **Table 4).** Example for the calculation of reproducibility of extraction

| Soil type | Activity of extracts (dpm) | | | A_{spike} | Q | CV_{eQ} |
|-----------|----------------------------|---------|---------|--------------------|-------|-----------|
| W33/A | 7 942.8 | 7 811.3 | 7 946.4 | 7 529.4 | 1.049 | |
| W33/B | 8 057.9 | 8 080.6 | 8 016.6 | 7 594.7 | 1.060 | |
| W33/C | 8 410.9 | 8 491.3 | 8 515.7 | 7 661.0 | 1.106 | |
| W33/D | 7 913.0 | 7 966.3 | 7 976.6 | 7 387.6 | 1.076 | |
| W33/E | 8 205.3 | 8 166.4 | 8 199.6 | 7 683.1 | 1.066 | |
| | | | | Average Q | 1.072 | 0.020 |

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672 **Table 5.** Reproducibility of extraction of soil samples

| | n | \bar{Q}_{rec}^a | CV_e^{673} |
|---------------|---|-------------------|---------------------|
| Y97 | 5 | 0.996 | 0.085 |
| X65 | 4 | 1.101 | 0.005 |
| W33 | 5 | 1.072 | 0.020 |
| V01 | 5 | 0.998 | 0.005 |
| V02 | 4 | 1.045 | 0.004 |
| U129 | 5 | 0.907 | 0.019 |
| Grand average | | 1.009 | 0.0054 ^b |

674 ^a: average recovery of ¹⁴C atrazine after extraction with acetone

675 ^b: calculated from pooled variances excluding two outlier values of 29

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678 **Table 6.** Efficiency of extraction of ¹⁴C-labeled test compounds from soil

| | | Percentage recovery of residues ^a | | | |
|-----|----------------|--|--------------|-----------------|-------|
| | | Carbofuran | Chlorpyrifos | Chlorfenvinphos | DDT |
| X65 | Q _E | 86.60 | 73.80 | 69.00 | 90.00 |
| | Q _S | 13.00 | 28.50 | 32.60 | 9.70 |
| | Q _T | 99.6 | 102.3 | 101.6 | 99.7 |
| V01 | Q ^E | 76.38 | 93.40 | 85.58 | 75.73 |
| | Q _S | 21.92 | 8.20 | 12.72 | 24.57 |
| | Q _T | 98.3 | 101.6 | 98.3 | 100.3 |
| V02 | Q _E | 76.27 | 83.10 | 82.84 | 64.44 |
| | Q _S | 25.8 | 16.2 | 15.1 | 34.8 |
| | Q _T | 102.1 | 99.3 | 97.9 | 99.2 |

679 ^a: calculated for dry soil as an average of results of 3 replicate tests;

680 Q_E: average ¹⁴C activity expressed as parent residue found in the Ac-EtAc combined
681 extract;

682 Q_S: average ¹⁴C activity expressed as parent residue remaining in the soil after extraction
683 with Ac-EtAc solvent system;

684 Q_T: average total ¹⁴C recovered activity expressed as parent residue.

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690 **Table 7.** Reproducibility^a of optimized method applied for soils (X65, Y97, W33)

| Spike Compound | Level 1 (LOQ) | | Level 2 (20LOQ) | | Level 3 (100LOQ) | |
|-------------------|----------------|-----------------|-----------------|-----------------|------------------|-----------------|
| | \bar{Q}_{L1} | CV _Q | \bar{Q}_{L2} | CV _Q | \bar{Q}_{L3} | CV _Q |
| Azinphos Ethyl | 91.2 | 0.15 | 87.8 | 0.13 | 81.4 | 0.10 |
| Chlorfenvinphos | 99.9 | 0.12 | 79.4 | 0.13 | 89.8 | 0.11 |
| Chlorpyrifos | 86.6 | 0.15 | 80.5 | 0.12 | 83.1 | 0.09 |
| Dimethenamid | 102.7 | 0.20 | 75.9 | 0.10 | 92.6 | 0.11 |
| Oxyfluorfen | 84.5 | 0.13 | 76.3 | 0.10 | 88.5 | 0.11 |
| Pendimethalin | 110.8 | 0.11 | 75.8 | 0.11 | 83.0 | 0.13 |
| Prometryn | 100.9 | 0.20 | 87.0 | 0.18 | 100.5 | 0.13 |
| Propazine | 96.7 | 0.15 | 77.4 | 0.14 | 87.1 | 0.11 |
| Terbutylazine | 110.0 | 0.15 | 106.6 | 0.18 | 110.7 | 0.21 |
| Terbutryn | 113.7 | 0.17 | 85.8 | 0.16 | 102.7 | 0.12 |

691 ^a: reproducibility was determined from the results of 15 tests performed by 3 analysts on
 692 different days

693 \bar{Q}_{L1} : average recovery obtained from reproducibility study;

694 CV_Q: relative standard deviation of recovery values

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| Compound | Water solubility (mg/l) (20-25 °C) | Vapour pressure mPa (25 °C) | Henry constant Pa m ³ mol ⁻¹ | log K _{ow} |
|-----------------|---------------------------------------|--------------------------------|---|---------------------|
| Azinphos-ethyl | 4-5 | 0.32* | 2.5×10 ⁻² | 3.18 |
| Dimetenamid | 1200 | 36.7 | 8.32×10 ⁻³ | 2.15 |
| Chlorfenvinphos | 121 (Z isomer) | 1.0 | NA | 3.85 (Z) |
| | 7.3 (E isomer) | | | 4.22 (E) |
| Chlorpyrifos | 1.5 | 2.7 | 6.76 ×10 ⁻¹ | 4.7 |
| Oxyfluorfen | 0.116 | 0.0267 | 8.33×10 ⁻² | 4.47 |
| Pendimethalin | 0.33 | 1.94 | 2.728 | 5.2 |
| Promertyn | 33 | 0.165 | 1.2×10 ⁻³ | 3.1 |
| Propazine | 5.0 | 0.0039* | 1.97×10 ⁻⁴ | 3.01 |
| Terbuthylazine | 9 | 0.09 | 2.3×10 ⁻³ | 3.4 |
| Terbutryn | 22 | 0.225 | 1.5×10 ⁻³ | 3.65 |

700 * measured at 20 °C; NA: not available;

702 **Table S2.** Operation conditions of gas chromatographs

| | Varian 3800 | Aglient 6890A |
|-----------------------------|---|---|
| Detector: | TSD 300 °C | NPD 320 °C |
| Injector: | PTV Mod.1079 / high performance liner 280 °C | Split/Splitless in splitless mode |
| Column: | CP-Sil-8CB Low Bleed MS (Varian) 25 m * 0.32 mm * 0.25 µm, | HP5UI 30 m- 0.25 µm |
| Retention gap | 2.5 m * 0.32 mm methyl deactivated | 2.5 m * 0.32 mm methyl deactivated |
| Temperature program: | Start: 60 °C, 0 min raise 1: 25 °C/min to 160 °C raise 2: 4 °C/min to 200 °C raise 3: 20 /min to 270 °C, hold for 3.4 min Total run time: 21 min | 80 °C 1 min 32.7 °C/min to 170 °C 0 min 10 °C/min to 310 °C 1+5 min 19 min |
| Carrier gas: | He, 4 mL/min | He 1.2 mL/min |
| Gas supply for detector: | Make up: N ₂ , 26 mL/min air: 175 mL/min hydrogen: 4.3 mL/min | Make up: N ₂ 60 mL air: 34 mL/min hydrogen: 3 mL/min |

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