

## ecoinvent: Materials and Agriculture

# Developments in Wood and Packaging Materials Life Cycle Inventories in ecoinvent

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### Abstract

**Goal, Scope and Background.** This paper gives an overview on how the wood and packaging material production is inventoried in ecoinvent. Packaging materials have been a very important topic in the area of Life Cycle Assessment for more than twenty years. Wood is the most important renewable material and regenerative fuel used worldwide, and an important raw material for paper / board. Several methodological problems arising when inventorying wood for material and energetic uses in a generic database are discussed in more detail. Within the ecoinvent project, the Swiss data base for life cycle inventory data, two reports are dedicated to these two important topics – report No. 9 for wood and report No. 11 for packaging materials.

**Methods.** The whole wood chain has been modeled in a consistent way. This allows one to use this data for LCAs of building materials, bioenergy or paper production. The data represent average technologies used in Central Europe in the year 2000. A revenue-based co-product allocation approach is used for the different outputs. Correction factors are introduced for the consistent modeling of mass-based, material inherent wood properties such as solar energy, carbon uptake and land use. For packaging materials, the datasets represent European average data for the most often used materials as well as specific datasets for the production of actual packaging boxes and containers.

**Results and Discussion.** For wood, revenue-based allocation and the use of the correction factors for mass-related wood properties are shown and explained. For packaging materials, the importance of the raw material wood to the total load is shown. Furthermore trends in the data inventories for board packaging materials over the last two decades are discussed: mainly due to the increased comprehensiveness of the data, higher cumulative emissions can be observed.

**Conclusion.** For wood, the database ecoinvent provides consistent datasets for the entire chain from forestry to intermediate products such as timber, different types of wood-based boards, chips, pellets, etc. For packaging materials, the number of datasets of basic materials has been extended. A modular concept for actual packaging container datasets allows the user an easy modeling of various types of packaging containers/boxes. In the area of paper and board, a comprehensive database for the production of various types of pulp, paper and board is provided, which is representative for the average European production situation.

**Outlook.** Since wood is only limited and representative data for Europe is therefore not included, an update in the near future would be reasonable. Possible further extensions in the future could include various, final wooden products. For the data on paper/board, different levels of quality are observed, requiring a selective up-date of these data. Future extensions could include datasets for the import of pulp from overseas – especially from South America and Canada.

**Keywords:** Board; ecoinvent; life-cycle inventories; packaging materials; paper; revenue-based allocation; Switzerland; wood

### Introduction

The packaging area has been growing enormously since the 1960s. Due to the fact that packaging material has already fulfilled its function at the beginning of the use phase of the respective product turning into waste, the environmental relevance of packaging materials got very important. Thus, packaging materials were one of the main focus in the early years of LCA (BUS 1984) and inventory data in this area already have a long tradition in Switzerland, cumulating in its present version as part of the ecoinvent database. One of the raw materials often used for packaging – and at the same time the most important raw material for the production of paper and board – is wood. In fact, it is the world's most important renewable material and regenerative fuel (Bowyer 1995, Eldag 1980, Schulz 1993, Sutton 1993). Wood is widely used as construction material, and as raw material for derived timber products (e.g. plywood, particle board, etc.). In the following two sections of this paper, the areas of packaging materials and wood and their respective integration into ecoinvent are described. A detailed overview of the database ecoinvent is given in Frischknecht et al. (2004a).

The overall structure of the wood part of the ecoinvent v1.1 database and how the most crucial wood-related methodological issues were solved are presented in Section 1. The special properties of wood as a naturally grown material, which is not a priori suited for industrial production and in most cases is partly used as material and partly for energy conversion, require specific methodological consideration for establishing life cycle inventories of the wood chain.

Section 2 then deals with the packaging area in ecoinvent, emphasising on the paper/board sector. In this sector, ecoinvent distinguishes several types of products, e.g. market pulps, graphical papers, packaging papers (kraft, corrugated) and several types of cardboard. A more detailed overview of the structure of the paper/board sector within the database ecoinvent is shown in Fig. 3. Further, a comparison of this most up-to-date data with older data from Habersatter et al. (1998), one standard packaging inventory in the era before ecoinvent, as well as the influence of the wood chain to the total load of paper, are shown and discussed.

## 1 Wood

A broad variety of wood products – materials and fuels – and production processes are integrated in ecoinvent. Data v1.1 from direct forestry products such as roundwood and industrial wood to different types of wood-based boards, chips, pellets and further co-products such as bark, sawdust, etc. (Table 1). This poses specific methodological problems with regard to a consistent modelling of the whole wood chain.

### 1.1 Goal and scope

It is the goal of the datasets on wood in ecoinvent to provide consistent generic life cycle inventory data for LCAs of wood in its applications as a building material, fuel and starting material for other industrial products such as paper and cardboard. Data represent the most common technologies in central Europe around the year 2000.

### 1.2 Functional unit and reference flow

The choice of the functional unit and the reference flow of wood products in a generic LCI database is influenced by the units commonly used for the products, but also by the important properties of the products. Since all possible uses

of wood from massive beams and boards to chips and pellets for incineration are included, the description of the product needs to provide sufficient information for all of them. Neither mass nor volume is by itself a suitable reference for the definition of reference flow of wood products. Density and heating value depend very much on the water content of wood, which varies in the range of >180% for (industrial) wood in the forest to 10% (w/w, relative to dry wood mass) for technically dried wood along the production chain. Thus, information on volume, dry matter content and heating value is needed to model applications of wood as construction material and fuel: volume is how construction wood is sold and bought, dry matter content is important to model the emissions of burning wood and the heating value is needed for the energy output of incineration processes. Since transportation is modelled in ton-km, it is also necessary to know the apparent mass. Thus, besides volume or mass, the water content of wood needs to be indicated for every wood fraction represented in a database.

In ecoinvent Data v1.1, these special properties of wood are taken into account by choosing volume as a reference for the definition of the reference flows while indicating the water content of every fraction in the name of the product or process. Fig. 1 shows the first steps in the wood chain as it is modelled in ecoinvent, and how the products and processes are named.

Volume is chosen as a reference flow instead of dry matter content, because it is the relevant unit for wood on the market for many products. Unlike apparent mass, volume is independent of the moisture content (above 30%) and thus relates linearly to the dry matter content of wood. Below 30% moisture content, species-specific volume changes of wood are associated with the drying process. A shrinkage ratio of 9% (v/v) for spruce and 12% (v/v) for beech are assumed in ecoinvent for the technical drying processes (Werner et al. 2003, according to Frühwald et al. 1996).

Table 1: Classes of wood products in ecoinvent Data v1.1

Class	Examples of products	Short description of data	Main sources
Direct forest products	Round wood, fuel wood	Data on beech and spruce used for generic hard- and softwood, respectively. Sustainable forest management assumed.	BFS/BUWAL 2000, Frischknecht et al. 1996, Frühwald et al. 1996, Schweinle 2000, 2001, Werner 2002
Sawn timber	Massive boards, air/kiln dried, raw/planed	Old, but comprehensive data regarding energy consumption. No information on direct emissions (where dust might be an issue).	Frühwald et al. 1996, Hurst 1996, Ressel 1986
Boards from round wood	Plywood, laminated timber element, wood wool board	Old, but comprehensive data regarding material and energy consumption. Some information on direct VOC emissions.	Frühwald et al. 1996, Hess 1996, Ressel 1986, Schniewind 1989
Boards from industrial residue wood	Particleboard, OSB, fibre boards, cement bonded particleboard	Recent and comprehensive data for average material and energy input. Specific products might differ considerably. Some information on VOC and dust emissions	Frühwald et al. 2000, Wegener et al. 1994
Fuel	Chips, pellets, charcoal	Data of unknown quality on energy consumption and no data on direct emissions from the processes.	Anonymous 1996, Bergmair 1996, BFS/BUWAL 2000, Frischknecht et al. 1996
Chemical wood protection		Confidential data for specific products of three classes. Other products might differ considerably.	Hillier 1997, Künninger & Richter 2000
Infrastructure	Production plants, chainsaws, choppers	Chainsaw based on measurements, other data partly based on expert's guesses. Only to be used to model infrastructure as a background process.	Bergmair 1996, Frischknecht et al. 1996, expert guesses

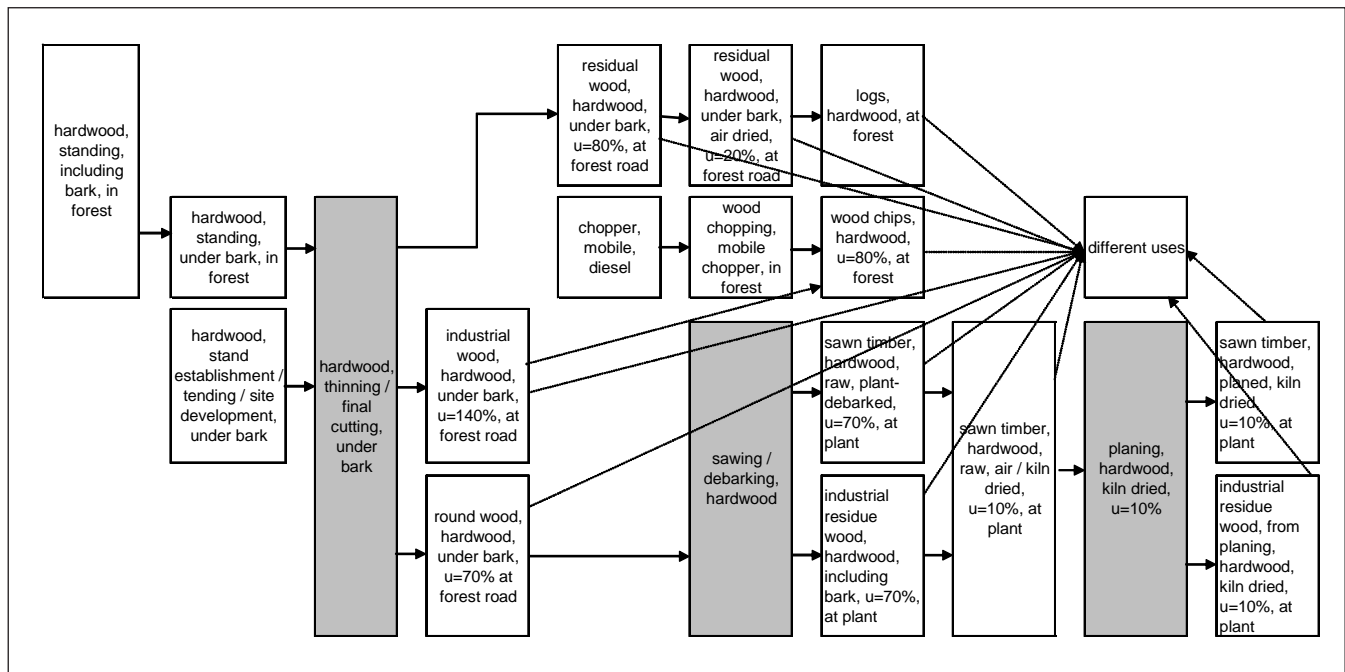


Fig. 1: Overview of hardwood production as modelled in ecoinvent Data v1.1. Grey fields represent multi-output processes. Process and product names contain information on the moisture content and on the reference function. 'Under bark' indicates that the volume relates to the wood without bark, while 'including bark' obviously indicates that the volume includes wood and bark

The volume of wood products can be defined in different ways; ecoinvent uses the most common definition that is used for the commercialisation of a specific product – an information which is given in the ecoinvent meta-information on the processes. For most products, the volume of wood refers to 'wood under bark', i.e. to the wood without the bark. This means, that 1 m<sup>3</sup> of 'round wood at forest road' actually corresponds to more than 1 m<sup>3</sup> because the bark is not included in the functional unit. On the other hand, wood chips are usually measured in bulked volume. Thus, 1 m<sup>3</sup> of wood chips from industry refers to less than 1 m<sup>3</sup> wood under bark, whereas all wood-based boards are inventoried referring to their final volume in m<sup>3</sup>.

### 1.3 Co-product allocation

Wood as a naturally grown material, irregular in shape and structure, is not a priori suited for industrial processing or use in technical applications. As a consequence, the management and processing of wood generates a variety of co-products throughout the wood processing chain from its cultivation in managed forests, its extraction, sawing, processing to intermediate and finished products, on to its recycling, incineration or final disposal. Co-products generated are, e.g. residues from thinning, bark, sawdust, shavings, chips and fibres, side-cuts, wood waste and waste of intermediate products from wood and wood-based industries (Jungmeier et al. 2002a, Jungmeier et al. 2002b, Werner 2002). In many cases, the economic value of these co-products is considerably lower than the economic value of the main product or can even be negative, i.e. they are considered as waste that has either to be disposed of or recycled. As revenues (value) are in fact the driving force behind any economic process (Huppel 1992),

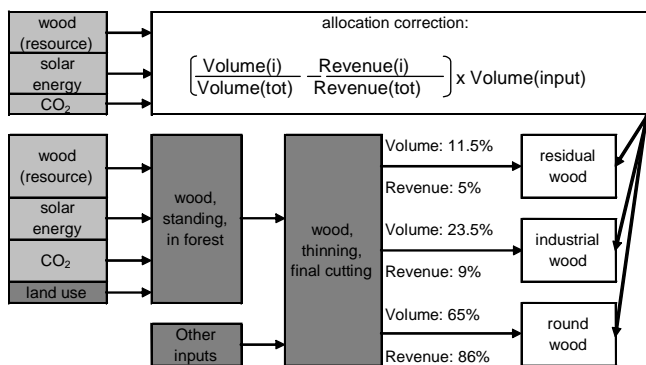
multi-output processes are allocated based on revenues of the different co-products. Revenue-based allocation also takes into account the fact that an established industrial demand exists for almost all co-products of the wood chain.

It has been argued against economic allocation factors that prices are subject to large fluctuations due to market influence, which would make the LCI results rather short living or even arbitrary. Although true for prices as such, price ratios of different co-products – which are used for revenue-based co-product allocation – are much more stable over time than the individual prices. In addition to that, LCI-data has to be updated within a few years to depict the industrial development, which is of course also true for revenue-based allocation factors.

Revenues are of course only one option among several possible allocation factors such as volume ( $u=0$ ), volume (undried), volume (technically dried  $u=12\%$ ) or mass ( $u=0$ ), mass (undried), mass (technically dried  $u=12\%$ ). Results can differ considerably for the different allocation factors in dependency of the species to be considered (Werner 2002). For the socio-economic reasons outlined above, revenues are chosen as the default allocation factor for the wood chain.

### 1.4 Consequences of the revenue-based, co-product allocation for resource use, carbon flows and energy content

Revenue-based allocation poses a specific problem for the material inherent characteristics of the biomass, carbon and energy content, three parameters associated with the (oven dry) mass of wood (de Feyter 1995). As a consequence of the revenue-based allocation and given the large price differences for co-products, carbon bound in wood and its en-



**Fig. 2:** Allocation and allocation correction in ecoinvent. Because only part of the inputs to the first unit process in the wood chain need to be allocated by volume (light grey), while the others are allocated by economic revenue (dark grey), the introduction of allocation correction processes is necessary

energy content would be attributed to the main product in a much higher proportion than what would correspond to its mass. Thus, the main product would 'profit' from a proportionally much more negative GWP (i.e. storing of carbon) and mathematically produce much more thermal energy when incinerated than it actually does. On the other hand, the co-product with a much lower price would appear to carry neither the benefit of bound carbon nor to have any energy content. It is obvious that such an allocation would not lead to a consistent model of the wood chain.

In ecoinvent, this issue is solved by allocating wood as a resource from the ecosphere and the carbon and energy content of wood based on volume (under bark). In a system of subsequent unit processes, this can only be done by using allocation correction factors. **Fig. 2** shows the definition, calculation and use of these factors. The calculation procedure of the allocation correction factors (formula see **Fig. 2**) implies that their sum in one process is always zero (with the exception of the drying processes where an allocation correction process is inventoried to correct the effect of the shrinkage of the wood). **Table 2** gives a numerical example of the use and effect of the allocation correction factors. The roundwood input inventoried in the sawing process in ecoinvent is 1.71 m<sup>3</sup>. The outputs are 1m<sup>3</sup> sawn timber and 0.71 m<sup>3</sup> industrial residue wood. Allocation by revenue al-

locates 96% to sawn timber. Thus, the modelled CO<sub>2</sub> uptake of 1 m<sup>3</sup> sawn timber would be 1338 kg instead of the 815 kg which is really bound in the wood. The allocation correction factors are now calculated to compensate this mismatch.

**1.5 Industrial residual wood as input**

Some production processes use industrial residual wood such as sidings, chips or sawdust as input material, for example particleboard production. In many LCA studies, industrial residual wood as input is modelled by applying the cut-off procedure, i.e. industrial residual wood as an input does not carry any environmental burdens from former production processes. Given the fact that co-products with a positive market price are attributed environmental burdens throughout the wood chain in ecoinvent, such a procedure would result in an inconsistent application of the co-product allocation rule (Werner 2002), which is not in conformity with the 100% rule stated in ISO 14,041, clause 6.5.3.

In ecoinvent, this issue is solved with the definition of two industrial residual wood processes as an input for both hardwood and softwood (m<sup>3</sup>). In these processes, the different industrial residual wood fractions that are generated throughout the wood chain in ecoinvent are summarised, weighted by the actual production volumes of the main products. Moisture content of these industrial residual woods is calculated as the weighted average of the moisture content of the input fractions.

**1.6 Thermal energy surplus from internal production processes**

Many large wood processing companies are connected to a district heating network providing their thermal energy surplus to neighbouring companies or residential buildings. As a consequence, wood input and emissions from the wood incineration process have to be allocated to the thermal energy requirements of the wood processing company and to the energy surplus that is used by third parties.

The wood part of ecoinvent where this issue is relevant – board production – relies on secondary literature. Primary data collection was made only for a few processes, for which no secondary literature was available or where secondary literature had to be complemented. As a consequence, the allocation procedures applied in the secondary literature were

**Table 2:** Example for use and calculation of allocation correction factors

Sawing, softwood, forest-debarked	Inventoried volume [m <sup>3</sup> ]	CO <sub>2</sub> uptake [kg]	Allocation by revenue		
			Allocation factor	CO <sub>2</sub> uncorrected [kg]	CO <sub>2</sub> corrected [kg]
Input					
Round wood, softwood, debarked, u=70% at forest road	1.71	1394			
Softwood, allocation correction, 1 (allocated to sawn timber)	-0.64	-523			
Softwood, allocation correction, 2 (allocated to residue wood)	0.64	523			
Output					
Sawn timber, softwood, raw, forest-debarked, u=70%, at plant	1	815	0.96	1338	815
Industrial residue wood, softwood, forest-debarked, u=70%, at plant	0.71	579	0.04	55.75	579

adopted. Generally, the wood fraction and emissions corresponding to the surplus energy were subtracted from the plant data to only account for the thermal energy required for the production processes.

### 1.7 Modelling of land use

Land use is inventoried in ecoinvent according to a methodology developed based on results of the 14th discussion forum at ETHZ in 2001. Basically, the land occupation (in m<sup>2</sup>a) is inventoried independently to the transformation (in m<sup>2</sup>) from one occupation form to the other. More information is provided in Althaus et al. (2004b) and Frischknecht et al. (2004b). Since land use inventories in ecoinvent need to be appropriate, not only for forest but for all kinds of land use, the recommendations developed by the working group 'land use' of COST E9 (is in accordance with Doka et al. 2001) could not be fully implemented. Land use of the forest is inventoried as occupation of intensively used forest and transformation from extensively to intensively used forest. Because it is not possible to determine when this transformation took place (many forests in Europe were intensively used 2000 years ago during the Roman Empire), the time span to normalise the transformation to the product output was chosen one rotation cycle. Thus, the transformation value is calculated assuming that the land is transformed to harvest just once. This worst case estimation is in accordance with Doka et al. (2001). If the land transformation is interpreted as a major environmental issue in an LCA study, this assumption has to be questioned.

## 2 Packaging Materials

Life Cycle Inventories of packaging materials have a long tradition in Switzerland – starting with the first report of the Swiss EPA already twenty years ago (BUS (1984)). This report has been updated and extended several times (Habersatter 1991, Habersatter et al. (1998), cumulating in its present version which is part of the database ecoinvent. In all these Swiss packaging reports, a variety of different packaging base materials is covered – e.g. plastics, glass, board/paper, aluminium and steel.

### 2.1 Goal and scope

It is the goal of the datasets on packaging materials in ecoinvent to provide consistent generic life cycle inventory data for basic packaging materials, containers and boxes representing average European and/or Swiss technology. Across all the various stages of databases, the amount of basic materials grew constantly – achieving around 40 different materials in the latest database – whereas the amount of treatment processes got smaller in ecoinvent compared with all the precedent reports.

In this paper here, only the board/paper sector, as one of the most important packaging material sectors – economically and within ecoinvent – is treated. More details about the remaining packaging materials can be found in Hischier (2004) (glass, plastics) and in Althaus et al. (2004a) (aluminium, steel).

### 2.2 Structure of the paper/board sector within ecoinvent

A clear structure along the life-cycle of the paper production has been established within the framework of ecoinvent (see part III in Hischier 2004) – i.e. starting with datasets for forestry processes, passing various market pulp and paper/board production datasets, the consumption phase and ending up with waste paper collection datasets. The actual packaging box production is in separate datasets. An overview of the structure and the relation between the different datasets is given in Fig. 3.

According to Fig. 3, ecoinvent distinguishes datasets for Swiss and European conditions in most of the examined steps. This offers the choice of a large variety of datasets to meet the aim of the user's study. The broadest choice exists for the paper/board production. The paper/board consumption mix (processes with the extension 'at regional storage') is modelled based on statistical information from the Confederation of European Paper Industries (CEPI 2001) using only the inventories of graphical paper in ecoinvent Data v1.1 and an additional transport from the production to the consumption area (i.e. to Switzerland). Similar datasets for other regions and further types of paper/board can be created by following the methodology used and documented in Hischier (2004).

All datasets of the paper / board sector are established according to the general ecoinvent methodology (Frischknecht et al. 2004c). According to this methodology, no credits are given for recycled materials. DIP (deinked pulp) for example only accounts for the environmental load of the production of DIP out of waste paper; waste paper as input materials does not carry any environmental burden from its former paper/board life. There is a methodological difference between the European and the Scandinavian wood data, as latter datasets are not modelled as multi-output processes and do

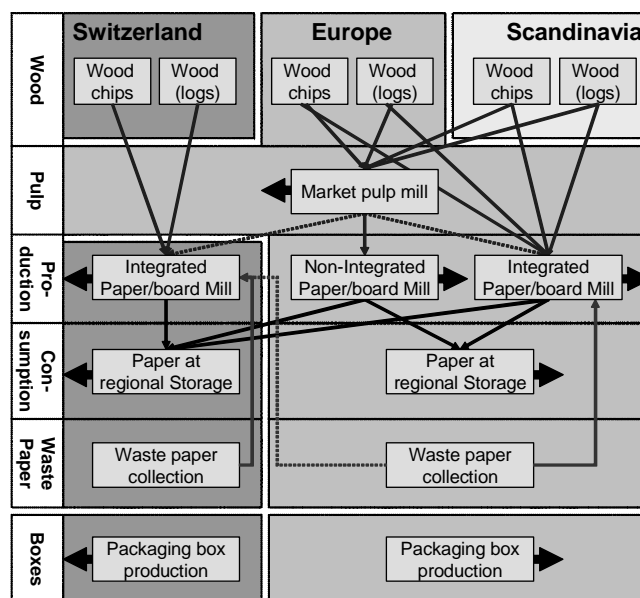


Fig. 3: Overview of the structure of the paper and board sector within the database ecoinvent. The bold arrows indicate the datasets for most common use of the users of the database ecoinvent

not contain the correction factors explained in Section 1. Scandinavian wood is reported as a single output process which is derived based on an unspecified allocation procedure. This is mainly due to the fact that these datasets are based on only one source providing no more information.

Specific datasets for the production of packaging containers and boxes are included as well. These datasets contain only the additional inputs and outputs, whereas the amount of raw material (e.g. corrugated board, aluminium, HDPE) is not included. This type of inventories allows a broader and much more flexible application of these packaging box datasets.

### 2.3 Comparison with former paper/board inventories

The precursor database for LCI data of paper and board in Switzerland is Habersatter et al. (1998), also called SRU250. **Table 3** gives an overview of the amount of datasets in SRU 250 and ecoinvent Data v1.1.

The main advantage of the new ecoinvent data compared to the data published in SRU250 across all types of paper/board is, that the respective dataset is based on a broader base of information in ecoinvent – literature sources, company information either by personal information from specific mills or from their Corporate Environmental Reports (CER) as well as other LCI inventory databases (FEFCO et al. 2000, KCL 2002). As far as possible, datasets based on the information of just one producer / paper mill was avoided.

The most important changes can be observed for graphical paper. The old inventories were all based on information of one single Swiss company; ecoinvent contains average European datasets for the various types of graphical paper. In addition to that, ecoinvent contains much more different datasets for graphical paper. This is due to two different

facts: (i) specific datasets in function of the available information are established, and (ii) different datasets for the production and the consumption situation are established. All in all, ecoinvent Data v1.1 contains a much more adequate set of data with regard to the coverage of the European paper/board market compared to all former Swiss inventories.

### 2.4 Case study packaging paper

As an example in the packaging area, the dataset 'kraft paper, unbleached, at plant (RER)' of ecoinvent Data v1.1 is shown here. First, this dataset is compared with the respective dataset in SRU 250, the dataset 'kraftpaper unbleached'. In a second step, the importance of the precursory wood chain (explained in details in Section 1) to the kraft paper dataset is shown.

Established as a European average dataset, the packaging paper dataset in ecoinvent contains the following production steps: wood handling process, chemical pulping, paper production on the paper machine, on-site energy production as well as the respective chemical recovery cycles and the internal waste water treatment plant. In **Fig. 4**, these different parts together with the most important material and energy flows are summarized. As can be seen there, the production of any kind of auxiliary materials is not included in the corrugated base paper production.

SRU 250 contains a dataset 'kraftpaper unbleached' based on the information from just 1 Swiss company. As this company is a non-integrated paper mill, the dataset contains the following production steps: paper production, on-site energy production as well as the respective chemical recovery cycles and the internal waste water treatment plant – but no production process of the used chemical pulp.

**Table 3:** Overview of the coverage of the paper/board sector within ecoinvent and the former Swiss packaging inventories (ecoinvent from Hischier (2004), SRU250 from Habersatter et al. (1998))

Type of paper/board	ecoinvent Data v1.1		SRU250	
	No. of data	Data sources	No. of data	Data sources
(i) Pulp				
Chemical pulp	6	KCL EcoData, literature, company information	8	3 represent average data from Sweden / remaining based each on 1 plant (CH, D, Can, S)
Mechanical pulp	3	Literature, company information	3	1 from 1 CH plant / other 2 based on various sources
Deinked pulp	– <sup>1)</sup>		1	1 Swiss plant
(ii) Graphical Paper				
Wood-containing paper	11	KCL EcoData, literature, company information	3	Each based on 1 Swiss plant
Woodfree paper	8	KCL EcoData, company information	2	1 based on 1 Swiss plant, the other on 2 Swiss plants
Recycling paper	2	(literature), company information	2	Each based on 1 Swiss plant
(iii) Packaging				
Packaging paper	2	KCL EcoData, company information	6	1 represents average from S / remaining based each on 1 plant (CH, A)
Corrugated board	14	FEFCO report	9	Each based on 1 to several plants (CH, D, A, S)
Carton board	6	KCL EcoData, literature, company information	5	3 based on 1 plant each (CH, S) / 2 represent average from various plants

<sup>1)</sup> according to European statistics, there is no market for deinked pulp, thus no such dataset in ecoinvent

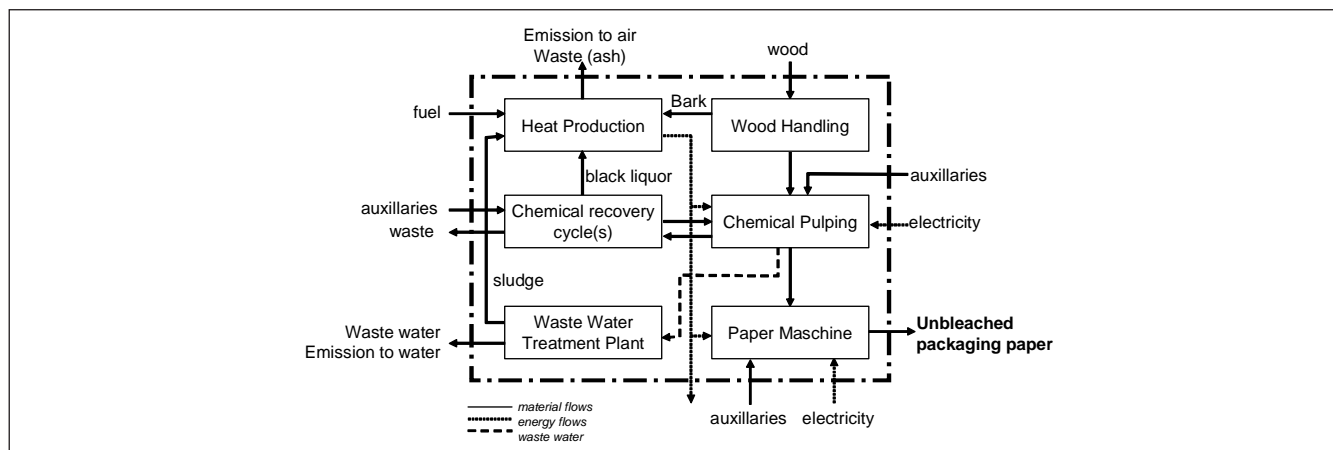


Fig. 4: Schematic overview of an unbleached kraft paper mill and its main material and energy flows

In **Table 4**, a selection of cumulative LCI results and the LCIA parameter 'cumulative energy demand (CED)' of the two examined datasets (unbleached kraft paper according to Hischier (2004), unbleached kraftpaper according to SRU 250) are shown. The cumulative energy demand (CED) is used there for the representation of the LCI input values of the various energy carriers.

As shown in Table 4, the dataset in ecoinvent Data v1.1 reports lower values for CED factors of fossil, nuclear and water compared with the old data, but a much higher value for the CED of biomass. This high amount is due to the fact that in ecoinvent, the amount of internally burned wooden waste is part of the material input (wood), whereas in SRU

250 waste as fuel often was not taken into account as an input. The wood input is responsible for 99% of this biomass CED value. Most emissions to air and water show higher values compared to the data in SRU 250. This is mainly due to the fact that the various materials used for the production of kraft paper are based on more detailed inventories within ecoinvent, resulting in higher overall values. One exception is the fossil carbon dioxide emissions into air, as these CO<sub>2</sub> emissions are mainly caused by use of fossil fuels within the kraft paper chain.

In **Table 5**, the amount of the precursory wood chain to the total load of the dataset 'kraft paper' is shown for the same selection of cumulative LCI results and the LCIA parameter 'cumulative energy demand (CED)' as in Table 4.

**Table 4:** Cumulative energy demand (CED) and selected LCI results from the two unbleached kraft paper datasets (ecoinvent from Hischier (2004), SRU250 from Habersatter et al. (1998)). The values represent the cumulative result over the whole chain per kg of paper and the relative changes from SRU 250 to ecoinvent

	ecoinvent data v1.1	SRU 250	
<b>Cumulative Energy Demand (in MJ-Eq)</b>			
non-renewables, fossil	1.12E+1	1.90E+1	-40.9%
non-renewables, nuclear	4.22E+0	1.22E+1	-65.4%
renewables, water	1.77E+0	3.56E+0	-50.2%
renewables, wind, solar, geothermal	1.56E-1	0	-
renewables, biomass	4.34E+1	3.00E-1	+ 14'400%
<b>Air emissions (in kg)</b>			
carbon monoxide, fossil	2.28E-3	9.80E-4	+132.2%
carbon dioxide, fossil	7.60E-1	1.06E+0	-28.3%
NM VOC	1.13E-3	1.22E-3	-7.0%
nitrogen oxides	5.09E-3	5.11E-3	-0.4%
sulphur dioxide	3.35E-3	2.62E-3	+27.8%
particulates, > 10 um	6.99E-4	1.78E-3	+76.6%
particulates, < 10 and > 2.5 um	4.13E-4		
particulates, < 2.5 um	8.09E-4		
<b>Water emissions (in kg)</b>			
BOD	2.35E-3	7.67E-4	+206.5%
COD	4.47E-3	1.11E-2	-59.7%
chloride	7.01E-3	3.22E-3	+117.7%

**Table 5:** Cumulative energy demand (CED) and selected LCI results from the unbleached kraft paper dataset in ecoinvent (Hischier (2004)) and their amount of the precursory wood chain (expressed in% of the respective total per kg of paper)

	ecoinvent data v1.1	Thereof from wood
<b>Cumulative Energy Demand (in MJ-Eq)</b>		
non-renewables, fossil	1.12E+1	8.3%
non-renewables, nuclear	4.22E+0	2.8%
renewables, water	1.77E+0	2.5%
renewables, wind, solar, geothermal	1.56E-1	0.8%
renewables, biomass	4.34E+1	99.0%
<b>Air emissions (in kg)</b>		
carbon monoxide, fossil	2.28E-3	52.7%
carbon dioxide, fossil	7.60E-1	7.5%
NM VOC	1.13E-3	43.7%
nitrogen oxides	5.09E-3	24.1%
sulphur dioxide	3.35E-3	3.5%
particulates, > 10 um	6.99E-4	6.2%
particulates, < 10 and > 2.5 um	4.13E-4	6.7%
particulates, < 2.5 um	8.09E-4	16.0%
<b>Water emissions (in kg)</b>		
BOD	2.35E-3	11.0%
COD	4.47E-3	6.3%
chloride	7.01E-3	10.0%

As shown in Table 5, the precursory wood chain contributes in a very broad range to the total load of the paper production – achieving 99% for CED of biomass. This value confirms that almost all burned biomass is internally produced waste, originating from the wood input. For all remaining CED values, the wood chain has only a minor influence. Most emissions to air and water show values from the wood chain of 10% and less, i.e. are not due to the input of wood, but from the energy consumption and/or the various auxiliaries used in the paper production process. There are three exceptions, where wood is relevant by being responsible for a quarter and more of the respective load: fossil carbon monoxide, NMVOC and nitrogen oxides. The first two are to a large extent due to the cutting activities (datasets 'power sawing' and 'diesel, burned in building machine') of the wood in the forest, while the latter one originates from the burned fuels for cutting (dataset 'diesel, burned in building machine') and for transporting the wood out of the forest (dataset 'transport, lorry 16t').

### 3 Conclusion and Outlook

The ecoinvent v1.1 database provides consistent datasets of wood-based products over the entire wood chain. The methodology used for modelling multi-output processes allows the use of these datasets in LCA studies with focus on material and energetic use including different end of life options. The LCI data represents a broad spectrum of wooden basic products. Unfortunately, the inventory data for many wood products is based on data of a limited number of companies and cannot be considered to be representative for Europe. Furthermore, some data needs actualization. In addition to that, various wooden final products like window frames, doors, floors, poles, railway sleepers, palisades, etc. could be included in a future version of ecoinvent (Künniger & Richter 1997, Künniger & Richter 1998, Künniger & Richter 2000, Richter et al. 1996, Werner & Richter 1997, Werner et al. 1997).

For packaging materials, further processes and materials could be included compared to the former packaging inventories in Switzerland, whereas for the production of specific packaging containers / boxes a new modular concept allows the user to easily model various types of packaging containers/boxes. With regard to paper/board, ecoinvent v1.1 contains a comprehensive set of production datasets for the various types of market pulp, paper and board representing the average European production situation. Different levels of data quality are observed between the paper/board datasets. An update of those datasets with a lower data quality is desirable. A future version of ecoinvent could also contain datasets of the imports of pulp from overseas, especially from South America and Canada.

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