

Post-print version of 'Jato-Espino, D., Yiwo, E., Rodriguez-Hernandez, J.,
Canteras-Jordana, J.C. (2018). "Design and application of a Sustainable Ur-
ban Surface Rating System (SURSIST)". **Ecol. Indic.**, 93: 1253-1263.
DOI: 10.1016/j.ecolind.2018.06.022'.

AUTHORS' POST-PRINT

Design and application of a Sustainable Urban Surface Rating System (SURSIST)

Daniel Jato-Espino^{a,*}; Ebenezer Yiwo^a; Jorge Rodriguez-Hernandez^a; Juan Carlos Canteras-Jordana^b

^a GITECO Research Group, Universidad de Cantabria, Av. de los Castros 44, 39005, Santander, Spain

^b Ecology Research Group, Department of Water and Environmental Sciences and Techniques, Universidad de Cantabria, Av. de los Castros 44, 39005, Santander, Spain

E-mail addresses: jatod@unican.es (D. Jato-Espino); me.joe22@yahoo.com (E. Yiwo); rodrighj@unican.es (J. Rodriguez-Hernandez); canteraj@unican.es (J.C. Canteras-Jordana)

* Corresponding author. Tel.: +34 942203943; Fax: +34 942201703.

Abstract

Urban surfaces reflect the economic, environmental and social idiosyncrasy of cities, playing a crucial role in the sustainable development of modern civilizations. Thus, the planning and efficient management of the skin of urban areas provides an opportunity to facilitate the fulfilment of the needs of present and future generations. However, there is a lack of specific tools to evaluate the contribution of these surfaces to achieving the Sustainable Development Goals (SDGs), which is the current framework adopted by the United Nations to measure progress towards sustainability. Consequently, this paper describes the design and application of a Sustainable Urban Surface Rating System (SURSIST) aimed at producing a composite sustainability index to measure the contribution of the land cover of entire cities to meeting the SDGs. SURSIST was based on a series of indicators proposed in accordance to the targets forming the SDGs, which were processed by combining CORINE Land Cover (CLC) maps with the Analytic Hierarchy Process (AHP) and the Technique for Order of Preference by Similarity to Ideal Solution (TOPSIS). The application of SURSIST to the Spanish cities of Santander and Valencia during the time period from 1990 to 2006 demonstrated the progressive decrease in sustainability experienced by their urban surfaces due to the increased presence of impermeable covers. The replacement of a moderate part of the built-up area present in both cities in 2006 by greenspace proved to be a solution for recovering the degree of sustainability lost from 1990.

Keywords

40 Land cover; Rating system; Sustainability Indicators; Sustainable Development Goals;
41 Urban planning; Urban surfaces

42

43 **1. Introduction**

44

45 The increasing trend in world population, which is forecasted to grow up to 9 billion
46 in 2050 (UNEP, 2012), is entailing changes in the urban skin of cities as a result of human-
47 related activities. About 70% of this increased population is expected to live in urban
48 areas by 2050 (Tucci, 2001), producing alterations in land covers, which are progressively
49 evolving from natural spaces to built-up surfaces to satisfy human needs and consumption
50 habits. This process is favoring the depletion of natural resources, as well as the disposal
51 of waste required for producing goods and services to the air, land and water. In the end,
52 these circumstances endanger meeting the objectives of sustainable development, which
53 seek to fulfil the needs of future generations.

54 The principles established in the United Nations Conference on Sustainable Develop-
55 ment (Rio+20), which crystallized in the Sustainable Development Goals (SDGs), high-
56 lighted that human developments must harmonize economic growth, environmental pro-
57 tection and social inclusion (Diaz-Sarachaga et al., 2017). The consideration of sustaina-
58 bility plays a key role in balancing the variety of factors involved in the path towards
59 urban progress. The adoption of equitable approaches is crucial, since developed areas
60 are hotspots in terms of resource consumption, waste and pollution generation, environ-
61 mental degradation and social inequality. Addressing all these issues requires compre-
62 hensive urban planning and design strategies to ensure quality of life in cities.

63 In this context, the concept of sustainable rating system emerged as a set of indicators
64 to evaluate sustainability through the scoring of a series of best practices (Hart, 2006).
65 One of the main virtues of these systems is their capability to jointly assess a wide variety
66 of different indicators, regardless of the units in which they are measured (€ kg CO_2 , m^3 ,
67 linguistic scores, etc.). The application of sustainable rating systems enables determining
68 overall indices revealing the degree of sustainability achieved, for which they might use
69 a series of theoretical methods related to the branch of operations research (Singh et al.,
70 2007).

71 The development of sustainable rating systems initially focused on the assessment of
72 buildings and infrastructures. Rating tools for buildings emerged more than twenty years
73 ago (Häkkinen, 2007) in the form of the following systems: Building Research Establish-
74 ment Environmental Assessment Method (BREEAM), Comprehensive Assessment Sys-
75 tem for Building Environmental Efficiency (CASBEE) and Leadership in Energy and
76 Environmental Design (LEED). The acceptance of these tools led to the development of
77 other rating systems for evaluating major infrastructures, such as Civil Engineering En-
78 vironmental Quality (CEEQUAL), ENVISION and Infrastructure Sustainability (IS) Rat-
79 ing Tool (Diaz-Sarachaga et al., 2016). Diaz-Sarachaga et al. (2017) recently developed

80 a new Sustainable Infrastructure Rating System for Developing Countries (SIRSDEC)
81 based on the consideration of the Millennium Development Goals (MDGs), which ex-
82 pired in 2015 and were superseded by the SDGs.

83 Some of the aforementioned rating systems evolved to provide specific frameworks
84 for assessing the degree of sustainability of urban developments, especially ENVISION
85 and LEED, which has released up to three different versions related to this matter (Neigh-
86 bourhood Development, Cities and Communities). However, there is a lack of ad-hoc
87 tools aimed at measuring the contribution of urban surfaces to sustainable development.
88 Some efforts have been undertaken to explore the implications of LULC changes for sus-
89 tainability (Hassan and Nazem, 2016; Li et al., 2001; Mwavu and Witkowski, 2008) and
90 ecosystem services (García-Nieto et al., 2018; Santos et al., 2018; Zhou et al., 2017), but
91 these studies do not provide a rating system for measuring how the land cover of a city
92 contributes to the SDGs over time either.

93 As a result of all these considerations, the aim of this research was to develop a Sus-
94 sustainable Urban Surface Rating System (SURSIST) to determine the extent to which the
95 land cover configuration of a city contributed to sustainability. To this end, a list of indi-
96 cators aligned with the specific targets to be met in the SDGs was proposed to evaluate
97 urban surfaces in terms of sustainable development. The sequential processing of these
98 indicators yielded a composite score indicating the degree of sustainability of the urban
99 skin of a city. The usefulness of SURSIST was tested through two different case studies
100 corresponding to Santander and Valencia (Spain), which provided evidence of the reduc-
101 tion in sustainability experienced by the land cover of both cities over time and the op-
102 portunity represented by greenspace to bring it back.

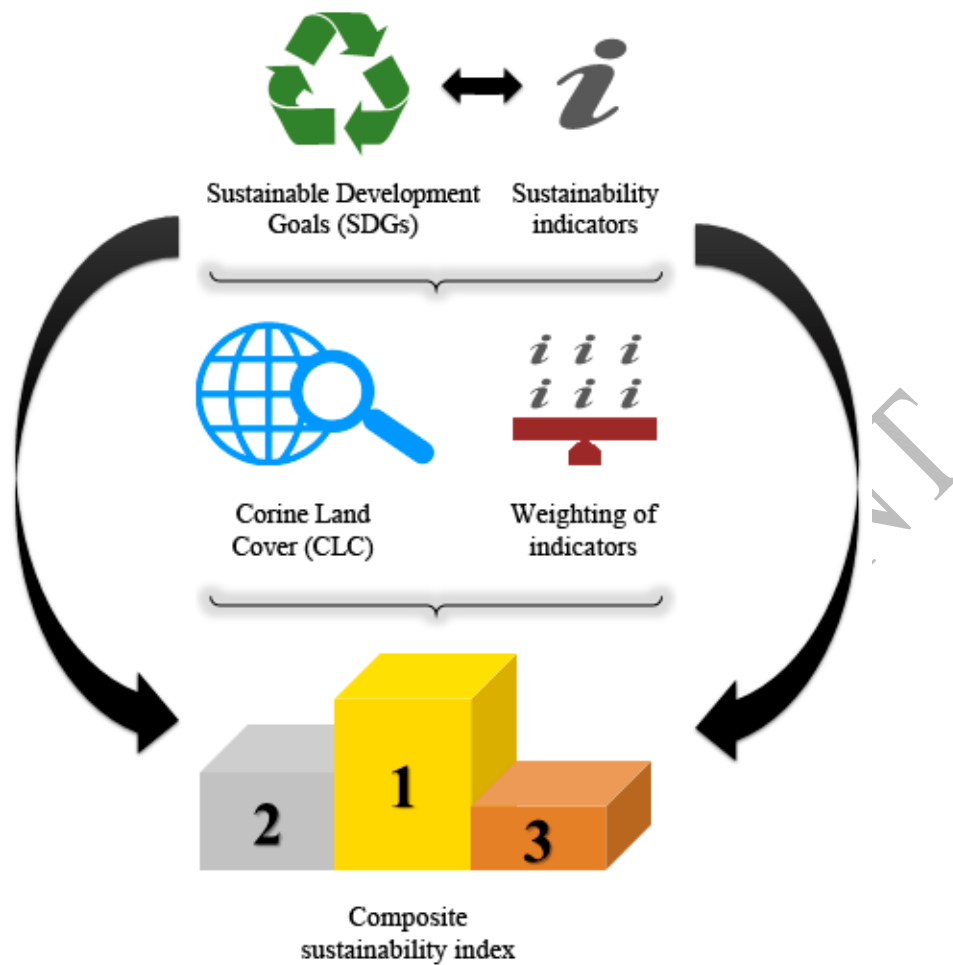
103

104 **2. Methodology**

105

106 The design of the Sustainable Urban Surface Rating System (SURSIST) was carried
107 out as a sequence of steps as represented in Fig. 1. The first phase consisted of conceiving
108 a list of indicators representing the potential contribution of urban surfaces to sustainabil-
109 ity, based on the Sustainable Development Goals (SDGs). Then, these indicators were
110 characterized and weighted with the support of the CORINE Land Cover (CLC) project
111 and the Analytic Hierarchy Process (AHP), respectively. The last step concerned the cre-
112 ation of a composite index to measure the overall contribution of the urban skin of a city
113 to achieve the SDGs using the Technique for Order of Preference by Similarity to Ideal
114 Solution (TOPSIS).

115



116
117 **Fig. 1.** Outline of the proposed methodology for the development of the Sustainable Urban Surface
118 Rating System (SURSIST)
119

120 2.1. Sustainable Development Goals (SDGs)

121
122 The SDGs, which were approved in the United Nations Conference on Sustainable
123 Development held in Rio de Janeiro in 2012 (UN-DESA, 2012), emerged to promote the
124 implementation of actions for fulfilling several objectives related to the sustainability of
125 people and the planet over the next years. The involvement of all countries and relevant
126 stakeholders is crucial for ensuring the achievement of the SDGs, whose main lines of
127 action are as follows:

- 128 • End of poverty, hunger and gender inequalities, with a focus on protecting human
129 health and ensuring the welfare of people.
- 130 • Mitigation of impacts caused by Climate Change to safeguard air, land, water and
131 biodiversity from degradation.
- 132 • Efficient management of natural resources and implementation of responsible pro-
133 duction and consumption practices.
- 134 • Promotion of the prosperity of people through peaceful and inclusive societies,
135 decent work and economic growth.

136 The New Urban Agenda adopted in the United Nations Conference on Housing and
137 Sustainable Urban Development (Habitat III) (UN, 2016) pointed out to sustainable urban
138 development as a trigger for global sustainable development. The skin of urban spaces
139 plays an essential role in the sustainability of cities, since it strongly influences their de-
140 gree of development and environmental, economic and social condition. Table 1 shows a
141 list of targets included in the United Nations SDGs, which were suggested due to their
142 close relationship to the planning of urban land cover. According to Table 1, the efficient
143 management of the surfaces of a city can contribute to achieving up to 18 targets grouped
144 into 12 SDGs.

145 This fact highlighted the need for designing a rating system for translating the links
146 between the urban skin and the SDGs into a semi-quantitative measure of the degree of
147 sustainability of a whole city in terms of its land cover. Due to its orientation to the SDGs,
148 the creation of such system can be very helpful in monitoring the fulfilment of the chal-
149 lenges posed by the United Nations to ensure leaving a better planet for future genera-
150 tions.

151

AUTHORS' POST-PRINT

Table 1. Proposed list of Sustainable Development Goals (SDGs) and targets to which urban surfaces might contribute

Sustainable Development Goal (SDG)	Target
1 No Poverty	5 By 2030, build the resilience of the poor and those in vulnerable situations and reduce their exposure and vulnerability to climate-related extreme events and other economic, social and environmental shocks and disasters
2 Zero Hunger	3 By 2030, double the agricultural productivity and incomes of small-scale food producers, in particular women, indigenous peoples, family farmers, pastoralists and fishers [...]
3 Good Health and Well-being	6 By 2020, halve the number of global deaths and injuries from road traffic accidents 9 By 2030, substantially reduce the number of deaths and illnesses from hazardous chemicals and air, water and soil pollution [...]
6 Clean Water and Sanitation	3 By 2030, improve water quality by reducing pollution, eliminating dumping and minimizing release of hazardous chemicals and materials, halving the proportion of untreated wastewater and substantially increasing recycling and safe reuse globally 4 By 2030, substantially increase water-use efficiency across all sectors and ensure sustainable withdrawals and supply of freshwater to address water scarcity and substantially reduce the number of people suffering from water scarcity
7 Affordable and Clean Energy	3 By 2030, double the global rate of improvement in energy efficiency
8 Decent Work and Economic Growth	4 Improve progressively, through 2030, global resource efficiency in consumption and production and endeavour to decouple economic growth from environmental degradation [...]
9 Industry, Innovation and Infrastructure	1 Develop quality, reliable, sustainable and resilient infrastructure, including regional and transborder infrastructure, to support economic development and human well-being, with a focus on affordable and equitable access for all
11 Sustainable Cities and Communities	2 By 2030, provide access to safe, affordable, accessible and sustainable transport systems for all, improving road safety, notably by expanding public transport, with special attention to the needs of those in vulnerable situations [...] 4 Strengthen efforts to protect and safeguard the world's cultural and natural heritage 5 By 2030, significantly reduce the number of deaths and the number of people affected and substantially decrease the direct economic losses relative to global gross domestic product caused by disasters, including water-related disasters [...] 6 By 2030, reduce the adverse per capita environmental impact of cities, including by paying special attention to air quality and municipal and other waste management
12 Responsible Consumption and Production	2 By 2030, achieve the sustainable management and efficient use of natural resources 4 By 2020, achieve the environmentally sound management of chemicals and all wastes throughout their life cycle, in accordance with agreed international frameworks, and significantly reduce their release to air, water and soil [...]
13 Climate Action	1 Strengthen resilience and adaptive capacity to climate-related hazards and natural disasters in all countries
14 Life Below Water	1 By 2025, prevent and significantly reduce marine pollution of all kinds, in particular from land-based activities, including marine debris and nutrient pollution
15 Life on Land	9 By 2020, integrate ecosystem and biodiversity values into national and local planning, development processes, poverty reduction strategies and accounts

154 2.2. Selection of sustainability indicators

155

156 The conceptualization of SURSIST was undertaken with the relationship to the SDGs
 157 as main objective, in order to include the aspects highlighted by the United Nations for
 158 achieving sustainable development. Hence, a particular set of indicators was proposed to
 159 rate the degree of sustainability of urban surfaces. Table 2 describes the contribution of
 160 each of these indicators to sustainability, as well as the SDGs and targets addressed
 161 through their consideration (see Table 1).

162

163 **Table 2.** List of sustainability indicators to assess the contribution of urban surfaces to sustainability

ID	Indicator	Contribution to sustainability	SDGs (targets)
I_1	Albedo coefficient	Attenuation of Global Warming	1 (5), 13 (1)
I_2	Impact on water pollution	Improvement of water quality	3 (9), 6 (3), 14 (1)
I_3	Threshold runoff	Flood mitigation	1 (5), 6 (4), 11 (5), 13 (1)
I_4	Energy from biomass	Energetic efficiency	7 (3), 8 (4), 12 (2)
I_5	Carbon sequestration	Air purification	3 (9), 11 (6), 12 (4)
I_6	Naturalness index	Safeguard of natural assets	11 (4)
I_7	Injury crashes	Decrease in road traffic accidents	3 (6)
-	-	Enhancement of accessibility	9 (1), 11 (2)
I_8	Species number	Generation of ecosystems	15 (9)
I_9	Agricultural land	Food production	2 (3)
I_{10}	Noise level	Noise pollution abatement	3 (9)

164

165 The Albedo coefficient represents the fraction of solar energy reflected from the Earth
 166 back into space, providing a measure of the reflectivity of the Earth surface. This concept
 167 is extremely related to Global Warming, since this phenomenon is influenced by the rate
 168 of heat exchanges between the Earth surface and the atmosphere. Some authors like Bretz
 169 and Akbari (1997) and Taha (1997) have highlighted the efficiency of land cover to re-
 170 duce surface and air temperatures near the ground. Therefore, the Albedo coefficient can
 171 have an important role in reducing the vulnerability of urban areas to climate-related haz-
 172 ards, as specified in SDGs 1 (5) and 13 (1).

173

174 Water purification is a key aspect in ensuring access to safe drinking water, which
 175 positively impacts on the health of the entire world. Providing reliable water services to
 176 the 27% of urban dwellers in the developing world who lack it is an essential long-term
 177 goal that will yield great health and economic benefits (UN-Water, 2010). If the land is
 178 not depleted by human activities and forestation is encouraged, water pollution might be
 179 reduced. Hence, communities engulfed with green environment are more likely to have
 180 clean water readily available without resorting to complex large scale treatment pro-
 181 cesses. For these reasons, the correlation of land cover to water quality parameters relates
 182 to the purification of water through SDGs 6 (3) and its subsequent positive impacts on
 the protection of living beings, represented by SDGs 3 (9) and 14 (1).

183 Built and paved surfaces hinder the infiltration of water into the soil and contribute to
184 runoff accumulation in urban areas. The increasing rate of built surfaces has resulted in
185 dramatic changes in stormwater runoff, which consequently favours the occurrence of
186 flooding phenomena in urban areas. Instead, vegetation coverage can trap runoff, acting
187 as a watershed and facilitating the percolation of water into the soil mass. Up to four
188 SDGs and targets, namely 1 (5), 6 (4), 11 (5) and 13 (1), focus on enhancing the resilience
189 to natural disasters and improving the management of water resources.

190 The fourth indicator stands for the potential generation of energy from biomass de-
191 rived from green and plant-related surfaces. Croplands, forestlands, and pasture and green
192 areas provide an opportunity to intensify and increase the production of energy per unit
193 land (Gallagher, 2006). Furthermore, the creation of energy from these surfaces is a nat-
194 ural phenomenon that does not require complex chains of processes. Consequently, this
195 indicator is also aligned with the principles of sustainable development, due to its contri-
196 bution to energy efficiency through SDGs 7 (3), 8 (4) 12 (2).

197 Carbon dioxide proportions in the atmosphere are one of the main responsible to both
198 Climate Change and particle and ozone concentrations, such that finding methods to cap-
199 ture carbon contents emitted by human-related activities becomes essential. Land
200 Use/Land Cover (LULC) change from natural to built-up areas is widely recognized as a
201 net source of greenhouse gas emissions at the global scale. Deforestation, land clearing
202 and other forms of LULC change driven by increasing population are main sources for
203 carbon. Ross et al. (2016) pointed out that LULC classes and changes should be investi-
204 gated to determine the carbon content of surfaces. SDGs 3 (9), 11 (6) and 12 (4) concern
205 the eco-friendly management of releases to air.

206 The concept of naturalness is spatiotemporal, since the identification of characteristics
207 making an item natural might vary according to the specifics of the situation (Lie, 2016).
208 Still, naturalness cannot be neglected in the quest for achieving the SDGs, because land-
209 scape pattern as a tool for ecological sustainability must be pivoted on this concept. Ac-
210 cording to Renetzeder et al. (2010), a series of landscape metrics served to assess natu-
211 ralness of Austrian and European landscapes as a proxy for their sustainability. Thus, this
212 aspect contributes to safeguard the natural heritage through SDG 11 (4).

213 Mobility is a key factor for ensuring sustainable development, since it enables people
214 to get access to products and services that are necessary for their daily lives. The integra-
215 tion of LULC policies has been recognized as an effective approach to guarantee a desired
216 level of connectivity between urban areas (Cervero, 2003). However, the intensification
217 in the development of urban surfaces has also been found to be positively correlated to
218 the occurrence of crashes (Ivan et al., 2000), due to their high degree of vehicle admissi-
219 bility. These dual considerations concern both SDG 3 (6), which seeks to reduce the num-
220 ber of injuries from vehicle accidents, and SDGs 9 (1) and 11 (2), intended to improve
221 accessibility for all.

222 Biodiversity offsets seek to balance the needs of development and nature conserva-
223 tion, such that the loss of biodiversity caused by development can be compensated by an
224 equivalent increase at a different geographic locality. If the gains from the offset equal
225 the losses of development, there is no net loss of biodiversity (Buschke, 2017). A better
226 understanding of the connections between biodiversity and the abiotic environment along
227 changing land use is crucial in developing sustainable measures to conserve biodiversity
228 under global change (Tukiainen et al., 2017). The generation of ecosystems and its impact
229 on the integration of biodiversity on planning was related to SDG 15 (9).

230 The agricultural productivity of land is essential for the socioeconomic development
231 and wellbeing of humans (Olesen and Bindi, 2002). The trends in LULC change world-
232 wide are characterized by both the increase in built-up surfaces and the reduction of arable
233 land. This decrease in cultivated land not only limits land productivity, but also affects
234 food security (Zhang et al., 2014). The results presented by Jin et al. (2015) indicated that
235 the conversion of cultivated land into built-up surfaces greatly impacted on the level and
236 spatial pattern of agricultural productivity. These aspects help meeting SDG 2 (3), since
237 the presence of agricultural land increases the rate of food productivity.

238 Estimating and controlling urban noise pollution have been identified as major chal-
239 lenges for the environmental planning and management of cities (Xie et al., 2011). The
240 LULC alterations derived from the intensification of urban development can produce a
241 series of environmental impacts, including noise pollution (King et al., 2012). Nuisances
242 like annoyance, sleep disturbance and other health effects caused by noise exposure might
243 be influenced by the degree of development of the surrounding urban surfaces. Conse-
244 quently, these considerations can be associated with SDG 3 (9), which addresses illnesses
245 from air pollution.

246

247 **2.3. Characterization of indicators**

248

249 The characterization of the indicators listed in Table 2 was carried out considering the
250 types of urban surfaces to be assessed using SURSIST. This was accomplished with the
251 support of the CORINE Land Cover (CLC) program as a geographical framework. The
252 CLC project is a scale 1:100,000 land cover database for the European Union, driven by
253 the European Environment Agency (EEA), which analyses information collected through
254 remote sensing (Stathopoulou and Cartalis, 2007). Table 3 shows the 3-level hierarchical
255 classification provided by the CLC project.

256 Hence, one of the cornerstones in the design of SURSIST consisted of crossing the
257 data included in Table 2 with those shown in Table 3, in order to determine the response
258 presented by each land cover type in terms of the proposed list of sustainability indicators.
259 In the end, this task enabled finding out which urban surfaces contributed the most to the
260 degree of sustainability of cities, facilitating the identification of opportunities to increase
261 it through the replacement of some covers by others.

Table 3. CORINE Land Cover (CLC) nomenclature (EEA, 1997)

Level 1	Level 2	Level 3
1 Artificial surfaces	11 Urban fabric	111 Continuous urban fabric
		112 Discontinuous urban fabric
	12 Industrial, commercial and transport units	121 Industrial or commercial units
		122 Road and rail networks and associated land
123 Port areas		
124 Airports		
13 Mine, dump and construction sites	131 Mineral extraction sites	
	132 Dump sites	
	133 Construction sites	
14 Artificial, non-agricultural vegetated areas	141 Green urban areas	
	142 Sport and leisure facilities	
2 Agricultural areas	21 Arable land	211 Non-irrigated arable land
		212 Permanently irrigated land
		213 Rice fields
	22 Permanent crops	221 Vineyards
		222 Fruit trees and berry plantations
		223 Olive groves
	23 Pastures	231 Pastures
	24 Heterogeneous agricultural areas	241 Annual crops associated with permanent crops
		242 Complex cultivation patterns
		243 Land principally occupied by agriculture, with significant areas of natural vegetation
		244 Agro-forestry areas
	3 Forest and semi natural areas	31 Forests
312 Coniferous forest		
313 Mixed forest		
32 Scrub and/or herbaceous vegetation associations		321 Natural grasslands
		322 Moors and heathland
		323 Sclerophyllous vegetation
		324 Transitional woodland-shrub
33 Open spaces with little or no vegetation		331 Beaches, dunes, sands
		332 Bare rocks
		333 Sparsely vegetated areas
	334 Burnt areas	
	335 Glaciers and perpetual snow	
4 Wetlands	41 Inland wetlands	411 Inland marshes
		412 Peat bogs
	42 Maritime wetlands	421 Salt marshes
		422 Salines
		423 Intertidal flats
5 Water bodies	51 Inland waters	511 Water courses
		512 Water bodies
	52 Marine waters	521 Coastal lagoons
		522 Estuaries
		523 Sea and ocean

264 2.4. Weighting of indicators

265

266 The weights of the indicators listed in Table 2 were determined according to the num-
 267 ber of SDGs and targets to which they were related. Instead of undertaking this process
 268 through direct allocation, the relative importance of the indicators was calculated using
 269 the Analytic Hierarchy Process (AHP). This course of action was adopted to give more
 270 importance to those indicators proving to be capable of addressing a higher number of the
 271 concerns posed by the SDGs. The fact that the United Nations lack a weighting system to
 272 prioritize some targets over others, as demonstrated through its SDG Index (Sachs et al.,
 273 2017), guaranteed the convenience of the proposed approach.

274 The AHP method, created by Saaty (1990), is based on quantifying a list of linguistic
 275 comparisons through a pairwise numerical scale. In this case, these qualitative compari-
 276 sons were derived from the degree of contribution of the indicators to achieve the SDGs.
 277 Table 4 represents the adapted scale proposed, which compares two indicators I_i and I_j
 278 according to the number N of targets in the SDGs they address.

279

280

Table 4. Adapted pairwise comparison scale for weighting the sustainability indicators

I_i	I_j	Importance (I_i with respect to I_j)	Numerical value
N	$N + 3$	Much less important	1/7
N	$N + 2$	Less important	1/5
N	$N + 1$	Slightly less important	1/3
N	N	Equally important	1
$N + 1$	N	Slightly more important	3
$N + 2$	N	More important	5
$N + 3$	N	Much more important	7

281

282 The p comparisons made according to this scale are arranged in the form of a matrix,
 283 such that the consistency of the pairwise comparison is measured through the maximum
 284 eigenvalue of the matrix (λ_{max}). Hence, the matrix is completely consistent when $\lambda_{max} =$
 285 p , whilst it becomes increasingly inconsistent as the eigenvalue grows according to Eq.
 286 (1):

287

$$C.R. = \frac{C.I.}{R.I.} = \frac{\lambda_{max} - p}{p - 1} < 0.1 \quad (1)$$

288

289 where $C.R.$ is the consistency ratio, $C.I.$ is the consistency index and $R.I.$ is the ran-
 290 dom consistency index, which represents an average $C.I.$ for a large number of randomly
 291 generated matrices of the same order.

292

2.5. Composite sustainability index

This step aimed at producing a composite index indicating the degree of contribution of the urban surfaces of an entire city to achieving the SDGs. This index was built from the aggregation of the ratings of indicators across the land cover types shown in Table 3. This aggregation process was carried out with the support of the Technique for Order of Preference by Similarity to Ideal Solution (TOPSIS) as a theoretical framework.

The TOPSIS method was conceived by Hwang and Yoon (1981) to find the alternative providing the best solution to problems characterized by having multiple criteria. According to the working principles of TOPSIS, this preferred alternative does not only keep the shortest distance from a positive ideal solution (A^+), but also the longest distance from a negative ideal solution (A^-). The application of the TOPSIS method is sequential and consists of the following steps:

1. Establish a rating matrix indicating the ratings r_{ij} of a set of different alternatives A_i ($i = 1, 2, \dots, m$) across the indicators I_j ($j = 1, 2, \dots, n$).

	I_1	I_2	...	I_n
A_1	r_{11}	r_{12}	...	r_{1n}
A_2	r_{21}	r_{22}	...	r_{2n}
...
A_m	r_{m1}	r_{m2}	...	r_{mn}

2. Normalize the rating matrix according to the following expressions, which determine the normalized rating n_{ij} for the alternative A_i with respect to the indicator I_j :

$$n_{ij} = \frac{r_{ij}}{\max_i z_j}, \quad \text{if } I_j \text{ is a benefit indicator} \quad (3)$$

$$n_{ij} = \frac{\min_i z_j}{r_{ij}}, \quad \text{if } I_j \text{ is a cost indicator} \quad (4)$$

The application of Eqs. (3) and (4) prevent rank reversal and enable obtaining the best alternative to the problem in “absolute” terms (García-Cascales and Lamata, 2012), since they account for the maximum and minimum achievable ratings ($\max_i z_j$ and $\min_i z_j$) in the space of alternatives. In those cases in which an indicator includes negative ratings, they were transformed into positive values using the formula proposed by Ginevicius and Podvezko (2007):

$$\bar{r}_{ij} = r_{ij} + \left| \min_i r_{ij} \right| + 1 \quad (5)$$

323

324

3. Build the weighted normalized rating matrix as follows:

325

$$v_{ij} = w_j \times n_{ij} \quad (6)$$

326

327

where v_{ij} is the weighted normalized rating for the alternative A_i with respect to the indicator I_j and w_j is the weight of the I_j , such that $\sum_{j=1}^n w_j = 1$.

328

329

330

4. Calculate the positive ideal solution (A^+) and negative ideal solution (A^-):

331

$$A^+ = w_j \quad (7)$$

332

$$A^- = w_j * \frac{n_{ij}(\min z_j)}{n_{ij}(\max z_j)} \quad (8)$$

333

334

where $n_{ij}(\min z_j)$ and $n_{ij}(\max z_j)$ correspond to the normalized ratings of $\max z_j$ and $\min z_j$.

335

336

337

5. Determine the Euclidean positive and negative distance (d_i^+ and d_i^-) from each alternative to A^+ and A^- :

338

339

$$d_i^+ = \sqrt{\sum_{j=1}^n (v_{ij} - v_j^+)^2} \quad (9)$$

340

$$d_i^- = \sqrt{\sum_{j=1}^n (v_{ij} - v_j^-)^2} \quad (10)$$

341

342

where v_j^+ and v_j^- are the positive and negative ideal weighted normalized rating for the indicator j , respectively.

343

344

345

6. Compute the relative closeness RC_i from the alternative to the ideal solution:

346

$$RC_i = \frac{d_i^-}{d_i^+ + d_i^-} \quad (11)$$

347

such that $0 \leq RC_i \leq 1$, since $d_i^+ \geq 0$ and $d_i^- \geq 0$.

3. Results and discussion: case studies of Santander and Valencia, Spain

The results of this research stemmed from the application of the Sustainable Urban Surface Rating System (SURSIST) to two case studies focused on the cities of Santander and Valencia, located in the north and east coasts of Spain (see Fig. 2). Santander and Valencia had 172,656 and 790,201 inhabitants and covered 34.76 and 134.65 km² by 2016, respectively. Both cities orient their economic activity to the tertiary sector, with about 70% and 84% of the population working on service-related activities (INE, 2016).

A notable difference between these two cities lies in their weather. Santander is characterized by an Oceanic climate consisting of temperate summers and cool winters, with a not very broad range of temperatures and moist conditions all year round. In contrast, Valencia has a dry Mediterranean climate that is translated into warm summers and temperate winters, experiencing high temperatures in August and intense rainfall events in autumn. As a result, Santander and Valencia belong to the types Cfb and Bsh in the Köppen classification, respectively (Chazarra et al., 2011).

Hence, the aim of this section was to measure the sustainability of both cities in terms of their urban surfaces. This analysis undertaken throughout a time horizon of 16 years (from 1990 to 2006), in order to get insight into the evolution of the sustainability of the urban skin of Santander and Valencia over the years. Their demographic and climatic differences further increased the interest in the results obtained, since they enabled testing the implementation of SURSIST through rather dissimilar cases.

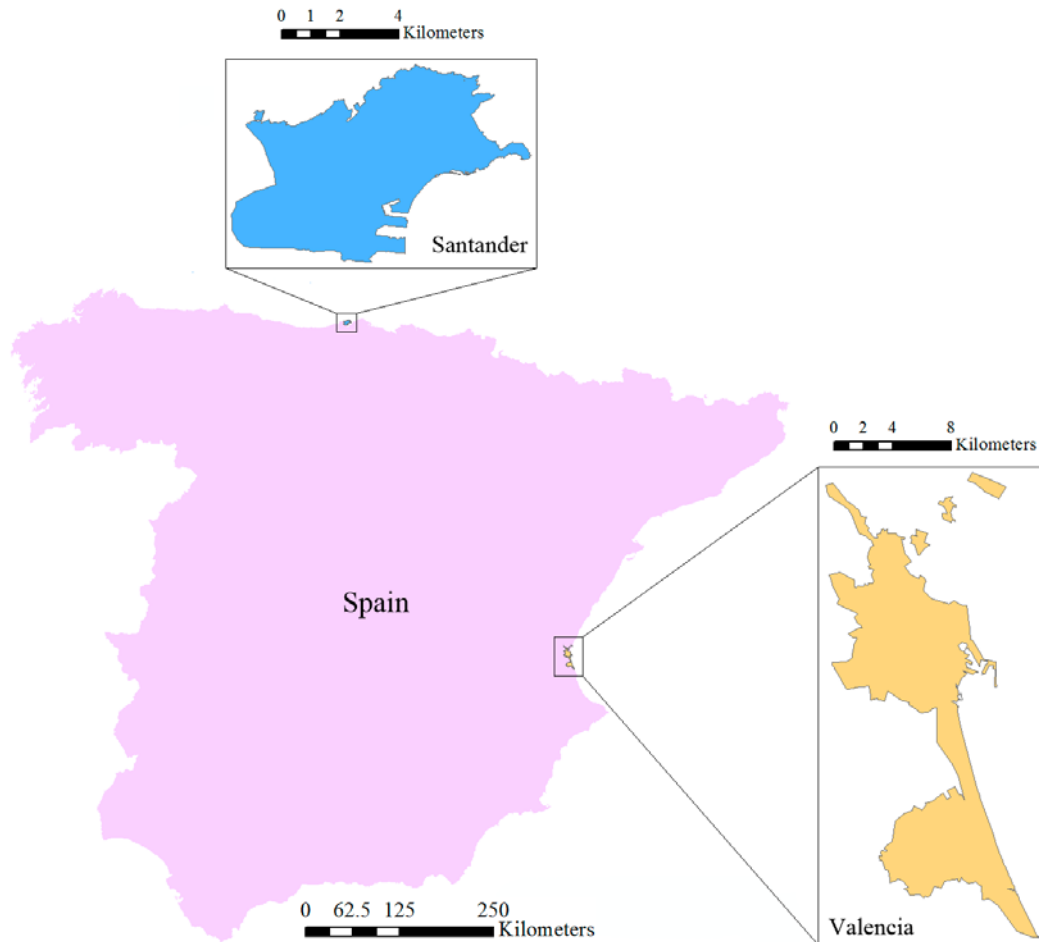


Fig. 2. Location of Santander and Valencia in relation to Spain

373

374

375

376 3.1. Characterization of indicators

377

378

379

380

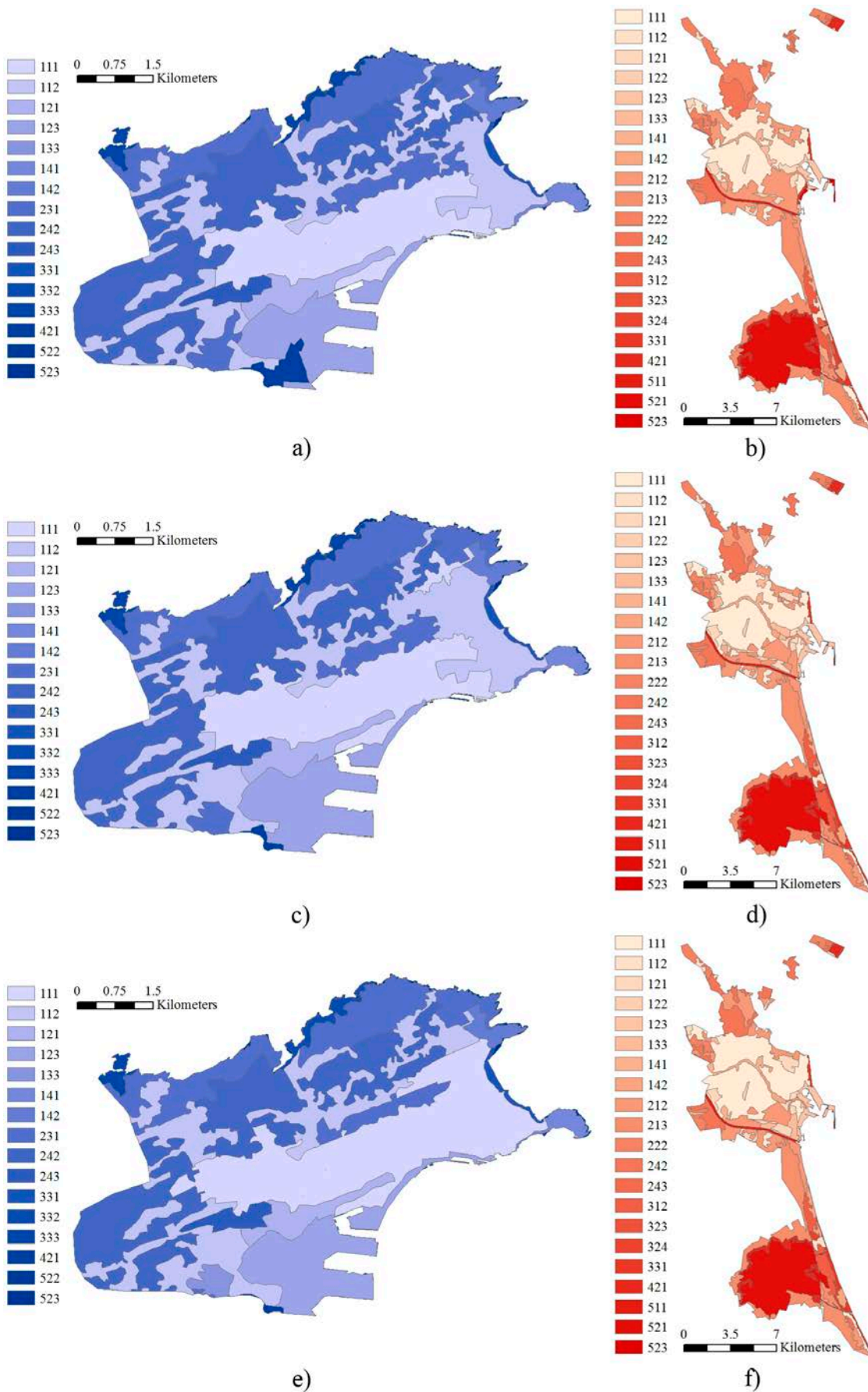
381

382

383

384

Land cover maps of Santander and Valencia were drawn with the support of ArcGIS 10.1 (ESRI, 2013), using data extracted from the CORINE Land Cover (CLC) project for the years 1990, 2000 and 2006 (IGN, 2017) according to the classification shown in Table 3. Fig. 3 illustrates these maps, which provide evidence of the increasing degree of development experienced by both cities over the years. This evolution stemmed from the growth in the area covered by built-up surfaces, especially through the categories 111 and 112 (continuous and discontinuous urban fabric).



385
386
387

Fig. 3. Land cover maps derived from the CORINE Land Cover (CLC) project (level 3) a) Santander 1990 b) Valencia 1990 c) Santander 2000 d) Valencia 2000 e) Santander 2006 f) Valencia 2006

388 Although the CLC project is also available in Spain for 2012, the time horizon of the
389 study was limited to 2006, because the 2012 map was prepared from the changes observed
390 in the Spanish Land Use and Land Cover Information System (SIOSE) between 2005 and
391 2011. Consequently, the variation in the process of production for the CLC map corre-
392 sponding to 2012 invalidated its use for comparative purposes. Furthermore, the SIOSE
393 project is a finer alternative to the CLC project (scale 1/25,000) (SIOSE, 2012), whose
394 increased level of detail hindered the collection of information to characterize the indica-
395 tors in Table 2.

396 Table 5 shows the ratings of each urban surface present in Santander and/or Valencia
397 across the indicators listed in Table 2 according to the codification of the CLC project
398 (see Table 3). I_1 corresponded to the values of the Albedo coefficient for urban surfaces
399 found in different sources (Coakley, 2003; Dobos, 2005; Wei et al., 2001). The determi-
400 nation of I_2 was undertaken from the correlation coefficients between land cover and
401 water quality presented by Wang et al. (2014). Such coefficients, which originally ranged
402 from -1 to 1, were normalized to remain between 0 and 1. The values of threshold runoff
403 (mm) needed for rating I_3 were approved by law through the Spanish Official State Ga-
404 zette (BOE, 2016). I_4 was evaluated using the report prepared by the Intergovernmental
405 Panel on Climate Change (IPCC), which provided an estimate of the global potential en-
406 ergy from biomass (EJ) associated with different land covers (IPCC, 2000). The potential
407 for carbon sequestration (Gg CO₂ eq./year) expressed by I_5 was rated based on the report
408 emitted by the Basque Government (Artetxe Arrien et al., 2014), which calculated this
409 attribute for six groups: forests, crops, pastures, wetlands, settlements and others. I_6 was
410 quantified using a range of 0 to 1 from several studies (Machado, 2004; Rojas et al., 2013;
411 Sepúlveda-Zúñiga et al., 2012) intended to produce a naturalness index for different land
412 cover categories, depending on the degree of disturbance produced by humans. I_7 con-
413 tributed twice to the SDGs, since this factor was an indicator of both traffic accidents and
414 accessibility (see Table 2). Hence, this indicator was represented by the number of crashes
415 per ha reported by Kim and Yamashita (2002). Biodiversity was measured through I_8 ,
416 which represented the mean standardized species number per m² for specific land cover
417 types, including vascular plants, moss and mollusks (Koellner and Scholz, 2008). The
418 opportunity for food production of urban surfaces was valued as a binary indicator (I_9)
419 according to the Food and Agriculture Organization of the United Nations (FAOSTAT),
420 which defines agricultural land as that devoted to crops, pastures, mowing, meadows or
421 vegetable garden (Lutzenberger et al., 2014). Finally, I_{10} provided an indicator of the
422 noise level (dBA) associated with different land covers. Its characterization was based on
423 the consideration of noise source areas and transmission loss areas, as well as the noise
424 environments they both produced (Caswell and Jakus, 1977).

425

426
427**Table 5.** Ratings for the CORINE Land Cover (CLC) codes (level 3) included in Santander and/or Valencia across the sustainability indicators

CLC	I_1	I_2	I_3	I_4	I_5	I_6	I_7	I_8	I_9	I_{10}	
111	0.10	0.0	1	0		-74	0.00	0.200	12.7	0.0	76
112	0.12	0.0	14	0		-74	0.00	0.175	18.0	0.0	55
121	0.15	0.0	4	0		-74	0.00	0.161	20.6	0.0	75
122	0.08	0.0	1	0		-74	0.00	0.208	28.6	0.0	79
123	0.15	0.0	1	0		-74	0.00	0.119	17.8	0.0	60
133	0.15	0.0	14	0		-74	0.00	0.119	15.5	0.0	40
141	0.21	0.3	23	525		193	0.25	0.044	24.6	1.0	40
142	0.15	0.0	32	0		-74	0.00	0.044	17.8	0.0	60
212	0.18	0.1	23	672		49	0.25	0.005	12.6	1.0	35
213	0.18	0.1	25	672		49	0.25	0.005	4.9	1.0	35
222	0.18	0.1	31	672		49	0.25	0.005	21.7	1.0	35
231	0.19	0.3	44	672		193	0.50	0.005	27.9	1.0	35
241	0.18	0.1	41	672		49	0.25	0.005	20.6	1.0	35
242	0.18	0.1	39	672		49	0.25	0.005	20.6	1.0	35
243	0.19	0.3	24	672		49	0.25	0.005	34.3	1.0	35
312	0.16	1.0	47	437		2,869	1.00	0.000	22.9	0.0	35
321	0.19	0.3	29	525		193	0.50	0.005	30.8	1.0	35
322	0.17	0.3	31	525		193	0.75	0.000	28.7	0.0	35
323	0.17	0.3	29	525		193	0.75	0.000	30.5	0.0	35
324	0.17	0.3	34	525		193	0.75	0.000	32.2	0.0	35
331	0.30	0.3	152	0		-1	0.50	0.000	9.1	0.0	35
332	0.17	0.3	2	0		-1	0.25	0.000	17.5	0.0	35
333	0.17	0.3	20	0		-1	0.25	0.002	33.6	0.0	35
421	0.10	0.5	2	0		0	0.75	0.000	31.5	0.0	35
511	0.10	0.5	0	0		0	0.50	0.000	14.1	0.0	35
512	0.10	0.5	0	0		0	0.50	0.000	14.1	0.0	35
521	0.10	0.5	0	0		0	0.50	0.000	14.1	0.0	35
522	0.10	0.5	0	0		0	0.50	0.000	14.1	0.0	35
523	0.07	0.5	0	0		0	0.50	0.000	14.1	0.0	35
Units	[0, 1]	Score	mm	EJ	Gg CO ₂ eq./yr.	Score	Crashes/ha	Std. species/m ²	Binary	dBA	

428

429 **3.2. Weighting of indicators**

430

431 The next step in the application of SURSIST concerned the calculation of the weights
432 of the sustainability indicators proposed in [Table 1](#). In other words, this enabled deter-
433 mining their relative importance in the evaluation of the contribution of the urban surfaces
434 forming the skin of Santander and Valencia to achieve the SDGs. This task was carried
435 out through the implementation of the Analytic Hierarchy Process (AHP).

436 Prior to the calculation of weights, I_7 was partitioned into two sub-indicators (I_{7a} and
437 I_{7b}) to represent its duality, since this indicator contribute to the SDGs through two con-
438 flicting factors, such as decrease in road traffic accidents (cost sub-indicator) and en-
439 hancement of accessibility (benefit sub-indicator) (see [Table 1](#)). Then, the indicators were

440 compared to each other based on the number of targets in the SDGs to which they were
 441 related, as schematized in Table 4. For instance, the first two indicators were related to 2
 442 (I_1) and 3 (I_2) targets, respectively, such that $N = 2$ in this case (see Table 2). Therefore,
 443 I_1 (N) was slightly less important than I_2 ($N + 1$), which corresponded to a numerical
 444 value of $1/3$, according to Table 4. The application of this process to the remaining com-
 445 binations of pairs of indicators (I_i with respect to I_j) yielded the values forming the com-
 446 parison matrix, which were used to determine the weights listed in Table 6 through the
 447 AHP method.

448

449 **Table 6.** Weights obtained for the proposed sustainability indicators according to the number of
 450 targets in the Sustainable Development Goals (SDGs) they addressed

Indicator	I_1	I_2	I_3	I_4	I_5	I_6	I_{7a}	I_{7b}	I_8	I_9	I_{10}
SDGs (targets)	2	3	4	3	3	1	1	2	1	1	1
Weight	0.071	0.146	0.277	0.146	0.146	0.029	0.029	0.071	0.029	0.029	0.029

451

452 The consistency of these weights was ensured by the value of $C.R.$ obtained (0.015)
 453 for the comparison matrix using Eq. (1). These results highlighted the importance granted
 454 by the United Nations to the management of a critically scarce resource like water and
 455 the mitigation of water-related disasters favored by Climate Change (I_3). The second level
 456 of importance corresponded to protecting natural resources, either directly by controlling
 457 the pollution that affects both air (I_5) and water (I_2) or indirectly through the search for
 458 potential sources to improve the energetic efficiency (I_4) of urban areas. Attenuating ur-
 459 ban warming (I_1) and guaranteeing an adequate degree of accessibility (I_{7b}) were the next
 460 factors in reaching higher weights. The remaining aspects under consideration (I_6 , I_{7a} , I_8 ,
 461 I_9 and I_{10}) only related to one SDG each and, therefore, received the lowest degree of
 462 importance.

463

464 3.3. Composite sustainability index

465

466 The characterization and weighting of indicators provided all the inputs required to
 467 apply the Technique for Order of Preference by Similarity to Ideal Solution (TOPSIS).
 468 Hence, the TOPSIS method was used to calculate the composite sustainability indices for
 469 Santander and Valencia in the time horizon from 1990 to 2006. The first task required to
 470 implement the TOPSIS method consisted of building the rating matrix as indicated in Eq.
 471 (2). The multiplication of the ratings obtained in Table 5 by the weighted sum of the areas
 472 covered by each type of land cover in Santander and Valencia, as illustrated in Fig. 3,
 473 yielded the rating matrix shown in Table 7. Based on the ratings achieved by both cities
 474 in 1990, 2000 and 2006 and those corresponding to single land cover types (see Table 5),
 475 the values of $\max_i z_j$ and $\min_i z_j$ included in Table 7 were suggested to establish extreme
 476 ratings that might be reached in rather sustainable and unsustainable environments.

477
478**Table 7.** Rating matrix and maximum and minimum achievable ratings for the proposed sustainability indicators

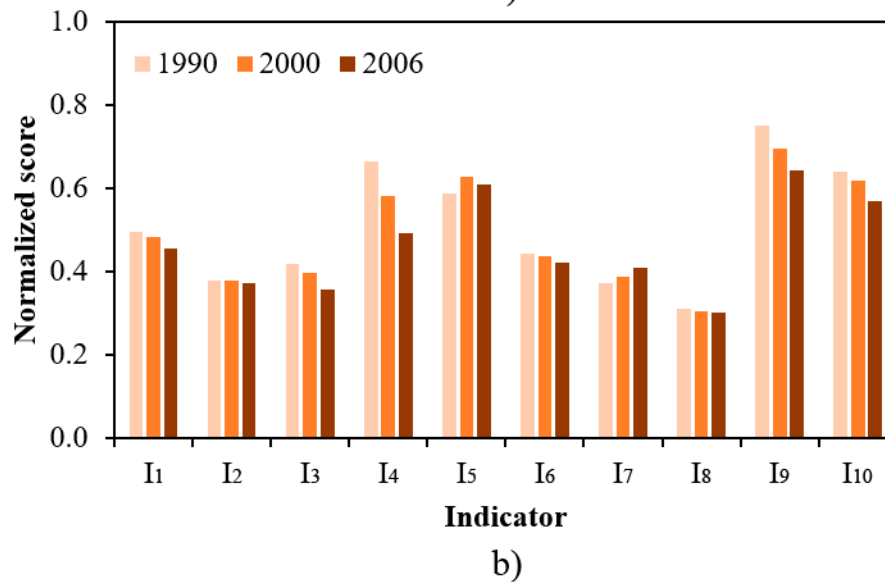
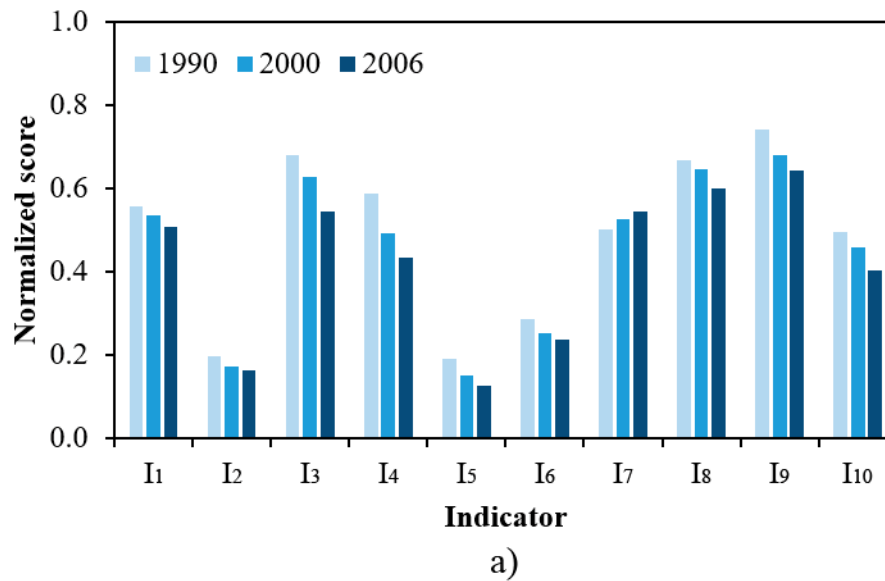
Indicator	Santander			Valencia			$\max_i z_j$	$\min_i z_j$
	1990	2000	2006	1990	2000	2006		
I_1	0.150	0.148	0.146	0.145	0.143	0.141	0.19	0.10
I_2	0.098	0.087	0.082	0.189	0.189	0.186	1.0	0.0
I_3	22.615	21.57	19.874	17.382	16.926	16.134	29	9
I_4	297.367	272.805	258.189	316.231	295.077	273.449	400	150
I_5	8.227	1.297	-3.023	78.132	84.875	81.541	150	-25
I_6	0.172	0.152	0.142	0.267	0.261	0.253	0.60	0.00
I_7	0.09	0.098	0.103	0.054	0.058	0.065	0.150	0.030
I_8	20.028	19.676	19.001	14.654	14.552	14.539	25.0	10.0
I_9	0.445	0.408	0.386	0.451	0.417	0.386	0.6	0.0
I_{10}	50.099	51.203	52.913	45.771	46.472	47.96	70	35

479

480 To provide a visual representation of the values reached by both Santander and Va-
 481 lencia across each indicator, Fig. 4 depicts their normalized ratings according to the max-
 482 imum and minimum feasible ratings shown in Table 7. The overall trend in both cases
 483 pointed out to a clear decrease in the scores achieved per indicator over the years. The
 484 only exception to this trend was I_7 , due to the improved accessibility provided by the
 485 progressive increase in built-up surfaces observed in Fig. 3.

486 The greatest differences between both cities were found in the presence of forests and
 487 semi natural areas, which was much higher in Valencia and resulted in positive impacts
 488 on the indicators related to air and water quality and naturalness (I_2 , I_5 and I_6). In contrast,
 489 Santander proved to have a better ratio of permeable zones and surfaces favoring mobility
 490 to the overall area of the city (I_3 and I_7), as well as a more adequate land configuration in
 491 terms of presence of species (I_8).

492



493

494

495

496

497

498

499

500

501

502

503

504

505

506

507

Fig. 4. Normalized scores for the sustainability indicators in a) Santander b) Valencia in 1990, 2000 and 2006

Once the original ratings in Table 7 were normalized and transformed when necessary through Eqs. (3), (4) and (5), the weighted normalized matrix was built using Eq. (6) to incorporate the degree of importance allocated to each indicator (see Table 6) into the process. Then, the calculation of the Euclidean distances d_i^+ and d_i^- as formulated in Eqs. (9) and (10) provided the inputs required for determining the relative closeness (RC_i) from the alternatives to the ideal solution through Eq. (11). Table 8 displays the values of RC_i reached by each alternative, which in this context were the combinations of cities (Santander and Valencia) and years (1990, 2000 and 2006).

As with most of the indicators in Fig. 4, the values of RC_i achieved suggested a decrease in the level of sustainability of both cities over the years. The main reason behind this situation lied in the developments experienced by Santander and Valencia throughout

508 the 16 years covered by the time period under consideration, which resulted in a progres-
 509 sive substitution of natural and green areas by built-up surfaces, as shown in Fig. 3. These
 510 changes were negative for the achievement of most of the SDGs listed in Table 2. On the
 511 one hand, the increasing presence of impermeable surfaces facilitated the occurrence of
 512 urban warming, flooding, vehicle crashes and air, noise and water pollution. On the other
 513 hand, these areas also hindered the generation of ecosystems, energy and food.

514 An important outcome to extract from Table 8 is related to the order of magnitude of
 515 the values of RC_i obtained, since they were very far from an ideal solution in sustainable
 516 terms ($RC_i = 1$). This circumstance indicated that there is much room for improvement
 517 in the design of urban land cover planning strategies toward the achievement of the SDGs.
 518 Still, the fact that several targets and indicators in Table 1 and Table 2 are in conflict to
 519 each other, in that the satisfaction of some of them results in the dissatisfaction of some
 520 others, is a challenge with which urban planners and decision-makers have to deal for
 521 conceiving solutions as comprehensive as possible.

522

523 **Table 8.** Relative closeness (RC_i) from the degree of sustainability of Santander and Valencia over
 524 the years in terms of their surfaces to the ideal solution

Year	RC_i	
	Santander	Valencia
1990	0.451	0.467
2000	0.415	0.460
2006	0.373	0.433

525

526 As a proof of potential solutions that might be implemented to enhance the sustaina-
 527 bility of the urban skin of Santander and Valencia, a strategy consisting of replacing part
 528 of the built-up cover in both cities by urban greenspace (e.g. gardens, urban crops, green
 529 roofs or grass pavers) was proposed to retrieve the existing value of RC_i at the beginning
 530 of the study period (1990). According to the nomenclature used by CLC project (see Ta-
 531 ble 3), this course of action was equivalent to substitute a portion of the surface corre-
 532 sponding to the categories 111 and 112 (continuous and discontinuous urban fabric) by
 533 141 (green urban areas), resulting in a new fictional scenario (2006*) whose differences
 534 from 2006 are indicated in Table 9.

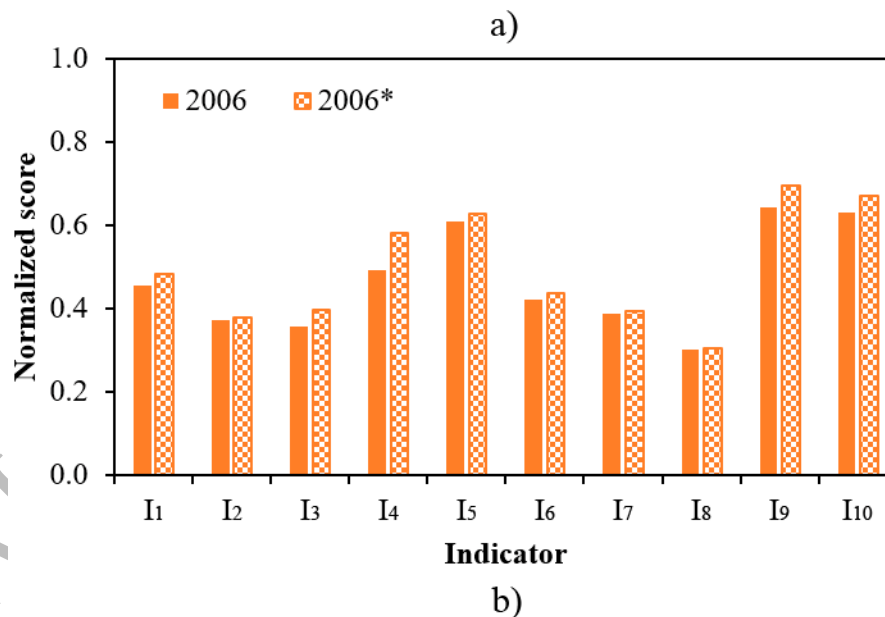
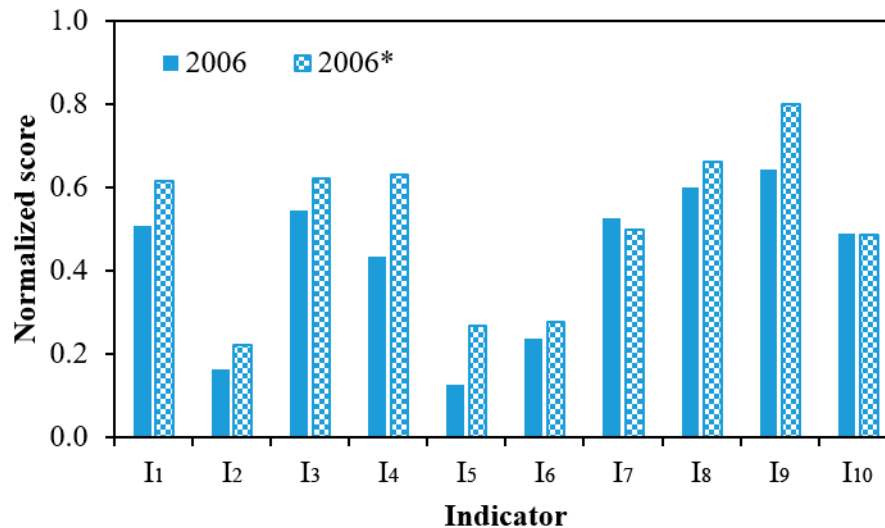
535

536 **Table 9.** Quantitative differences between the 2006 and 2006* scenarios in Santander and Valencia based
 537 on the areas covered by the CORINE Land Cover (CLC) categories 111, 112 and 141

CLC	Santander		Valencia	
	2006	2006*	2006	2006*
111	899.651	706.226	3366.206	2938.697
112	669.455	525.522	89.117	77.799
141	32.348	369.706	202.327	641.153

538

539 Therefore, the 2006* scenario incorporated these changes into the geographic config-
 540 uration of both cities in 2006, in order to increase the values of RC_i reached that year in
 541 Santander and Valencia up to those achieved in 1990 (see Table 8). The variations re-
 542 quired to restore the original conditions of sustainability resulted in the bar charts repre-
 543 sented in Fig. 5, which provide a comparison of the normalized ratings per indicator for
 544 2006 and 2006*.
 545



546
 547 **Fig. 5.** Normalized scores for the sustainability indicators in a) Santander b) Valencia in 2006 and in an
 548 hypothetical 2006 (2006*) in which a small portion of the built-up surface is replaced by greenspace
 549

550 The proposed changes produced an improvement in the scores of all indicators except
 551 I_7 , which decreased in relation to Fig. 4 due to loss of accessibility caused by the reduction
 552 in the built-up fabric. The modifications to convert the level of sustainability of the urban
 553 surfaces in Santander and Valencia in 2006 into that in 1990 required replacing 21.5%
 554 and 12.7% of their impermeable areas by greenspace. In overall terms, these changes

555 affected 9.4% and 3.2% of the whole areas occupied by Santander and Valencia, respec-
556 tively. The differences in the magnitude of the actions to be taken in both cities were
557 caused by the more pronounced decrease in sustainability experienced by Santander,
558 whose RC_i reduction ratio doubled that determined for Valencia. Still, the amount of land
559 cover involved in the proposed intervention strategies was limited in relation to the total
560 surface of both cities, which proves that small and medium-scale actions for greening the
561 skin of urban areas might have positive impacts for their sustainability

562 From the point of view of future urban planning and design of restoration and reha-
563 bilitation strategies, the nature of these green areas for enhancing the contribution of ur-
564 ban surfaces to achieving the SDGs should be strongly related to the concept of Green
565 Infrastructure (GI). GI provide a holistic opportunity to mitigate the harmful impacts of
566 both urbanization and Climate Change on cities. These techniques are multipurpose treat-
567 ment practices capable of delivering a wide variety of benefits related to sustainable de-
568 velopment, including temperature reduction, flood attenuation, runoff purification, car-
569 bon sequestration, job creation, food production and generation of both ecosystems and
570 spaces for social recreation.

571

572 **4. Conclusions**

573

574 This research conceived, developed and applied a Sustainable Urban Surface Rating
575 System (SURSIST) to measure the sustainability of the urban surfaces of an entire city
576 through the Sustainable Development Goals (SDGs) established by the United Nations.
577 SURSIST was founded on a combination of CORINE Land Cover (CLC) maps with the
578 Analytic Hierarchy Process (AHP) and the Technique for Order of Preference by Simi-
579 larity to Ideal Solution (TOPSIS). This framework consists of a series of measurable in-
580 dicators that can be easily extrapolated across European cities, by virtue of its theoretical
581 simplicity and the widespread availability of the data required. These indicators were se-
582 lected based upon their correlation to the targets addressed by the SDGs and further char-
583 acterized according to the land cover types included in the CLC project.

584 The results obtained through the application of SURSIST to the Spanish cities of San-
585 tander and Valencia ensured the usefulness of the proposed approach, demonstrating how
586 the level of sustainability of the urban skin of both cities progressively decreased with
587 time as a result of an increase in the built-up fabric, which disregarded a series of SDGs
588 related to the protection of the environment and the responsible use of natural resources.
589 The implementation of intervention strategies aimed at replacing part of the impermeable
590 cover caused by human-related activities by greenspace proved to restore the degree of
591 contribution to the SDGs provided by the urban surfaces of Santander and Valencia in
592 2006 to that in 1990, with moderate alterations in the whole area covered by both cities.
593 This retrieval of the original situation was related to the fact that the presence of green

594 areas benefited the sustainability indicators that contributed the most to the SDGs, which
595 concerned flood mitigation, air and water protection and energetic efficiency.

596 Therefore, SURSIST provides an easy-to-use rating system to evaluate the extent to
597 which the land cover of urban areas can help meeting the SDGs. Its application is intended
598 to facilitate handling of the decision-making processes required to design efficient urban
599 planning strategies to ensure the sustainability of future generations. Despite SURSIST
600 has been tested through two case studies consisting of cities with very different demog-
601 raphy, area and climate, future research should be devoted to applying this framework to
602 other cities. Hence, the proposed approach might be validated and enable identifying
603 global trends regarding the transformation of the Earth's surface provoked by human de-
604 velopment and how this affects the accomplishment of the SDGs. Possible actions to be
605 taken in the future should consider the implementation of Green Infrastructure (GI), since
606 these technologies are comprehensive measures capable of attenuating the negative im-
607 pacts of urbanization on the sustainability of urban surfaces by providing diverse eco-
608 nomic, environmental and social benefits.

609

610 **Acknowledgments**

611

612 This paper was possible thanks to the research project SUPRIS-SUReS (Ref.
613 BIA2015-65240-C2-1-R MINECO/FEDER, UE), financed by the Spanish Ministry of
614 Economy and Competitiveness with funds from the State General Budget (PGE) and the
615 European Regional Development Fund (ERDF).

616

617 **References**

618

- 619 Artetxe Arrien, A., del Hierro Cerezo, O., Pinto Tobalina, M., Gartzia Bengoetxea, N.,
620 Arias González, A., 2014. Sumideros de carbono de la Comunidad Autónoma del
621 País Vasco - Capacidad de secuestro y medidas para su promoción, 1st ed. Servicio
622 Central de Publicaciones del Gobierno Vasco, Vitoria (Spain).
- 623 BOE, 2016. Orden FOM/298/2016, de 15 de febrero, por la que se aprueba la norma 5.2
624 - IC drenaje superficial de la Instrucción de Carreteras. [WWW Document]. Off.
625 State Gaz. URL <https://www.boe.es/buscar/doc.php?id=BOE-A-2016-2405>
626 (accessed 5.29.18).
- 627 Bretz, S.E., Akbari, H., 1997. Long-term performance of high-albedo roof coatings.
628 *Energy Build.* 25, 159–167.
- 629 Buschke, F.T., 2017. Biodiversity trajectories and the time needed to achieve no net loss
630 through averted-loss biodiversity offsets. *Ecol. Modell.* 352, 54–57.
631 <https://doi.org/10.1016/j.ecolmodel.2017.02.021>
- 632 Caswell, S.J., Jakus, K., 1977. Role of Land Use Planning in Noise Control, in:
633 *Proceedings of the Conference on Metropolitan Physical Environment.* pp. 242–253.

- 634 Cervero, R., 2003. Growing smart by linking transportation and land use: Perspectives
635 from California. *Built Environ.* 29, 66–78.
636 <https://doi.org/10.2148/benv.29.1.66.53948>
- 637 Chazarra, A., Mestre Barceló, A., Pires, V., Cunha, S., Mendes, M., Neto, J., 2011. Atlas
638 Climático Ibérico - Iberian Climate Atlas. Agencia Estatal de Meteorología
639 (AEMET) and Instituto de Meteorologia de Portugal, Madrid (Spain) and Lisboa
640 (Portugal).
- 641 Coakley, J.A., 2003. Reflectance and albedo, surface, in: Holton, J.R., Curry, J.A., Pyle,
642 J.A. (Eds.), *Encyclopedia of Atmospheric Sciences*. Academic Press, Cambridge,
643 Massachusetts (U.S.), pp. 1914–1923.
- 644 Diaz-Sarachaga, J.M., Jato-Espino, D., Alsulami, B., Castro-Fresno, D., 2016. Evaluation
645 of existing sustainable infrastructure rating systems for their application in
646 developing countries. *Ecol. Indic.* 71. <https://doi.org/10.1016/j.ecolind.2016.07.033>
- 647 Diaz-Sarachaga, J.M., Jato-Espino, D., Castro-Fresno, D., 2017. Methodology for the
648 development of a new Sustainable Infrastructure Rating System for Developing
649 Countries (SIRSDEC). *Environ. Sci. Policy* 69.
650 <https://doi.org/10.1016/j.envsci.2016.12.010>
- 651 Dobos, E., 2005. Albedo, in: Lal, R. (Ed.), *Encyclopedia of Soil Science*. CRC Press,
652 New York (U.S.), pp. 64–66. <https://doi.org/10.1081/E-ESS.120014334>
- 653 EEA, 1997. CORINE Land Cover - Technical Guide, European Environmental Agency.
- 654 ESRI, 2013. ArcGIS Desktop: Release 10.1. Environmental Systems Research Institute,
655 Redlands, California (U.S.).
- 656 Gallagher, P.W., 2006. Energy Production with Biomass: What Are the Prospects?
657 *Choices* 21(1), 21–26.
- 658 García-Cascales, M.S., Lamata, M.T., 2012. On rank reversal and TOPSIS method. *Math.*
659 *Comput. Model.* 56, 123–132. <https://doi.org/10.1016/j.mcm.2011.12.022>
- 660 García-Nieto, A.P., Geijzendorffer, I.R., Baró, F., Roche, P.K., Bondeau, A., Cramer, W.,
661 2018. Impacts of urbanization around Mediterranean cities: Changes in ecosystem
662 service supply. *Ecol. Indic.* 91, 589–606.
663 <https://doi.org/10.1016/j.ecolind.2018.03.082>
- 664 Ginevicius, R., Podvezko, V., 2007. Some problems of evaluating multicriteria decision
665 methods. *Int. J. Manag. Decis. Mak.* 8, 527.
666 <https://doi.org/10.1504/IJMDM.2007.013415>
- 667 Häkkinen, T., 2007. Assessment of indicators for sustainable urban construction. *Civ.*
668 *Eng. Environ. Syst.* 24, 247–259. <https://doi.org/10.1080/10286600701315880>
- 669 Hart, M., 2006. *Guide to Sustainable Community Indicators*, 2nd ed. Hart Environmental
670 Data, North Andover, Massachusetts (U.S.).
- 671 Hassan, M.M., Nazem, M.N.I., 2016. Examination of land use/land cover changes, urban
672 growth dynamics, and environmental sustainability in Chittagong city, Bangladesh.
673 *Environ. Dev. Sustain.* 18, 697–716. <https://doi.org/10.1007/s10668-015-9672-8>

- 674 Hwang, C.-L., Yoon, K., 1981. Methods for Multiple Attribute Decision Making, in:
675 Multiple Attribute Decision Making. Lecture Notes in Economics and Mathematical
676 Systems. Springer-Verlag, New York (U.S.), pp. 58–191.
677 https://doi.org/10.1007/978-3-642-48318-9_3
- 678 IGN, 2017. Centro de Descargas del CNIG [WWW Document]. URL
679 <http://centrodedescargas.cnig.es/CentroDescargas/index.jsp> (accessed 8.1.17).
- 680 INE, 2016. España en cifras 2016. Instituto Nacional de Estadística, Madrid (Spain).
- 681 IPCC, 2000. Land-Use, Land-Use Change, and Forestry. Cambridge University Press,
682 Cambridge (U.K.).
- 683 Ivan, J.N., Wang, C., Bernardo, N.R., 2000. Explaining two-lane highway crash rates
684 using land use and hourly exposure. *Accid. Anal. Prev.* 32, 787–795.
685 [https://doi.org/10.1016/S0001-4575\(99\)00132-3](https://doi.org/10.1016/S0001-4575(99)00132-3)
- 686 Jin, G., Li, Z., Wang, Z., Chu, X., Li, Z., 2015. Impact of land-use induced changes on
687 agricultural productivity in the Huang-Huai-Hai River Basin. *Phys. Chem. Earth* 79–
688 82, 86–92. <https://doi.org/10.1016/j.pce.2015.01.005>
- 689 Kim, K., Yamashita, E., 2002. Motor vehicle crashes and land use empirical analysis from
690 Hawaii. *Transp. Res. Rec.*
- 691 King, G., Roland-Mieszkowski, M., Jason, T., Rainham, D.G., 2012. Noise levels
692 associated with urban land use. *J. Urban Heal.* 89, 1017–1030.
693 <https://doi.org/10.1007/s11524-012-9721-7>
- 694 Koellner, T., Scholz, R.W., 2008. Assessment of land use impacts on the natural
695 environment: Part 2: Generic characterization factors for local species diversity in
696 Central Europe. *Int. J. Life Cycle Assess.* 13, 32–48.
697 <https://doi.org/10.1065/lca2006.12.292.2>
- 698 Li, X., Peterson, J., Liu, G.-J., Qian, L., 2001. Assessing regional sustainability: The case
699 of land use and land cover change in the middle Yiluo catchment of the Yellow River
700 Basin, China. *Appl. Geogr.* 21, 87–106. [https://doi.org/10.1016/S0143-6228\(00\)00020-5](https://doi.org/10.1016/S0143-6228(00)00020-5)
- 701
- 702 Lie, S.A.N., 2016. *Philosophy of Nature: Rethinking naturalness*, 1st ed. Routledge,
703 London (U.K.).
- 704 Lutzenberger, A., Brillinger, M., Pott, S., 2014. Global Land-Use Analysis. *Globalands*
705 3711 93 10, 16–20.
- 706 Machado, A., 2004. An index of naturalness. *J. Nat. Conserv.* 12, 95–110.
707 <https://doi.org/10.1016/j.jnc.2003.12.002>
- 708 Mwavu, E.N., Witkowski, E.T.F., 2008. Land-use and cover changes (1988-2002) around
709 Budongo Forest Reserve, NW Uganda: Implications for forest and woodland
710 sustainability. *L. Degrad. Dev.* 19, 606–622. <https://doi.org/10.1002/ldr.869>
- 711 Olesen, J.E., Bindi, M., 2002. Consequences of climate change for European agricultural
712 productivity, land use and policy. *Eur. J. Agron.* 16, 239–262.
713 [https://doi.org/10.1016/S1161-0301\(02\)00004-7](https://doi.org/10.1016/S1161-0301(02)00004-7)

- 714 Renetzeder, C., Schindler, S., Peterseil, J., Prinz, M.A., M^ull^{er}, S., Wr^obka, T., 2010.
715 Can we measure ecological sustainability? Landscape pattern as an indicator for
716 naturalness and land use intensity at regional, national and European level. *Ecol.*
717 *Indic.* 10(1), 39–48. <https://doi.org/10.1016/j.ecolind.2009.03.017>
- 718 Rojas, C., Pino, J., Jaque, E., 2013. Strategic Environmental Assessment in Latin
719 America: A methodological proposal for urban planning in the Metropolitan Area of
720 Concepción (Chile). *Land use policy* 30, 519–527.
721 <https://doi.org/10.1016/j.landusepol.2012.04.018>
- 722 Ross, C.W., Grunwald, S., Myers, D.B., Xiong, X., 2016. Land use, land use change and
723 soil carbon sequestration in the St. Johns River Basin, Florida, USA. *Geoderma Reg.*
724 7(1), 19–28. <https://doi.org/10.1016/j.geodrs.2015.12.001>
- 725 Saaty, T.L., 1990. How to make a decision: The analytic hierarchy process. *Eur. J. Oper.*
726 *Res.* 48, 9–26. [https://doi.org/10.1016/0377-2217\(90\)90057-I](https://doi.org/10.1016/0377-2217(90)90057-I)
- 727 Sachs, J., Schmidt-Traub, G., Kroll, C., Durand-Delacre, D., Teksoz, K., 2017. SDG
728 Index and Dashboards Report 2017. Bertelsmann Stiftung and Sustainable
729 Development Solutions Network (SDSN), New York (U.S.).
- 730 Santos, A., Fernandes, M.R., Aguiar, F.C., Branco, M.R., Ferreira, M.T., 2018. Effects
731 of riverine landscape changes on pollination services: A case study on the River
732 Minho, Portugal. *Ecol. Indic.* 89, 656–666.
733 <https://doi.org/10.1016/j.ecolind.2018.02.036>
- 734 Sepúlveda-Zúñiga, E., Parra, L.E., Benítez, H.A., Rojas-Quezada, C., 2012. State of
735 vegetational naturalism and heterogeneity in pond wetlands and their effect on
736 macrolepidoptera diversity (insecta: Lepidoptera). *Shil. Rev. Lepidopterol.* 40, 155–
737 170.
- 738 Singh, R.K., Murty, H.R., Gupta, S.K., Dikshit, A.K., 2007. Development of composite
739 sustainability performance index for steel industry. *Ecol. Indic.* 7, 565–588.
740 <https://doi.org/10.1016/j.ecolind.2006.06.004>
- 741 SIOSE, 2012. Manual de Metadatos de SIOSE, Sistema de Información de Ocupación
742 del Suelo en España.
- 743 Stathopoulou, M., Cartalis, C., 2007. Daytime urban heat islands from Landsat ETM+
744 and Corine land cover data: An application to major cities in Greece. *Sol. Energy*
745 81, 358–368. <https://doi.org/10.1016/j.solener.2006.06.014>
- 746 Taha, H., 1997. Urban climates and heat islands: Albedo, evapotranspiration, and
747 anthropogenic heat. *Energy Build.* 25(2), 99–103. [https://doi.org/10.1016/S0378-7788\(96\)00999-1](https://doi.org/10.1016/S0378-7788(96)00999-1)
- 749 Tucci, C.E.M., 2001. Urban drainage in specific climates. United Nations Educational,
750 Scientific and Cultural Organization (UNESCO), Paris (France).
- 751 Tukiainen, H., Alahuhta, J., Field, R., Ala-Hulkko, T., Lampinen, R., Hjort, J., 2017.
752 Spatial relationship between biodiversity and geodiversity across a gradient of land-
753 use intensity in high-latitude landscapes. *Landsc. Ecol.* 32, 1049–1063.

- 754 <https://doi.org/10.1007/s10980-017-0508-9>
- 755 UN, 2016. United Nations Conference on Housing and Sustainable Urban Development
756 (Habitat III), in: United Nations. pp. 3–25.
- 757 UN-DESA, 2012. United Nations Conference on Sustainable Development, Rio+20
758 [WWW Document]. URL <https://sustainabledevelopment.un.org/rio20> (accessed
759 8.1.17).
- 760 UN-Water, 2010. Combating waterborne disease at the household level: The International
761 Network to Promote Household Water Treatment and Safe Storage, UN-Water
762 Decade Programme on Advocacy and Communication (UNW-DPAC). World
763 Health Organization, Geneva (Switzerland).
- 764 UNEP, 2012. Greening the Economy Through Life Cycle Thinking - Ten Years of the
765 UNEP/SETAC Life Cycle Initiative, United Nations Environmental Programme.
- 766 Wang, G., Yinglan, A., Xu, Z., Zhang, S., 2014. The influence of land use patterns on
767 water quality at multiple spatial scales in a river system. *Hydrol. Process.* 28, 5259–
768 5272. <https://doi.org/10.1002/hyp.10017>
- 769 Wei, X., Hahmann, A.N., Dickinson, R.E., Yang, Z.-L., Zeng, X., Schaudt, K.J., Schaaf,
770 C.B., Strugnell, N., 2001. Comparison of albedos computed by land surface models
771 and evaluation against remotely sensed data. *J. Geophys. Res. Atmos.* 106, 20687–
772 20702. <https://doi.org/10.1029/2001JD900218>
- 773 Xie, D., Liu, Y., Chen, J., 2011. Mapping Urban environmental noise: A land use
774 regression method. *Environ. Sci. Technol.* 45, 7358–7364.
775 <https://doi.org/10.1021/es200785x>
- 776 Zhang, Q., Wallace, J., Deng, X., Seto, K.C., 2014. Central versus local states: Which
777 matters more in affecting China's urban growth? *Land use policy* 38, 487–496.
778 <https://doi.org/10.1016/j.landusepol.2013.12.015>
- 779 Zhou, X.-Y., Lei, K., Meng, W., 2017. An approach of habitat degradation assessment
780 for characterization on coastal habitat conservation tendency. *Sci. Total Environ.*
781 593–594, 618–623. <https://doi.org/10.1016/j.scitotenv.2017.03.212>