

Original article

Breakdown of the species-area relationship in exotic but not in native forest patches

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ABSTRACT

We studied the pattern of bird species richness in native and exotic forest patches in Hungary. We hypothetized that species-area relationship will depend on forest naturalness, and on the habitat specialization of bird species. Therefore, we expected strong species-area relationship in native forest patches and forest bird species, and weaker relationship in exotic forest patches containing generalist species. We censused breeding passerine bird communities three times in 13 forest patches with only native tree species, and 14 with only exotic trees in Eastern Hungary in 2003. Although most bird species (92%) of the total of 41 species occurred in both exotic and native forests, the species-area relationship was significant for forest specialist, but not for generalist species in the native forests. No relationship between bird species and area was found for either species group in the forest with exotic tree species. The comparison of native versus exotic forest patches of similar sizes revealed that only large (>100 ha) native forests harbor higher bird species richness than exotic forests for the forest specialist bird species. There is no difference between small and medium forest patches and in richness of generalist species. Thus, the species-area relationship may diminish in archipelago of exotic habitat patches and/or for habitat generalist species; this result supports the warning that the extension of exotic habitats have been significantly contributing to the decline of natural community patterns. © 2008 Published by Elsevier Masson SAS.

1. Introduction

Forests are invaluable for human beings. They are important for recreation and human well being, and provide numerous ecosystem services, and are vital for the maintenance of the majority of biological diversity on Earth (Lacaze, 2000; Dirzo and Raven, 2003; Ozanne et al., 2003; Lewis, 2006). These roles need to be considered when we try to manage forests. For example, the use of wood for heating and timber in construction or paper industry needs fast growing trees, planted in a regular pattern for easy management, without other tree or scrub species. These plantations often are of non-native species and thus are inferior for recreation, and inappropriate for the maintenance of native biodiversity (Koch and Skovsgaard, 1999; Carnus et al., 2006; Gentry et al., 2006). For example, in Hungary 20% of forest cover is the black locust (Robinia pseudoacacia), originally from North America, and 15% is black pine (Pinus nigra), a European species, but not native to

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Hungary (Mátyás, 1997). There are similar patterns for other 115 116 countries (e.g., EEA, 2006).

What are the effects of exotic (i.e. introduced or non-117 native) trees and forests on native wildlife? Bird species 118 usually prefer native habitats over exotic patches (Ramos, 119 1996; Ortega et al., 2006). Native bushes are superior foraging 120 sites for birds (French et al., 2005). Nesting on exotic bushes 121 in urban parks resulted in higher rate of nest failure than 122 nesting on native bushes (Schmidt and Whelan, 1999; 123 Borgmann and Rodewald, 2004). Exotic forest plantations 124 can lure settling birds into such suboptimal habitat, where 125 nest failure is higher than in native forests (Remes, 2003).

126 Knowledge of the effects of exotic tree plantations on birds 127 at the community level is limited. Several studies aimed to 128 compare communities in native and exotic forests, using 129 simple parameters. Species richness and abundance is less 130 variable than species composition in native versus exotic 131 forests (Hausner et al., 2002; Johnson and Freedman, 2002; 132 Steverding and Leuschner, 2002; Bakker and Higgins, 2003), 133 although some studies have failed to detect any differences 134 (Donald et al., 1998; Fleishman et al., 2003; Wilson et al., 135 2006). The underlying mechanism of community differences 136 between native and exotic forests may be the predation pres-137 sure on avian broods (Barber et al., 2001; Carignan and Villard, 2002), or the selective habitat preferences of species (Lerner 138 and Stauffer, 1998). However, we have been unable to find 139 any comparison of the species-area relationship (SAR) in ex-140 otic versus native forests. This is surprising, because SAR is 141 a basic rule of ecology, stating that species richness increases 142 with increasing sample area (Rosenzweig, 1995; Báldi and 143 McCollin, 2003; Drakare et al., 2006). Thus, the important 144 question is whether a fundamental ecological relationship is 145 altered by one of the most peculiar and pervasive human 146 activities of modern times - to introduce plant species to areas 147 outside their native ranges. In this study we investigated if 148 exotic forests harbour fewer species than native patches. We 149 compared the species-area relationship of bird assemblages 150 in native versus exotic forest patches in Eastern Hungary. We 151 hypothetized that SAR will depend on forest naturalness, 152 and on the habitat specialization of species. Naturalness prob-153 ably acts via heterogeneity, which is higher in native than in 154 exotic forests (Thompson et al., 2003; Bartha et al., 2006). For-155 est patches are islands for forest specialist species, but less so 156 for generalist species, which may occur in the surrounding 157 landscape. This may mask the general species-area relation-158 ship (e.g. Magura et al., 2001). Therefore, we expect significant 159 SAR in native forest patches and forest specialist bird species, 160 and weak, if any in exotic forest patches with generalist species. Further, we expect similar species richness values in 161 small native and exotic forest patches, where no real interior 162 habitat is available. However, different species richnesses are 163 expected in large patches due to difference in the species-area 164 relationships. Such finding may have important nature con-165 servation consequences for the maintenance of biodiversity. 166

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2. Study area and methods

170 The study area is located in Eastern Hungary on the Szatmár-171 Bereg plain (Fig. 1) (N 48°05′55″, E 22°30′22″). The plain is covered by pastures and agricultural land with scattered forest patches. The natural vegetation of Central-European plains is forest, but the millennia of human activities (e.g., cattle grazing) modified it into primarily grassland areas with small forest patches (Standovár and Primack, 2001). Roughly 80% of the region is farmland. The surroundings was similar for all studied patches, grassland and/or arable fields. Native and exotic tree species were present in the patches. We chose 13 forest patches with only deciduous native tree species (Quercus robur, Carpinus betulus, Quercus petraea, Populus canescens, Populus alba, with a few Acer campestre, Salix alba and Fraxinus angustifolia), and 14 with only deciduous exotic tree species (Robinia pseudoacacia, Populus canadensis, with few individuals of Quercus rubra, Acer negundo, Salix matsudana, Amorpha fruticosa and Fraxinus pennsylvanica). The age of studied forest patches was 35–70 years. Three area categories of forest fragments were established: small (<10 ha), medium (10 < 100 ha), and large (>100 ha) (Table 1). Within size class fragments with natural tree composition and those with exotic trees were distinguished. There was no significant difference between the mean area of these two groups within area category ($t_7 = 0.289$, p = 0.781 for small forests, $t_7 =$ -0.812, p = 0.443 for medium forests, and $t_7 = 2.853$, p =0.064 for large forests). Forest patch heterogeneity was estimated by eye as percent cover of shrubs and number of nest holes within the censused areas.

Breeding bird communities were censused in the forest patches in the breeding season of 2003. Three censuses were carried out during the season (April, May and early June), only under good weather conditions (no wind or rain), from sun rise to 9 a.m. (Moskát, 1987). We applied a standard point census technique (100 m radius, 5 min census time); all birds seen or heard were recorded. Sampling effort was standardized for all patches - since many patches were only a few hectares large, one point per patch at such patches was possible to census. We distinguished between forest specialist and habitat generalist species based on literature data (Snow and Perrins, 1998), considering local conditions.

The species-area relationship was established with the most frequently used log-log transformed model (Rosenzweig, 1995), using individual patch areas (not the categories) in the calculation. The number of forest specialist and generalist bird species between the two forest types (native and exotic) with similar size were examined by repeated measure analysis of variance (ANOVA). Forest type (native or exotic) was considered as factor and the time of the counting (April, May and June) were used as repeated measures. The data were normalized by log(x + 1) transformation. When the results of the ANOVA showed that there was difference in the species richness among the forest types, this was tested by a Tukey-type multiple comparison (Sokal and Rohlf, 1981). The analyses were carried out using the SPSS-PC program (SPSS, 1999).

3. Results

Shrub cover did not differ significantly between native and exotic forest patches ($t_7 = 0.324$, p = 0.755 for small forests, $t_7 = 0.040$, p = 0.969 for medium forests, and $t_7 = 0.828$,

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Fig. 1 – Location of the study area in Eastern Hungary. Triangles represent forest patches of exotic tree species and circles native tree species. The size of the mark represents three area categories of the patches (<10 ha, 10–100 ha, >100 ha).

p = 0.435 for large forests). There were more nest holes in native forest patches compared to the exotic patches in small and large forests, but not in the medium size category ($t_7 = 3.653$, p = 0.008 for small forests, $t_7 = 1.005$, p = 0.348 for medium forests, and $t_7 = 2.543$, p = 0.038 for large forests).

Altogether 41 bird species were observed during the censuses. Most species (92%) were observed in both the native and the exotic forest patches (Appendix). The species-area relationship was significantly positive in the native forest patches for forest specialist species, and positive but not significant for generalist species (Table 2). No species-area relationship were found for bird assemblages in exotic forest patches (Table 2).

Comparing bird species richness between native and exotic forest patches of similar size classes revealed that the small

and medium patches were not significantly different for either forest specialist or generalist species. In large forests, however, bird species richness was significantly higher in native than in exotic forests for forest specialist species, but not for generalist species (Table 3). This relationship between the native and exotic forest patches seems to be stable, at least within season, because the interaction term of time * naturalness was on no occasion significant (Table 3). The time effect alone indicated significant decline in species number from April to June in the small forest patches (for both forest and generalist species), and for forest specialist species in medium forest patches. The number of generalist species did not change within season in medium-sized forest patches, and none of the groups declined over the season in the large forest patches (Table 3).

Table 1 – Characteristics of fragments and the number of observed bird species in three size groups of native and exotic forest patches in Eastern Hungary

	Small		Med	ium	Large		
	Native	Native Exotic		Exotic	Native	Exotic	
Number of fragments	5	4	4	5	4	5	
Area (ha) \pm s.d.	$\textbf{7.1} \pm \textbf{1.86}$	$\textbf{6.6} \pm \textbf{2.92}$	$\textbf{41.0} \pm \textbf{28.94}$	$\textbf{58.7} \pm \textbf{34.99}$	$\textbf{425.0} \pm \textbf{206.16}$	130.0 ± 18.71	
Number of nestholes \pm s.d.	697 ± 344	55 ± 49	465 ± 447	204 ± 336	725 ± 284	172 ± 351	
Cover of shrubs (%) \pm s.d.	44 ± 27.9	$\textbf{38} \pm \textbf{32.3}$	33.75 ± 31.5	$\textbf{33} \pm \textbf{24.4}$	46 ± 35.0	28 ± 31.2	
Number of observed forest specialist species	12.2 ± 1.10	10.5 ± 1.73	13.8 ± 4.99	14.0 ± 5.61	$\textbf{16.5} \pm \textbf{1.26}$	12.4 ± 2.19	
Number of observed generalist species	$\textbf{3.2} \pm \textbf{1.64}$	$\textbf{3.3} \pm \textbf{2.06}$	1.8 ± 1.50	$\textbf{3.0} \pm \textbf{1.58}$	$\textbf{1.8}\pm\textbf{0.96}$	4.0 ± 1.58	

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Table 2 - Equations of the species-area relationship of forest specialist birds in patches of native and exotic trees in Eastern Hungary. Equations are given for forest specialist and generalist species separately. Asterisks indicate significant relationship

348		Equation	R	n	р
349	Native forests				
350	Number of forest	Y = 0.82 + 0.09X	0.66	13	0.02*
351	specialist species				
352	Number of generalist species	Y = 0.46 - 0.01X	-0.07	13	0.84
353	Exotic forests				
354	Number of forest	Y = 0.76 + 0.09X	0.40	14	0.16
355	specialist species	$Y = 0.47 \pm 0.03X$	0 10	14	0 73
356	Number of generalist species	1 = 0.47 + 0.052	0.10	1-1	0.75

4. Discussion

360 We studied bird communities in a landscape with patches of 361 natural and exotic tree species. The majority of the landscape 362 was farmland (arable fields and grasslands), with scattered 363 forest patches, including the studied patches. Although it is 364 known that different landscape matrix influences species di-365 versity even in similar forest patches (Lindenmayer et al., 2002; Watson et al., 2005), here we had the same landscape 366 type as matrix. Therefore, we supposed that the matrix effect 367 was similar for all patches, thus, did not bias the results. 368

There was no difference in bird species composition of 369 natural versus exotic forest patches. In a study of birds in 370 tree plantations versus native shrub patches in the Negev, 371 Israel, Shochat et al. (2001) found only a small overlap of 372 bird species. Probably the difference between exotic and 373 native patches in our study was not large enough to exclude 374 bird species, only to influence the frequency of occurrence. 375 The communities showed different responses to area; the spe-376 cies-area relationship explained the number of bird species 377 (i.e. significant positive species-area relationship) in native 378 forests, but not in exotic forest patches. A similar pattern 379 was found by Shochat et al. (2001) for native scrub fragments 380 versus planted forests; bird species richness depended on area 381 Q1 in the former, but not in the latter. Santos et al. (2006) 382 compared native oak and mature pine plantations in Spain, 383 and they found similar bird species richness in native and 384 plantation archipelagoes, and different species-area relation-385 ship, just in this study. However, all species-area regression 386 models were significant in their study (Santos et al., 2006).

387 The difference between native and exotic forests was 388 more pronounced, if species were classified according to their specificity to the forest habitat. Species richness of 389 generalists was not related to forest area at all. The species 390 - area relationship of forest specialist bird species was not 391 significant in exotic forest patches, but was significant 392 (positive) in native patches. The clear difference between 393 the response of specialist and generalist (including exotic) 394 species to fragmentation was described for several taxa, in-395 cluding birds (McCollin, 1993; Germaine et al., 1998), plants 396 (Abbott, 1992; Bakker and Higgins, 2003) and invertebrates 397 (Magura et al., 2001; Ostergard and Ehrlen, 2005; Ouin 398 et al., 2006). More generally, species traits are known to 399 confound the SAR (Ewers and Didham, 2006).

There may be two potential mechanisms to explain the presence of SAR in native, but not in exotic forest patches. First, there is usually a basic difference between exotic and native forests in spatial heterogeneity; native forests are more heterogeneous (Mátyás, 1997; Thompson et al., 2003). We also found some indications of such a trend (more nest holes in native forests), although others were in our case not statistically significant (shrub cover). Habitat heterogeneity and species richness have a positive correlation (Tews et al., 2004), thus, the more heterogeneous and complex habitat structure of native forests may promote forest specialist bird species rather than generalist species (MacNally et al., 2000). Second, native forests are more island-like patches, because their complex structure is more different from the surrounding landscape, than the exotic forests with a simple structure. Therefore, forest specialist species are probably restricted to the native forest "isolates", with the subsequent SAR, while exotic forests might not function as isolates. This landscape effect may also be responsible for the absence of a SAR in some cases (Estades and Temple, 1999; Wethered and Lawes, 2003; Lövei et al., 2006).

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Increased habitat fragmentation results in the relative increase of edge habitats due to the increase of edge/core ratio. Edges are favored by generalist and/or early successional species (Harris and Silva-Lopez, 1992; Imbeau et al., 2003; Lövei et al., 2006), therefore we expected a reverse relationship for generalist species with forest area: as forest area increases the proportion of edges, and the generalists species, is decreasing. Although these were significant neither for native (R = -0.1) nor for exotic (R = 0.1) forests, the trends of the regression coefficients (Table 2) support this hypothesis.

The SAR is a fundamental rule of ecology, but it still suffers from several biases (Báldi and McCollin, 2003). Here we demonstrated the key role of native/exotic species composition, that is the quality of habitats and the specialist/generalists character of target species on the SAR (Lövei et al., 2006). An archipelago of non-native habitats, and/or non-native species in the archipelago may lead to the breakdown of the SAR.

The age of forest patches may be relevant factor in determining species richness. In this study the age of patches were 35-70 years, that none of them were old growth. Humphrey (2005) and Santos et al. (2006), for example, showed that 100-200 year old plantations already conferring substantial benefits to many species. It is clear that for the studied forest archipelago and landscape only large (>100 ha) forest patches containing native tree species can preserve natural patterns. Such patches support forest specialist bird species during the whole breeding season. We warn, however, against using only species presence in forest patches for patch evaluation: on our study nearly all the observed species (92%) were present in both native and exotic forest patches. Therefore, simple presence-absence survey may be misleading, because it can not identify the superior value of native forests. Other studies also highlighted the subordinate role of exotic trees and bushes for nesting, foraging and community assemblage (Schmidt and Whelan, 1999; Remes, 2003; Borgmann and Rodewald, 2004; French et al., 2005). This evidence support the conclusion that the expansion of exotic trees and bushes via forestry practice and gardening will harm natural patterns and the underlying processes, hence accelerating the decline of birds in isolated forests.

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Table 3 – Results of the repeated measures ANOVA for the bird species richness in deciduous forests with similar size. Naturalness of the forests (native, exotic) comprised the factor and the time of the counting (April, May and June in 2003) were used as repeated measures. Results of the Tukey test indicate which forest category differs significantly (p < 0.05) from the other; for example 'Native > Exotic' indicates that the species richness was significantly higher in the native forests than in the exotic patches

Variable	Source	SS	df	MS	F	р	Tukey posteriori test
Small forests, Number of	Within-Subjects Effects						
forest specialist species	Time	0.130	2	0.065	18.100	0.000	
	Time \times Naturalness	0.004	2	0.002	0.558	0.585	
	Error	0.050	14	0.004			
	Between-Subjects Effects						
	Naturalness	0.013	1	0.013	0.653	0.446	
	Error	0.143	7	0.020			
Small forests, Number of	Within-Subjects Effects						
generalist species	Time	0.150	2	0.075	4.830	0.025	
	Time $ imes$ Naturalness	0.043	2	0.021	1.382	0.283	
	Error	0.217	14	0.016			
	Between-Subjects Effects						
	Naturalness	0.028	1	0.028	0.294	0.604	
	Error	0.662	7	0.095			
Medium forests, Number of	Within-Subjects Effects	0.090	2	0.040	7 025	0.007	
lorest specialist species	Time × Naturalness	0.000	2	0.040	0.474	0.007	
	Frror	0.003	2 14	0.003	0.474	0.032	
	LIIOI	0.077	11	0.000			
	Between-Subjects Effects						
	Naturalness	0.006	1	0.006	0.072	0.797	
	Error	0.625	/	0.089			
Medium forests. Number of	Within-Subjects Effects						
generalist species	Time	0.006	2	0.003	0.210	0.813	
8	Time × Naturalness	0.084	2	0.042	2.751	0.098	
	Error	0.213	14	0.015			
	Deturges Cubicata Effects						
	Between-Subjects Effects	0.011	1	0.011	0.055	0.000	
	Naturainess	0.011	1	0.011	0.255	0.629	
	LIIOI	0.297	/	0.042			
Large forests, Number of	Within-Subjects Effects						
forest specialist species	Time	0.021	2	0.011	3.024	0.081	
	$\textbf{Time} \times \textbf{Naturalness}$	0.005	2	0.003	0.734	0.498	
	Error	0.050	14	0.004			
	Between-Subjects Effects						
	Naturalness	0.162	1	0.162	10.228	0.015	Native > Exotic
	Error	0.111	7	0.016			
Large forests, Number of	Within-Subjects Effects						
generalist species	Time	0.000	2	0.000	0.005	0.995	
	Time \times Naturalness	0.031	2	0.016	0.458	0.642	
	Error	0.478	14	0.034			
	Between-Subjects Effects						
	Naturalness	0.077	1	0.077	1.012	0.348	
	Error	0.536	7	0.077			

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Appendix

List of observed bird species in taxonomic order from 27 forest patches in Eastern Hungary. Total number of observations during the three censuses in 2003 is given. Snow and Perrins (1998) was followed for nomenclature. Asterisks indicate forest specialist species

orest specialist species							
		Small		Medium		Large	
		Native	Exotic	Native	Exotic	Native	Exotic
Columbidae							
Columba palumbus	Woodpigeon	0	6	2	3	0	0
Streptopelia turtur	Turtle Dove	7	7	4	5	5	6
Upupidae							
Upupa epops	Ноорое	0	0	0	0	0	2
Picidae							
Dendrocopos major*	Great Spotted Woodpecker	7	2	3	6	8	9
Dendrocopos medius*	Middle Spotted Woodpecker	0	0	0	0	1	0
Dendrocopos minor*	Lesser Spotted Woodpecker	0	0	2	1	0	0
Dryocopus martius*	Black Woodpecker	1	0	1	4	3	0
Jynx torquilla*	Wryneck	6	4	1	3	3	9
Picus viridis*	Green Woodpecker	0	0	1	0	0	1
Motacillidae	-						
Anthus trivialis	Tree Pipit	6	4	3	12	2	2
Turdidae	-						
Erithacus rubecula*	Robin	12	4	8	12	8	5
Luscinia megarhynchos*	Nightingale	10	18	4	13	3	6
Turdus merula*	Blackbird	14	9	11	12	20	8
Turdus philomelos*	Song Trush	1	0	0	2	3	0
Sylviidae	5						
Locustella fluviatilis*	River Warbler	1	3	1	2	0	9
Phylloscopus collybita*	Chiffchaff	9	3	8	9	12	8
Phylloscopus sibilatrix*	Wood Warbler	6	5	6	7	12	1
Phylloscopus trochilus*	Willow Warbler	0	1	1	0	2	0
Svlvia atricapilla*	Blackcap	22	13	18	19	18	13
Sylvia curruca	Lesser Whitethroat	0	1	0	0	0	0
Muscicapidae		°,	-	°,	Ũ	Ū	°,
Muscicana striata*	Spotted Flycatcher	0	0	1	2	2	1
Ficedula sn *	Flycather sp	0	0	0	0	1	0
Aegithalidae	rijeamer op.	Ŭ	°,	°,	Ũ	-	
Aeaithalos caudatus*	Long-tailed Tit	0	0	1	1	0	0
Paridae	Long tanea In	Ŭ	°,	-	-	Ŭ	
Parus ater*	Coal Tit	0	0	0	0	3	0
Parus caeruleus*	Blue Tit	3	2	3	6	11	4
Parus major*	Great Tit	17	8	11	12	14	9
Parus nalustris*	Marsh Tit	0	0	0	0	0	1
Sittidae		č	Ŭ	Ŭ	v	v	1
Sitta europaea*	Nuthatch	1	0	4	2	3	2
Certhiidae		-	Ŭ		2	5	2
Certhia sp.*	Treecreeper	10	8	4	8	5	10
Oriolidae		10	U	•	Ũ	5	10
Oriolus oriolus*	Golden Oriol	9	9	4	12	8	14
Laniidae		2	5	•	12	Ŭ	- 1
Lanius collurio	Red-backed Shrike	1	2	0	3	0	4
Corvidae		1	2	0	5	Ū	1
Garrulus alandarius*	lav	1	0	2	3	2	1
Sturnidae	Juy	T	0	2	5	2	1
Sturnus vulgaris	Starling	21	21	17	Q	10	48
Passaridaa	Staring	21	21	1/	0	10	40
Passar montanus	Tree Sparrow	0	0	0	1	0	0
Fussel monunus	The sparrow	0	0	0	1	0	0
Cardualis cardualis	Goldfinch	0	2	0	1	0	0
Cardualis curuuells	Groopfinch	0	2	0	1	0	2
Coccothrauston coccothrauston*	Hawfinch	3	12	0	2	۲ 11	/
Even and a soloho*	Choffmah	0	13	0	24	11	4
	Chalinch	29	23	28	24	33	3/
Serinus serinus	Serin	0	0	0	0	0	1

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		Small		Medium		Large		
		Native	Exotic	Native	Exotic	Native	Exotic	
Emberizidae								
Emberiza citrinella Miliaria calandra	Yellowhammer	5	2	2	7	2	10	
	Com Building	1	0	0	0	0	0	
A E F E R E N C E S Abbott, I., 1992. Biogeography of grasses (Poaceae) on islands of southwestern Australia. Aust. J. Ecol. 17, 289–296. Bakker, K.K., Higgins, K.F., 2003. Avian use of natural versus planted		ha Ar Harris coi	habitat, and residential development in Greater Tucson, Arizona. Ecol. Appl. 8, 680–691. Harris, L.D., Silva-Lopez, G., 1992. Forest fragmentation and the conservation of biological diversity. In: Fiedler, P.L., Jain, S.K.					
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