- 1 Conservation genomics reveals possible illegal trade routes and
- 2 admixture across pangolin lineages in Southeast Asia
- Helen C. Nash a*, Wirdateti b, Gabriel W. Low a, Siew Woh Choo c, d, Ju Lian Chong e,
- 4 Gono Semiadi ^b, Ranjeev Hari ^{c,f}, Muhammad Hafiz Sulaiman ^e, Samuel T. Turvey ^g,
- 5 Theodore A. Evans a, h, Frank E. Rheindt a
- ^a Department of Biological Sciences, National University of Singapore, 14 Science
- 7 Drive 4, 117543, Singapore
- 8 b Research Center for Biology, Indonesian Institute of Sciences (LIPI), JI Raya
- 9 Jakarta-Bogor Km 45, Cibinong 16911, Indonesia
- ^c Genome Informatics Research Laboratory, High Impact Research (HIR) Building,
- 11 University of Malaya, 50603 Kuala Lumpur, Malaysia
- d Department of Biological Sciences, Science Building B, Xi'an Jiaotong-Liverpool
- University, 111 Ren'ai Road, Suzhou Dushu Lake Science and Education Innovation
- District, Suzhou Industrial Park, Suzhou, P. R. China, 215123
- e Department of Biological Sciences, Faculty of Science & Technology, Universiti
- Malaysia Terengganu, 21030 Kuala Terengganu, Terengganu, Malaysia
- ^f Centre for Bioinformatics, School of Data Sciences, Perdana University, 43400
- 18 Serdang, Selangor, Malaysia
- 19 g Institute of Zoology, Zoological Society of London, Regent's Park, London NW1
- 20 4RY, UK

23

- ^h School of Biological Sciences, The University of Western Australia (M092), 35
- 22 Stirling Highway, Crawley, WA 6009, Australia

*Corresponding author (helencatherinenash@yahoo.co.uk, +65 84313054)

Abstract

25

The use of genome-wide genetic markers is an emerging approach for informing 26 evidence-based management decisions for highly threatened species. Pangolins are 27 the most heavily trafficked mammals across illegal wildlife trade globally, but 28 Critically Endangered Sunda pangolins (Manis javanica) have not been widely 29 studied in insular Southeast Asia. We used > 12,000 single nucleotide polymorphic 30 markers (SNPs) to assign pangolin seizures from illegal trade of unknown origin to 31 possible geographic sources via genetic clustering with pangolins of known origin. 32 Our SNPs reveal three previously unrecognized genetic lineages of Sunda pangolins, 33 possibly from Borneo, Java and Singapore/Sumatra. The seizure assignments 34 35 suggest the majority of pangolins were traded from Borneo to Java. Using mitochondrial markers did not provide the same resolution of pangolin lineages, and 36 37 to explore if admixture might explain these differences, we applied sophisticated tests of introgression using > 2,000 SNPs to investigate secondary gene flow 38 39 between each of the three Sunda pangolin lineages. It is possible the admixture which we discovered is due to human-mediated movements of pangolins. Our 40 findings impact a range of conservation actions, including tracing patterns of trade, 41 repatriation of rescue animals, and conservation breeding. In order to conserve 42 genetic diversity, we suggest that, pending further research, each pangolin lineage 43 should as a precaution be protected and managed as an evolutionarily distinct 44 conservation unit. 45

46

47

48

Keywords: SNPs; mitochondrial markers; gene flow; illegal wildlife trade; population assignment; conservation breeding

49

50

1. Introduction

The importance of using evidence-based conservation to inform effective management decisions is increasingly recognized by conservation researchers and practitioners (Sutherland et al. 2004; Segan et al. 2010; Nash et al. 2016). The use of genetics to inform evidence-based management decisions for highly threatened species can improve conservation outcomes (Allendorf et al. 2010; Corlett 2016;

- Pierson et al. 2016). For example genetic tools have been used to define species
- 57 delimitations accurately (e.g. in passerine birds, Lohman et al. 2010; in crocodiles;
- 58 Shirley et al. 2014), trace illegal wildlife trade (e.g. in sharks, Clarke et al. 2006; in
- elephant ivory, Wasser et al. 2008), assess population viability (e.g. for the Komodo
- dragon Ciofi et al. 1999; and sturgeon Schueller and Hayes 2011), and to inform
- conservation breeding, which includes captive breeding and genetic rescue of wild
- populations (e.g. Florida panther, Johnson et al. 2010; Burmese roofed turtle, Cilingir
- et al. 2017).
- The conservation of pangolins would benefit from the application of genetic tools.
- There are eight pangolin species which are insectivorous, scaly mammals, native to
- Asia and Africa. Populations of all species are in decline due to habitat clearing and
- 67 high levels of poaching driven by demand for traditional medicines and meat
- (Challender et al. 2014a). Consequently, pangolins are considered to be the most
- 69 heavily trafficked mammals across illegal wildlife trade globally (Challender et al.
- 70 2015). The Sunda pangolin (*Manis javanica*) is distributed across several countries
- in Southeast Asia. 'Sunda' refers to Sundaland, which is a biogeographical region
- including the Malay Peninsula and Indonesian archipelago. Sunda pangolins are
- 13 listed as Critically Endangered in the IUCN Red List, especially populations in
- Indonesia (Challender et al. 2014b). The need for increased genetic research is
- highlighted as a priority activity for pangolins in the global conservation action plan of
- the IUCN Species Survival Commission's Pangolin Specialist Group (Challender et
- 77 al. 2014a).
- A broad variety of genetic methods and different genetic markers can help to inform
- 79 conservation actions for pangolins (Alacs et al. 2010; Hassanin et al. 2015; Tan et al.
- 2016). Among these methods, genome-wide markers such as single nucleotide
- polymorphisms (SNPs) are a powerful tool to provide detailed information about
- 82 population structure, sometimes with greater resolution than other markers including
- microsatellites (Malenfant et al. 2014). Moreover, next-generation sequencing
- methods such as double digest restriction site-associated DNA sequencing
- 85 (ddRADseq) have been used successfully to investigate population structure of
- mammals (Knowles et al. 2016).

There are existing genetic studies for the Sunda pangolin. A whole-genome of the Sunda pangolin was recently sequenced and published (Choo et al. 2016), providing the genomic infrastructure for further genetic research to inform robust conservation actions and management plans. The population genetic structure of Sunda pangolins across Indonesia has previously been investigated using the mitochondrial (mtDNA) control region (Wirdateti & Semiadi, 2017), with this study suggesting that there might be more than one pangolin lineage across Indonesia. However, mtDNA is only a single marker, prone to biases such as selective sweeps and introgression (Ballard & Whitlock, 2004). Wider-scale techniques such as ddRADseq could instead be used to look more rigorously at Sunda pangolin populations with genome-wide genetic markers. The illegal trade of wildlife generates billions of USD per year, and a worryingly high proportion of this trade includes pangolins (UNODC, 2016). Unsustainably large seizures of pangolins have occurred across insular Southeast Asia, for example, >5 tonnes in Medan, Indonesia (WCS News Releases, 2015), and tonnes of scales and meat destined for countries such as China (Cheng et al. 2017). A paucity of wild

seizures of pangolins have occurred across insular Southeast Asia, for example, >5 tonnes in Medan, Indonesia (WCS News Releases, 2015), and tonnes of scales and meat destined for countries such as China (Cheng et al. 2017). A paucity of wild samples of known geographic origin has hindered genetic assignments to investigate the sources of illegally traded pangolins (Zhang et al. 2015), and further widespread sampling of pangolins of known geographic origin with full chains of custody and expert identification is required. Meanwhile, the origin of illegally traded pangolins is an urgent issue which needs to be addressed.

In this study, we used ddRADseq to generate genome-wide genetic markers to assign pangolins of unknown origin from seizures of illegal trade to genetic clusters with wild samples of known origin in Sundaland. This is the first application of next-generation sequencing methods to the population assignment of pangolin seizures. We conducted analyses for > 12,000 SNPs and two mitochondrial genes. Our results are consequential to a range of conservation actions for pangolins, such as tracing illegal wildlife trade, repatriation of rescued pangolins, and conservation breeding.

2. Methods

2.1. Sample Collection

We aimed to collect wild Sunda pangolin samples of known origin from across 117 insular Southeast Asia (Figure 1), and with the Indonesian Institute of Sciences (LIPI) 118 we also acquired pangolin samples from seizures of illegal wildlife trade of unknown 119 origin from across Indonesia (Figure 1). In total, we obtained 97 Sunda pangolin 120 samples between July 2008 and January 2016. The total included 89 Indonesian 121 tissue samples from the muscle of dead pangolins, eight of which had a known origin 122 from wild populations in Java (Jember), Sumatra (Lampung) and Kalimantan 123 (Pangkalanbun, Indonesian Borneo), plus seven blood samples from wild 124 125 Singaporean pangolins, and one tissue sample from a dead pangolin in Sarawak (Malaysian Borneo). Pangolin specialists were present at each site for the collection 126 of wild samples and were able to confirm via morphological features, such as hind 127 scale counts (Gaubert 2011), that the wild samples were all Sunda pangolins. LIPI 128 also maintain official national records and documentation of their samples. 129 Veterinarians at Wildlife Reserves Singapore (WRS) collected the blood samples 130 from anaesthetized Singaporean pangolins. Similarly, veterinarians at Kadoorie 131 Farm and Botanic Garden in Hong Kong, China, collected a blood sample from one 132 Chinese pangolin (Manis pentadactyla) in July 2014, which was used as an outgroup. 133 2.2. DNA extraction and ddRADseq library preparation 134 We extracted DNA at LIPI, the National University of Singapore (NUS), the Universiti 135 Malaysia Terengganu, and Kadoorie Farm and Botanic Garden, using Qiagen 136 DNeasy Blood & Tissue Kits. Chinese and Sunda pangolin DNA were exported to 137 Singapore with appropriate Convention on International Trade in Endangered 138 Species of Wild Fauna and Flora (CITES) permits in 2014 and 2015/16 respectively. 139 To measure DNA yields we used fluorometric quantitation of double-stranded DNA 140 via Qubit 2.0™. For tissue samples with low yields, we re-extracted DNA using 141 phenol chloroform. 142 To obtain genome-wide markers (SNPs) with next-generation sequencing, we 143 modified a protocol for ddRADseq (Peterson et al. 2012) following Tay et al. (2016) 144 and used the restriction enzymes EcoRI-HF and MspI because they worked well for 145 other taxa (Garg et al. 2016; Ng et al. 2017). During optimisation of the RADseq 146 protocol, we selected a Sera-Mag® bead ratio that produced DNA fragments within a 147 range of 250-650 base pairs (bp) (Appendix A). We used additional control samples 148

- with molecular grade water instead of DNA throughout all procedures to confirm
- there had been no contamination. Singapore Centre on Environmental Life Sciences
- 151 Engineering checked the quality of each DNA library and sequenced each library in
- two lanes of one flowcell of an Illumina HiSeg 2500 Rapid Sequencing Run to
- produce 2 x 150 bp paired end reads. We spiked both lanes with 5% PhiX to
- increase the quantity of data obtained.
- 155 2.3. Bioinformatic Analysis
- 156 2.3.1. Identification of SNPs
- To call SNPs across our reads, we first checked the quality of the 150 bp paired end
- reads with FastQC version 0.11.5 (Andrews 2016). We used a Phred Score of 20 as
- our quality threshold, which meant we had to truncate reads to 135 bp for further
- analysis (Appendix B). We then demultiplexed the reads using process_radtags in
- STACKS version 1.35 (Catchen et al. 2013), which grouped the uniquely labelled
- reads of each sample (Appendix C). We indexed the nuclear genome sequence of a
- Sunda pangolin (Choo et al. 2016) using Burrows-Wheeler Alignment Tool version
- 0.7.1. (Li et al. 2013), and we aligned our reads to it using bwa_memscript (Li et al.
- 2013) (Appendix D). We used Samtools version 1.4 (Li et al. 2009) to convert SAM
- to BAM files, and to sort the BAM files (Appendix E). Our ref_map.pl pipeline in
- STACKS included pstacks, cstacks and sstacks, with a minimum stacks depth of 5,
- to call SNPs in each sample, and match loci across populations according to
- alignment positions (Catchen et al. 2013) (Appendix F). In addition, we ran
- population analysis in ref_map.pl, with a minimum of 90% of individuals in a
- population required to process a locus for that population. For the population labels
- we assigned Singaporean versus non-Singaporean samples. Later, we checked if a
- default model of no population substructure changed the overall results, which it did
- 174 not (Appendix G).
- For further quality control of our SNP calling, we removed SNPs with 10% or more
- missing data, and individuals with more than 15% of loci missing, using PLINK
- version 1.9 (Purcell et al. 2007). We also tested higher missing data cut-offs to see if
- we could retain additional samples for analysis but it was not feasible. We pruned
- SNPs that were correlated (Appendix H), which applied a sliding window of 25 bp
- and removed correlated SNPs of R-squared ≥ 0.9 using a normal distribution curve

- across each window. We also used PGDSpider version 2.1.0 (Lischer & Excoffier
- 2012) and BayeScan version 2.1 (Foll & Gaggiotti 2008) to further confirm that no
- loci were under selection (Appendix I).
- 184 2.3.2. Population genomic analysis
- 2.3.2.1. Principal Component Analysis and Fst estimation
- As a preliminary analysis to explore how many genetic clusters there were across
- our samples, we applied Principal Component Analysis (PCA) to the remaining 83
- Sunda pangolin samples to compare principal components of 12,150 SNPs using the
- SNPRelate package in R version 3.2 (R Core Team, 2016) (Figure 2a, and Appendix
- J). We also used SNPRelate to investigate pairwise Fst between clusters to get a
- sense of the extent of genetic differentiation (Appendix J), which applied the method
- of Weir & Cockerham (1984) to estimate Fst.
- 2.3.2.2. Bayesian clustering approaches and Network Analyses
- 194 Informed by the PCA results, we next applied a Bayesian clustering approach, using
- 195 STRUCTURE and CLUMPP, which required an a priori understanding of the
- potential number of clusters. For STRUCTURE (Pritchard et al. 2000), we tested K =
- 197 1 to K = 7 to investigate whether there might be 1 to 7 genetic clusters across our 83
- samples (Appendix K). We then used CLUMPP version 1.1.2 to determine the
- optimal alignment of clusters (Jakobsson & Rosenberg 2007) (Appendix L).
- We were aware that STRUCTURE sometimes generates erroneous results due to
- uneven sample sizes between subpopulations (Puechmaille 2016), so we double-
- 202 checked our genetic cluster results by using Network Analyses in NetView version
- 1.0, available in RStudio version 0.99.903 (RStudio Team 2015). The Network
- Analyses were based on a genetic distance matrix which we made from PLINK
- version 1.9 for our 83 samples using the 12,150 SNPs. We investigated three
- 206 network algorithms: Fast-greedy, Infomap and Walktrap (Appendix M). Network
- 207 Analyses provide a range of results which are all valid clustering arrangements
- 208 (Appendix M). We visualized the clustering arrangement with the highest genetic
- distances as a mutual k-nearest neighbour graph in RStudio (Figure 2b).

During the course of our research, new methods to understand population genetic 210 structure became available, so in addition we applied fineRADstructure package v0.2 211 (Malinksy et al. 2016) to quantify the ancestry sources in each population (Figure 2d). 212 FineRADstructure utilizes a fineSTRUCTURE MCMC clustering algorithm (Lawson 213 et al. 2012) to infer a co-ancestry matrix, which is a summary of nearest neighbour 214 haplotype relationships across the dataset (Malinksy et al. 2016). We used our 215 haplotypes.tsv file generated from the above populations analysis in ref_map.pl, and 216 we converted it to a FineRADstructure input using Python scripting contributed by 217 218 Emiliano Trucchi (Appendix S, Malinksy et al. 2016). The conversion script was run in PyCharm 2017.3 (Professional Edition), using 1 as the maximum number of SNPs 219 per locus, and 50% as the maximum percentage of missing loci to be included in the 220 PCA. The amount of missing data to allow was based on the missing data plot from 221 fineRADstructure (Appendix S) in order to exclude only samples with high missing 222 data. The clustered fineRADstructure co-ancestry matrix for 80 samples was 223 visualized in RStudio (Figure 2c, and Appendix S). 224

2.3.3. Phylogenetic Analyses 225

227

228

230

231

232

234

235

236

237

239

- 2.3.3.1. Phylogenomics using SNPs 226
- We felt the consistent emerging trend across our genetic clustering results warranted some further investigation to try and better understand the evolutionary trajectory of 229 these lineages. In order to generate nucleotide sequences containing SNPs for the construction of maximum likelihood phylogenies in RAxML version 8.2.9 (Stamatakis, 2014), we aligned demultiplexed 135bp sequence reads in pyRAD version 3.0.64 (Eaton, 2014). This included the Chinese pangolin as an outgroup. SNPs were called 233 using the ddrad option with the following parameters: clustering threshold of 95%, minimum cluster coverage of five, maximum of six low-quality sites per locus, minimum of 79 samples present in a final locus, and maximum of three individuals with a shared heterozygous site per locus (Appendix N). A total of 2,365 SNPs was generated and inputted in RAxML version 8.2.9 with the following parameters: GTRGAMMA option provided, 1000 rapid bootstraps inferences, and a final 238 maximum likelihood search (Figure 3a, and Appendix O). We tested other runs with increasing and decreasing amounts of missing data, but the bootstrap support

values did not improve, so we only present the concatenation method with 2,365

242 SNPs (Figure 3a).

243

247

248

249

250

253

257

261

262

265

268

2.3.3.2. Phylogenetics using mitochondrial DNA

244 It was useful to compare the genetic clustering results of our genome-wide markers 245 with mitochondrial DNA (mtDNA) markers which are commonly used to assign

species in illegal wildlife trade, so we sequenced two mtDNA coding genes,

cytochrome b (Cytb) and cytochrome oxidase c subunit 1 (CO1). We didn't have

sufficient DNA remaining from every pangolin sample, but we were able to include

the majority, 59 pangolins. We conducted Sanger sequencing at NUS and LIPI using

primer sequences provided by LIPI (Appendix Q). Sequences were aligned and then

251 Cytb and CO1 sequences were manually concatenated in MEGA 7.0 (Kumar et al.

252 2016). The complete concatentation sequence consisted of 1575 base pairs (*Cytb* =

787 bp, CO1 = 788 bp), and the sequences are available on GenBank (respective

accession numbers MG825495-MG825551 and MG825552-MG825610). We added

255 the mitogenome of a Chinese pangolin (GenBank KT445978.1) and one additional

Sunda pangolin (NC_026781.1), then used MEGA 7.0 (Kumar et al. 2016) to

construct a phylogenetic tree with maximum likelihood (ML), with a General Time

Reversible Model, partial deletion and 1000 bootstraps (Figure 4a). In order to

facilitate comparison of results, the first letter of the mtDNA tree labels reflect the

genetic clustering result from the SNPs, e.g. J = Java, and the seizure location is

also given at the end of each label. We used asterisks beside the labels to indicate

when the seizure location and genetic cluster result from the SNPs were the same

263 geographic area (Figure 4a). We also generated a haplotype network from the

264 concatenated mtDNA sequences using the Median-Joining method in PopART

(Leigh & Bryant 2015) (Figure 4b). The colour of each sample label represents the

SNP cluster results to further aid comparison between the mtDNA and SNP results

267 (Figure 4b).

2.3.4. Tests for Introgression

The contrasts between our genetic clustering results from the SNPs (Figure 2)

versus the topology of the mtDNA tree/haplotype network (Figure 4) raised our

suspicion that potential introgression might explain these differences. It was not

272 possible to test every sample for introgression because the computational run time is

prohibitive. Instead we selected a few anomalous results from the mtDNA tree, and we applied ABBA BABA tests to investigate secondary gene flow in those samples from our wild samples of known origin (Zinenko et al. 2016). The ABBA BABA tests required nucleotide sequences containing SNPs so we used our PHYLIP files from pyRAD including 2,365 SNPs (Appendix R). Our first ABBA BABA test used a wild sample from Borneo, MZBR 1163, as group A; an anomalous sample, MZBR 1040, which falls within the Bornean cluster of SNPs, but has suspected introgression from Java based on the mtDNA result, as group B; and a wild sample from Java, MZBR 1184, as group C; the outgroup was our Chinese pangolin. This tested MZBR 1040 for suspected introgression from the Javan lineage. Our second and third tests used a wild sample from Java, MZBR 1184, as group A; and an anomalous sample, MZBR 0270, which falls within the Javan cluster of SNPs, but has suspected introgression from both Singapore/Sumatra and Borneo based on the mtDNA result, as group B; group C was initially a wild sample from Singapore/Sumatra, rescue 1, which we then switched to MZBR 1163 a wild sample from Borneo in another run; the outgroup was always our Chinese pangolin. This initially tested MZBR 0270 for suspected introgression from the Singapore/Sumatra lineage, and next tested the same sample for suspected introgression from the Bornean lineage. The ABBA BABA results were summarised in a simple cartoon figure (Figure 3b). These ABBA BABA tests provided examples of introgression.

293 2.3.5. Tracing illegal trade

273

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

- The genetic clustering results from each genome-wide population genomic analysis,
- including PCA, STRUCTURE, NetView and fineRADstructure, all gave compatible
- results. Hence, any of those methods could be used to assign the seized pangolin
- samples of unknown origin to the wild samples of known origin (Figure 2 and
- 298 Appendix L). The assignment results were the same for every method, except that
- fineRADstructure only used 80/83 samples due to differences in sample filtering
- (Figure 2c). The additional three results came from only PCA (Figure 2a),
- 301 STRUCTURE (Appendix L) and NetView (Figure 2b). The phylogenetic trees
- (Figures 3 and 4) were not used for the illegal trade assignments due to the
- presence of introgression and the poor resolution of the trees.

There is perhaps some further grey support for our genetic assignments according to 304 the locations of the illegal seizures. Therefore, we highlighted in bold cases where 305 the sample seizure location is similar to its genetic cluster result, and the wild 306 samples of known origin labelled as WILD are also highlighted in bold (Figure 5). 307 All necessary research permits and ethics approvals were granted prior to 308 commencement of this project. In particular, NUS Institutional Animal Care and Use 309 Committees (IACUC) approved research methods, and we obtained CITES permits 310 311 for all pangolin samples. 3. Results 312 3.1. ddRAD sequencing and SNP discovery 313 314 Our collection of samples included 97 Sunda pangolins and 1 Chinese pangolin. Following library preparation, 2 Sunda pangolin tissue samples did not yield 315 316 sufficient DNA fragments within the 250-650 bp range, so only 96 samples underwent Illumina sequencing (Appendix P). In total we obtained 49.67 GB of data 317 318 from 96 samples across two lanes (reads one = 24.12 GB, reads two = 25.55 GB), and there were no lane differences in the results. 319 After demultiplexing our reads, 7 Sunda pangolin samples were discarded due to low 320 sequencing coverage (Appendix P). Across the remaining 89 samples 60,197 SNPs 321 were called via the STACKS pipeline. In PLINK 43,439 loci were removed due to 322 having ≥10% missing data. Only 83 samples met our <15% missing data 323 requirement; the Chinese pangolin sample was excluded at this stage. A further 324 4,608 loci were removed due to correlation. Consequently, for further analysis we 325 had 83 Sunda pangolin samples with 12,150 SNPs. BayeScan did not detect 326 selection. 327 3.2. Population Genomic Structure 328 3.2.1. Principal Component Analysis and Fst estimation 329 330 The default settings of SNPRelate removed a further 223 SNPs prior to PCA of 11,927 SNPs. The PCA results suggested that we were dealing with three distinct 331

genetic clusters of Sunda pangolins, possibly from Sumatra/Singapore, Java, and

Borneo according to the genetic cluster labels of wild pangolins of known origin

332

334	(Figure 2a). The first two eigenvectors of PCA held the largest percentage of
335	variance among the population, principal component 1 = 4.47 % and principal
336	component 2 = 4.38 %. The mean Fst between clusters is: Java versus
337	Sumatra/Singapore = 0.0684, Borneo versus Java = 0.0556, Borneo versus
338	Sumatra/Singapore = 0.0446, and across all three clusters = 0.0545. The mean Fst
339	values increased when we removed samples showing signals of introgression (see
340	section 3.3.2 for details of introgression), for example, the mean Fst across all three
341	clusters increased when MZBR 0270 was removed, mean Fst = 0.0556 (Appendix J)
342	3.2.2. Bayesian clustering and Network Analyses
343	The results from STRUCTURE, Structure Harvester and CLUMPP supported the
344	PCA results, similarly indicating that three genetic clusters across the 83 Sunda
345	pangolins is the most likely arrangement (Appendix L). Moreover, all of the clustering
346	approaches across PCA (Figure 2a), STRUCTURE (Appendix L), NetView (Figure
347	2b) and fineRADstructure (Figure 2c) produced compatible clustering results.
348	3.2.2.1. Additional insight to population substructure
349	The co-ancestry matrix from fineRADstructure used 46, 274 loci to illustrate the
350	levels of co-ancestry across 80 Sunda pangolin samples (Figure 2c and Appendix S)
351	with yellow representing the lowest levels of co-ancestry. The matrix confirmed that
352	there are three main clusters, Borneo, Java and Sumatra/Singapore, and revealed
353	additional differences in co-ancestry among clusters/samples. The black colour
354	represents the highest level of co-ancestry and suggests that MZBR 1030 and 1031
355	might be highly related pangolins from Sumatra. The purple coloured MZBR 1166
356	and 1167 could be closely related pangolins from Borneo. The deep red colour
357	represents samples MZBR 1184, 1185, 1190, 1063 and 1060 and suggests these
358	might be close relatives perhaps from a similar area in Java. We were aware of the
359	relationships between MZBR 1030 and 1031, also 1166 and 1167, from the NetView
360	analyses (Appendix M). However, we were not aware about the groupings of
361	pangolins labelled with deep red colours which show very high levels of co-ancestry.
362	FineRADstructure provided more compelling detail about population substructure
363	than our other population genomic analyses.

3.3. Phylogenetic trees

3.3.1. Phylogenomic tree using SNPs

365

Phylogenomic analysis with the concatenation method RAxML used 2,365 SNPs, 366 and although the Javan samples in the RAxML tree did not form a monophyletic 367 group (Figure 3a), the results were otherwise compatible with our population 368 clustering results from PCA, STRUCTURE, NetView and fineRADstructure (Figures 369 2a-c and Appendix L). Overlaid coloured shading illustrates the population genomic 370 clustering results on top of the RAxML tree to aid comparison (Figure 3a). The 371 RAxML trees generally had low bootstrap support. A few samples also had high 372 373 missing data, including MZBR 1189, 1183 and 1179 (Appendix O). We tested a wide variety of parameters in both pyRAD and RAxML to try to improve the bootstrap 374 375 support, for example, minimum coverage for a cluster (5 to 10), maximum number of sites with quality score less than 20 (4 to 6), clustering threshold (0.85 to 0.95), and 376 377 minimum samples in a final locus (70 to 81), however, the results did not improve. 3.3.2. Mitochondrial DNA tree and the issue of introgression 378 The results of the mtDNA tree (Figure 4a) and the mtDNA haplotype network (Figure 379 4b) were congruent. The mtDNA tree was not well resolved (Figure 4a) and the 380 haplotype network clearly shows the differences to the SNP results (Figure 4b). Our 381 ABBA BABA tests of secondary gene flow using 2,365 SNPs (Appendix R) 382 demonstrated that there are signals of introgression across our samples, for example, 383 MZBR 0270 showed a signal of introgression from both Singapore/Sumatra and 384 Borneo. The test result for MZBR 1040 was not statistically significant. The 385 statistically significant ABBA BABA test results are summarised in a simple cartoon 386 diagram (Figure 3b). 387 3.4. Tracing illegal trade 388 Based on the congruent genetic clustering results from PCA (Figure 2a), 389 STRUCTURE (Appendix L), NetView (Figure 2b) and fineRADstructure (Figure 2c), 390 which all showed three key distinct clusters which were geographically labelled by 391 the wild samples of known origin, we conclude that 20 of our samples possibly 392 originated from Java, 21 are possibly from Sumatra/Singapore, and 42 are possibly 393 from Borneo (Figure 5). These three key clusters are represented by black squares 394 (Figure 5) beneath which a full list of each sample's seizure location or wild location 395

has been provided. The samples highlighted in bold do not appear to have been translocated overseas, their seizure location is similar to their genetic origin. The pie charts (Figure 5) show the trade route at point of seizure of the samples, they do not include the wild samples. The arrows help to clarify the pie charts by pointing in the geographic direction of the trade.

Among the 49 pangolins seized from illegal wildlife trade in Java (Figure 1), only 18 may have been sourced within Java, while 23 may instead have been imported from Borneo, and 8 from Sumatra. The 6 pangolins seized in Medan, North Sumatra (Figure 1), may have originated in Sumatra as these samples group with the Sumatra/Singapore cluster. All of the 16 pangolins seized in Borneo (Figure 1) likely originated there. Only one pangolin across our dataset seems to have been traded from Borneo into Sumatra.

4. Discussion

396

397

398

399

400

401

402

403

404

405

406

407

408

409

410

411

412

413

414

415

416

417

418

419

420

421

422

423

424

425

426

427

Our study is the first application of next-generation sequencing methods to the critically endangered Sunda pangolin. The SNP data from each analysis that we have employed suggests there are three key genetic clusters across our samples, which are likely from Borneo, Java and Singapore/Sumatra. The majority of trade across our samples seems to be from Borneo into Java. The presence of introgression across our samples likely explains the poor bootstrap support of the phylogenetic trees. Only the population genomic analyses, PCA, STRUCTURE, NetView and fineRADstructure, provided sufficient clarity to genetically match the pangolins of unknown origin from illegal trade to the wild samples of known origin. The most compelling detail about population substructure was generated by the fineRADstructure method, and we recommend the use of this programme for further research. FineRADstructure not only confirmed our three main clusters, Borneo, Java and Singapore/Sumatra, it revealed further insights into the substructure of these clusters, such as highly related pangolins and other shared co-ancestry. There might be a geographic basis for some of the further substructures which fineRADstructure identified, such as the deep red clusters showing very high levels of shared co-ancestry (Figure 2c), but we require further wild samples of known origin to geographically label these clusters to more concise geographic areas. Currently, the deep orange coloured group within the Singapore/Sumatra cluster is

the only subgroup which we can geographically label. The wild Singaporean samples 428 label that cluster as Singaporean. 429 Genetic data are essential to help trace illegal wildlife trade and create forensic 430 genetic databases (Ogden et al. 2009; Wasser et al. 2015; Ogden & Linacre 2015). 431 Although we understand that without the use of trained conservation dogs it is very 432 difficult to capture rare and nocturnal pangolins, which cannot be easily baited, we 433 urge further sampling of wild pangolins of known origin to help inform geographic 434 population assignments. In Singapore our wild sampling was more extensive than 435 436 elsewhere in Southeast Asia due to a 24hr /7 day per week rescue service for wildlife, which is managed by ACRES, as well as rehabilitation of pangolins at Singapore Zoo. 437 438 We hope that other partnerships such as this between NGOs and researchers could facilitate improved pangolin sampling elsewhere. It is important that during the 439 440 collection of reference samples full chains of custody are documented with expert verification. 441 The possible directions of illegal trade which we have revealed in this study (Figure 5) 442 are similar to patterns of illegal trade of other wild species, for example, it has been 443 documented that birds are also poached in large numbers across Indonesia and 444 transported to markets in Jakarta and Java for sale or onwards distribution (Chng et 445 al. 2015). There is some further grey support for our population assignments from 446 our seizure locations, for example, from the local police seizures in Borneo all of 447 those seized pangolins clustered within our probable Bornean cluster (Figure 5). 448 449 The distinct genetic clustering of populations from Borneo, Java and Singapore/Sumatra matches divergence patterns seen in other vertebrate taxa 450 (Wilson & Reeder, 2005; Leonard et al. 2015). It is possible that our Sunda pangolin 451 452 lineages might not only be genomic but also ecologically differentiated, since the habitats on Java are primarily monsoonal dryland and savannah, whereas the 453 habitats on the other sampled areas are primarily rainforest (Sundevall, 1843; 454 Whitten et al. 1996). Studies of ecology and morphology across the three Sunda 455 pangolin lineages revealed in this study should be conducted to help inform their 456 taxonomic classification (Gaubert & Antunes 2005), and phenotypic inquiry may 457 uncover traits that support the genomic divergences shown by our data. 458

The genetic divisions were not well detected by our mtDNA data, partly due to 459 admixture and introgression. It is possible that genetic introgression explains the 460 differences between our tentative RAxML (Figure 3a) and mtDNA trees (Figure 4a). 461 Also mtDNA is a single marker that only reflects the maternal history (Ballard & 462 Whitlock, 2004). We suggest that for the population assignment of illegally traded 463 pangolins the use of genome-wide SNP markers can provide higher resolution than 464 mtDNA markers alone. We suggest that future population clustering results using 465 SNP data could be obtained quickly with just fineRADstructure (Malinsky et al. 2016) 466 467 and more georeferenced samples of known origin are needed. Sunda pangolins are also distributed across Thailand, Myanmar, Cambodia, Laos 468 and Vietnam, and it is possible that greater substructure exists across their full 469 geographic range (Zhang et al. 2015). Protecting genetic diversity is important for the 470 471 resilience and survival of species (Pierson et al. 2016), and this genetic diversity must therefore be considered in conservation management plans, not least to inform 472 473 conservation breeding (Hua et al. 2015), including captive breeding and genetic rescue, and repatriation of rescued pangolins. Selection of pangolin release 474 475 locations should be mindful of the appropriate genetic source population (Challender et al. 2014b). Currently, there is no consideration of genetic substructure when 476 pangolin seizures of unknown origin are released into the wild, and the impacts of 477 introducing and mixing pangolins from different lineages into the same area are not 478 fully understood. The genetic introgression uncovered in our study might possibly be 479 due to human influences, such as translocation of pangolins outside of their natural 480 range (Pantel & Chin, 2009). Conservation breeding records also need to document 481 and consider the highest resolution of population genetic structure to ensure that 482 genetic diversity is well-managed (Allendorf et al. 2010). We wish to emphasize the 483 importance of genetic screening of all individuals involved in releases and 484 conservation breeding. 485 As stated in the IUCN Pangolin Specialist Group action plan (Challender et al. 2014), 486 487 all pangolin species require further genetic investigation. In our study we chose the double-digest form of RADseq instead of other forms, because this method might be 488 more reproducible for other pangolin researchers to follow, as well as less expensive 489

for researchers with smaller budgets (Andrews et al. 2016).

The findings of our research impact a range of conservation actions, including tracing illegal wildlife trade, delimitation of pangolin conservation units, repatriation of rescued pangolins to appropriate locations, and conservation breeding. To maximise the resolution of genetic tracing of pangolins, we recommend that genome-wide markers are used in combination with mtDNA genes. Our findings provide a new baseline to help begin to understand Sunda pangolin populations, and we hope these findings will inspire future research and management actions that can support effective conservation of pangolins in Asia.

Acknowledgments

All authors contributed equally to this work. All authors discussed the results and
implications and commented on the manuscript at all stages. We thank the
Indonesian Institute of Sciences (LIPI), Lee Kong Chian Natural History Museum
(LKCNHM), Agri-Food & Veterinary Authority of Singapore (AVA), Wildlife Reserves
Singapore (WRS), Department of Wildlife and National Parks Peninsular Malaysia
(DWNP), Universiti Malaysia Terengganu (UMT) and Kadoorie Farm and Botanic
Garden (KFBG) for assistance with sample collection and the arrangement of
relevant permits and permissions. Special thanks are given to Yulianto (LIPI) and H.
Zhang (KFBG) who helped to extract DNA, and S. Oh (WRS), A. Ali (WRS), S. Luz
(WRS), P. Lee (WRS), C. F. Maosheng (LKCNHM), M. Chua (LKCNHM), R. Meier
(LKCNHM), C.Y. Gwee (NUS), G. Ades (KFBG) and A. Grioni (KFBG). We also
thank R. Asher (University of Cambridge) for his encouragement and advice, and the
IUCN-SSC Pangolin Specialist Group and Singapore Pangolin Working Group who
provided logistical support throughout this research. The Rheindt Lab at NUS shared
a ddRADseq protocol. L. Wijedasa gave us base maps for the figures. S. Thompson
helped with coding for fineRADstructure. Funding was provided by the Dennis Gould
Foundation; H.C.N is supported by a SINGA PhD Research Scholarship at NUS;
S.T.T is supported by a Royal Society University Research Fellowship (UF130573).
We also acknowledge internal departmental funding at the Department of Biological
Sciences at NUS; and the Research Center for Biology-LIPI Competitive Project
3400.001.002.021 SEAMEO BIOTROP DIPA 060.12/PSRP/SPK-PNLT/2014.

521	Literature Cited
522 523	Alacs EA, et al. 2010. DNA detective: A review of molecular approaches to wildlife forensics. Forensic Science, Medicine, and Pathology. 6 (3): 180–194.
524 525	Allendorf FW, Hohenlohe PA, Luikart G. 2010. Genomics and the future of conservation genetics. Nature Reviews Genetics. 11 (10): 697–709.
526 527	Andrews KR, et al. 2016. Harnessing the power of RADseq for ecological and evolutionary genomics. Nature Reviews Genetics. 17 : 81–92.
528 529 530	Andrews S. 2016. FastQC: a quality control tool for high throughput sequence data. Version 0.11.5 Available at: http://www.bioinformatics.babraham.ac.uk/projects/fastqc
531 532	Ballard JWO, Whitlock MC. 2004. The incomplete natural history of mitochondria. Molecular Ecology. 13 (4): 729–744.
533 534	Catchen J, et al. 2013. Stacks: an analysis tool set for population genomics. Molecular Ecology. 22 (11): 3124–40.
535536537	Challender DWS, Waterman C, Baillie JEM. 2014a. Scaling up pangolin conservation. IUCN SSC Pangolin Specialist Group Conservation Action Plan. Zoological Society of London, London, UK.
538539540	Challender DWS, et al. 2014b. Manis javanica. The IUCN Red List of Threatened Species 2014: e.T12763A45222303. Available at: http://dx.doi.org/10.2305/IUCN.UK.2014-2.RLTS.T12763A45222303.en .
541542543	Challender DWS, Harrop SR, MacMillan DC. 2015. Understanding markets to conserve trade-threatened species in CITES. Biological Conservation. 187 : 249–259.
544 545	Cheng W, Xing S, Bonebrake TC. 2017. Recent Pangolin Seizures in China Reveal Priority Areas for Intervention. Conservation Letters. doi:10.1111/conl.12339
546547548	Chng SCL, Eaton JA, Krishnasamy K, Shepherd CR, Nijman V. 2015. In the Market for Extinction: An inventory of Jakarta's bird markets. TRAFFIC. Petaling Jaya, Selangor, Malaysia.

550	immunity. Genome Research. 26 (10): 1312-1322.
551 552 553	Çilingir FG, Rheindt FE, Garg KM, Platt K, Platt SG, Bickford DP. 2017. Conservation genomics of the endangered Burmese roofed turtle. Conservation Biology. 28: in press.
554 555 556	Ciofi C, et al. 1999. Genetic divergence and units for conservation in the Komodo dragon Varanus komodoensis. Proceedings of the Royal Society B. 266 : 2269-2274.
557 558 559	Clarke S, et al. 2006. Identification of shark species composition and proportion in the Hong Kong shark fin market based on molecular genetics and trade records. Conservation Biology. 20 : 201-211.
560 561	Corlett RT. 2016. A Bigger Toolbox: Biotechnology in Biodiversity Conservation. Trends in Biotechnology. 35 (1): 55-65.
562 563	Drummond AJ, et al. 2012. Bayesian phylogenetics with BEAUti and the BEAST 1.7. Molecular Biology and Evolution. 29 (8): 1969–1973.
564 565 566	Earl DA, vonHoldt BM. 2012. STRUCTURE HARVESTER: A website and program for visualizing STRUCTURE output and implementing the Evanno method. Conservation Genetics Resources. 4 (2): 359–361.
567 568	Eaton DAR. 2014. PyRAD: assembly of de novo RADseq loci for phylogenetic analyses. Bioinformatics. 30 (13): 1844-1849.
569 570 571	Evanno G, Regnaut S, Goudet J. 2005. Detecting the number of clusters of individuals using the software STRUCTURE: A simulation study. Molecular Ecology. 14 (8): 2611–2620.
572 573 574	Foll M, Gaggiotti O. 2008. A genome-scan method to identify selected loci appropriate for both dominant and codominant markers: A Bayesian perspective Genetics. 180 (2): 977–993.
575 576	Garg KM, et al. 2016. Genome-wide data help identify an avian species-level lineage that is morphologically and vocally cryptic. Molecular Phylogenetics and Evolution 102: 97–103

578 579 580	Gaubert P. 2011. Family Manidae. In: Wilson, D.E., Mittermeier, R.A. (Eds.), Handbook of the Mammals of the World, Vol. 2: Hoofed Mammals. Lynx Edicions, Barcelona, Spain.
581 582 583	Gaubert P, Antunes A. 2005. Assessing the taxonomic status of the Palawan pangolin Manis Culionensis (Pholidota) using discrete morphological characters. Journal of Mammalogy. 86 (6): 1068–1074.
584 585 586	Hassanin A, Hugot JP, van Vuuren BJ. 2015. Comparison of mitochondrial genome sequences of pangolins (Mammalia, Pholidota). Comptes Rendus Biologies. 338 (4): 260–265.
587 588	Hua L, et al. 2015. Captive breeding of pangolins: Current status, problems and future prospects. ZooKeys. 507 : 99–114.
589 590 591	Jakobsson M, Rosenberg NA. 2007. CLUMPP: A cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. Bioinformatics. 23 (14): 1801–1806.
592 593	Johnson WE, et al. 2010. Genetic restoration of the Florida panther. Science 329 : 1641-1645.
594 595 596	Knowles LL, et al. 2016. Quantifying the similarity between genes and geography across Alaska's alpine small mammals. Journal of Biogeography. 43 (7): 1464–1476.
597 598 599	Kumar S, Stecher G, Tamura K. 2016. MEGA7: Molecular Evolutionary Genetics Analysis Version 7.0 for Bigger Datasets. Molecular Biology and Evolution. 33 (7): 1870-4.
600 601	Lawson DJ, Hellenthal G, Myers S, Falush D. 2012. Inference of Population Structure using Dense Haplotype Data. PLoS Genetics. 8 (1): e1002453.
602 603	Leigh JW, Bryant D. 2015. POPART: full-feature software for haplotype network construction. Methods in Ecology and Evolution. 6 : 1110–1116.
604 605	Leonard JA, et al. 2015. Phylogeography of vertebrates on the Sunda Shelf: a multi- species comparison. Journal of Biogeography. 42 : 871–879.

606 607	Li H, et al. 2009. The Sequence Alignment/Map format and SAMtools. Bioinformatics, 25 (16): 2078–2079.
608	Li H, et al. 2013. Burrows-Wheeler Alignment Tool v0.7.1. Available at:
609	http://bio-bwa.sourceforge.net/bwa.shtml
610	Lischer HEL, Excoffier L. 2012. PGDSpider: An automated data conversion tool for
611	connecting population genetics and genomics programs. Bioinformatics. 28(2):
612	298–299.
613	Lohman DJ, et al. 2010. Cryptic genetic diversity in "widespread" Southeast Asian
614	bird species suggests that Philippine avian endemism is gravely underestimated.
615	Biological Conservation. 143: 1885-1890.
616	Malenfant RM, Coltman DW, Davis CS. 2014. Design of a 9K illumina BeadChip for
617	polar bears (Ursus maritimus) from RAD and transcriptome sequencing.
618	Molecular ecology resources. 15 (3): 587-600.
619	Malinsky M, Trucchi E, Lawson D, Falush D. 2016. RADpainter and
620	fineRADstructure: population inference from RADseq data. BioRxiv 057711.
621	(pre-print). doi: https://doi.org/10.1101/057711.
622	Nash HC, Wong MHG, Turvey ST. 2016. Using local ecological knowledge to
623	determine status and threats of the critically endangered Chinese pangolin
624	(Manis pentadactyla) in Hainan, China. Biological Conservation. 196: 189-195.
625	Ng NS, Wilton PR, Prawiradilaga DM, Tay YC, Indrawan M, Garg KM, Rheindt FE.
626	2017. The effects of Pleistocene climate change on biotic differentiation in a
627	montane songbird clade from Wallacea. Molecular Phylogenetics and Evolution.
628	114 : 353-366.
629	Ogden R, Dawnay N, McEwing R. 2009. Wildlife DNA forensics - Bridging the gap
630	between conservation genetics and law enforcement. Endangered Species
631	Research. 9 (3): 179–195.
632	Ogden R, Linacre A. 2015. Wildlife forensic science: A review of genetic geographic
633	origin assignment. Forensic Science International: Genetics. 18: 152-159.

634635636	Pantel S, Chin SY. 2009. Proceedings of the Workshop on Trade and Conservation of Pangolins native to South and Southeast Asia. TRAFFIC Southeast Asia. Petaling Jaya, Selangor, Malaysia.
637 638 639	Peterson BK, et al. 2012. Double digest RADseq: An inexpensive method for de novo SNP discovery and genotyping in model and non-model species. PLoS ONE. 7 (5).
640 641	Pierson JC, et al. 2016. Genetic factors in threatened species recovery plans on three continents. Frontiers in Ecology and the Environment. 14 (8): 433–440.
642 643	Pritchard JK, Stephens M, Donnelly P. 2000. Inference of population structure using multilocus genotype data. Genetics. 155 (2): 945–959.
644 645 646	Puechmaille SJ. 2016. The program structure does not reliably recover the correct population structure when sampling is uneven: Subsampling and new estimators alleviate the problem. Molecular Ecology Resources. 16 (3): 608–627.
647 648 649	Purcell S, et al. 2007. PLINK: A tool set for whole-genome association and population-based linkage analyses. American Journal of Human Genetics. 81 (3) 559–575.
650	Rambaut A, et al. 2014. Tracer v1.6. Available at: http://beast.bio.ed.ac.uk/Tracer
651 652	Rambaut A, Drummond AJ. 2016. TreeAnnotator v2.4.2. Institute of Evolutionary Biology, University of Edinburgh.
653 654 655	R Core Team 2016. R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing, Vienna, Austria. Available at: http://www.R-project. org/
656 657	RStudio Team 2015. RStudio: Integrated Development for R. RStudio, Inc., Boston, MA. Available at: http://www.rstudio.com/
658 659 660 661	Schueller AM, Hayes DM. 2011. Minimum viable population size for lake sturgeon (Acipenser fulvescens) using an individual-based model of demographics and genetics Canadian Journal of Fisheries and Aquatic Sciences. 68 : 62-73.

002	Segan DB, et al. 2010. Osing conservation evidence to guide management.
663	Conservation Biology. 25(1): 200-202.
564	
665	Shirley MH, et al. 2014. Rigorous approaches to species delimitation have significant
566	implications for African crocodilian systematics and conservation. Proceedings
667	of the Royal Society B. 281 : 20132483.
568	Stamatakis A. 2014. RAxML version 8: a tool for phylogenetic analysis and post-
569	analysis of large phylogenies. Bioinformatics. 30.
570	Sundevall CJ. 1843. Wer sicht der gattung Manis. Kongl. Svenska vetenskaps-
571	akademiens handlingar. 1842: 245-283.
572	Sutherland WJ, et al. 2004. The need for evidence-based conservation. Trends in
573	Ecology and Evolution. 19: 305-308.
574	Tan TK, et al. 2016. PGD: a pangolin genome hub for the research community.
575	Database (Oxford). Available at:
576	http://database.oxfordjournals.org/content/2016/baw063.full.
577	Tay YC, Chng MWP, Sew WWG, Rheindt FE, Tun KPP, Meier R. 2016. Beyond the
578	Coral Triangle: high genetic diversity and near panmixia in Singapore's
579	populations of the broadcast spawning sea star Protoreasternodosus. Royal
580	Society Open Science. 3: 160253.
581	United Nations Office on Drugs and Crime (UNODC). 2016. World Wildlife Crime
582	Report: Trafficking in protected species. Available at:
583	https://www.unodc.org/documents/data-and-
584	analysis/wildlife/World_Wildlife_Crime_Report_2016_final.pdf
585	Wasser SK, et al. 2008. Combating the illegal trade in African elephant ivory with
586	DNA forensics. Conservation Biology. 22: 1065-1071.
587	Wasser SK, et al. 2015. Genetic assignment of large seizures of elephant ivory
588	reveals Africa's major poaching hotspots. Science. 349 (6243): 84–88.
589	WCS News Releases. April 2015. https://newsroom.wcs.org/News-
590	Releases/articleType/ArticleView/articleId/6715/April-27-Indonesian-National-

691	Police-Seize-Major-Shipment-of-Pangolins-Arrest-Smuggler.aspx
692 693	Whitten T, Soeriaatmadja RE, Afiff SA. 1996. The Ecology of Java and Bali. The Ecology of Indonesia Series, II. Periplus Editions (HK) Limited.
694 695	Wilson DE, Reeder DM. (editors). 2005. Mammal Species of the World. A Taxonomic and Geographic Reference (3rd ed). Johns Hopkins University Press
696 697 698	Wirdateti, Semiadi G. 2017. Genetic Variation of Confiscated Pangolins of Sumatra, Java, and Kalimantan based on Control Region Mitochondrial DNA. Jurnal Veteriner. 18 (2): 181-191.
699	Zhang H, et al. 2015. Molecular tracing of confiscated pangolin scales for
700	conservation and illegal trade monitoring in Southeast Asia. Global Ecology and
701	Conservation. 4 : 414–422.
702	Zinenko O, et al. 2016. Hybrid origin of European Vipers (Vipera magnifica and
703	Vipera orlovi) from the Caucasus determined using genomic scale DNA markers
704	BMC Evolutionary Biology. 16 :76.
705	

Figure Captions

- Fig. 1 Locations of seizures of illegally traded pangolins and wild Sunda pangolins of
- known origin across Indonesia (Java, Kalimantan and Sumatra), Malaysia and
- Singapore that were used in this study. Numbers in brackets represent the number
- of pangolins sampled per location
- Fig. 2 Population genetic clusters. The pangolin samples within each geographically
- labelled cluster are similar across each figure. (a) PCA of 11,927 SNPs across 83
- Sunda pangolins. Principal component 1 = 4.47 % and principal component 2 =
- 4.38 %. (b) Network Analysis with the highest genetic distance of clusters using
- 12,150 SNPs across 83 Sunda pangolins from the Walktrap model. (c) Clustered
- fineRADstructure coancestry matrix using 46, 274 loci across 80 Sunda pangolins
- Fig. 3 (a) Maximum likelihood phylogeny in RAxML using 2,365 SNPs. The coloured
- overlaid shading illustrates the clustering results from other methods, PCA, Structure,
- 719 NetView and fineRADstructure to facilitate comparison with those results: blue =
- Borneo, green = Singapore/Sumatra, no colour = Java. (b) The directions of detected
- 721 introgression between pangolin lineages indicated by dashed arrows, from ABBA
- 722 BABA tests using 2,365 SNPs
- Fig. 4 (a) mtDNA phylogeny, with concatenated *Cytb* and *CO1* data. Sample labels
- begin with a key of the SNP cluster result to facilitate comparison: B = Borneo, J =
- Java, S = Singapore/Sumatra. Seizure locations are given following the sample
- name. Asterisks indicate samples where seizure location and SNP cluster result
- were the same geographic area. (b) mtDNA haplotype network. The colour of
- sample label represents the SNP cluster result to facilitate comparison: blue =
- Borneo, green = Singapore/Sumatra, purple = Java, black = mtDNA only. Wild
- samples of known origin are labelled as WILD.
- Fig. 5 Inferred directions of the illegal trade of pangolins. Varied population genomic
- analyses using > 12,000 SNPs all provided compatible results (PCA, Structure,
- NetView and fineRADstructure). A full list of each sample's seizure location or wild
- location is provided below each genetic cluster. The pangolins highlighted in bold
- were not translocated overseas, their seizure location is similar to their genetic origin.
- Pie charts show the trade route at point of seizure of the pangolin (they do not

include the wild samples). Arrows show the direction of trade, with the largest arrow reflecting the highest volume of trade, and the dashed arrows reflecting less trade











