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1 **THE INFLUENCE OF SEX, PARASITISM, AND ONTOGENY ON THE**  
2 **PHYSIOLOGICAL RESPONSE OF EUROPEAN EEL (*Anguilla anguilla*)**  
3 **TO AN ABIOTIC STRESSOR**

4 *Short title: Interaction of biotic and abiotic factors on stress response of eels*

5  
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22  
23 **What is already known:**

24 The consequences of different biotic factors and their interaction on the physiological stress  
25 response of eels to abiotic stressors have long been assumed. Yet, very few studies have  
26 explored these relationships using empirical research. Such information is crucial to develop  
27 effective management practices needed to assist with the recovery of the European eel, currently  
28 classified as an Endangered species.

29 **What this study adds:**

30 This study revealed the importance of considering the role of biotic factors (in this case: sex,  
31 parasitism and ontogeny) acting together to influence the stress response of the European eel to  
32 abiotic stressors. To our knowledge this is the first physiological study that simultaneously  
33 examines these different biotic factors. Furthermore, this study is highly relevant as there is a  
34 paucity of information on the influence of biotic factors on the physiological response of the  
35 European eel and other fish species to different abiotic stressors.

37 **ABSTRACT**

38 Migration of adult European eel (*Anguilla anguilla*) from freshwater feeding grounds to oceanic  
39 spawning grounds is an energetically demanding process and is accompanied by dramatic  
40 physiological and behavioural changes. Humans have altered the aquatic environment (e.g.  
41 dams) and made an inherently challenging migration even more difficult; human activity is  
42 regarded as the primary driver of the collapse in eel populations. The neuroendocrine stress  
43 response is central in coping with these challenging conditions, yet, little is known about how  
44 various biotic factors such as sex, parasites, and ontogeny influence (singly and via interactions)  
45 the stress response of eel. In this study, mixed effect models and linear models were used to  
46 quantify the influence of sex, parasitism (*Anguillicola crassus*), life-stage (yellow and silver  
47 eels), and silvering stage on the stress response of eels when exposed to a standardized handling  
48 stressor. The physiological response of eels to a standardized abiotic stressor (netting  
49 confinement in air) was quantified through measurements of blood glucose and plasma cortisol.  
50 The relationships between biotic factors and the activity of gill  $\text{Na}^+/\text{K}^+$  - ATPase was also  
51 examined. Analyses revealed that in some instances a biotic factor acted alone while in other  
52 cases several factors interacted to influence the stress response. Blood glucose concentrations  
53 increased following exposure to the standardized stressor and remained elevated after 4 hours.  
54 Variation in plasma cortisol concentrations following exposure to the stressor were found to be  
55 time-dependent, which was exacerbated by the life-stage and parasitism condition. Males and  
56 non-parasitized silver eels had the highest  $\text{Na}^+/\text{K}^+$ -ATPase activity. Silvering stage was strongly  
57 positively correlated with  $\text{Na}^+/\text{K}^+$ -ATPase activity in female eels. Collectively, these findings  
58 confirm that the factors mediating stress responsiveness in fish are complicated and aspects of  
59 inherent biotic variation cannot be ignored.

60

- 61 **Keywords:** silver eel, yellow eel, stress response, *Anguillicola crassus*, cortisol, glucose,
- 62 Na<sup>+</sup>/K<sup>+</sup>-ATPase activity

## 63 INTRODUCTION

64 In freshwater and marine ecosystems fish are often exposed to natural and anthropogenic  
65 stressors (Arthington et al. 2016). To compensate for the challenge imposed by a stressor, fish  
66 undergo a series of biochemical and physiological changes (i.e., the stress response; Wendelaar  
67 Bonga 1997; Gorissen, and Flik 2017). The glucocorticoid stress response is an essential  
68 mediator of allostasis that maintains stability (homeostasis) or facilitates adaptation to changing  
69 conditions (McEwen and Wingfield 2003; Angelier 2013), therefore promoting the survival and  
70 recovery of individuals (Sapolsky et al. 1999). The stress response is characterized by the  
71 production and release of glucocorticoid steroid hormones (i.e., cortisol in fish) shortly after  
72 the perception of the stressor (Axelrod and Reisine 1984). In the short term, this stress response  
73 is adaptive, providing the fuel (i.e., glucose) needed to respond to a stressor (Mommsen et al.  
74 1999; Barton 2002). However, if the stressor persists, the action of glucocorticoids can occur at  
75 the expense of other life-history components through a reduction in the amount of energy  
76 available for essential functions (Korte et al. 2005). In fish, stress can negatively affect growth,  
77 health (immunocompetence), reproduction, and welfare, and ultimately result in mortality  
78 (Schreck 1981, 2000; Barton 2002; Fuzzen et al. 2011).

79 For diadromous fish species, the transition from life in freshwater (FW) to seawater (SW)  
80 is a very important and a challenging period usually characterized by high levels of mortality  
81 (Bruijs et al. 2009; Piper et al. 2015). The European eel (*Anguilla anguilla*), a catadromous  
82 species, undertakes an outward migration of ~5000–6000 km to spawning grounds in the  
83 Sargasso Sea (van Ginneken et al. 2005; Aarestrup et al. 2009), which is known as the longest  
84 spawning migration among all the species of eels (Aoyama 2009) and is performed without  
85 feeding (Righton et al. 2012). Before migrating to SW eel's life is spent feeding in freshwater  
86 (for up to 25 years) to store enough fat (>20% of the body mass; Tesch 2003) (yellow eel stage)  
87 to fuel migration that may take many months (Righton et al. 2016), as well as, to provide

88 sufficient energy to produce offspring. After attaining an adequate lipid reserve, eels start lipid  
89 mobilization (EELREP 2005; Trischitta et al. 2013) and sexual maturation, metamorphosing  
90 into “silver eels”. During this stage, eels stop feeding, and begin the long migration back to the  
91 Sargasso Sea for spawning (Righton et al. 2012). Males (on average 40 cm) usually start their  
92 migration in August while females (on average bigger than 40 cm) leave later, during October  
93 and December (Tesch 2003).

94 Spawning migration of eels is a complex and energetically demanding process during  
95 which eels are very vulnerable to natural and anthropogenic challenges that can impair their  
96 migratory capacity as they transition from freshwater to saline water (Gollock et al. 2005,  
97 Iversen et al. 2013, Trischitta et al. 2013, Wilson 2013). Durif et al. (2005) described five  
98 different stages of the silvering process in female eels according to their physiological changes  
99 as they prepare for their spawning migration: a growth phase (I and II) a pre-migration phase  
100 (II) and two migration phases (IV and V). In part due to their catadromous lifestyle, European  
101 eel populations have seen marked declines throughout their natural range in the past few  
102 decades and are currently classified as Critically Endangered (Jacoby and Gollock 2014) and  
103 listed under Appendix I-III of the Convention on International Trade in Endangered Species of  
104 Wild Fauna and Flora (CITES 2013). Several factors are thought to have contributed to these  
105 declines including barriers to migration, habitat loss, parasites (e.g. *Anguillicola crassus*),  
106 disease, climate change, bioaccumulation of toxins, predation, changes in ocean currents and  
107 overfishing (Dekker 2003; Knights 2003; Van Ginneken et al. 2005; Belpaire et al. 2009;  
108 Geeraerts and Belpaire 2010; Durif et al. 2011; Kettle et al. 2011; Wahlberg et al. 2014). The  
109 drastic decline of European eel populations has hastened the implementation of management  
110 measures aimed at restoring stocks by preventing mortality during migration (European Union  
111 implemented the Eel Recovery Plan 2007- Council Regulation No. 1100/2007/EC and the  
112 International Council for the Exploitation of the Sea -ICES 2014).

113 Despite the extensive body of literature that has explored the stress response of fish in  
114 general (reviewed in Schreck 2010; Pankhurst 2011), to our knowledge no studies have  
115 specifically explored how biotic characteristics acting in concert may influence the stress  
116 response and recovery in European eel, as analysed in this study. The main goal of this study  
117 was to analyze how individual factors such as sex, parasitic load (non-parasitized and  
118 parasitized with *A. crassus*), and ontogenetic phase (yellow, silver, and different silvering  
119 stages) interact to influence the physiological response to a standardized handling and air  
120 exposure stressor. To determine which biotic characteristics are associated with the stress  
121 response, we used mixed effect and linear models to quantify the physiological responses of  
122 eels. We measured blood parameters (i.e., plasma cortisol and body glucose) immediately  
123 (baseline), 1 hour (stress response) and 4 hours (recovery period) post-exposure to the stressor.  
124 We also tested for relationships between biotic factors and the activity of gill  $\text{Na}^+/\text{K}^+$  - ATPase  
125 given the important role of this gill enzyme in diadromous species. Moreover, plasma cortisol  
126 is also associated with branchial  $\text{Na}^+/\text{K}^+$ -ATPase activity, which plays a central role in whole-  
127 body osmoregulation (Towle 1981; Sancho et al 1997) such that stress has the potential to also  
128 influence osmoregulatory processes.

129

## 130 **MATERIAL AND METHODS**

### 131 *Animals and experimental design*

132 European eels were caught during downstream migration between October and November  
133 of 2014 in a trap located in River Gudenå at Vestbirk hydropower station, at a downstream trap  
134 in Flade Sø and by electrofishing at Bygholm Å and Lake Stigsholm, Denmark. The eels (N=72,  
135 mean total length ( $L_t$ )  $\pm$  S.D. =  $51.9 \pm 8.3$  cm, mean total weight ( $W_t$ )  $\pm$  S.D. =  $249.7 \pm 127.3$   
136 g) were transported and held in three 8000L holding tanks (water temperature 12-15°C) at the  
137 National Institute of Aquatic Resources, Technical University of Denmark, in Silkeborg,

138 Denmark, until the experiments were carried out (holding time of between 5 and 9 days). To  
139 minimize stress during holding and facilitate recovery from capture, transportation and  
140 handling, shelter was provided for the eels. This shelter was comprised of 3.0 and 4.5 cm  
141 diameter by 70 cm long PVC pipes that were placed in the holding tanks. These pipes also  
142 limited the influence of removal of an individual for treatment on the remaining eels in the  
143 holding tank since a single pipe could be removed without disturbing the other eels. Overall,  
144 57 females eels (mean  $L_t \pm S.D. = 54.9 \pm 6.1$  cm, mean  $W_t \pm S.D. = 289.7 \pm 111.9$  g) and 15  
145 males (mean  $L_t \pm S.D. = 40.5 \pm 3.4$  cm, mean  $W_t \pm S.D. = 105.13 \pm 27.2$  g), were tested. Each  
146 eel received the same experimental treatment. First, an eel was removed from the holding tank  
147 by netting a PVC pipe on either end and lifting it from the tank, with minimal disturbance. A  
148 blood sample was then collected within 3-min of capture to act as baseline sample of plasma  
149 cortisol and blood glucose (as per Lawrence et al. 2018). Next, the eel was exposed to a  
150 standardized stressor in the form of a 10-min air exposure, before being moved into an  
151 individual 80-L holding tank with 20-L of water. To measure the magnitude of the stress  
152 response in each eel, blood samples were collected again at one and four hours after their  
153 baseline sample. Eels were not anaesthetised during this procedure because it has been shown  
154 to influence gill  $Na^+/K^+$ -ATPase activity (Toni et al. 2014) – another parameter measured in  
155 this study (details provided in the *Plasma and Gill sample analysis* section) and would have  
156 confounded our ability to measure the stress response. Anesthesia can influence the stress  
157 response in a number of ways – both muting it and also serving as a stressor itself (there is a  
158 significant metabolic demand associated with clearing anesthetics; Neiffer and Stamper 2009).  
159 We acknowledge that the blood sampling at the 1 hr time point would have served as a stressor  
160 that had the potential to influence the stress levels measured at the 4 hr time point but all fish  
161 were handled similarly and this occurred during a period when the stress response was already  
162 at its peak. Stress associated with sampling during the first blood sampling period was simply



163 part of the standardized stressor while stress associated with sampling during the final time  
164 point was irrelevant given that no further sampling would occur. Blood sampling without  
165 anaesthesia is relatively common in the study of stress physiology in wild fish (e.g., Cooke et  
166 al., 2005) including studies that involve repeated sampling of individuals (e.g., Cook et al.  
167 2012). To minimise disturbance of fish during blood sampling, this procedure was always  
168 conducted by the same operator. Fish were euthanized via decapitation using a sharp knife. All  
169 applicable international, national, and/or institutional guidelines for the care and use of animals  
170 were followed. Animal care approval for this study falls under the Danish Animal Experiment  
171 Inspectorate (licence number: 2013-15-2934-00808).

172

### 173 *Individual condition*

174 At the end of each experiment eels were sacrificed and measured for body mass, total length,  
175 body width at maximum body depth, body height at maximum body depth, pectoral fin length  
176 and horizontal and vertical eye diameters. These measurements were used to distinguish males  
177 from females and to calculate three morphometric indices: eye index, fin index and Fulton's  
178 condition factor (Durif et al. 2005; Bolger and Connolly 1989). These indices were used  
179 together with the external morphological characteristics of silver-phase eel (presence of black  
180 corpuscles in the lateral line; dark dorsal part of the body and lighter “silver” ventral region;  
181 and snout shape and dark coloration of the extremities of the pectoral fins and tail), as selective  
182 criteria to distinguish between the yellow and silver phases, as well as to determine the different  
183 silvering stage (stage I to V; Pankhurst 1982; Durif et al. 2005). The swimbladder of each eel  
184 was also removed and any *A. crassus* present in the swimbladder lumen were removed and  
185 enumerated.

186

### 187 *Plasma and Gill sample analysis*

188 Blood samples were obtained by puncture of the caudal vasculature using pre-heparinised  
189 (10 000 USP units/ml heparin sodium: Sandoz, Canada), needles (25 G 1/2'') and 1 ml syringes  
190 (BD Plastipak, 1ml) and the blood was stored briefly in ice. The total sampling time never  
191 exceeded 3 min. The volume of blood removed for each sample was approximately 0.2 ml.  
192 After each blood sample was obtained, sub-samples were removed for immediate determination  
193 of blood glucose concentrations using a glucose meter (Accuchek, Roche Diagnostics; Stoot et  
194 al. 2014) and the remainder of the sample was centrifuged for 10 min at 4,000 RPM to separate  
195 plasma from the blood cells. The aliquoted plasma was immediately frozen in liquid N<sub>2</sub> and  
196 then stored frozen at -80°C for later analysis. Individual plasma cortisol concentrations (ng/mL)  
197 were determined according to the radioimmunoassay procedure described in Pottinger and  
198 Carrick (2001) with two minor adjustments. The antibody used in this study was IgG-F-2 rabbit  
199 anti-cortisol (IgG Corp; Nashville, TN, USA) and tracer ([1,2,6,7]<sup>3</sup>H-cortisol, 2.59 TBq/mmol;  
200 Perkin-Elmer, U.K.) was added in a 25 µL aliquot of buffer at the same time as the antibody  
201 was dispensed.

202 Measurement of gill Na<sup>+</sup>/K<sup>+</sup>-ATPase activity followed procedures outlined by McCormick  
203 (1993). Gill filaments from the second right gill arch were removed from each eel, placed in a  
204 tube containing ice-cold SEI buffer (300 mM sucrose, 20 mM Na<sub>2</sub>EDTA, 50 mM imidazole,  
205 pH 7.3) frozen in N<sub>2</sub> and stored at -80°C until analysed. Gill homogenates were centrifuged at  
206 1000 g for 1 min and the supernatant was assayed for ATPase activity in the presence and  
207 absence of 0.5 mM ouabain. Each assay was run in triplicate. Protein content was measured by  
208 the Lowry (1951) method modified for a plate reader. The difference between the two  
209 determined activities (with and without ouabain) was calculated as the Na<sup>+</sup>/K<sup>+</sup>-ATPase activity.

210

211 *Statistical Analysis*

212 Data were analysed for normality using the Shapiro-Wilcoxon test. To meet the normality  
213 requirements of parametric analysis, cortisol and glucose data were log(x) transformed (log-  
214 cortisol (logC) and log-glucose (logG), respectively.

215 Response variables logC and logG were fitted with linear mixed effects models (LME)  
216 with individual fish as a random factor and time (baseline, 1 hour and 4 hours), sex (male or  
217 female), life-stage (yellow or silver) and parasite condition (non-parasitized vs parasitized with  
218 *A. crassus*) as fixed effects. Silvering stages (I to V) could not be compared independently due  
219 to the small number of individual females in each stage; therefore, individuals were grouped in  
220 three groups according to their similarities of development (after Durif et al. 2005). Group 1  
221 included all the individuals belonging to the silvering stage I and II, group 2 had individuals in  
222 stage III and group 3 had individuals in stage IV and V. To understand the effects of silvering  
223 stage on logC and logG, a new LME model was run with silvering condition included as a fixed  
224 effect and sex and stage (redundant factor) removed as possible predictors. Only females  
225 (N=57: silver N=35; yellow N=22) were used in this analysis as the number of silver males was  
226 very low for a statistical analysis (N=15: silver N=6; yellow N=9)

227 Linear models (LM) were used to assess the effect of sex, life-stage, parasite condition and  
228 silvering stages on gill Na<sup>+</sup>/K<sup>+</sup>-ATPase activity. Data were analysed using the *nlme* function  
229 implemented in the R statistical environment (package version 3.1-117, R core team ; Pinheiro  
230 et al, 2017). To compare model fits objectively, and determine which was the most appropriate,  
231 an information theoretic approach was performed to compare models using Akaike's  
232 information criterion (AIC; Akaike 1974; Burnham and Anderson, 2002). Models were  
233 validated by examining histograms of the normalized residuals, plotting the normalized  
234 residuals against fitted values. The final models were refitted using maximum likelihood (ML).  
235 Mean values are reported together with standard error (mean ± S.E) and results were considered  
236 significant for  $\alpha < 0.05$ .

237

## 238 **RESULTS**

### 239 *Parasitism*

240 Overall, 20 eels were parasitized with *A. crassus*. The number of parasites in the eels varied  
241 between 1 and 11 individuals per specimen, and it was different according to the : sex (females:  
242 N= 19, males N=1), life-stage (yellow eels N= 7, silver eels N=13) and silvering condition  
243 (life stage I: N= 7, life stage II, N= 8, Life stage III: N= 4). The different number of parasites  
244 in each silvering stage resulted in different levels of Glucose, Cortisol and Gill Na<sup>+</sup>/K<sup>+</sup>-ATPase  
245 activity (Table 1).

246

### 247 *Glucose*

248 The final model for blood glucose (logG) contained time and parasite condition as the main  
249 explanatory factors (logG~time\*parasitism; AIC : -222.05, dF: 8). Blood glucose varied  
250 significantly with time (Table 2), and increased from  $4.5 \pm 0.3$  mmol/L (mean  $\pm$  S.E) in  
251 unstressed eels at 0 h to  $7.6 \pm 0.4$  mmol/L at 1 hour after the stressor and to  $10.2 \pm 0.5$  mmol/L  
252 4 hours following the stressor (Fig. 1). Temporal variation of glucose was similar between  
253 parasitized and non-parasitized eels, with parasitized eels exhibiting slightly higher overall  
254 glucose levels ( $7.5 \pm 0.4$  mmol/L) than non-parasitized eels ( $7.1 \pm 0.3$  mmol/L; Fig. 1), although  
255 this difference was not significant.

256 Variation of plasma glucose in female eels was best explained by a model that included  
257 both time and the interaction of parasite condition with silvering (logG~  
258 time+parasitism\*silvering; AIC: -145.53, dF: 8). Plasma glucose levels at 1h ( $7.7 \pm 0.5$   
259 mmol/L) and 4h ( $10.2 \pm 0.6$  mmol/L) after the stressor were significantly different from the  
260 values in unstressed eels ( $4.5 \pm 0.4$  mmol/L) (Table 2). Parasitism and life-stage were also  
261 important covariates in explaining the variation of plasma glucose in female eels, improving

262 the statistical model, nevertheless their effects were not statistically significant (Table 2).  
263 Indeed, similar number of parasites in each silvering stage led to approximately the same values  
264 on plasma glucose on eels. Although, minor, still, blood glucose increased the number of *A.*  
265 *crassus* existent in each eel, in particularly in eels parasitized with more than 4 individuals.  
266 (Table 1). Overall, mean glucose levels in non-parasitized female eels were lower ( $6.8 \pm 0.4$   
267 mmol/L) than those in parasitized female eels ( $7.9 \pm 0.4$  mmol/L).

268

### 269 *Plasma cortisol*

270 Plasma cortisol levels varied significantly with time (Table 2) and were also dependent on  
271 parasite condition and life stage of eels (logC~time\*parasitism + parasitism\*life-stage; AIC:  
272 29.42, dF: 10). Mean plasma cortisol levels significantly increased in the first hour after the  
273 stressor from  $29.19 \pm 4.0$  ng/mL to  $57.84 \pm 3.48$  ng/mL after which they decreased to levels  
274 slightly higher than those in unstressed eels ( $37.73 \pm 3.6$  ng/mL) (Table 2). Although non-  
275 parasitized eels exhibited higher levels of cortisol overall ( $48.6 \pm 3.8$  ng/mL) than parasitized  
276 eels ( $33.1 \pm 2.4$  ng/mL) (Table 2), net changes in variation was larger in parasitized eels (Fig.  
277 2a). This was particularly evident in the first hour where mean plasma cortisol concentrations  
278 rose significantly from baseline levels of  $16.9 \pm 2.0$  ng/mL to  $54.2 \pm 3.2$  ng/mL (Table 2, Fig.  
279 2a). Overall, non-parasitized silver eels had higher plasma cortisol levels ( $58.6 \pm 6.7$  ng/mL)  
280 when compared to non-parasitized yellow eels ( $39.15 \pm 3.5$  ng/mL). Nonetheless, parasitism  
281 strongly influenced cortisol response in silver eels, which had the lowest levels of cortisol found  
282 ( $30.7 \pm 2.5$  ng/mL) (Fig. 2b, Table 2).

283 In female eels, plasma cortisol concentrations were found to vary with silvering stage and  
284 the interaction between time and parasitism (logC~silvering+time\*parasitism; AIC : 36.06, dF:  
285 10). Female eels belonging to the maximum silvering stage group exhibited higher levels of  
286 plasma cortisol ( $58.02 \pm 6.3$  ng/mL), when compared to eels of the second ( $27.9 \pm 2.3$  ng/mL)

287 and first group ( $35.5 \pm 3.3$  ng/mL) (Table 2). Parasitized female eels exhibited the lowest levels  
288 of plasma cortisol ( $32.3 \pm 2.8$  ng/mL) when compared to non-parasitized eels ( $49.1 \pm 5.0$   
289 ng/mL). The variation of plasma cortisol in female parasitized eels was found to increase with  
290 the number of *A. crassus* (Table 1). When parasitized with more than 4 individuals eels had an  
291 increased on plasma cortisol levels. Nevertheless when in the last silvering stage, even a small  
292 number *A. crassus* appears to elicit a strong increase of plasma cortisol. Variation in plasma  
293 cortisol was also time dependent (Table 2); plasma cortisol significantly increased from  $27.9 \pm$   
294  $5.0$  ng/mL to  $56.5 \pm 4.1$  ng/mL in the first hour following the stressor, decreasing to values  
295 close to the baseline levels after 3h ( $38.6 \pm 4.5$  ng/mL) (Table 2). This temporal variation was  
296 found to be related to the parasitism status of the individual (Table 2). After exposure to a  
297 stressor, parasitized eels exhibited a stronger increase in cortisol levels when compared to non-  
298 parasitized eels (Table 2). This variation was clearly evident in the first hour following  
299 disturbance (non- parasitized eels:  $61.9 \pm 7.1$  ng/mL, parasitized eels:  $53.9 \pm 3.5$  ng/mL) (Table  
300 2). However, by the 4h time point, plasma cortisol levels had recovered to near the levels seen  
301 in non-parasitized eels (4h:  $44.4 \pm 8.0$  ng/mL, baseline:  $39.9 \pm 9.7$  ng/mL), but not in parasitized  
302 eels (4h:  $26.2 \pm 3.0$  ng/mL, baseline:  $14.7 \pm 2.2$  ng/mL) (Table 2).

303

#### 304 *Gill Na<sup>+</sup>/K<sup>+</sup>-ATPase activity*

305 Gill Na<sup>+</sup>/K<sup>+</sup>-ATPase activity varied between individuals of different sexes, life stages and  
306 parasitism levels (Na<sup>+</sup>/K<sup>+</sup>-ATPase activity~ sex + life-stage\*parasitism; AIC: 293.97, dF: 6).  
307 Males exhibited higher levels of Na<sup>+</sup>/K<sup>+</sup>-ATPase ( $8.71 \pm 0.8$  μmol ADP/mg protein/h) than  
308 females ( $6.02 \pm 0.5$  μmol ADP/mg protein/h) (Table 2). Na<sup>+</sup>/K<sup>+</sup>-ATPase levels were found to  
309 vary with the life-stage of eels, with the highest values found in silver eels ( $7.97 \pm 0.6$  μmol  
310 ADP/mg protein/h) when compared to yellow eels ( $4.74 \pm 0.5$  μmol ADP/mg protein/h) (Fig.  
311 3, Table 2). Within life stages the variation of Na<sup>+</sup>/K<sup>+</sup>-ATPase activity was conditioned by the

312 parasitism level, particularly in silver eels where non-parasitized individuals exhibited  
313 significantly higher Na<sup>+</sup>/K<sup>+</sup>-ATPase activity (10.26 ± 0.8 μmol ADP/mg protein/h) than  
314 parasitized silver eels (7.22 ± 0.5 μmol ADP/mg protein/h) (Fig. 3, Table 2).

315 In female eels, gill Na<sup>+</sup>/K<sup>+</sup>-ATPase activity increased through silvering stage (Table 2).  
316 The highest values of Na<sup>+</sup>/K<sup>+</sup>-ATPase activity were found in the third silvering group (7.74 ±  
317 0.8 μmol ADP/mg protein/h), decreasing in the second group (5.79 ± 0.6 μmol ADP/mg  
318 protein/h) and were lowest in the first group (4.02 ± 0.5 μmol ADP/mg protein/h). Gill Na<sup>+</sup>/K<sup>+</sup>-  
319 ATPase activity in female parasitized eels increased with the number of *A. crassus*, in particular  
320 in eels parasitized with more than 4 *A. crassus* (Table 1).

321

## 322 **DISCUSSION**

323 In this study we examined the effect of sex, parasite burden, and ontogeny, alone and in  
324 combination, on the stress response and Na<sup>+</sup>/K<sup>+</sup>-ATPase activity of European eels when  
325 exposed to a standardized handling stressor. To our knowledge, this is the first study to examine  
326 the impact of the interaction of different biotic factors on the physiological response of eels.  
327 The results of this study revealed a physiological response to our experimental handling stressor  
328 with the extent of the response modulated by biotic factors. Interestingly, in some instances  
329 biotic factors acted alone while in other cases several factors interacted to influence the  
330 physiological response.

331 Eels subjected to the stressor exhibited significantly higher concentrations of glucose  
332 throughout the 4 h duration of the study, with the most significant increase observed during the  
333 first hour after disturbance. The prolonged elevation of glucose reflects a mobilisation of energy  
334 to provide short-term support for immediate coping activities to promote survival. Parasitized  
335 and non-parasitized eels showed similar levels of glucose, a result consistent with Gollock et  
336 al. (2004) and their study on parasite-mediated stress responses to handling stressors in

337 European eel. Moreover, we observed that the number of *A. crassus* in female parasitized eels  
338 led to slightly higher concentrations of blood glucose, but this was not significantly different  
339 between the three silvering stages.

340 As expected, eels also exhibited a strong cortisol response to stress. Cortisol significantly  
341 increased in the first hour after exposure to the stressor followed by a decrease in the next 3  
342 hours. This is noteworthy given that we repeatedly sampled fish such that there would have  
343 been some level of stress associated with blood sampling at the 1 hour time point. Despite that,  
344 cortisol recovery was still evident at the 4 hour time point. When considering both males and  
345 females, the temporal variation in cortisol was similar in both parasitized and non-parasitized  
346 eels; parasitized eels exhibited a stronger response in terms of increment of cortisol when  
347 compared to non-parasitized eels. This finding suggests that parasitic state plays an important  
348 role in the stress response of eels. The similarity on the variation and levels of cortisol between  
349 this study and Gollock et al. (2004), as well as, the fact that eels used in this study were also  
350 wild and may have been infected by *A. crassus* for a long period of time, support the argument  
351 of Gollock et al. (2004) that the results obtained can reflect an adaptation to the effects of  
352 chronic parasitism. Moreover, Sures et al. (2001) found that there is a strong stress response of  
353 eels to the larval and young adult stages of *A. crassus*, but no chronic response to older adults.  
354 Although we have not analyzed the life stage of *A. crassus* infecting the tested specimens, it is  
355 possible that the tested eels could have been in an early onset of infection. The environmental  
356 characteristics of the system where eels lived (water temperature, water salinity) also played an  
357 important role on the results obtained as it is known that the spread, extent and intensity of  
358 infestation by *A. crassus* is dependent on water salinity and the age and size of the fish (Sures  
359 et al. 2001; Lefebvre and Crivelli 2012). Differences in plasma cortisol levels between non-  
360 parasitized and parasitized specimens was strongly evident on female silver eels, even if overall  
361 there were no significant differences were found between the two life-stages. This evidence that



362 there is a synergistic influence of multiple stressors on the stress response. Female eels  
363 categorized as being in the last stage of silvering (III) exhibited the highest levels of plasma  
364 cortisol which may have some implications for reproductive function. Moreover, eels on the  
365 third silvering stage were found to be more susceptible to the presence of parasites, as the  
366 highest levels of plasma cortisol were found even when the number of parasites was low (<4  
367 individuals). Parasitism on the last silvering stage may negatively influence migration and  
368 reproduction of the eels.

369 High levels of cortisol for prolonged periods of time have been shown to play a role in  
370 energy mobilization as result of its lipolytic effect increasing free fatty acid levels, reduce  
371 growth rate by increasing the pituitary gonadotropin, reduce immune function, and disrupt fish  
372 reproduction function by depressing sex steroid levels (Huang et al. 1999). The implications of  
373 high levels of cortisol for prolonged periods during exposure to chronic or frequent intermittent  
374 acute stressors on eel reproduction are therefore potentially important. The morphological and  
375 physiological transformation of yellow eels to the silver phase and the initiation of their  
376 spawning migration is only triggered when the levels of lipids is >20% of the body mass (Palstra  
377 et al. 2010; van den Thillart et al. 2009). As such, elevations in cortisol have the potential to  
378 influence both maturation and spawning migrations. Cortisol is also known to be related to SW  
379 adaptation of fish helping them to acclimate to a hyperosmotic environment (SW) by increasing  
380 hypoosmo-regulatory capacity (Mommsen et al. 1999). Cortisol mediates SW-acclimation by  
381 stimulating the gill chloride cell proliferation and Na<sup>+</sup>/ K<sup>+</sup>-ATPase activity ensuring the  
382 transmembrane transfer of the cations Na<sup>+</sup> and K<sup>+</sup> and affecting the transepithelial movements  
383 of cations in gills (Madsen 1990a; Sancho et al 1997; McCormick 1995). The stimulatory role  
384 of cortisol on gill Na<sup>+</sup>/K<sup>+</sup>-ATPase activity in the American eel (*Anguilla rostrata*) was  
385 previously shown by Butler et al. (1972), on their study of the effects of environmental salinity  
386 and adrenocortical steroids on Na<sup>+</sup>/ K<sup>+</sup>-ATPase activity. Also, studies on salmonids showed

387 that gill  $\text{Na}^+/\text{K}^+$ -ATPase activity responds positively to injections of cortisol in Atlantic  
388 salmon, *Salmo salar* (Bisbal and Specker 1991), rainbow trout, *Oncorhynchus mykiss* (Madsen  
389 1990a), and sea trout *Salmo trutta* (Madsen 1990b; Fontainhas-Fernandes et al. 2003).

390 Gill  $\text{Na}^+/\text{K}^+$ -ATPase activity was significantly higher in silver eels than in yellow eels.  
391 This result was particularly evident in non-parasitized silver eels since parasitized silver eels  
392 appeared to have suppressed  $\text{Na}^+/\text{K}^+$  - ATPase activity. Despite the low values of  $\text{Na}^+/\text{K}^+$  -  
393 ATPase activity in female parasitized eels, it was observed that these values increased with  
394 silvering stage as well as with the number of *A. crassus*. It is known that  $\text{Na}^+/\text{K}^+$ -ATPase  
395 activity plays a crucial role in the osmoregulation of eels, thus the suppression of such protein  
396 will limit the success of eels in salt water and therefore compromise their migration,  
397 reproduction and concomitant survival. Control and mitigation of the levels of *A. crassus* in  
398 eels, in particularly on the latest stages of maturation of this species are critical and must be  
399 developed. Such conservation measures will contribute to the reduction of the decline of the  
400 European eel, currently classified as critically. Considering the well-known effects of cortisol  
401 on  $\text{Na}^+/\text{K}^+$ -ATPase activity (McCormick, 1995), the highest levels of this parameter on non-  
402 parasitized silver eels may have been related to the high levels of cortisol found in these  
403 specimens. Furthermore, non-parasitized yellow eels exhibited the lowest levels of gill  $\text{Na}^+/\text{K}^+$   
404  $\text{K}^+$ -ATPase activity. Once again, parasitism acting synergistically with other biotic factors  
405 affect ion regulation via an indirect effect on gill  $\text{Na}^+/\text{K}^+$  regulation.

406 The highest levels of  $\text{Na}^+/\text{K}^+$  - ATPase activity were found in males and can be a  
407 consequence of different stages of sexual maturation achieved by the specimens. Considering  
408 that males initiate their migration earlier than females (Palstra et al. 2007; 2010), and that the  
409 experiments were carried out at the end of October through the beginning of November, our  
410 findings may then be related with the fact that most males could have been in a more advanced  
411 silvering stage than the females. Although expected, this is an interesting result as it indicates

412 that the success of spawning of males become more susceptible and/or compromised by  
413 environmental conditions (e.g. parasite load) earlier than females. Gill  $\text{Na}^+/\text{K}^+$  - ATPase activity  
414 increased with silvering stage. Nevertheless, this variation was exacerbated in non-parasitized  
415 eels, which exhibited an elevation of Gill  $\text{Na}^+/\text{K}^+$  - ATPase activity between silvering stage 2  
416 and 3. Again, this points toward the idea that parasitized eels can be adapted to deal with stress,  
417 and therefore their response to stress would be less severe.

418

## 419 **CONCLUSIONS**

420 This paper documented a strong glucose and cortisol response of European eels to a holding  
421 stressor (netting confinement in air) that was mediated by the interaction of several biotic  
422 factors. Such biotic interactions were also found to play an important role in the variation of  
423  $\text{Na}^+/\text{K}^+$ -ATPase activity. Because we assessed the role of multiple biotic factors  
424 simultaneously we had the ability to test their influence alone and in combination which is a  
425 robust approach relative to examining them individually (e.g., just parasite burden) which has  
426 been the typical approach in the literature thus far. Indeed, we revealed that the stress response  
427 of eels was found to differ between life stage, sex and parasitism condition, as well as, with the  
428 number of parasites. Parasitism, mainly when acting together with other biotic stressors, plays  
429 an important role in the physiological response of the eels to stressors and presumably has the  
430 potential to influence the maturation, reproductive and osmoregulatory processes in this  
431 species. Future studies that examine the influence of biotic factors acting alone and interacting  
432 under different abiotic conditions are needed to better understand the role of stress in the  
433 different life stages, sex, health conditions and other physiological characteristics of wild fish.

434

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444

#### 445 **REFERENCES**

446 Aarestrup K., F. Økland, M.M. Hansen, D. Righton, P. Gargan, M. Castonguay, L. Bernatchez,  
447 P. Howey, H. Sparholt, M.I. Pedersen, and R.S. McKinley. 2009. Oceanic spawning migration  
448 of the European Eel (*Anguilla anguilla*). *Science* 325: 1660.

449 Angelier F. and J.C. Wingfield. 2013. Importance of the glucocorticoid stress response in a  
450 changing world: theory, hypotheses and perspectives. *General and Comparative*  
451 *Endocrinology*, 10th International Symposium on Avian Endocrinology 190: 118–128.

452 Aoyama J. 2009. Life history and evolution of migration in Catadromous Eels (Genus  
453 *Anguilla*). *Aqua Biosci. Monogr.* 2:1–42.

454 Arthington A.H., N.K. Dulvy, W. Gladstone, and I. J. Winfield. 2016. Fish conservation in  
455 freshwater and marine realms: status, threats and management. *Aquat. conserv. mar. freshw.*  
456 *ecosys* 26:838-857.

457 Axelrod J. and T. D. Reisine. 1984. Stress hormones: Their interaction and regulation. *Science*  
458 224: 452-459.

459 Barton B.A. 2002. Stress in fishes: a diversity of responses with particular reference to changes  
460 in circulating corticosteroids. *Integr. Comp. Biol.* 42: 517–525.

461 Belpaire C.G.J., G. Goemans, C. Geeraerts, P. Quataert, K. Parmentier, and P. Hagel, and J. De  
462 Boer. 2009. Decreasing eel stocks: survival of the fattest? *Ecol. Freshw. Fish* 18: 197–214.

463 Bisbal GA. and J. L. Specker. 1991. Cortisol stimulates hypo-osmoregulatory ability in Atlantic  
464 salmon, *Salmo salar* L. *J. Fish Biol.* 39: 421–432.

465 Boetius I. and J. Boetius. 1980. Experimental maturation of female silver eels, *Anguilla*  
466 *anguilla*. Estimates of fecundity and energy reserves for migration and spawning. *Dana* 1: 1–  
467 28.

468 Bolger T. and P.L. Connolly. 1989. The selection of suitable indices for the measurement and  
469 analysis of fish condition. *J. Fish Biol.* 34:171-182.

470 Bruijs M. and C. Durif. 2009. Silver eel migration and behaviour. In *Spawning Migration of*  
471 *the European Eel*, ed. G. Van den Thillart, S. Dufour and J.C. Rankin, 65–95. New York:  
472 Springer.

473 Burnham K.P. and D.R. Anderson. 2002. *Model Selection and Multimodel Inference* Springer  
474 Science, Business Media

475 Butler D.G. and F.J. Carmichael. 1972. (Na<sup>+</sup>/ K<sup>+</sup>)-ATPase activity in eel (*Anguilla rostrata*)  
476 gills in relation to changes in environmental salinity: Role of adrenocortical steroids, *Gen.*  
477 *Comp. Endocrinol.* 19: 421-427.

478 CITES. 2013. *Convention on International Trade in Endangered Species of Wild Fauna and*  
479 *Flora*. Bangkok: CITES, 1973–2013.

480 Cook K.V., C.M. O'Connor, S.H. McConnachie, K.M. Gilmour, and S.J. Cooke. 2012.  
481 Condition dependent intra-individual repeatability of stress-induced cortisol in a freshwater  
482 fish. *Comp. Biochem. Physiol., Part A* 161:337-343

483 Cooke S.J., G.T. Crossin, D. Patterson, K. English, S.G. Hinch, J.L. Young, R. Alexander, M.C.  
484 Healey, G. Van Der Kraak, and A.P. Farrell. 2005. Coupling non-invasive physiological and  
485 energetic assessments with telemetry to understand inter-individual variation in behaviour and  
486 survivorship of sockeye salmon: development and validation of a technique. *J. Fish Biol.*  
487 67:1342-1358.

488 Dekker W. 2003. Did lack of spawners cause the collapse of the European eel, *Anguilla*  
489 *anguilla*? *Fish. Manag. Ecol.* 10: 365–376.

490 Durif C., S. Dufour, and P. Elie. 2005. The silvering process of the eel: a new classification from  
491 the yellow resident stage to the silver migrating stage. *J Fish Biol* 66:1–19

492 Durif C. M. F., J. Gjosaeter, and L.A. Vollestad. 2011. Influence of oceanic factors on *Anguilla*  
493 *anguilla* (L.) over the twentieth century in coastal habitats of the Skagerrak, southern Norway.  
494 *Proc. R. Soc. Lond. B Biol. Sci.* 278: 464–473.

495 EELREP. 2005. Estimation of the reproduction capacity of European eel. Final report.  
496 Available via <http://www.fishbiology.net/eelrepsum.html>. Accessed 13 Oct 2009

497 Fontáinhas-Fernandes A., E. F. Gomes, M. A. Reis-Henriques, and J. Coimbra. 2003. Effect of  
498 cortisol on some osmoregulatory parameters of the teleost, *Oreochromis niloticus* L., after  
499 transference from freshwater to seawater. *Arq Bras Med Vet Zootec* 55:562–567.

500 Fuzzen M., N. J Bernier, and G. Van Der Kraak. 2011. Stress and reproduction. *Hormones and*  
501 *reproduction in vertebrates*, 1: 103-117.

502 Geeraerts C. and C. Belpaire 2010. The effects of contaminants in European eel: a review.  
503 Ecotoxicology 19: 239–266.

504 Gollock M.J., C.R. Kennedy, E.S. Quabius, and Brown J.A. 2004. The effect of parasitism of  
505 European eels with the nematode, *Anguillicola crassus* on the impact of netting and aerial  
506 exposure. Aquaculture 233: 45–54.

507 Gollock M.J., C.R. Kennedy, and J.A. Brown. 2005. European eels, *Anguilla anguilla* (L.),  
508 infected with *Anguillicola crassus* exhibit a more pronounced stress response to severe hypoxia  
509 than uninfected eels. J. Fish. Dis. 28: 429–436.

510 Gorissen, M. and G. Flik. 2016. 3- The endocrinology of the stress response in fish: an  
511 adaptation-physiological view. Fish physiology. 35 :75-111.

512 Howey P., H. Sparholt, M.I. Pedersen, and R.S. McKinley. 2009. Oceanic spawning migration  
513 of the European Eel (*Anguilla anguilla*). Science 325: 1660.

514 Huang Y.S., K. Rousseau, M. Sbaihi, N. Le Belle, M. Schmitz, and S. Dufour. 1999. Cortisol  
515 Selectively Stimulates Pituitary Gonadotropin  $\beta$ -Subunit in a Primitive Teleost, *Anguilla*  
516 *anguilla*. Endocrinol. 140: 1228–1235.

517 ICES. 2014. Report of the joint EIFAAC/ICES/GFCM working group on eel. 3–7 November  
518 2014, Rome, Italy.

519 Iversen M.H., F. Økland, E.B. Thorstad, and B. Finstad. 2013. The efficacy of Aqui-S vet. (iso-  
520 eugenol) and metomidate as anaesthetics in European eel (*Anguilla anguilla* L.), and their  
521 effects on animal welfare and primary and secondary stress responses. Aquacult. Res 44: 1307-  
522 1316.

523 Jacoby D. and M. Gollock. 2014. *Anguilla anguilla*. The IUCN red list of threatened species.  
524 Version 2014.3. [www.iucnredlist.org](http://www.iucnredlist.org).

525 Kettle A.J., L. Asbjorn Vollestad, and J. Wibig. 2011. Where once the eel and the elephant were  
526 together: decline of the European eel because of changing hydrology in southwest Europe and  
527 northwest Africa? *Fish Fish.* 12: 380–411.

528 Knights B. 2003. A review of the possible impacts of long-term oceanic and climate changes  
529 and fishing mortality on the recruitment of anguillid eels of the Northern Hemisphere. *Sci. Total*  
530 *Environ.* 310: 237–244.

531 Korte S. M., J.M. Koolhaas, J.C. Wingfield, and B.S. McEwen. 2005. The Darwinian concept  
532 of stress: benefits of allostasis and costs of allostatic load and the trade-offs in health and  
533 disease. *Neurosc Biobehav R.* 29: 3-38.

534 Lawrence, M.J., S. Jain-Schlaepfer, A.J. Zolderdo, D.A. Algera, K.M. Gilmour, A.J. Gallagher,  
535 and S.J. Cooke. 2018. Are 3-minutes good enough for obtaining baseline physiological  
536 samples from teleost fish? *Can. J. Zool.* 00:000-000.

537 Lefebvre F. and A.J. Crivelli. 2012. Salinity effects on anguillicolosis in Atlantic eels: a natural  
538 tool for disease control. *Mar Ecol-Prog Ser.* 471:193-202.

539 Lowry O.H., N.J. Rosebrough, A.L. Farr, and R.J. Randall. 1951. Protein measurement with  
540 the Folin Phenol reagent. *J. Biol. Chem.* 193: 265-275

541 Madsen S.S. 1990a. Enhanced hypoosmoregulatory response to growth hormone after cortisol  
542 treatment in immature rainbow trout, *Salmo gairdneri*. *Fish Physiol. Biochem.* 8: 271–279.



543 Madsen S.S. 1990b. The role of cortisol and growth hormone in seawater adaptation and  
544 development of hypoosmoregulatory mechanisms in sea trout parr (*Salmo trutta trutta*). Gen.  
545 Comp. Endocrinol. 79:1–11.

546 McCormick S.D. 1993. Methods for nonlethal gill biopsy and measurement of Na<sup>+</sup>, K<sup>+</sup>-ATPase  
547 activity. Can. J. Fish. Aquat. Sci. 50: 656–658.

548 McCormick S.D. 1995. Hormonal control of gill Na<sup>+</sup>, K<sup>+</sup>-ATPase and chloride cell function.  
549 In C. M. Wood and T.J. Shuttleworth (eds.), Fish physiology, Vol. XIV, Ionoregulation:  
550 Cellular and molecular approaches, pp. 285–315. Academic Press, New York.

551 McEwen B.S. and J.C. Wingfield. 2003. The concept of allostasis in biology and bio-medicine.  
552 Horm. and Behav. 43: 2–15.

553 Mommsen T.P., M.M. Vijayan, and T.W. Moon. 1999. Cortisol in teleosts: dynamics,  
554 mechanisms of action, and metabolic regulation. Rev. Fish Biol. Fish. 9: 211–268.

555 Neiffer D. L. and M.A. Stamper. 2009. Fish sedation, anesthesia, analgesia, and euthanasia:  
556 considerations, methods, and types of drugs. ILAR journal, 50: 343-360.

557 Palstra A.P., D.F.M. Heppener, V.J.T. van Ginneken, C. Szekely, G.E.E.J. van den Thillart.  
558 2007. Swimming performance of silver eels is severely impaired by the swim-bladder parasite  
559 *Anguillicola crassus*. J. Exp. Mar. Biol. Ecol. 352: 244–256.

560 Palstra A. P. and G.E.E.J. van den Thillart. 2010. Swimming physiology of European silver  
561 eels (*Anguilla anguilla* L.): energetic costs and effects on sexual maturation and reproduction.  
562 Fish Physiol. Biochem. 36: 297–322.

563 Pankhurst N.W. 1982. Relation of visual changes to the onset of sexual maturation in the  
564 European eel *Anguilla anguilla* (L.). J. Fish Biol. 21: 127–140.

565 Pankhurst N.W. 2011. The endocrinology of stress in fish: an environmental perspective.  
566 *General and Comparative Endocrinology* 170: 265–275.

567 Pelster B. 2015. Swimbladder function and the spawning migration of the European eel  
568 *Anguilla anguilla*. *Frontiers in Physiology* 5: 486.

569 Pinheiro J., D. Bates, S. DebRoy S., D. Sarkar and R Core Team (2017). nlme: Linear and  
570 Nonlinear Mixed Effects Models. R package version 3.1-131, [https://CRAN.R-](https://CRAN.R-project.org/package=nlme)  
571 [project.org/package=nlme](https://CRAN.R-project.org/package=nlme).

572 Piper A.T., C. Manes, F. Siniscalchi, A. Marion, R.M. Wright, and P.S. Kemp. 2015. Response  
573 of seaward migration European eel (*Anguilla anguilla*) to manipulated flow fields. *Proc. R. Soc.*  
574 *B.* 282.

575 Pottinger T.G. and T.R. Carrick. 2001. Stress responsiveness affects dominant-subordinate  
576 relationships in rainbow trout. *Hor. Behav.* 40: 419-427

577 Righton D., K. Aarestrup, D. Jellyman, P.Sébert, G. van den Thillart, and K. Tsukamoto. 2012.  
578 The *Anguilla* spp. migration problem: 40 million years of evolution and two millennia of  
579 speculation. *J. Fish Biol.* 81: 365–386.

580 Righton D., H. Westerberg, E. Feunteun, F. Økland, P. Gargan, E. Amilhat, Julian Metcalfe, J.  
581 Lobon-Cervia, N. Sjöberg, J. Simon, A. Acou, M. Vedor, A. Walker, T. Trancart, U. Brämick,  
582 and K. Aarestrup. 2016. Empirical observations of the spawning migration of European eels:  
583 The long and dangerous road to the Sargasso Sea. *Science Advances*, 2, e1501694.

584 Sancho E., M.D. Ferrando, and E. Andreu. 1997. Inhibition of gill Na,K -ATPase activity in  
585 the eel, *Anguilla anguilla*, by fenitrothion. *Ecotoxicol. Environ Saf.* 38: 132–136.

586 Sapolsky R.M. 1999. Glucocorticoids, stress, and their adverse neurological effects: relevance  
587 to aging. *Exp Gerontol* 34: 721–732.

588 Schreck C.B. 1981. Stress and compensation in teleostean fishes: Response to social and  
589 physical factors. In A. D. Pickering (ed.), *Stress and fish*, pp. 295–321. Academic Press, New  
590 York.

591 Schreck C.B. 2000. Accumulation and long-term effects of stress in fish. In G. P. Moberg and  
592 J. A. Mench (eds.), *The biology of animal stress*, pp. 147–158. CABI Publishing, Wallingford,  
593 U.K

594 Schreck C.B. 2010. Stress and fish reproduction: The roles of allostasis and hormesis. *Gen.*  
595 *Comp. Endocrinol.* 165, 549-556.

596 Stoot, L. J., N.A. Cairns, F. Cull, J. J. Taylor , J. D. Jeffrey, F. Morin, J.W. Mandelman, T.D.  
597 Clark and S.J. Cooke .2014. Use of portable blood physiology point-of-care devices for basic  
598 and applied research on vertebrates: a review. *Conserv. Physiol.* 2:1.

599 Sures B., K. Knopf, W. Kloas. 2001. Induction of stress by the swimbladder nematode  
600 *Anguillicola crassus* in European eels, *Anguilla anguilla*, after repeated experimental infection.  
601 *Parasitology* 123: 179–184. Tesch F.W. 2003 *The eel*, 5th edn. Oxford, UK: Blackwell Science.

602 Toni C., A.G. Becker, L.N. Simões, C.G. Pinheiro, L.L. Silva, B.M. Heinzmann, and B.O.  
603 Caron. 2014. Fish anesthesia: effects of the essential oils of *Hesperozygis ringens* and *Lippia*  
604 *alba* on the biochemistry and physiology of silver catfish (*Rhamdia quelen*). *Fish Physiol.*  
605 *Biochem.* 40: 701–714.

606 Towle D.W. 1981. Role of Na<sup>+</sup>, K<sup>+</sup>-ATPase in ionic regulation by marine and estuarine animals.  
607 *Mar. Biol. Lett.* 2:107–122.

608 Trischitta F., Y. Takei, and P. Sébert. 2013. Eel Physiology Editora CRCPress, pp. 378.

609 van den Thillart G. and S. Dufour. 2009. How to estimate the reproductive success of European  
610 silver eels, in: van de Thillart, G. et al. (Ed.) Spawning migration of the European eel:  
611 reproduction index, a useful tool for conservation management. Fish & Fisheries Series, 30: pp.  
612 3-9.

613 Wendelaar Bonga S.E. 1997. The stress response in fish. *Physiol. Rev.* 77: 591–625

614 van Ginneken V., E. Antonissen, U.K. Müller, R. Booms, E. Eding, J. Verreth, and G. van den  
615 Thillart. 2005. Eel migration to the Sargasso: remarkably high swimming efficiency and low  
616 energy costs. *J. Exp. Biol.* 208: 1329–1335.

617 Wahlberg M., H. Westerberg, K. Aarestrup, E. Feunteun, P. Gargan, and D. Righton. 2014.  
618 Evidence of marine mammal predation of the European eel (*Anguilla anguilla* L.) on its marine  
619 migration. *Deep-Sea Res I* 86: 32–38.

620 Wilson J.M. 2013. Stress physiology. In: Trischitta, F, Takei, Y, Sebert, P (Eds). Eel  
621 Physiology. CRC Press, Boca Raton, pp. 320-359. ISBN: 9781466598270

622 Wilson J.M., J.C. Antunes, P.D. Bouça, and J. Coimbra. 2004. Osmoregulatory plasticity of the  
623 glass eel of *Anguilla anguilla*: freshwater entry and changes in branchial ion-transport protein  
624 expression. *Can. J. Fish. Aquat. Sci* 61: 432–442.

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629 **FIGURE LEGENDS**

630

631 Figure 1: Variation of glucose on non- parasitized and parasitized eels through time (baseline,  
632 1 and 4 hours) (mean  $\pm$  standard error).

633

634 Figure 2: Variation of plasma cortisol on non- parasitized and parasitized eels: (a) among time  
635 (baseline, 1 and 4 hours) (mean  $\pm$  standard error) and (b) between life-stages (yellow and silver).

636 The solid black line represents the median (50th percentile) and the bottom and top box edges  
637 are the 25th (Q1) and 75th (Q3) percentile, respectively. The bottom whisker is the  
638  $\max(\min(x), Q1-1.5*IQR)$  with  $IQR=Q3-Q1$ , whereas the top whisker is the  
639  $\min(\max(x), Q3+1.5*IQR)$ .

640

641 Figure 3: Gill  $Na^+/K^+$ -ATPase activity variation of non-parasitized and parasitized yellow and  
642 silver eels. The solid black line represents the median (50th percentile) and the bottom and top  
643 box edges are the 25th (Q1) and 75th (Q3) percentile, respectively. The bottom whisker is the  
644  $\max(\min(x), Q1-1.5*IQR)$  with  $IQR=Q3-Q1$ , whereas the top whisker is the  
645  $\min(\max(x), Q3+1.5*IQR)$ .

646

**TABLES**

Table 1. Glucose, Cortisol and Na<sup>+</sup>/K<sup>+</sup>-ATPase activity on female parasitized eels in different silvering conditions according parasite range (mean ± standard error).

Silvering stage	Parasite range	N	Variable		
			Cortisol	Glucose	Na <sup>+</sup> /K <sup>+</sup> -ATPase activity
I	(0,4]	5	13.04± 2.69	4.08 ± 1.07	4.76 ± 0.63
	(4,8]	2	26.60 ± 13.30	4.35 ± 1.65	3.53 ± 1.02
II	(0,4]	4	9.60 ± 3.81	4.15 ± 0.57	7.06 ± 0.50
	(4,8]	3	11.90 ± 1.83	3.96 ± 0.94	4.89 ± 1.80
	(8,12]	1	14.80 ± -	4.30 ± -	7.56 ± -
III	(0,4]	3	21.53 ± 8.38	4.60 ± 1.80	7.95 ± 1.42
	(4,8]	1	7.80 ± -	9.60 ± -	10.05 ± -

**TABLES**

Table 2. Statistical outputs from linear mixed effects models: random effects model (Glucose and Cortisol) and fixed-effects ( $\text{Na}^+/\text{K}^+$ -ATPase activity). *P* values of significant parameters are indicated.

Variables	Parameter	Value	Std.Error	t-value	p-value
Glucose					
a) General					
	Time (1h)	0.269	0.022	12.165	<0.0001
	Time (4h)	0.410	0.022	18.356	<0.0001
	Parasitized	0.060	0.048	1.239	0.220
	Time (1h) x Parasitized	0.018	0.034	-0.520	0.604
	Time (4h) x Parasitized	0.041	0.034	-1.186	0.238
b) Females in different silvering stages					
	Time (1h)	0.260	0.019	13.352	<0.0001
	Time (4h)	0.388	0.020	19.747	<0.0001
	Parasitized	0.044	0.088	0.511	0.612
	Silvering stage II	0.035	0.097	-0.361	0.720
	Silvering stage III	-0.075	0.080	-0.944	0.351
	Parasitized x Silvering stage II	-0.050	0.134	-0.379	0.707
	Parasitized x Silvering stage III	0.134	0.127	1.057	0.297
Cortisol					
a) General					
	Time (1h)	0.315	0.050	6.260	<0.0001
	Time (4h)	0.095	0.050	1.881	0.062

Parasitized	-0.139	0.097	-1.428	0.159
Life-stage (silver)	0.117	0.073	1.601	0.115
Time (1h) x Parasitized	0.255	0.078	3.281	0.001
Time (4h) x Parasitized	0.139	0.078	1.783	0.077
Parasitized x Life-stage (silver)	-0.246	0.115	-2.140	0.037

b) Females in different silvering stages

Silvering stage II	-0.069	0.087	-0.791	0.433
Silvering stage III	0.142	0.082	1.717	0.093
Parasitized	-0.285	0.091	-3.120	0.003
Time (1h)	0.299	0.063	4.731	<0.0001
Time (4h)	0.095	0.063	1.509	0.135
Time (1h) x Parasitized	0.325	0.093	3.498	0.0008
Time (4h) x Parasitized	0.188	0.093	2.014	0.047

Na<sup>+</sup>/K<sup>+</sup>-ATPase activity

a) General

Sex (males)	2.717	0.755	3.599	0.0006
Parasitized	1.051	0.972	1.082	0.284
Life-stage (silver)	6.046	0.818	7.395	<0.0001
Parasitized x Life-stage (silver)	-3.506	1.308	-2.681	0.0095

b) Females in different silvering stages

Silvering stage II	1.973	0.937	2.106	0.0412
Silvering stage III	5.163	0.874	5.908	<0.0001
Parasitized	0.211	0.758	0.278	0.782

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Figure 1

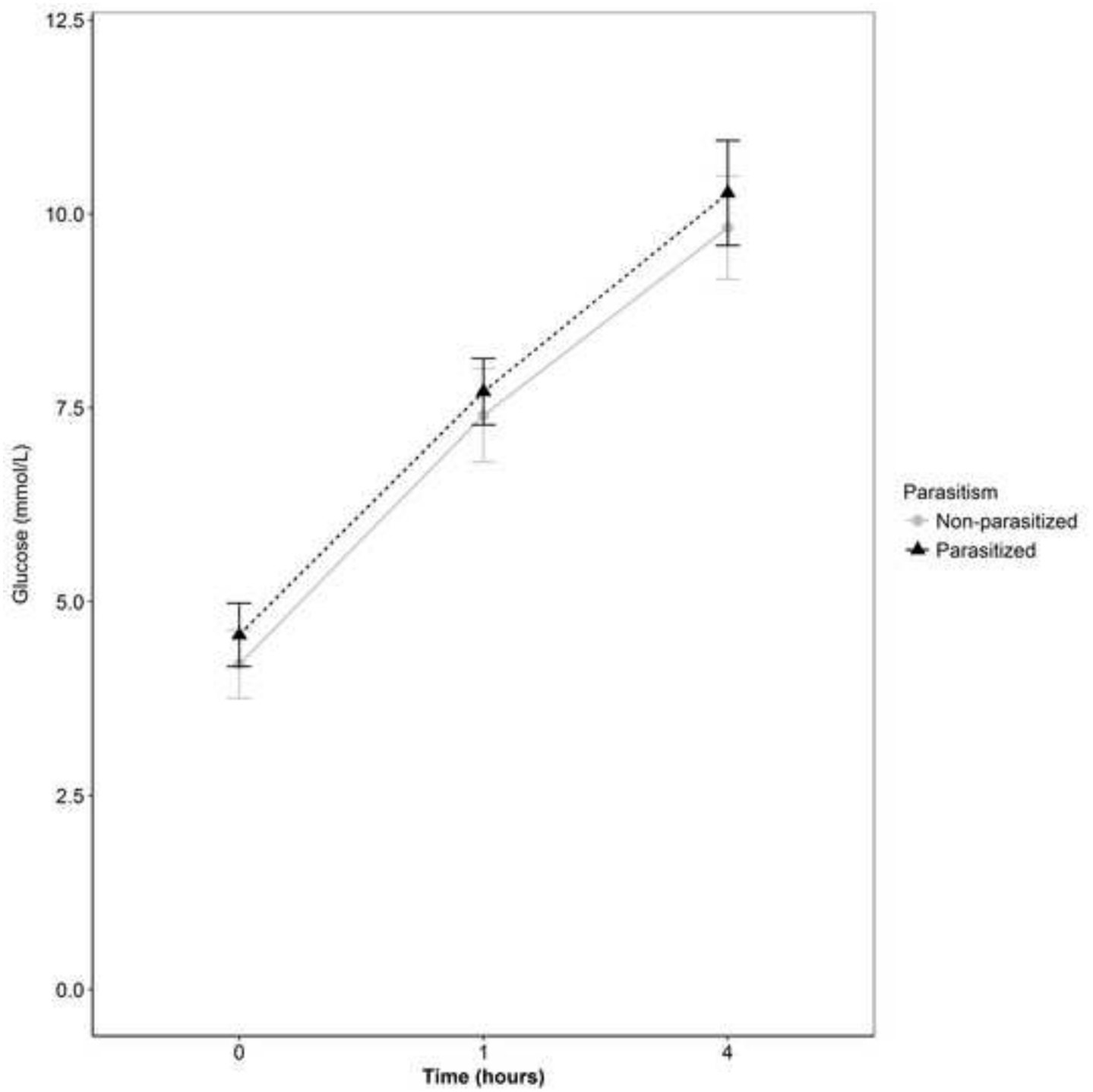


Figure 2a

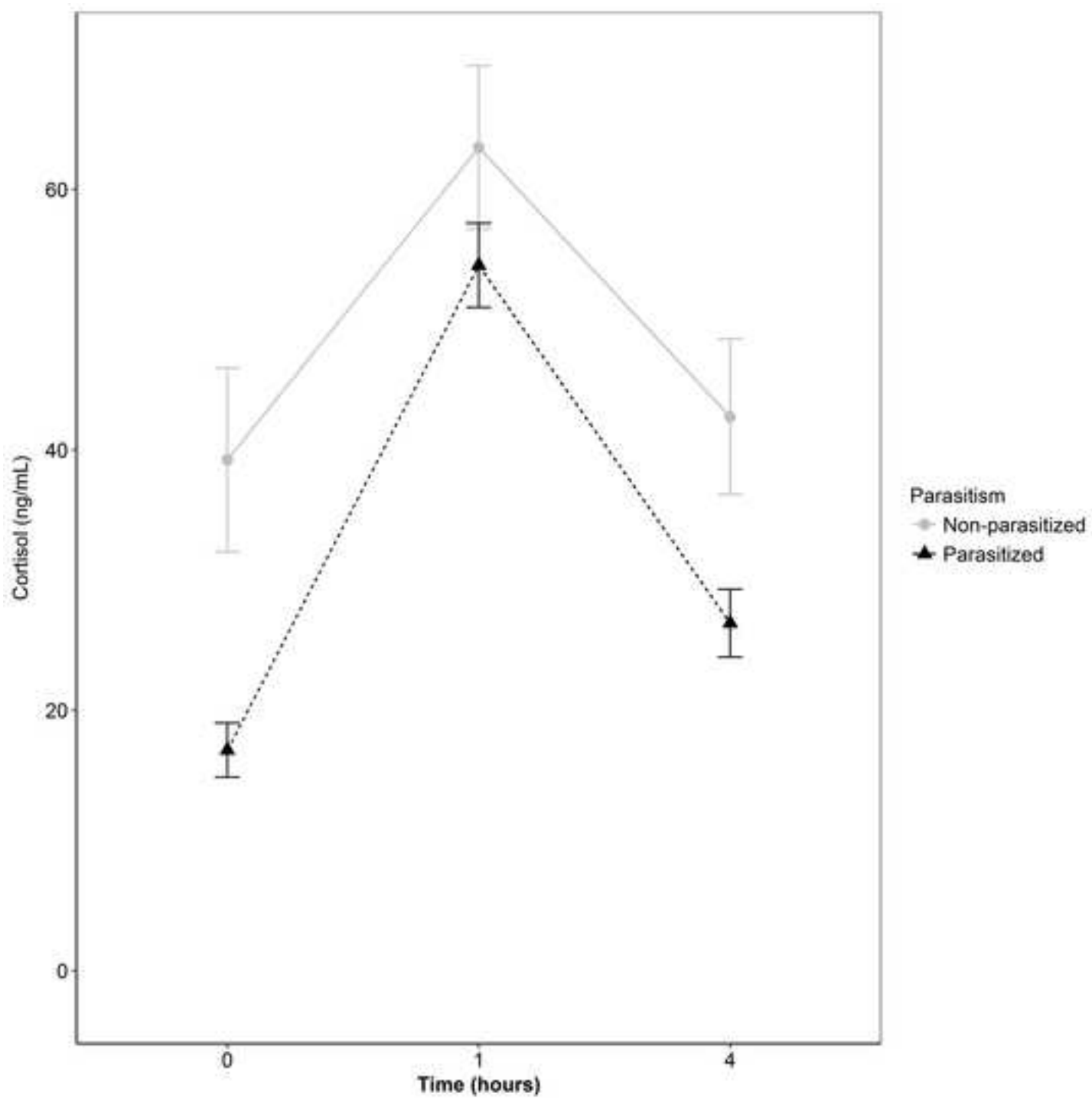


Figure 2b

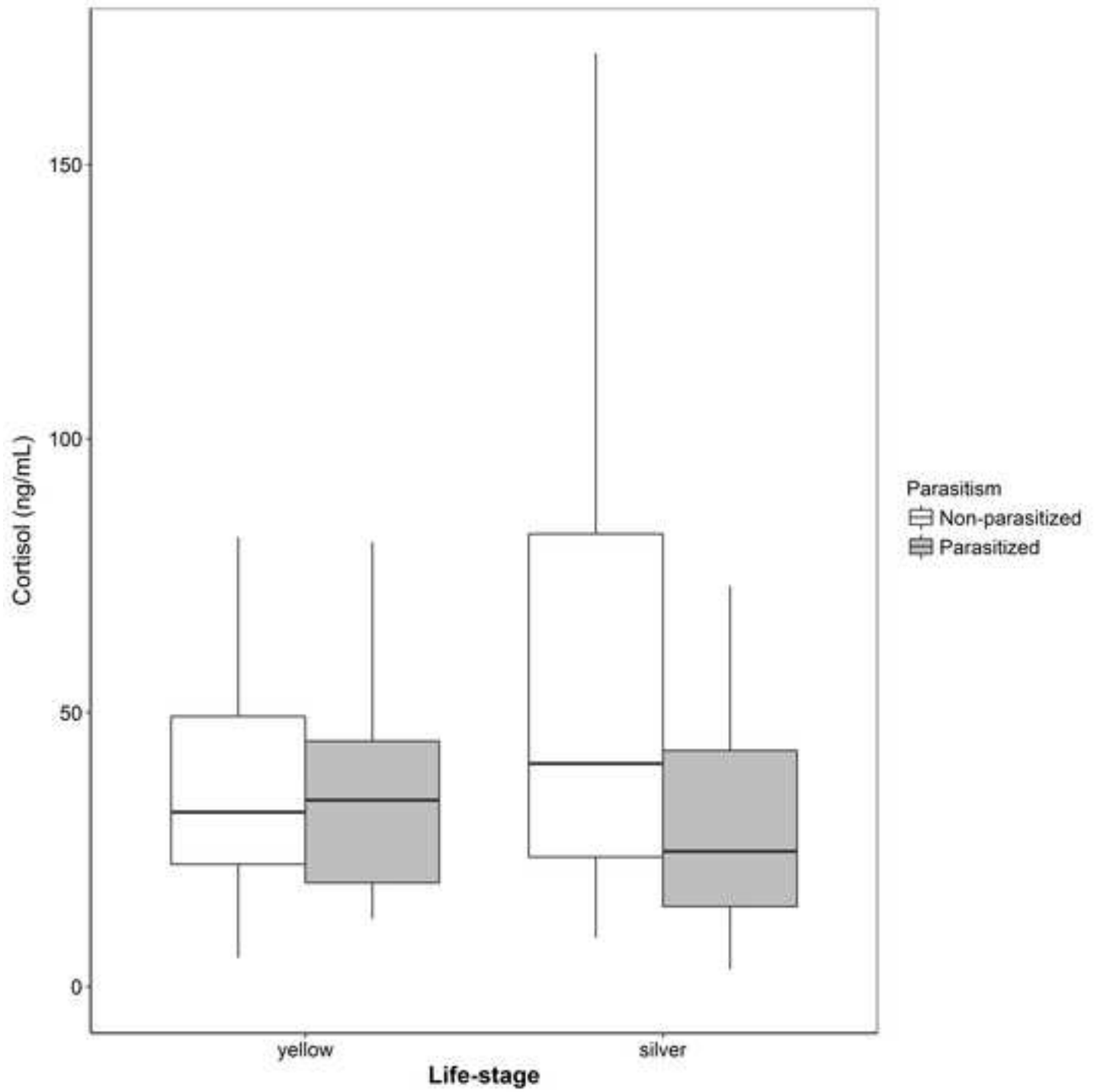


Figure 3

