

A geostatistical model for combined analysis of point-level and area-level data using INLA and SPDE

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Abstract

In this paper a Bayesian geostatistical model is presented for fusion of data obtained at point and areal resolutions. The model is fitted using the INLA and SPDE approaches. In the SPDE approach, a continuously indexed Gaussian random field is represented as a discretely indexed Gaussian Markov random field (GMRF) by means of a finite basis function defined on a triangulation of the region of study. In order to allow the combination of point and areal data, a new projection matrix for mapping the GMRF from the observation locations to the triangulation nodes is proposed which takes into account the types of data to be combined. The performance of the model is examined and compared with the performance of the method RAMPS via simulation when it is fitted to i) point, ii) areal, and iii) point and areal data to predict several simulated surfaces that can appear in real settings. The model is applied to predict the concentration of fine particulate matter ($PM_{2.5}$), in Los Angeles and Ventura counties, United States, during 2011.

Keywords: Misalignment, Data fusion, Model-based geostatistics, INLA, SPDE

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1. Introduction

Spatial and spatio-temporal data arise in a wide range of scientific disciplines, including the environmental, epidemiological, geographical and ecological fields [1]. Data are typically observed either at points in space (point data), or over areal units such as counties or postal codes (areal data). Examples include air pollution measurements taken at a set of ambient stations, temperature and precipitation measurements from weather stations, and population sizes from census tracts. In epidemiology, point data arise when the locations at which cases of disease occur are available, and areal data are often reported when point data are aggregated over geographical subregions of the region of study due to ethical concerns over data use and patient confidentiality [2].

Spatially misaligned data are becoming increasingly common due to advances in both data collection and management, as well as to the ability to merge data from large databases such as disease registries. When information is available from multiple sources on different scales, data may be fused to examine just one variable, such as disease counts recorded in different administrative units. Here the aim is interpolation [3]. Alternatively, we might wish to relate one variable to other variables that are available at different spatial resolutions and alignments. An example is determining whether the risk of an adverse outcome provided at zip level is related to exposure to an environmental pollutant measured at a network of stations, after adjusting for population at risk and other county level demographic information. Here the aim is regression [3].

In this paper we will focus on the data fusion problem which seeks to learn about a particular variable by combining data that are available at different spatial scales. Others have previously developed Bayesian models enabling fusion of data obtained at areal and point-referenced resolutions via the use of latent point-level processes [4], hierarchical downscaling [5], modelling data conditional on the resolution [6], and the use of algorithms such as the reparameterized and marginalized posterior sampling (RAMPS) [7].

The previous approaches use Bayesian predictive inference implemented via

Markov chain Monte Carlo (MCMC) based methods. These methods have made a great impact on statistical practice by making Bayesian inference tractable for complex models but they also present a wide range of problems in terms of convergence and computational time [8]. In this paper we propose general and flexible hierarchical Bayesian models to analyze spatially misaligned data. In order to fit the models, we resort to the Integrated Nested Laplace approximation (INLA) [9] and the Stochastic Partial Differential Equation (SPDE) [10] approaches which are a computationally effective alternative to MCMC for Bayesian inference. In order to allow the combination of data at different spatial resolutions, we propose a new projection matrix for mapping the GMRF in the SPDE method which takes into account how the different types of data are collected. This new approach is fast and flexible.

The outline of the paper is as follows. First, we present flexible models for handling spatial misaligned data in fusion problems. Then, we briefly introduce the INLA and SPDE approaches for Bayesian inference, and present the projection matrix that allows the combination of point and areal data. In Section 3, a simulation study is carried out to compare the performance of the model when estimating several simulated surfaces using point, areal, and point and areal data combined. Then, in Section 4 we evaluate the model in comparison to the RAMPS alternative method for data fusion by applying the methods to several simulated data scenarios. In Section 5, we present an application of the model to real data showing spatial misalignment. In this application, we obtain the spatial distribution of fine particulate matter ($PM_{2.5}$), in Los Angeles and Ventura counties, United States, during 2011. Finally, the conclusions are presented.

2. Models and Inference

2.1. Models

The models proposed assume that there is a spatially continuous variable underlying all observations that can be modeled using a Gaussian random field

process. This process is denoted by $S = \{S(\mathbf{x}) : \mathbf{x} \in D \subset \mathbb{R}^2\}$, has mean function $E[S(\mathbf{x})] = 0$ and stationary covariance function $\text{Cov}(S(\mathbf{x}), S(\mathbf{x}')) = \Sigma(\mathbf{x} - \mathbf{x}')$. Conditionally on S , point data Y_i observed at a finite set of sites, say $\mathbf{x}_i \in D$, $i = 1, 2, \dots, I$, are mutually independent with

$$Y_i | S(\mathbf{x}_i) \sim N(\mu(\mathbf{x}_i) + S(\mathbf{x}_i), \tau^2),$$

where $\mu(\mathbf{x}_i)$ represents the large scale structure. Areal data observations arise as block averages in blocks $B_j \subset D$, $j = 1, 2, \dots, J$,

$$Y(B_j) = |B_j|^{-1} \int_{B_j} (\mu(\mathbf{x}) + S(\mathbf{x})) d\mathbf{x}, \quad |B_j| > 0,$$

where $|B_j| = \int_{B_j} 1 d\mathbf{x}$ denotes the area of B_j .

These models can also accommodate explanatory covariates by including them in the large scale part of the model. Moreover, the models can also be extended to include random effects that can deal with other sources of variability.

2.2. Inference

We fit the models by using the INLA [11, 9] and SPDE approaches [10] which can be easily applied using the R package R-INLA [12]. INLA uses a combination of analytical approximation and numerical integration to do approximate Bayesian inference in latent Gaussian models which includes a large class of models ranging from generalized linear mixed to spatial and spatio-temporal models.

The combination of INLA and SPDE permits analysis of point-level data. In the SPDE approach, the continuously indexed Gaussian field S is represented as a discretely indexed Gaussian Markov random field (GMRF) by means of a finite basis function defined on a triangulation of the region of study. Specifically,

$$S(\mathbf{x}) = \sum_{g=1}^G \psi_g(\mathbf{x}) S_g,$$

where, $\psi_g(\cdot)$ denotes piecewise polynomial basis functions on each triangle, $\{S_g\}$ are zero-mean Gaussian distributed weights, and G are the number of vertices in the triangulation.

In the SPDE approach, the covariance function of the Gaussian field S is required to belong to the Matérn family which represents a very flexible class of covariance functions that appears naturally in many scientific fields [13]. Specifically, for locations \mathbf{x}_i and $\mathbf{x}_j \in \mathbb{R}^2$, the Matérn covariance function is defined as

$$\text{Cov}(S(\mathbf{x}_i), S(\mathbf{x}_j)) = \frac{\sigma^2}{2^{\nu-1}\Gamma(\nu)} (\kappa \|\mathbf{x}_i - \mathbf{x}_j\|)^\nu K_\nu(\kappa \|\mathbf{x}_i - \mathbf{x}_j\|).$$

Here, K_ν is the modified Bessel function of second kind and order $\nu > 0$. The integer value of ν determines the mean square differentiability of the process and it is usually fixed since it is poorly identified in applications. σ^2 denotes the variance and $\kappa > 0$ is related to the range ρ , the distance at which the spatial correlation is close to 0.1 [14].

2.3. Approximation of integrals

In practice, the integrals appearing in the models proposed may not be available in a closed-form. In our approach, we will approximate them using the representation of the continuous Gaussian random field as a GMRF provided by the SPDE approach. Thus,

$$\int_{B_j} S(\mathbf{x}) d\mathbf{x} \approx \sum_{g=1}^G A_{jg} S_g,$$

where G is the number of vertices in the triangulation, $\{S_g\}$ are zero-mean Gaussian distributed weights, and A is a $J \times G$ sparse matrix that maps the GMRF from the J observation locations to the G triangulation nodes.

The matrix A specified in the SPDE approach is designed to deal with point-referenced data. To adapt this approach to our problem, we need to make some modifications that allow accommodation of both point and areal data. Specifically, for the construction of A , we need to differentiate between point and areal observations. If we consider observations taken at point locations in the study region, the projection matrix A can be chosen in the same way as in the SPDE approach. Thus, the row i in A corresponding to an observation at point \mathbf{x}_i , will possibly have three non-zero values at the columns that represent the

vertices of the triangle that contains the point. If \mathbf{x}_i is within the triangle, these
90 values are equal to the barycentric coordinates. That is, they are proportional
to the areas of each of the three subtriangles defined by the point \mathbf{x}_i and the
triangle's vertices, and sum to 1. If \mathbf{x}_i is equal to a vertex of the triangle, row
 i will have just one non-zero value equal to 1 at that vertex. Intuitively, if we
assume that each triangulation vertex has a weight given by the GMRF S , the
95 value of $S(\mathbf{x})$ at a location that lies within one triangle, is the projection of the
plane formed by the triangle vertices weights at location \mathbf{x} .

On the other hand, the model specifies that a particular observation in an
area B and the process S are linked through the mean value of the random
field in the entire area: $|B|^{-1} \int_B S(\mathbf{x}) d\mathbf{x}$, where $|B|$ denotes the area of B . As
100 a result, the rows of A corresponding to a particular observation in an area will
have non-zero values in all vertices inside the area and will be equal to $1/H$,
where H is the number of vertices within the area. Here we need to note that we
approximate the integral of the process in the area by an average of all vertices
weights inside the area. Therefore, to minimize the error of the approximation,
105 it is important to construct a fine triangulation of the domain.

The R code for the combined analysis of point-level and area-level data using
this approach is provided in the Appendix.

3. Simulation study

In this section we carry out a small simulation study to assess the perfor-
110 mance of the method when predicting different spatial surfaces combining data
that have been obtained at several configurations of points and areas in the
region of study. First, we generate several spatial surfaces that may reproduce
some of the situations that can appear in real settings. Then, for each of the
surfaces, the model is fitted using point and areal measurements of the surfaces
115 taken at different configurations. Finally, the merits of the model in each of
the simulated situations are evaluated. The rest of this section describes the
geographic region and the data configurations we decide to use throughout the

simulations, the models used to generate the spatial surfaces, the models fitted, and the results of the simulation study.

120 *3.1. Simulated data*

We are interested in testing the method in a range of situations that can appear in real settings. To do so, we decide to simulate several spatial surfaces of phenomena that can have continuous values in \mathbb{R} . These data may represent, for example, concentration levels of some air pollutant, and may be measured
 125 directly at points where monitoring stations are located, or may be obtained at cells of a regular grid produced by numerical models. In our simulation, we consider the unit square as the study region and take observations at different configurations of randomly generated points and regular grids and vertical bands over the region of study.

130 The different data scenarios are created by varying the number of points and areas in the data sets to be combined. Specifically, data sets of 10, 15, 30, 60 and 100 points are combined with data sets of 4, 16 and 100 squared areas, or 2, 4 and 10 vertical areas. We also create scenarios with no point data and scenarios with no areal data. Examples of such configurations are shown in
 135 Figure 1.

We construct four surfaces for simulated data. For locations \mathbf{x}_i in the unit square, observations Y_i in \mathbb{R} are simulated as follows:

$$Y_i = \mathbf{z}_i\boldsymbol{\beta} + S(\mathbf{x}_i), \quad i = 1, \dots, n,$$

where $\mathbf{z}_i = (1, z_i)$ denotes the vector of the intercept and covariates, $\boldsymbol{\beta} = (\beta_0, \beta_c)'$ is the coefficient vector, and S is a zero-mean Gaussian field with Matérn covariance function with variance σ^2 and range ρ . In the simulations, we set the intercept $\beta_0 = 0$, σ^2 equal to 4 or 1 and ρ equal to 0.7 or 0.1. Moreover,
 140 we use as a covariate a geographic trend with $\beta_c = 2$. The trend covariate is calculated as $(x_i^2 - \overline{x^2})$, where x_i^2 is the second coordinate of location \mathbf{x}_i , and $\overline{x^2}$ is the mean of the second coordinate over the study region. In this way, we are generating a surface where the values increase from south to north. This could

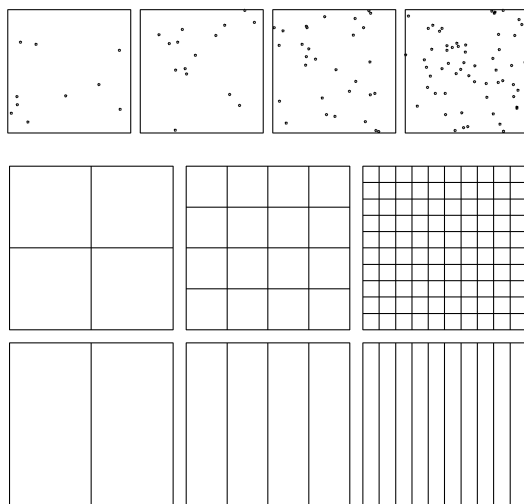


Figure 1: Examples of point and areal configurations used in the simulation study.

represent a surface that reflects changes in temperature or other environmental
 145 covariates that are related with latitude. The values of the parameters for each
 of the simulated surfaces are presented in Table 1. Examples of the simulated
 surfaces are shown in Figures 2 to 5.

	covariate	β_c	σ^2	ρ
US1	-	-	4	0.7
US2	-	-	1	0.1
US3	trend	2	4	0.7
US4	trend	2	1	0.1

Table 1: Parameters of the models used to generate the surfaces in the simulation study.

3.2. Fitted models

Let Y_i , $i = 1, \dots, n + m$, denote the simulated observations at points \mathbf{x}_i ,
 $i = 1, \dots, n$, and areas B_i , $i = n + 1, \dots, n + m$. The fitted models assume a
 Normal likelihood with mean μ_i for the first level, and the following structure

for μ_i in the second level:

$$\mu_i = \mathbf{z}_i \boldsymbol{\beta} + S(\mathbf{x}_i), \quad i = 1, \dots, n,$$

$$\mu_i = \mathbf{z}_i \boldsymbol{\beta} + |B_i|^{-1} \int_{B_i} S(\mathbf{x}) d\mathbf{x}, \quad i = n + 1, \dots, n + m.$$

Here, $\mathbf{z}_i = (1, z_i)$ denotes the vector of the intercept and the covariate, $\boldsymbol{\beta} = (\beta_0, \beta_c)'$ is the coefficient vector, and S is a zero-mean Gaussian field with Matérn covariance function with parameters σ^2 and ρ . If the data are simulated without a covariate, the fitted models do not incorporate the covariate term. If they are simulated using a covariate, the models incorporate the effect of the same covariate.

The models are fitted assuming the following prior distributions. The model parameter ν is set fixed to 1 in the Matérn function implying a continuous domain Markov field. We assign a flat improper prior to the intercept β_0 , and a zero-mean Gaussian distribution with precision equal to 0.001 for the effect of the covariate. Finally, $S \sim N(0, Q^{-1})$ where Q is a sparse precision matrix depending on hyperparameters κ and σ^2 .

3.3. Results

For each simulated pattern, we generate point and areal data and predict the simulated surface applying the model to i) point, ii) areal, and iii) point and areal data combined. We generate 100 surfaces from each simulated scenario to have stable results. The merits of the model in each situation are assessed using the mean squared errors (MSE) of the predictions. The MSE for each simulated data set is calculated as

$$\text{MSE} = \left(\frac{1}{R} \sum_{\mathbf{x} \in R} (u(\mathbf{x}) - \hat{u}(\mathbf{x}))^2 \right)^{1/2},$$

where R denotes the number of locations in the study region, $u(\mathbf{x})$ is the value of the simulated surface at location \mathbf{x} , and $\hat{u}(\mathbf{x})$ is the prediction of $u(\mathbf{x})$. Figures 2 to 5 show the MSEs for each of the scenarios and combinations of data averaged over the 100 replications.

The results show that the MSE depends on the simulated surfaces and also on the types of data used to fit the model. We observe that when the model is applied using data obtained in just areas or points, the MSE decreases as the number of areas or points increases. There are some situations, however, where the decrease of MSE is very small. This is the case of scenario US2 where the data are generated using $\sigma^2 = 1$, $\rho = 0.1$ and with no covariates. We also see that in general, the combination of point and areal data provides better predictions than if the method is applied just to one type of data. There are a few exceptions however. For example, when there is a large amount of areal or point data then information from just one type of data could be enough to accurately predict the real process. In these situations, a joint analysis is not useful to improve the predictions obtained using a point or areal analysis. For example, if there are 100 squared areal observations and just a few point observations, the addition of point data to the analysis does not provide additional information to obtain better predictions.

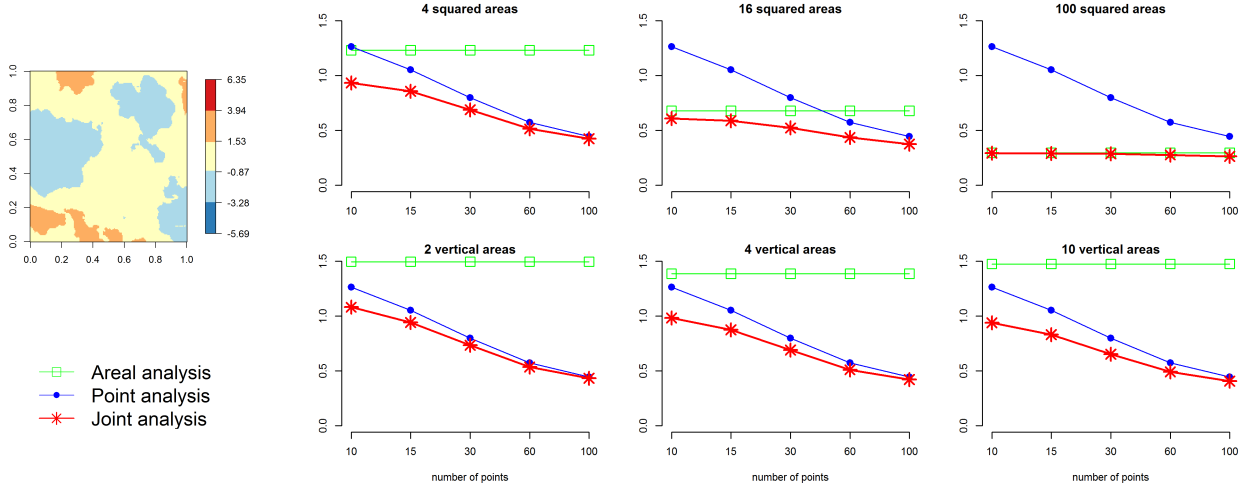


Figure 2: Scenario US1 results. First column: one of the 100 simulated surfaces. Second to fourth columns: MSEs of the predictions obtained for the simulated surfaces US1 averaged over 100 replications by type of analysis.

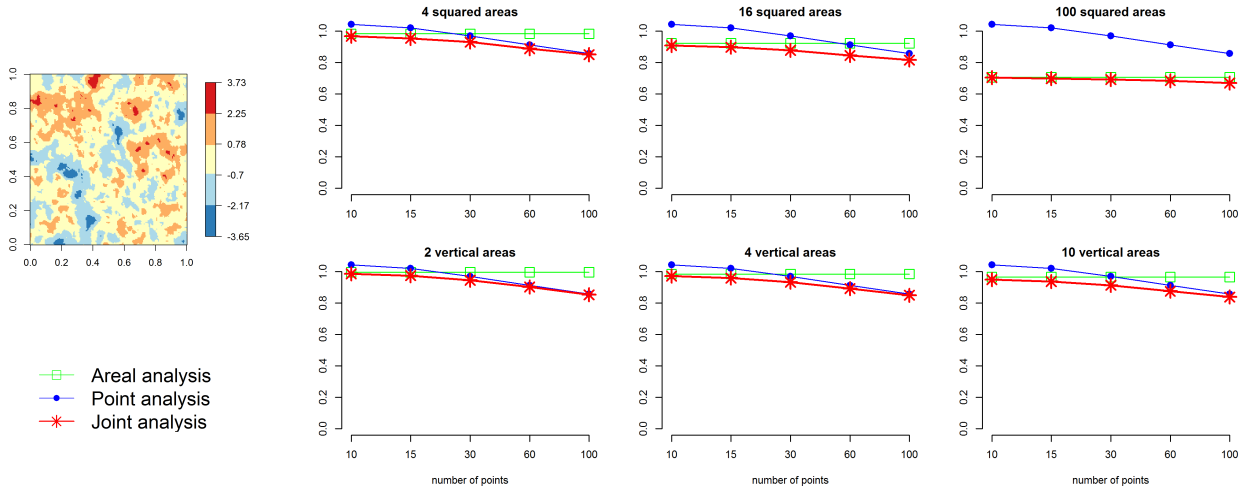


Figure 3: Scenario US2 results. First column: one of the 100 simulated surfaces. Second to fourth columns: MSEs of the predictions obtained for the simulated surfaces US2 averaged over 100 replications by type of analysis.

4. Performance evaluation in comparison with RAMPS

In this section, we compare the method presented with another existing method for data fusion. Specifically, we present a performance evaluation of our method in comparison with the reparameterized and marginalized posterior sampling (RAMPS) algorithm for complex Bayesian geostatistical models. We chose RAMPS for comparison because of its flexibility and the availability of an R package called `ramps` that implements all of its capabilities [15]. RAMPS enables joint modeling of areal and point data arising from the same underlying spatial process, and allows accommodation of non-spatial correlation and variance heterogeneity as well as spatial and/or temporal correlation. Specifically, an observation vector Y which may contain both point and areal data is modeled as follows:

$$Y = X\beta + W\gamma + KZ + \epsilon,$$

$$\gamma \sim N(0, \Sigma_\gamma), \quad Z \sim N(0, \Sigma_Z), \quad \epsilon \sim N(0, \Sigma_\epsilon),$$

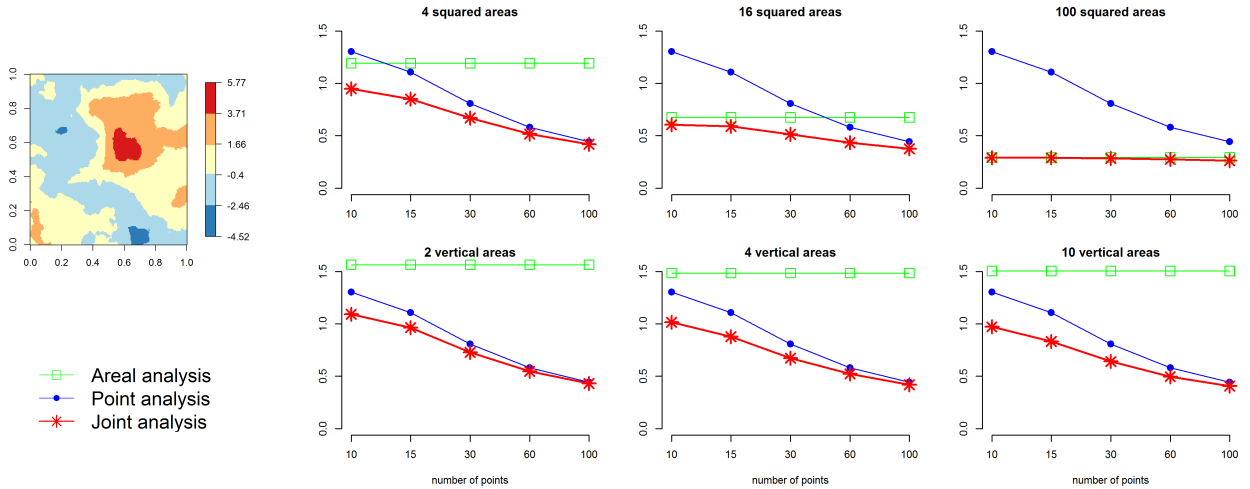


Figure 4: Scenario US3 results. First column: one of the 100 simulated surfaces. Second to fourth columns: MSEs of the predictions obtained for the simulated surfaces US3 averaged over 100 replications by type of analysis.

where β is a vector of regression coefficients, γ is a vector of non-spatial random effects, Z is an vector of spatial random effects, ϵ is a vector of measurements errors, and the matrices X , W , and K are design matrices for fixed effects, non-spatial random effects, and spatial random effects, respectively. The model is fitted using an algorithm that involves reparameterizing the variance parameters, reformulating the means structure, marginalizing the joint posterior distribution, and applying the slice sampling MCMC method based on simplexes.

Here, we simulate four surfaces with different characteristics and obtain point and areal observations. Then, our method and RAMPS are applied to predict the simulated surfaces using all the observations combined. Finally, the performance of the methods is evaluated by means of the MSE, the parameter estimates and the run time.

The four spatial surfaces are generated on $[0, 1] \times [0, 1]$ using a Gaussian model with a Matérn covariance structure with variance σ^2 , range ρ and overall mean $\beta_0 = 0$. For surfaces S1 and S3, we set $\sigma^2 = 4$ and $\rho = 0.7$, in surfaces S2 and S4, $\sigma^2 = 1$ and $\rho = 0.1$. Moreover, in surfaces S3 and S4 we use a

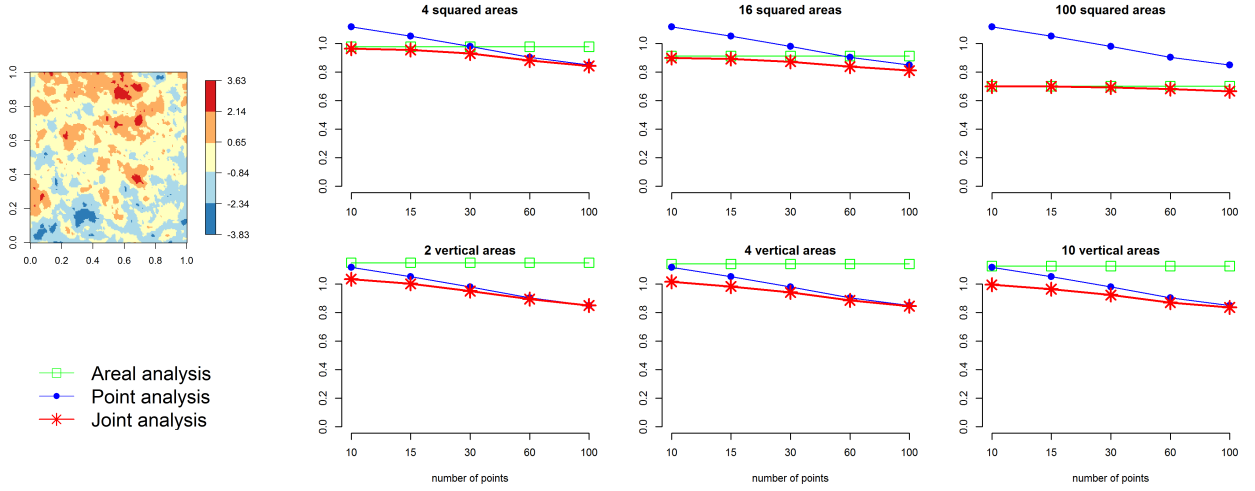


Figure 5: Scenario US4 results. First column: one of the 100 simulated surfaces. Second to fourth columns: MSEs of the predictions obtained for the simulated surfaces US4 averaged over 100 replications by type of analysis.

geographic trend covariate calculated as $(x_i^2 - \overline{x^2})$ and coefficient $\beta_c = 2$. The observations to be combined are obtained as the point values corresponding to 100 randomly generated locations and the average values in the cells of a 4×4 regular grid. The surfaces generated and the areal and point observations for each scenario are shown in Figure 6.

We apply our method and RAMPS to predict the simulated surfaces. Let Y_i , $i = 1, \dots, n+m$, denote the simulated observations at points \mathbf{x}_i , $i = 1, \dots, n$, and areas B_i , $i = n+1, \dots, n+m$. The model fitted assumes a Normal likelihood with mean μ_i expressed as

$$\mu_i = \mathbf{z}_i \boldsymbol{\beta} + S(\mathbf{x}_i), \quad i = 1, \dots, n,$$

$$\mu_i = \mathbf{z}_i \boldsymbol{\beta} + |B_i|^{-1} \int_{B_i} S(\mathbf{x}) d\mathbf{x}, \quad i = n+1, \dots, n+m.$$

Here, $\mathbf{z}_i = (1, z_i)$ is the vector of the intercept and the covariate, $\boldsymbol{\beta} = (\beta_0, \beta_c)'$ is the coefficient vector, and S is a zero-mean Gaussian field with Matérn covariance function with parameters σ^2 and ρ . This model is fitted to the data simulated using a covariate. If the data are simulated without covariate, the

fitted model does not incorporate the covariate term. Due to the large amount of time needed to fit the model using RAMPS, prediction is just done at 231 uniformly distributed locations.

210 The priors used when applying our method are the same as the ones employed in the simulation study in Section 3. When applying RAMPS, flat priors are used on the intercept β_0 and the covariate coefficient β_c . An inverse gamma prior with shape and scale parameters set to 0.01 is used for σ^2 , and an uniform prior on $(0, 2)$ is used for the range ρ . Using RAMPS, convergence is achieved
215 running a MCMC chain of 30,000 iterations and using a burn-in of 1,000 and a thinning rate of 30 iterations for each of the surfaces. Posterior means and 95% CIs are calculated with the remaining 966 iterations. We note that the use of different priors may result in different estimates and run times. However, showing results from even only these priors reveals the main differences in
220 performance between our method and RAMPS.

For each of the simulated surfaces and methods, we calculate the MSE, the posterior means and 95% CIs for the model parameters, and the run times. These values are shown in Table 2. We observe that lower MSEs are obtained with our method than with RAMPS in all surfaces except for surface S2, where
225 ours is 0.02 higher (0.89 with our method and 0.87 with RAMPS). The highest difference in MSE is obtained when the methods are applied to predict surface S3 which is simulated using a geographic trend as a covariate, $\sigma^2 = 4$ and $\rho = 0.7$. In S3, the MSE obtained with our method is equal to 0.46 compared to 1.75 with RAMPS. We see that neither our method nor RAMPS accurately
230 recover the true values of the parameters used in the simulations, both methods yield 95% CIs that contain the true values for most of the parameters. We also see that in the simulated surfaces S1 and S3, that is, when $\sigma^2 = 4$ and $\rho = 0.7$, the upper limits of the 95% CIs for σ^2 obtained with RAMPS are very high. With our method, however, narrower 95% CIs for σ^2 are obtained. Finally, we
235 note that RAMPS needs longer run times than our method. Specifically, for all surfaces RAMPS takes more than three hours whereas our method finishes in less than one minute.

5. Application to air pollution data

The methodology proposed provides a valuable tool in a wide range of re-
240 search fields. Here, we present an application where we obtain the spatial dis-
tribution of a common air pollutant: fine particulate matter ($PM_{2.5}$) in Los
Angeles and Ventura counties, United States, during 2011. We fit a spatial
model combining information for the variable of interest from point and areal
resolutions. We also model point and areal data separately to assess the differ-
245 ences in the predictions obtained.

Particulate matter, or PM, are a mixture of microscopic solids and liquid
droplets floating in the air that are considered harmful to public health and the
environment [16]. These particles are made up of a number of components such
as acids, chemicals, metals, soil and dust, and are emitted in the atmosphere
250 either directly from a source or as result of complicated chemicals reactions.
Particulate matter which are less than $2.5 \mu\text{m}$ in diameter ($PM_{2.5}$) pose one of
the greatest problems since they can get deep into the lungs and cause serious
health effects including increased respiratory symptoms, heart or lung diseases,
and even premature death [16].

255 Information on concentration (micrograms per cubic meter) for $PM_{2.5}$ in
Los Angeles and Ventura counties are available as direct measurements at loca-
tions of monitoring sites, and as estimates inferred from satellite-derived $PM_{2.5}$
sources at a raster grid. The monitoring data have been obtained from a set
of 14 sites sparsely located in the region at which the United States Environ-
260 mental Protection Agency (EPA) regularly measures $PM_{2.5}$ among other air
pollutants [17] We have used the mean of the daily measurements recorded in
year 2011 in each of the monitoring stations. The satellite-derived estimates rep-
resent three-year mean grids (2010-2012) of $PM_{2.5}$ concentrations derived from
a combination of MODIS (Moderate Resolution Imaging Spectroradiometer),
265 MISR (Multi-angle Imaging SpectroRadiometer) and SeaWiFS (Sea-Viewing
Wide Field-of-View Sensor) AOD (Aerosol Optical Depth) satellite retrievals
[18, 19]. The raster grid has a grid cell resolution of 6 arc-minutes (0.1 degree

or approximately 10 km at the equator). Figure 7 shows the concentration values in each of the monitoring stations and in the raster grid.

The model used to predict $\text{PM}_{2.5}$ values in the study region is specified as follows. The $\text{PM}_{2.5}$ concentration, Y_i , at each of the locations of the monitoring stations, \mathbf{x}_i , $i = 1, \dots, n$, and cells of the raster grid, B_i , $i = n + 1, \dots, n + m$, are modeled as Gaussian observations with mean μ_i :

$$Y_i \sim \text{Normal}(\mu_i, \sigma^2), \quad i = 1, \dots, n, n + 1, \dots, n + m,$$

$$\mu_i = \beta_0 + S(\mathbf{x}_i), \quad i = 1, \dots, n,$$

$$\mu_i = \beta_0 + |B_i|^{-1} \int_{B_i} S(\mathbf{x}) d\mathbf{x}, \quad i = n + 1, \dots, n + m,$$

270 where β_0 is the intercept and S is a zero-mean Gaussian field with Matérn covariance function with parameters σ^2 and ρ . We fit the model three times using areal, point and areal and point data. The model is fitted using the same priors as the ones employed in the simulation study in Section 3.

Although estimates differed by analysis the 95% CIs overlapped (Table 3).
 275 The most accurate predictions (tightest CIs) for model parameters were generally for the areal model, then our model, while using only the point data resulted in large uncertainty around the estimates.

Maps of the predictions obtained and the 95% CI showing the range of plausible values for each location are shown in Figure 8. Although all maps show
 280 a predicted $\text{PM}_{2.5}$ higher in the south, there are some differences depending on the analysis used. For example, with the joint analysis the predicted $\text{PM}_{2.5}$ is higher close to the city of Los Angeles than the predicted $\text{PM}_{2.5}$ obtained with the areal analysis. Also, by using both areal and point data, we are able to obtain more accurate predictions in the south where there is both point and
 285 areal information than the ones obtained using just one type of data.

6. Discussion

In this paper we have presented a joint Bayesian model to combine point and areal data. The model assumes that underlying all observations there is a

spatially continuous variable that can be modeled using a Gaussian random field
290 process. INLA and SPDE approaches were used to fit the data and represent
the continuously indexed Gaussian random field as a discretely indexed GMRF
by means of a basis function representation defined on a triangulation of the
region of interest. In order to allow the combination of point and areal data, we
proposed a new projection matrix for mapping the GMRF from the observation
295 locations to the triangulation nodes which takes into account the types of data.

The results show that the goodness of fit depends on the simulated surfaces
and also the types of data used to fit the model. In most situations we observe
that the combination of point and areal data provides better predictions than
if the method is applied to just one type of data, and this was consistent over
300 both simulated and real data. Our method was also demonstrably superior to
RAMPS by obtaining better predictions in much shorter run times on simulated
data.

Real data are messy, especially when attempting to use multiple sources
of information. Our method performed well when applied to monitored air
305 pollution data. When predicting the concentration of $PM_{2.5}$ in Los Angeles
and Ventura counties during 2011, the point analysis gave markedly different
results to the areal analysis (Figure 8). In part, this may be due to differences
in time periods, as well as the relative lack of monitoring stations in the north.
Combining these estimates using our method enabled more accurate predictions
310 of the concentration of $PM_{2.5}$, particularly in the south.

A limitation of the method proposed is that it is only applicable to Gaussian
data. Unfortunately, this does not include many important settings such as
disease mapping problems where data are typically modelled using Poisson or
Binomial distributions and non-identity link functions.

315 Models based on aggregated data contain the potential for ecological fallacy
which occurs when estimated associations obtained from analyses of variables
measured at an aggregated level lead to conclusions different from analyses
based on the same variables measured at the individual level. The resulting
bias, called ecological bias, is comprised of two effects: the aggregation bias due

320 to the grouping of individuals, and the specification bias due to the differential
distribution of confounding variables created by grouping [20], [21]. In many
situations, however, it is difficult to obtain sufficient point data to obtain con-
clusions and we should make the most of the information from the available
data regardless of their spatial resolution. For example, [22] show how the com-
325 bination of disease data from different sources can improve inferences from that
using a single data set, and demonstrate that analyses combining related data
at both the individual and aggregate level can reduce ecological bias and add
precision.

There are many situations where data are very hard to obtain and especially
330 in these cases it is very important to optimize the use of all available informa-
tion. The method proposed enables obtaining better predictions by combining
data obtained at different resolutions. However, we should be aware that bias
could arise if we combine data that are not completely comparable such as data
collected about different populations or at different times. In such situations we
335 may decide to use just one of the data sets or alternatively to adjust for bias in
the model.

A major advantage of the method presented is that the Bayesian framework
used could be easily extended to adequately model many problems of interest.
For example, the model may be extended to accommodate spatio-temporal data
as follows. Let us consider a spatio-temporal Gaussian process $S = \{S(\mathbf{x}, t) : \mathbf{x} \in D \subset \mathbb{R}^2, t \in T \subset \mathbb{R}\}$ with $E[S(\mathbf{x}, t)] = 0$ and stationary covariance
function $\text{Cov}(S(\mathbf{x}, t), S(\mathbf{x}', t')) = \Sigma(\mathbf{x} - \mathbf{x}', t - t')$. Data observed at locations
 $\mathbf{x}_i, i = 1, \dots, I$, and times $t_k, k = 1, \dots, K$ may be modeled as

$$Y_{i,k} | S(\mathbf{x}_i, t_k) \sim N(\mu(\mathbf{x}_i, t_k) + S(\mathbf{x}_i, t_k), \tau^2).$$

Then, observations in areas $B_j \subset D, j = 1, 2, \dots, J$, and periods of time $\tau_l \in T, l = 1, \dots, L$, are expressed averaging the process in space and also in time,

$$Y(B_j, \tau_l) = |B_j|^{-1} |\tau_l|^{-1} \int_{B_j} \int_{\tau_l} (\mu(\mathbf{x}, t) + S(\mathbf{x}, t)) d\mathbf{x} dt, \quad |B_j| > 0, \quad |\tau_l| > 0.$$

Also, it is possible to include covariates, and handling and representing different

sources of uncertainty, including sampling error, measurement error, as well as prediction errors at unsampled locations. Another advantage of the method is
340 that by using the approximate methods INLA and SPDE, we are able to obtain results quickly and avoid assessing the convergence and mixing properties of the chains generated by using MCMC-based methods. In addition, since this method is less computationally intensive we are able to deal with large data sets.

345 The combination of point-level and area-level referenced data is an important and not yet completely resolved methodological issue within the general area of spatial statistics. We think that the approach presented may be a helpful advance in this area by providing a useful tool that is applicable in a wide range of situations where information at different spatial resolutions is combined.

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References

- 355 [1] N. Cressie, *Statistics for spatial data*, Wiley, New York, 1993.
- [2] A. B. Lawson, Bayesian point event modeling in spatial and environmental epidemiology, *Statistical Methods in Medical Research* 21 (2012) 509–529.
- [3] S. Banerjee, B. P. Carlin, A. E. Gelfand, *Hierarchical modeling and analysis for spatial data*, Second Edition, Chapman and Hall/CRC, Boca Raton,
360 2014.
- [4] M. Fuentes, A. E. Raftery, Model evaluation and spatial interpolation by bayesian combination of observations with outputs from numerical models, *Biometrics* 66 (2005) 36–45.

- [5] V. J. Berrocal, A. E. Gelfand, D. M. Holland, A spatio-temporal downscaler
365 for outputs from numerical models, *J. Agric. Biol. Environ. Stat.* 15 (2010)
176–197.
- [6] C. K. Wikle, L. M. Berliner, Combining information across spatial scales,
Technometrics 47 (1) (2005) 80–91.
- [7] M. K. Cowles, J. Yan, B. Smith, Reparameterized and marginalized pos-
370 terior and predictive sampling for complex bayesian geostatistical models,
Tech. rep., The University of Iowa, Department of Statistics and Actuarial
Science (2007).
URL <http://www.stat.uiowa.edu/techrep/>
- [8] B. M. Taylor, P. J. Diggle, INLA or MCMC? A tutorial and comparative
375 evaluation for spatial prediction in log-Gaussian Cox processes, *Journal of
Statistical Computation and Simulation* 84 (10) (2014) 2266–2284.
- [9] H. Rue, S. Martino, N. Chopin, Approximate Bayesian inference for la-
tent Gaussian models by using integrated nested Laplace approximations,
Journal of the Royal Statistical Society, Series B 71 (2) (2009) 319–392.
- 380 [10] F. Lindgren, H. Rue, J. Lindström, An explicit link between gaussian fields
and gaussian markov random fields: the stochastic partial differential equa-
tion approach, *Journal of the Royal Statistical Society, Series B* 73 (4)
(2011) 423–498.
- [11] H. Rue, S. Martino, Approximate Bayesian inference for hierarchical Gaus-
385 sian Markov random fields models, *Journal of Statistical Planning and
Inference* 137 (2007) 3177–3192.
- [12] H. Rue, S. Martino, F. Lindgren, D. Simpson, A. Riebler, R-inla: Approx-
imate bayesian inference using integrated nested laplace approximations,
<http://www.r-inla.org/> (2013).
- 390 [13] P. Guttorp, T. Gneiting, Studies in the history of probability and statistics
xlix on the matérn correlation family, *Biometrika* 93 (4) (2006) 989–995.

- [14] M. Cameletti, F. Lindgren, D. Simpson, H. Rue, Spatio-temporal modeling of particulate matter concentration through the spde approach, *AStA Advances in Statistical Analysis*.
- 395 [15] B. J. Smith, J. Yan, M. K. Cowles, Unified geostatistical modeling for data fusion and spatial heteroskedasticity with r package ramps, *Journal of Statistical Software* 25 (10) (2008) 1–21.
- [16] United States Environmental Protection Agency (EPA), Particulate Matter (PM) Pollution, <https://epa.gov/pm-pollution> (2015).
- 400 [17] United States Environmental Protection Agency (EPA), Air Quality Data Mart, https://aqs.epa.gov/aqsweb/documents/data_mart_welcome.html (2015).
- [18] A. van Donkelaar, R. V. Martin, M. Brauer, B. L. Boys, Global annual pm2.5 grids from modis, misr and seawifs aerosol optical depth (aod),
405 1998-2012. palisades, ny: Nasa socioeconomic data and applications center (sedac), <http://dx.doi.org/10.7927/H4028PFS> (2015).
- [19] A. van Donkelaar, R. V. Martin, M. Brauer, B. L. Boys, Use of satellite observations for long-term exposure assessment of global concentrations of fine particulate matter, *Environmental Health Perspectives* 123 (2) (2015)
410 135–143.
- [20] W. S. Robinson, Ecological correlations and the behavior of individuals, *American Sociological Review* 15 (1950) 351–357.
- [21] C. A. Gotway, L. J. Young, Combining incompatible spatial data, *Journal of the American Statistical Association* 97 (459) (2002) 632–648.
- 415 [22] C. Jackson, N. Best, S. Richardson, Hierarchical related regression for combining aggregate and individual data in studies of socio-economic disease risk factors, *Journal of the Royal Statistical Society. Series A (Statistics in Society)* 171 (2008) 159–178.

Appendix

420 R code for combined analysis of point-level and area-level data using INLA
and SPDE.

```
# Point observations
# coop: matrix of point locations
# yp: vector of observed values at points
425 # xp: vector of covariate values at points

# Areal observations
# spol: SpatialPolygons object containing the areas
# cooa: matrix of the spatial coordinates of spol
430 # ya: vector of observed values in areas
# xa: vector of covariate values in areas

# Prediction points
# coopred: matrix with point locations for prediction
435 # ypred: vector of observed values in prediction points (NA)
# xpred: vector of covariate values in prediction points

# Mesh
# meshfit: fine triangulated mesh
440

# Matern SPDE model object
spde <- inla.spde2.matern(mesh=meshfit, alpha=2)

# Point observations
445 Ap <- inla.spde.make.A(mesh=meshfit, loc=coop)
stk.p <- inla.stack(tag='point',
                    data=list(y=yp),
                    A=list(Ap, 1),
```

```

effects=list(s=1:spde$n.spde, data.frame(b0=1, x=xp)))
450
# Areal observations
locin <- meshfit$loc[as.vector(which(!is.na(over(SpatialPoints(meshfit$loc), spol))))],]
block <- rep(0, nrow(locin))
for(i in 1:length(spol)){
455 block[as.vector(which(!is.na(over(SpatialPoints(locin), spol[i]))))] <- i
}
Aa <- inla.spde.make.A(mesh=meshfit, loc=locin, block=block, block.rescale="sum")
stk.a <- inla.stack(tag='areal',
                    data=list(y=ya),
460                    A=list(Aa, 1),
                    effects=list(s=1:spde$n.spde, data.frame(b0=1, x=xa)))

# Prediction points
Apred <- inla.spde.make.A(mesh=meshfit, loc=coopred)
465 stk.pred <- inla.stack(tag='pred',
                        data=list(y=ypred),
                        A=list(Apred, 1),
                        effects=list(s=1:spde$n.spde, data.frame(b0=1, x=xpred)))

470 # Stack
stk.full <- inla.stack(stk.p, stk.a, stk.pred)

# Fit model
formula <- y ~ 0 + b0 + x + f(s, model=spde)
475 res <- inla(formula, data=inla.stack.data(stk.full),
              control.predictor=list(compute=TRUE, A=inla.stack.A(stk.full)))

```

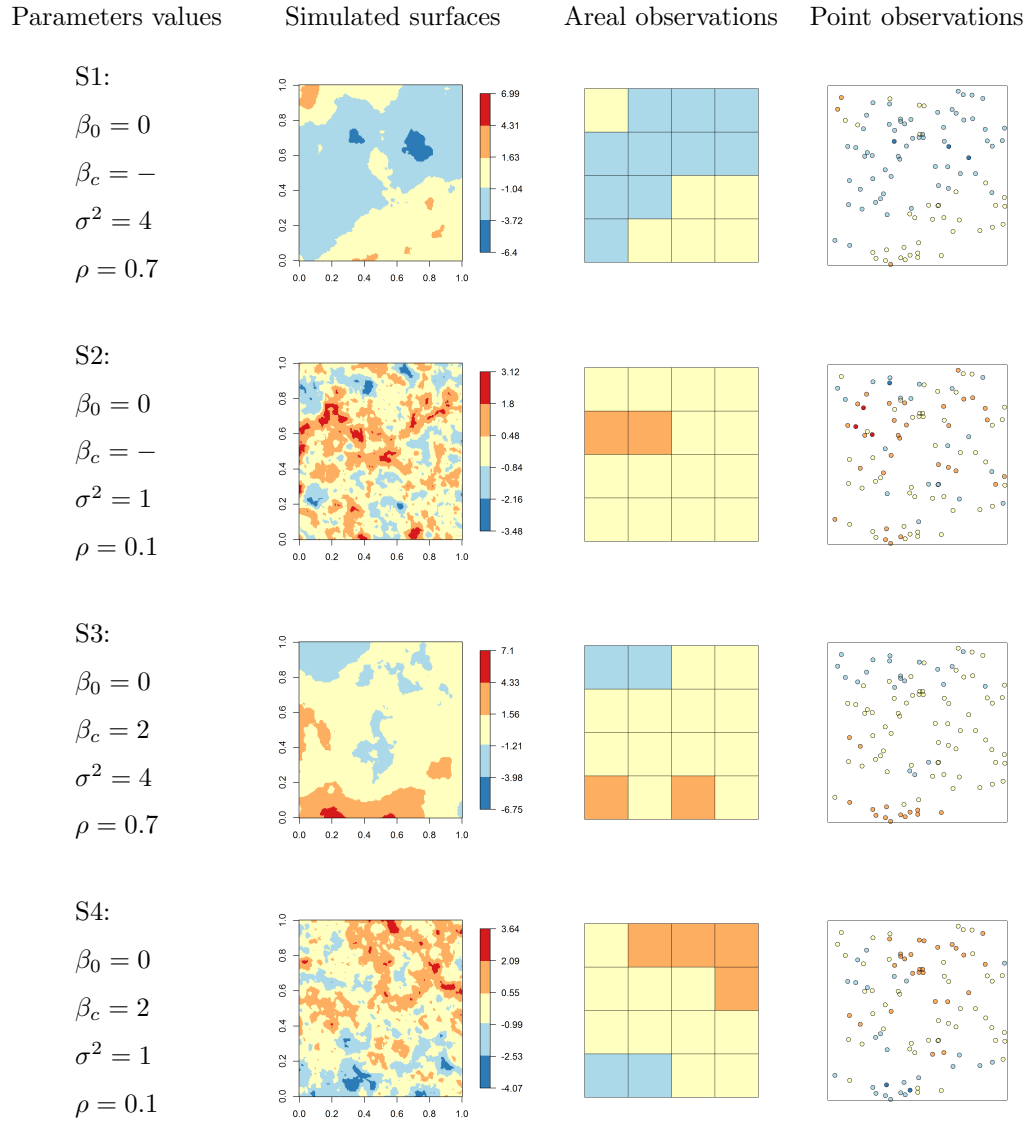


Figure 6: Simulated surfaces and point and areal observations used in the performance study.

Simulated surface	Method	MSE	β_0	β_c	σ^2	ρ	time
S1 ($\sigma^2 = 4, \rho = 0.7$)	Our method	0.41	-0.62 (-2.31, 1.28)	-	3.33 (1.52, 6.52)	0.50 (0.30, 0.75)	37.11 seconds
	RAMPS	0.95	-0.82 (-5.45, 3.95)	-	12.99 (1.66, 119.17)	0.33 (0.13, 1.40)	4.26 hours
S2 ($\sigma^2 = 1, \rho = 0.1$)	Our method	0.89	0.01 (-0.23, 0.23)	-	1.62 (1.00, 2.64)	0.07 (0.04, 0.11)	43.09 seconds
	RAMPS	0.87	0.00 (-0.21, 0.23)	-	0.78 (0.51, 1.11)	0.04 (0.03, 0.06)	3.55 hours
S3 ($\sigma^2 = 4, \rho = 0.7$)	Our method	0.46	0.31 (-3.29, 3.95)	-2.94 (-7.24, 2.00)	4.60 (1.03, 14.72)	0.80 (0.35, 1.68)	34.78 seconds
	RAMPS	1.75	-0.02 (-12.25, 10.19)	-5.83 (-9.99, -2.24)	35.38 (1.77, 139.96)	0.83 (0.19, 1.93)	4.71 hours
S4 ($\sigma^2 = 1, \rho = 0.1$)	Our method	0.81	-0.24 (-0.58, 0.10)	1.61 (0.51, 2.67)	0.98 (0.60, 1.52)	0.13 (0.07, 0.23)	50.32 seconds
	RAMPS	0.93	-0.14 (-0.40, 0.13)	1.73 (0.94, 2.57)	0.90 (0.63, 1.28)	0.04 (0.03, 0.06)	3.65 hours

Table 2: MSEs, posterior means and 95% CIs of the parameters, and run time for each of the simulated surfaces and methods.

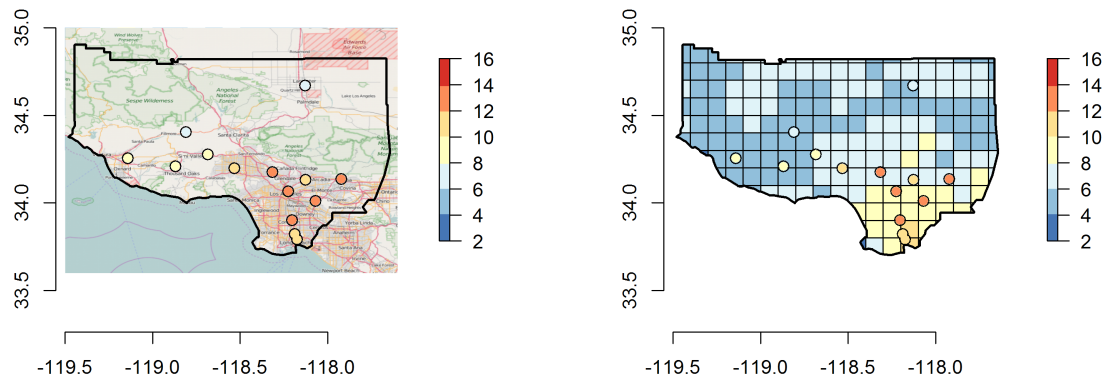


Figure 7: $PM_{2.5}$ concentration (micrograms per cubic meter) in monitoring stations (left) and raster grid together with monitoring stations (right) in Los Angeles and Ventura counties in 2011.

	β_0	σ^2	ρ
Joint analysis	7.31 (6.46 , 8.22)	2.72 (0.94 , 6.35)	1.20 (0.60 , 2.22)
Areal analysis	7.50 (7.09 , 7.91)	3.74 (1.65 , 7.50)	0.85 (0.52 , 1.34)
Point analysis	9.51 (5.42 , 12.94)	7.00 (1.78 , 20.2)	0.74 (0.31 , 1.55)

Table 3: Posterior means and 95% CIs of the model parameters in the air pollution study.

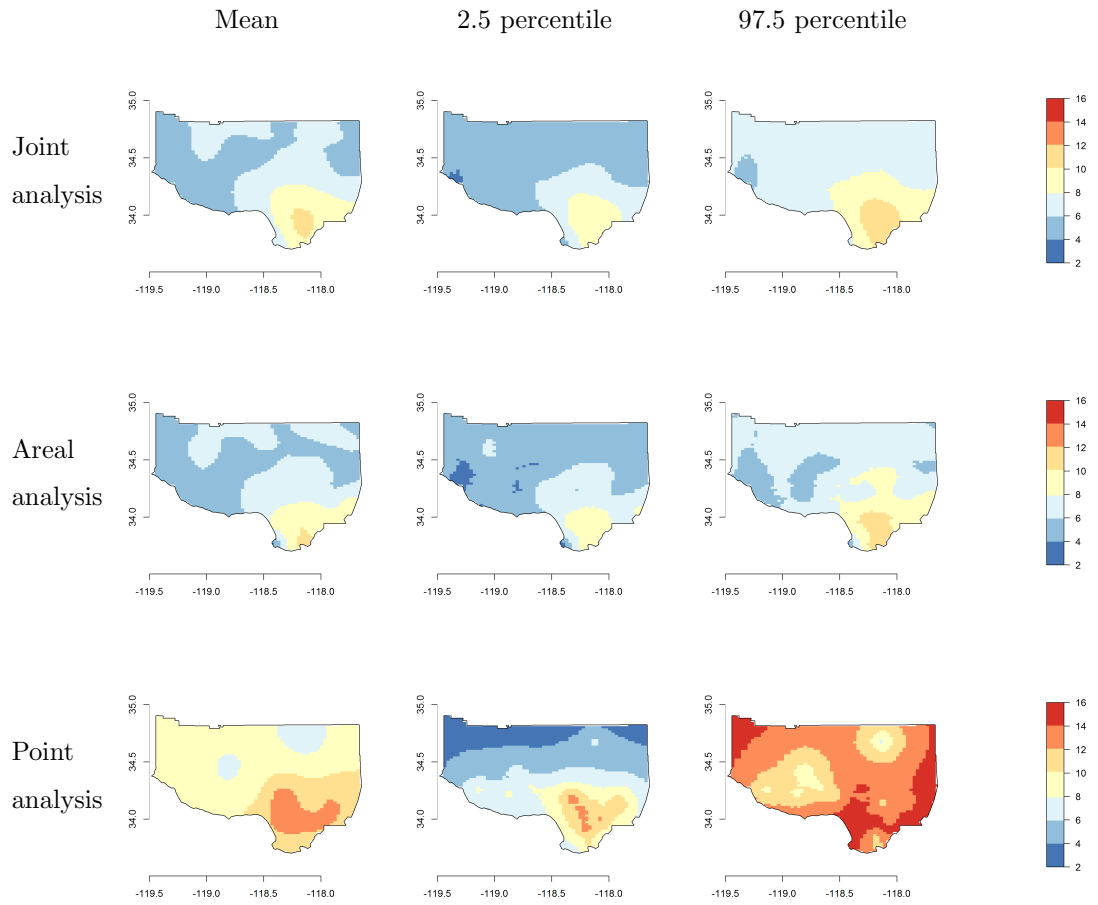


Figure 8: Posterior PM_{2.5} concentration (micrograms per cubic meter) by type of analysis, Los Angeles and Ventura counties, 2011.