# Response of boreal forest vegetation to the fertility status of the organic layer along a climatic gradient 

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#### Abstract

Relationships between boreal understorey vegetation and chemical composition of the organic layer, stand productivity and climatic factors were studied on 27 coniferous monitoring plots along a $\mathrm{S}-\mathrm{N}$ gradient in Finland using ordination techniques (fitting environmental vectors and surfaces to the NMDS pattern). Total N , exchangeable Ca and $\mathrm{Mg}, \mathrm{BS}$, pH and organic layer thickness formed the subset of variables with maximum correlation with plant community composition. Of the stand parameters, site index ( $H_{100}$ ), volume and basal area correlated best with the vegetation pattern. Our results support the hypothesis that site productivity can be predicted on the basis of the occurrence of understorey plants. This was especially true in herbaceous plants, where the number of species well depicted the site index on both pine and spruce plots. We present quantitative evidence that the response curves along an extensive nitrogen gradient varied between species representing dwarf shrubs, herbs, grasses, mosses and lichens.


## Introduction

The climatic and edaphic conditions are the primary factors affecting differentiation of the understorey vegetation in boreal forests. The implication of this relationship, which assumes that the composition of the understorey vegetation reflects the fertility and productivity status of the site, forms the basis for the Finnish forest site type classification (Cajander 1909, 1949, Ilvessalo 1922, Kuusipalo 1985), and is a well-established tool in practical forestry in all the Fennoscandian countries (Dahl et al. 1967, Hägglund and Lundmark 1977, Økland and Eilertsen 1993, Giesler et al. 1998). The same premise also underlies the widely used
central-European Ellenberg system (Ellenberg et al. 1991, Diekmann and Falkengren-Grerup 1998, Hill et al. 2000, Grandin 2004), which gives indicator values to plant species along gradients depicting e.g. soil fertility, water availability and climatic conditions. Both classification systems, despite being widely applied, still lack quantitative and experimental verification. Therefore, more knowledge is required about the dependence of the composition of the understorey vegetation on soil properties and climate.

The understorey vegetation and the organic layer of boreal forests develop concurrently and interact in many ways (Kuusipalo 1985). Vascular plants depend on the organic layer for their supply of nutrients and water (Nilsson and

Wardle 2005). On the other hand, the understorey vegetation has a considerable impact, through plant litterfall, on the composition, microbiological processes and nutrient cycling of the organic layer (Saetre 1999, Kanerva and Smolander 2007, Hilli et al. 2008). The overstorey trees also have a considerable influence on the understorey vegetation through shading and regulation of the moisture and nutrient levels (Kuusipalo 1985, Barbier et al. 2008).

Analysis of the extensive vegetation data of the Finnish national forest inventories (NFI) using multivariate methods (3rd NFI 1951-1953: Lahti and Väisänen 1987; 8th NFI 1985-1986: Tonteri et al. 1991a, 1991b) has shown that the main gradient in the understorey vegetation of upland forests in Finland ranks the habitats from "dry-and-poor" to "moist-and-rich", and expresses site fertility and forest productivity. Nieppola and Carleton (1991) found this relationship in a relatively large data covering mature Scots pine stands in southern Finland. Nieppola (1993) also investigated how the occurrence and number of species are related to site productivity. However, in these studies the fertility status of the site was determined indirectly using the site index ( $H_{100}$, dominant tree height at the age of 100 years). The chemical and physical properties of the soil have been determined for subjectively classified forest site types in many studies in Finland (e.g. Valmari 1921, Urvas and Erviö 1974, Sepponen 1985, Tamminen 2000). Only a few studies (Kuusipalo 1984, 1985) have analysed the relationships between forest vegetation, soil nutrients and site indices using a continuum approach and applied multivariate ordination and clustering methods.

According to Økland and Eilertsen (1993), the main task of research on forest vegetation is to identify the major vegetational gradients, the ecological complex gradients that cause them, and the spatial scales on which they operate. The species composition and structure of the vegetation varies along ecological gradients in accordance with the differing habitat requirements and tolerances of the individual species. Improved knowledge of the species-specific responses to many abiotic (e.g. moisture, nutrients, light and temperature) and biotic factors is needed in order to better understand the vegetation-environment
dynamics of boreal forests.
In this study, we analyse how the understorey vegetation is related to the large-scale variation in site fertility, stand characteristics and climatic factors on a set of coniferous, intensively monitored upland plots in Finland. Our data enable us to test how well the understorey vegetation reflects the nutrient level of the growth substrate using quantitative data on the vegetation, tree stand and soil chemistry that represents all the boreal bio-geographical zones, and the complete latitudinal gradient running through Finland. We hypothesise that site productivity, as expressed by the site index $\left(H_{100}\right)$, correlates positively with total nitrogen concentrations in the organic layer and can therefore be predicted on the basis of the occurrence of the understorey plant species. Our specific aims are:

- to identify the most important chemical variables in the organic layer and stand characteristics that correlate best with the vegetation pattern,
- to investigate whether the vegetation-environment relationship is linear or non-linear, and
- to determine the abundance (cover) curves of plant species representing different functional groups along the nitrogen, calcium and acidity gradients of the organic layer.


## Material and methods

## Study sites

The 14 Norway spruce (Picea abies) and 13 Scots pine (Pinus sylvestris) Level II plots are a part of the European intensive (Level II) monitoring plot network, established under the EUForest Focus and UN/ECE-ICP Forest monitoring programmes, and represent the hemi-boreal, southern, middle and northern boreal vegetation zones in Finland (Fig. 1). As the distance between the southernmost (Solböle No. 28) and northernmost plots (Kevo No. 22) is over 1100 km , the material is strongly affected by a southnorth climatic gradient. The plots were selected subjectively to be approximately representative of the prevailing climatic conditions, tree species


Fig. 1. The location of the plots and the distribution of the vegetation zones in Finland. $\mathrm{HB}=$ hemi-boreal, SB = southern boreal, $\mathrm{MB}=$ middle boreal, $\mathrm{NB}=$ northern boreal. $\mathrm{P}=\mathrm{Scots}$ pine plot and $\mathrm{S}=$ Norway spruce plot. The age of the stands (years) is given after the plot name.
and site types in Finland. In the final selection, however, factors related to the other monitoring activities (e.g. deposition, soil solution and meteorological monitoring) on the plots, such as logistical aspects and financial restraints, played a decisive role. The soil texture on the spruce plots is till and on the pine plots sorted glaciofluvial material, and the soil type on all the plots is podzol.

Twenty three of the 27 stands are located in semi-natural, commercial forests managed according to standard silvicultural regimes. With a few exceptions (primarily Nos. 2, 5 and 6), the stands were mature and represented a late phase of succession. Silvicultural thinnings, which would strongly affect the species composition, were not carried out in any of the stands during the period 1994-2000. Four of the stands (Nos. $19,20,21$ and 22) are located in strict nature conservation areas and have not been subjected
to forest management for several decades (Kokko et al. 2002). The mean age of these uneven-aged stands (130-180 years) is considerably older than most of managed stands (Fig. 1), and they also contain an appreciable admixture of other tree species. The basic stand characteristics are given in Derome et al. (2007).

## Estimation of plant species cover

The understorey vegetation was inventoried on the 27 plots during July-August in 2003. The bottom layer (mosses, liverworts and lichens) and the field layer (vascular plants, height $<50$ cm : herbs, grasses, sedges, dwarf shrubs and tree seedlings) were investigated. Altogether 16 sample quadrats, each $2 \mathrm{~m}^{2}(1.41 \times 1.41 \mathrm{~m})$, were marked out systematically ( $4 \times 4$ design) on the plot $(30 \times 30 \mathrm{~m})$ reserved for vegetation monitoring (Fig. 2). The cover of the individual plant species on the $2 \mathrm{~m}^{2}$ quadrats was assessed using the following scale: $0.01 \%$ (solitary or very sparsely growing shoots), $0.1 \%, 0.2 \%$, $0.5 \%, 1 \%, 2 \%, \ldots, 99 \%, 100 \%$. The mean cover of each species was calculated over an area of $400 \mathrm{~m}^{2}$ (the Common Sample Area), which is used in all the countries participating in the Forest Focus/ICP Forest monitoring programme. Species (mostly bryophytes and lichens) occurring on the monitoring area ( $400 \mathrm{~m}^{2}$ ), but not on the quadrats, were given the cover value of $0.01 \%$. Two to four botanists inventoried the vegetation, and field tests were carried out to calibrate their individual assessment levels.

## Chemical composition of the organic layer and particle size distribution of the underlying mineral soil

The thickness of the organic layer was measured at 120 points on each plot. Samples of the organic layer outside each vegetation quadrat were taken using a soil auger (diam. 58 mm ) in 2003. The 16 samples were combined to give one composite sample per plot. The samples were dried and milled to pass through a 1 mm sieve. pH was measured in water. Loss in weight on ignition (LOI) was determined by ashing
the samples in a muffle furnace at $550{ }^{\circ} \mathrm{C}$ for 3 hours. Exchangeable $\mathrm{K}, \mathrm{Ca}, \mathrm{Mg}, \mathrm{Zn}, \mathrm{Mn}$, Na and extractable P and S were determined by inductively coupled plasma atomic emission spectrometry (ICP/AES) after extraction with 1 M ammonium acetate ( pH 4.65 ) $+1 \%$ EDTA. Total N and C were determined on a CHN analyser. All concentrations are expressed on an organic matter basis (LOI) in order to reduce the variation caused by the unavoidable inclusion of mineral soil in the organic layer samples. Cation exchange capacity (CEC) and base saturation (BS) were determined using extraction with 0.1 $\mathrm{M} \mathrm{BaCl}{ }_{2}$, and titration to an end-point of 7.8 , on samples taken from the adjacent plot reserved for destructive sampling (e.g. soil and soil solution sampling). The particle-size distribution was determined gravimetrically on mineral soil samples taken from the subsoil (depth $20-50 \mathrm{~cm}$ ). The proportion of the finest particles $(<63 \mu \mathrm{~m})$, i.e. clay + silt, was calculated.

## Stand characteristics

Tree species, breast-height diameter (DBH) and tree height were measured on every tree with a DBH of at least 4.5 cm on each plot in 19992000. KPL software (Heinonen 1994) was used to transform the individual tree measurements into stand-level characteristics (basal area, dominant height, mean height, cubic volume). Taper curve functions by Laasasenaho (1982) were used for estimating the individual tree volumes. Ten dominant trees on the buffer area around each plot were cored in order to determine the age of the dominant tree layer. The site index $\left(H_{100}\right)$ was calculated using the functions of Gustavsen (1980) with dominant height and stand age as input variables.

## Weather and climatic factors

Plotwise values for the effective temperature sum (threshold $+5^{\circ} \mathrm{C}$ ) and annual precipitation in 2003 were derived from model interpolations based on measurements made by the Finnish meteorological institute (Ojansuu and Henttonen 1983). The de Martonne index indicating humid-


Fig. 2. The plot $\left(30 \times 30 \mathrm{~m}=900 \mathrm{~m}^{2}\right)$ used for the inventory of the understorey vegetation. Plant species cover was assessed on the small sample quadrats (16 $\times 2 \mathrm{~m}^{2}$ ). Additional plant species found growing outside the small quadrats were recorded within areas of $4 \times$ $100 \mathrm{~m}^{2}(\mathrm{~A}-\mathrm{D})$ forming the monitoring area.
ity (Tuhkanen 1980) was calculated using the following equation:

$$
\begin{equation*}
H=P /(T+10), \tag{1}
\end{equation*}
$$

where $P=$ annual precipitation (mm), and $T=$ annual mean temperature $\left({ }^{\circ} \mathrm{C}\right)$.

## Statistical analyses

We used global non-metric multidimensional scaling (NMDS) (Minchin 1987) in ordinating the vegetation data of the 27 plots in order to find the main compositional gradients ( R programme, MetaMDS in Vegan library, Oksanen 2007). A two-dimensional solution using the Bray-Curtis coefficients as a measure of dissimilarity in floristic composition between the sample plots was chosen for the final method. The third dimension did not provide any essentially new information and is not presented. Species scores were calculated as cover-weighted averages of the sample scores.

We divided the plots on the vegetation continuum into Cajanderian forest site types (for
short description, see e.g. Lahti and Väisänen 1987) according to the ordination pattern and occurrence of indicator species. We then calculated the site type specific means ( $\pm$ SE) for the cover percentages of the plant species separately for south (hemi-boreal, southern and middle boreal zones) and north (northern boreal zone) Finland (Fig. 1). Corresponding parameters were calculated for the climatic, organic layer, soil texture and stand variables. It was not possible to perform statistical tests between the site types owing to the small number of plots.

Maximum correlations between the environmental variables and the ordination pattern (plot scattergram) were calculated using a linear vector fitting procedure ("Envfit") of Vegan. The significance of the correlations was assessed with Monte Carlo tests ( 1000 permutations). In addition to the vectors, we fitted non-parametrically smoothed surfaces of the selected environmental variables on the ordination space in order to evaluate whether the vegetation-environment relationship was linear or non-linear. The smooth surfaces were fitted using generalized additive models (GAM, gaussian distribution of error) with thin plate splines (Wood 2000), and their significance was tested by permutation tests ( $n$ $=1000$ ) (cf. Virtanen et al. 2006). If the relationship was linear, the fitted surface was a plane and its $R^{2}$ was close to that of the vector; but if the relationship was non-linear, $R^{2}$ was higher for the surface than for the vector. The best subset of variables giving the maximum rank correlation with plant community dissimilarities (ordination pattern) was calculated using the "Bioenv" procedure in the Vegan library (Oksanen 2007).

We fitted linear regressions between the $H_{100}$ index and total N concentration in the organic layer separately for the pine and spruce plots. Pearson correlations were calculated between the $H_{100}$ index and the selected variables of the organic layer and climate, and between the number of species in different plant groups and the N concentration of the organic layer and the $H_{100}$ index. We modelled the cover distributions of the selected vascular and bryophyte species, the sum of lichens, as well as the number of species, in relation to the N gradient using GAMs (R programme, Mgcv library, Wood 2006). The cover of the vascular plants was also modelled
to the $\mathrm{C} / \mathrm{N}$ ratio, pH and exchangeable Ca concentration of the organic layer. We used the quasi-poisson distribution of the error and $\log$ as a link function in the models. The deviance test ( $\chi^{2}$-test) was applied to decide which of the possible error distributions to choose. The degree of freedom was automatically selected with generalized cross-validation in the GAMs, with an upper limit of 4 df . Corresponding models were constructed for the number and cover of species in relation to the $H_{100}$ index.

## Results

## Ordination analysis of the vegetation

The plots were located in accordance with the dominating tree species (spruce plots on the right and pine plots on the left) and the fertility level of the organic layer along the main compositional gradient (Fig. 3a). The most fertile herb-rich heath forests were located on the right, mesic heaths in the centre, followed by sub-xeric and xeric heaths on the left. The corresponding Cajanderian forest site types in the different boreal zones are given in Table 1. Within each site type the northern plots were located slightly to the left of the southern ones, indicating a more severe climate or lower fertility. The second gradient shifted from lower to higher latitudes, and separated the younger succession stages (lower part of the ordination space) from the older northern stands. Of the unmanaged plots, Evo No. 19 and Kevo No. 22 were slightly isolated, whereas Lieksa No. 20 and Oulanka No. 21 were located relatively close to the managed plots representing the corresponding forest types.

The arrangement of the species scores corresponded to the general fertility and succession pattern of the plots. The demanding bryophyte species (e.g. Brachythecium spp.) on the right were replaced by generalist bryophytes in the centre (e.g. Pleurozium schreberi and Dicranum polysetum), and by drought-tolerant lichens (e.g. Cladina rangiferina) on the left in the species ordination (Fig. 3b). For vascular plants, the focus on demanding herbs (e.g. Mainthemum bifolium) and grasses (e.g. Calamagrostis arundinacaea) was on the right, in generalist species

Fig. 3. Global non-metric multidimensional scaling (NMDS) of the vegetation data on the 13 Scots pine (circles and crosses) and 14 Norway spruce plots (triangles) in 2003. - a: Ordination of the sample plots. Site types: filled triangle $=$ herb-rich heath forest; open triangle or circle $=$ mesic heath forest; filled circle = sub-xeric heath forest; cross $=$ xeric heath forest. - b: Weighted averages of the species. The most abundant species labeled with names, other marked with crosses. Abbreviations of the species names are created from the first four letters of the generic and species names. BRAC_SUM = Brachythecium spp., HEPA_SUM = Hepatics species, other than Barbilophozia, BARB_SUM = Barbilophozia spp.


Table 1. Distribution of the plots according to site types and corresponding Cajanderian (1909) forest site types in the hemi-boreal, southern, middle and northern boreal zones.

| Site type | Hemi-boreal | Southern boreal | Middle boreal | Northern boreal |
| :--- | :--- | :--- | :--- | :--- |
| Herb-rich heath forest | OMT | OMT |  |  |
| Mesic heath forest | No. 28 | Nos. 11, 15, 17, 19, 23, 25 |  | HMT |
| Sub-xeric heath forest |  | MT |  | Nos. 3, 5, 7, 8, 21, 31 |
|  |  | Nos. 12, 24 | EVT | EMT |
| Xeric heath forest | Nos. 10,13,14,16 | No. 20 | Nos. 2, 4, 6, 22 |  |
|  |  | CT | ECT | UVET |
|  |  | No. 18 | No.9 | No. 1 |

OMT = Oxalis-Myrtillus type, MT = Myrtillus type, VT = Vaccinium type, CT = Calluna type, EVT = Empetrum-Vaccinium type, ECT = Empetrum-Calluna type, HMT = Hylocomium-Myrtillus type, EMT = Empetrum-Myrtillus type and UVET = Uliginosum-Vaccinium-Empetrum type.
(e.g. Vaccinium myrtillus) in the centre, and in the least demanding species adapted to nutrientpoor, acidic substrates (e.g. Calluna vulgaris) on the left. The abundance of moisture-demanding liverworts (e.g. Barbilophozia spp.) increased towards the northern old stands, which can provide suitable humidity and coarse woody debris for their growing substrates.

The mean values for the cover and number of species in the understorey vegetation and for the environmental variables on different site types, are given for the southern and northern plots in Tables 2 and 3, respectively. Southern herb-rich heath plots maintained a high number of vascular plants and bryophytes, while northern subxeric and xeric plots had species-rich bryophyte and lichen communities (Table 2).

## Explanatory variables

Vegetation differentiation was explained by several inter-correlated environmental variables. The maximum correlations between all the studied environment variables and the NMDS ordination configuration (Fig. 3a) are given in Table 4. Examples of different variables are presented as fitted vectors and surfaces on the ordination space in a panel figure (Fig. 4). The climatic, organic layer, soil texture and stand variables are henceforth treated separately.

## Climate

The combination of climatic variables (out of a maximum of four) giving the best rank correlation with the vegetation pattern was effective temperature sum, annual precipitation and humidity index ( $r=0.241$ ). Effective temperature sum ( $\max -\min =1462-763$ degree days, d.d.) and annual precipitation ( $\max -\mathrm{min}=714-$ 400 mm ) decreased from south to north, whereas humidity and altitude had the maximum values on the northern hills. All the climatic variables showed a non-linear relationship with the ordination pattern ( $R^{2}$ values for surfaces were higher than linear $R^{2}$ values) (Fig. 4 and Table 4), which may explain the relatively low rank correlation for the subset of variables. Temperature isoclines
for the spruce stands paralleled with dimension 1 ( $x$-axis), but those for the pine stands with dimension 2 ( $y$-axis), resulting in marked curvature on the fitted surface.

## Organic layer and soil texture

Total nitrogen ( N ), exchangeable calcium (Ca) and magnesium ( Mg ) concentrations, base saturation (BS), pH and thickness of the organic layer formed the subset of variables for the organic layer (out of a maximum of $14, \mathrm{Na}$ excluded) that gave the maximum rank correlation ( $r=$ 0.592 ) with the plant community composition. In general, the main vegetation gradient was strongly related to the total N concentration, which increased from the xeric and subxeric pine plots towards the herb-rich heath plots dominated by spruce (Fig. 4 and Table 3). The surface pattern for the $\mathrm{C} / \mathrm{N}$ ratio (not shown) was almost the same as that for N , but the direction of the values was the opposite (maximum in xeric plots). Of the macronutrients ( $\mathrm{N}, \mathrm{P}, \mathrm{K}, \mathrm{Ca}, \mathrm{S}$ and Mg ), only P and K had relatively low $R^{2}$ values for both the fitted vectors and surfaces. The surfaces of $P$ and K were clearly unimodal, with maximum values on the "high" altitude Oulanka and Pallasjärvi plots. The P/N ratio (not shown) had a similar surface pattern as P (Fig. 4), with the lowest values on the southern herb-rich and mesic heaths (Table 3). Calcium, $\mathrm{Mg}, \mathrm{BS}$ and pH increased towards the southern herb-rich heaths, i.e. from the lower left to the upper right in the ordination space. The pH values correlated strongly ( $p<0.001$ ) with Ca and BS (Table 5). Also extractable sulphur (S) (not shown) increased towards the herb-rich heaths, but from the upper left to the lower right. CEC (not shown) had a very similar surface pattern to that of BS. Of the micronutrients (Mn and Zn ), only Mn (not shown) correlated with the ordination pattern, following a linearly increasing trend from the lower left to the upper right. Manganese had strong positive correlation ( $p<$ 0.001 ) with $\mathrm{pH}, \mathrm{Ca}, \mathrm{BS}$ and the proportion of fine particles (Table 5). The Na concentrations were highest on the plots subjected to a marine influence (Nos. 1, 22 and 23).

The relationships between the vegetation and the variables most strongly depicting site fertility
Table 2. Means and standard errors of the mean (SE) of the cover of the selected plant species and the number of species in the species groups in the four site type classes. South = hemi-boreal, southern and middle boreal zone, North = northern boreal zone. $n=$ number of plots. "Other woody species" includes dwarf shrubs and shrubs and trees (height < 50 cm ). "Cladina sum" = reindeer lichens, "Cladonia sum" = cup lichens, and "other lichens" = Cetraria spp., Stereocaulon spp. and the leather lichens (Peltigera spp. and Nephroma arcticum), $=$ = species not found.

| Species group | Herb-rich heath forest South ( $n=7$ ) |  | Mesic heath forest |  |  |  | Sub-xeric heath forest |  |  |  | Xeric heath forest |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | South ( $n=2$ ) |  | North ( $n=6$ ) |  | South ( $n=5$ ) |  | North ( $n=4$ ) |  | South ( $n=2$ ) |  | $\begin{gathered} \text { North } \\ (n=1) \\ \text { Mean } \end{gathered}$ |
|  | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE |  |
| Dwart shrubs |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Calluna vulgaris | - |  | - |  | 0.03 | 0.03 | 1.53 | 1.15 | 2.81 | 2.41 | 4.39 | 1.76 | - |
| Empetrum nigrum | - |  | - |  | 0.95 | 0.86 | 0.20 | 0.17 | 9.79 | 4.42 | 0.81 | 0.81 | 26.20 |
| Vaccinium myrtilus | 10.81 | 4.10 | 22.22 | 16.28 | 43.59 | 4.52 | 16.03 | 4.04 | 9.79 | 1.69 | 5.64 | 3.21 | 0.39 |
| Vaccinium vitis-idaea | 1.51 | 0.72 | 6.67 | 4.08 | 10.02 | 1.65 | 19.45 | 5.35 | 13.81 | 2.84 | 17.22 | 5.78 | 19.56 |
| Other woody species | 2.39 | 0.90 | 2.69 | 0.14 | 1.69 | 0.71 | 1.09 | 1.01 | 1.47 | 0.52 | 0.25 | 0.18 | 0.52 |
| Number of species | 6.7 | 0.6 | 9.00 | 4.00 | 7.8 | 1.6 | 7.2 | 1.1 | 9.3 | 0.5 | 6.0 | 0 | 5 |
| Herbs |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Maianthemum bifolium | 6.60 | 1.80 | 3.65 | 1.40 | - |  | 0.36 | 0.32 | 0 | 0 | - |  | - |
| Oxalis acetosella | 3.96 | 1.12 | 0.02 | 0.02 | - |  | - |  | 0 | 0 | - |  | - |
| Trientalis europaea | 0.63 | 0.21 | 0.94 | 0.64 | 0.02 | 0.02 | 1.96 | 1.18 | 0 | 0 | - |  | - |
| Other herb species | 7.34 | 3.13 | 0.90 | 0.07 | 0.57 | 0.27 | 0.06 | 0.02 | 0 | 0 | 0.01 | 0 | - |
| Number of species | 17.0 | 2.7 | 10.5 | 0.5 | 4.3 | 0.5 | 4.6 | 1.0 | 1.0 | 0.6 | 0.5 | 0.5 | 0 |
| Grasses \& sedges |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Calamagrostis arundinacea | 1.37 | 1.01 | 0.26 | 0.13 | - |  | 0 | 0 | - |  | - |  | - |
| Deschampsia flexuosa | 2.35 | 1.68 | 1.98 | 1.30 | 1.74 | 0.56 | 0.75 | 0.46 | 0.29 | 0.23 | - |  | - |
| Luzula pilosa | 0.44 | 0.18 | 0.17 | 0.14 | 0.14 | 0.09 | 0.02 | 0.01 | - |  | - |  | - |
| Other grass species | 0.67 | 0.37 | 0.13 | 0.12 | - |  | 0.10 | 0.10 | - |  | - |  | - |
| Number of species | 6.3 | 1.2 | 4.5 | 0.5 | 1.8 | 0.4 | 1.8 | 0.7 | 0.8 | 0.3 | 0 | 0 | 0 |
| Bryophytes |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Dicranum polysetum | 0.73 | 0.29 | 6.69 | 2.19 | 0.60 | 0.19 | 17.16 | 7.28 | 1.64 | 0.68 | 8.98 | 7.28 | 0.01 |
| Dicranum scoparium | 2.01 | 0.53 | 4.41 | 1.64 | 2.56 | 0.57 | 0.30 | 0.17 | 1.93 | 0.78 | 0.32 | 0.12 | - |
| Hylocomium splendens | 8.67 | 4.09 | 9.09 | 0.71 | 40.36 | 5.19 | 8.66 | 4.69 | 1.39 | 1.25 | 0.08 | 0.08 | - |
| Pleurozium schreberi | 26.65 | 7.43 | 32.69 | 5.81 | 31.35 | 7.89 | 48.09 | 7.95 | 50.91 | 10.70 | 51.13 | 2.38 | 7.88 |
| Other mosses | 10.79 | 3.22 | 7.58 | 2.80 | 8.71 | 2.17 | 4.61 | 2.92 | 7.74 | 1.96 | 0.19 | 0.04 | 3.68 |
| Hepatics sum | 1.46 | 0.89 | 0.02 | 0.02 | 2.48 | 1.27 | 0 | 0 | 1.40 | 0.78 | 0.01 | 0 | 4.10 |
| Number of species | 22.9 | 2.5 | 19.5 | 0.5 | 18.8 | 1.5 | 8.2 | 1.4 | 20.8 | 2.3 | 8.5 | 2.5 | 20.0 |
| Lichens |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Cladina sum | - |  | 0.04 | 0.04 | 0.01 | 0 | 0.13 | 0.09 | 4.43 | 1.03 | 9.78 | 6.65 | 38.32 |
| Cladonia sum | 0.03 | 0.02 | 0.01 | 0 | 0.01 | 0 | 0.07 | 0.04 | 2.03 | 0.84 | 2.85 | 2.45 | 3.98 |
| Other lichens | - |  | 0 | 0 | 0.02 | 0.02 | 0.04 | 0.04 | 1.60 | 1.10 | 0.27 | 0.06 | 0.40 |
| Number of species | 1.4 | 0.6 | 2.0 | 1.0 | 4.5 | 1.8 | 6.4 | 1.7 | 19.0 | 3.4 | 17.5 | 1.5 | 21.0 |

Table 3. Means and standard error of the mean (SE) of the environmental variables characterising the climate of the plot, organic layer, soil texture and tree stand in the four site types. See Table 2 for further explanations. Exch. = exchangeable, Extr. = extractable, om = organic matter.

| Environmental variable | Herb-rich heath forest South ( $n=7$ ) |  | Mesic heath forest |  |  |  | Sub-xeric heath forest |  |  |  | Xeric heath forest |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | South ( $n=2$ ) |  | North ( $n=6$ ) |  | South ( $n=5$ ) |  | North ( $n=4$ ) |  | South ( $n=2$ ) |  | North $(n=1)$ <br> Mean |
|  | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE | Mean | SE |  |
| Altitude (m a.s.l.) | 94.1 | 25.6 | 86.5 | 56.5 | 265.0 | 11.0 | 116.8 | 22.1 | 206.0 | 40.2 | 69.0 | 21.0 | 105.0 |
| Eff. temp. sum (d.d.) | 1327.9 | 33.9 | 1322.5 | 24.5 | 911.8 | 26.2 | 1318.4 | 41.9 | 864.5 | 66.4 | 1276.5 | 113.5 | 763.0 |
| Annual precipitation (mm) | 588.0 | 29.2 | 530.8 | 2.3 | 521.0 | 14.8 | 618.2 | 30.1 | 466.6 | 41.5 | 564.9 | 88.7 | 399.7 |
| Humidity index | 41.8 | 2.2 | 37.0 | 0.2 | 50.2 | 0.3 | 44.9 | 2.9 | 45.9 | 2.7 | 42.3 | 3.8 | 40.5 |
| Organic layer |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Thickness (cm) | 3.6 | 0.4 | 3.4 | 1.4 | 3.7 | 0.4 | 3.1 | 0.1 | 2.5 | 0.6 | 2.5 | 1.1 | 1.1 |
| pH | 3.9 | 0.1 | 3.7 | 0.0 | 3.7 | 0.1 | 3.6 | 0.1 | 3.6 | 0.0 | 3.4 | 0.1 | 3.4 |
| Total $\mathrm{N}\left(\mathrm{g} \mathrm{kg}^{-1} \mathrm{om}\right)$ | 21.4 | 0.7 | 20.5 | 1.1 | 13.0 | 0.2 | 14.9 | 0.7 | 11.5 | 1.0 | 14.7 | 2.6 | 12.8 |
| Extr. P ( $\left.\mathrm{mg} \mathrm{kg}^{-1} \mathrm{om}\right)$ | 171.0 | 36.6 | 104.0 | 25.9 | 491.8 | 32.7 | 242.0 | 28.2 | 256.9 | 41.0 | 222.1 | 11.7 | 274.5 |
| Exch. $\mathrm{K}\left(\mathrm{mg} \mathrm{kg}^{-1} \mathrm{om}\right)$ | 978.9 | 69.8 | 831.6 | 88.6 | 1352.9 | 35.4 | 853.2 | 71.6 | 915.7 | 49.3 | 790.6 | 43.6 | 736.3 |
| Exch. $\mathrm{Ca}\left(\mathrm{mg} \mathrm{kg}^{-1} \mathrm{om}\right)$ | 3485.1 | 615.0 | 2018.4 | 12.5 | 2346.0 | 172.4 | 2205.9 | 158.3 | 1642.7 | 54.6 | 1662.3 | 641.2 | 1610.1 |
| Exch. $\mathrm{Mg}\left(\mathrm{mg} \mathrm{kg}{ }^{-1} \mathrm{om}\right)$ | 608.6 | 68.1 | 388.6 | 42.9 | 447.7 | 25.4 | 317.9 | 21.5 | 374.6 | 52.8 | 207.0 | 16.6 | 491.9 |
| Extr. S (mg kg ${ }^{-1}$ om) | 275.5 | 13.7 | 228.8 | 16.4 | 246.7 | 16.2 | 220.8 | 10.7 | 198.2 | 15.2 | 197.8 | 20.7 | 176.1 |
| Exch. Mn (mg kg ${ }^{-1}$ om) | 569.4 | 221.4 | 153.3 | 15.1 | 596.0 | 110.8 | 197.9 | 24.4 | 245.3 | 72.6 | 96.3 | 7.0 | 109.5 |
| Exch. Zn (mg kg $\left.{ }^{-1} \mathrm{om}\right)$ | 30.6 | 8.7 | 24.9 | 5.4 | 25.4 | 3.0 | 29.9 | 1.5 | 19.9 | 2.3 | 27.4 | 3.1 | 20.3 |
| Exch. $\mathrm{Na}\left(\mathrm{mg} \mathrm{kg}{ }^{-1} \mathrm{om}\right)$ | 101.9 | 21.7 | 86.5 | 17.9 | 46.2 | 1.7 | 65.9 | 9.6 | 72.6 | 22.5 | 66.9 | 0.5 | 135.4 |
| C/N ratio | 24.1 | 0.8 | 25.0 | 1.2 | 39.8 | 0.8 | 35.0 | 2.4 | 44.9 | 3.1 | 38.9 | 5.0 | 41.8 |
| $\mathrm{P} / \mathrm{N}$ ratio | 8.1 | 1.8 | 5.2 | 1.5 | 37.8 | 2.4 | 16.6 | 2.7 | 23.2 | 4.5 | 15.7 | 3.6 | 21.5 |
| CEC (meq kg ${ }^{-1}$ om) | 360.9 | 21.6 | 379.9 | 44.1 | 293.0 | 13.7 | 285.0 | 17.1 | 267.8 | 27.2 | 283.5 | 67.1 | 325.7 |
| BS\% | 68.9 | 5.4 | 67.5 | 6.0 | 60.4 | 2.5 | 55.7 | 2.0 | 49.5 | 1.3 | 43.7 | 7.1 | 54.7 |
| Mineral soil |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Fine particles (clay + silt) (\%) | \%) 41.3 | 10.9 | 21.3 | 20.3 | 36.3 | 3.6 | 9.2 | 4.8 | 31.7 | 8.7 | 17.0 | 13.9 | 13.4 |
| Stand |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of stems per ha | 711.6 | 108.5 | 733.5 | 22.5 | 1321.2 | 160.5 | 699.6 | 150.7 | 1152.0 | 234.4 | 472.5 | 94.5 | 300.0 |
| Volume ( $\mathrm{m}^{3} \mathrm{ha}^{-1}$ ) | 362.0 | 38.8 | 245.6 | 24.3 | 119.5 | 12.1 | 233.1 | 15.0 | 98.1 | 19.5 | 124.4 | 24.9 | 71.8 |
| Basal area ( $\mathrm{m}^{2}$ ha-1) | 32.6 | 3.3 | 26.7 | 1.2 | 19.2 | 1.3 | 23.8 | 1.9 | 16.6 | 2.7 | 15.0 | 0.8 | 12.5 |
| Stem diameter (cm) | 24.1 | 1.3 | 21.2 | 1.6 | 13.2 | 1.0 | 20.9 | 1.5 | 12.2 | 1.4 | 20.1 | 2.7 | 21.6 |
| Tree height (m) | 21.9 | 1.0 | 17.7 | 2.3 | 10.3 | 1.0 | 19.1 | 1.2 | 9.4 | 1.6 | 16.7 | 3.0 | 10.2 |
| Stand age (years) | 84.3 | 14.6 | 57.5 | 2.5 | 120.8 | 21.5 | 80.0 | 13.8 | 101.3 | 27.3 | 105.0 | 15.0 | 200.0 |
| Site's $H_{100}$ index (m) | 29.5 | 1.6 | 31.2 | 1.3 | 15.1 | 1.5 | 25.4 | 0.8 | 13.6 | 2.2 | 18.1 | 2.2 | 8.7 |

( $\mathrm{N}, \mathrm{C} / \mathrm{N}$ ratio and BS ) and which were related to soil acidity ( $\mathrm{pH}, \mathrm{Ca}$ ) were close to linear, whereas $\mathrm{K}, \mathrm{P}$ and Mg showed a non-linear response that was probably modified by local factors such as topography (altitude) or the parent soil material. The thickness of the organic layer had a low correlation with the vegetation pattern, although it increased linearly from the xeric pine plots to the moister spruce plots. Similarly, the proportion of fine particles (clay + silt) in the mineral soil was the highest in the herb-rich heath plots dominated by spruce, but there was great variation in the soil texture within the forest site types, and the correlation with the vegetation gradient remained relatively low.

## Tree stand

The vegetation gradient was differentiated by the overstorey tree species: the more fertile, moist sites dominated by spruce were located on the right, and the nutrient-poor, drier pine plots on the left in the ordination space. The subset of stand variables (out of a maximum of 7) giving maximum correlation with the variation in the plant community comprised the site index $\left(H_{100}\right)$, basal area and volume ( $r=0.460$ ). Tree species was not included in the model because it correlated too strongly with the other stand characteristics. The number of stems per ha did not explain the ordination pattern. Stand age increased linearly towards the northern plots in the upper part of the ordination space, but its correlation was lower than that for the variables depicting tree size (Fig. 4). Basal area and volume (not shown) had similar correlations with the vegetation pattern, and showed a nonlinear increasing trend towards the fertile site types. $H_{100}$ had an almost linear relationship with the main fertility gradient of the vegetation, and it increased from the upper left to the lower right in the same way as the total N concentration.

## Species responses to the fertility gradient and the site index

Different functional plant groups showed varying distributions in their cover and species richness
along the fertility gradient (Fig. 5). The cover distribution of bryophytes depended on the taxonomic group (Fig. 5a). The most common species, Pleurozium schreberi, had a high cover on almost all the plots, showing an increasing trend towards the northern N-poor sites. The curves of Hylocomium splendens and Dicranum polysetum were unimodal, with maximum cover on the left and in the centre of the N gradient, respectively. On the other hand, the cover of Dicranum majus increased slightly towards the N -rich and that of lichens to the N -poor end of the gradient.

Evergreen ericaceous dwarf shrubs dominated on the sites with a relatively low pH , low N

Table 4. Maximum linear correlations ( $R^{2}$ ) of the environmental variables with the NMDS ordination pattern (see Figs. 3 and 4). The variables are ranked from high to low correlation in each variable group. The significance of the correlations ( $p$ ) was calculated using the Monte Carlo test. Exch. = exchangeable, Extr. = extractable. $n=27$ plots.

| Variable | $R^{2}$ | $p<$ |
| :--- | :---: | :---: |
| Effective temperature sum | 0.493 | 0.001 |
| Annual precipitation | 0.333 | 0.008 |
| Altitude | 0.122 | 0.214 |
| Humidity index | 0.051 | 0.535 |
| Organic layer |  |  |
| Total N | 0.665 | 0.001 |
| C/N ratio | 0.660 | 0.001 |
| pH | 0.600 | 0.001 |
| BS | 0.581 | 0.001 |
| Extr. S | 0.533 | 0.001 |
| Exch. Mg | 0.515 | 0.001 |
| Exch. Ca | 0.422 | 0.001 |
| CEC | 0.361 | 0.001 |
| Exch. Mn | 0.272 | 0.026 |
| Thickness | 0.180 | 0.094 |
| P/N ratio | 0.150 | 0.141 |
| Exch. K | 0.089 | 0.324 |
| Exch. Zn | 0.073 | 0.441 |
| Extr. P | 0.057 | 0.475 |
| Exch. Na | 0.033 | 0.655 |
| Mineral soil |  |  |
| Fine particles (clay + silt) (\%) | 0.233 | 0.042 |
| Stand |  |  |
| Tree species | 0.683 | 0.001 |
| H100 | 0.674 | 0.001 |
| Height | 0.540 | 0.001 |
| Volume | 0.492 | 0.001 |
| Basal area | 0.476 | 0.001 |
| Diameter | 0.336 | 0.009 |
| Age | 0.295 | 0.020 |
| Number of stems | 0.025 | 0.736 |



Fig. 4. Non-metric multidimensional scaling (NMDS) ordination (same as in Fig. 3a) with fitted environmental vectors and non-linear GAM surfaces. The vectors show the direction and magnitude of the linear correlations of the environmental variables with the plot scores. The surfaces depict the smooth trends between the environmental variables and plot scores. Crosses indicate the position of the spruce plots, and circles of the pine plots.
and Ca concentrations and high $\mathrm{C} / \mathrm{N}$ ratio in the organic layer, whereas grasses and herbs were the most abundant at the opposite end of the fertility gradient (Fig. 5b-e). The deciduous dwarf shrub Vaccinium myrtillus had the maximum cover between these groups. The most common vascular plant species reached their maximum cover along the N gradient in the following order: Calluna vulgaris (lowest N) < Empetrum nigrum, Vaccinium vitis-idaea < Vaccinium myrtillus < Deschampsia flexuosa < Maianthemum bifolium (Fig. 5b and Table 3).

The $H_{100}$ values ranged from 8.2 to 26.7 m on the pine and from 10.2 to 34 m on the spruce
plots. The $H_{100}$ values increased with increasing total N concentrations in the organic layer on the plots of both species (Fig. 6a). The Pearson correlation between the $H_{100}$ values and N concentrations was higher for the spruce plots than for the pine plots (Table 6). On the pine plots the exchangeable Ca concentration had higher correlation than that for N . High $H_{100}$ values invariably occurred on the nitrogen-rich plots. The nitro-gen-poor plots, however, represented a rather wide range of site index values. $H_{100}$ values near and even below 10 occurred on the northernmost plots, reflecting the severe, near sub-arctic climatic conditions in northernmost Finland. When

|  | pH | N | P | K | Ca | Mg | S | Mn | Zn | Na | $\mathrm{C}: \mathrm{N}$ | P:N | CEC | BS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| N | 0.582 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| P | -0.248 | -0.596 |  |  |  |  |  |  |  |  |  |  |  |  |
| K | 0.259 | -0.234 | 0.792 |  |  |  |  |  |  |  |  |  |  |  |
| Ca | 0.829 | 0.503 | -0.140 | 0.286 |  |  |  |  |  |  |  |  |  |  |
| Mg | 0.595 | 0.535 | -0.039 | 0.321 | 0.708 |  |  |  |  |  |  |  |  |  |
| S | 0.459 | 0.566 | 0.092 | 0.395 | 0.435 | 0.579 |  |  |  |  |  |  |  |  |
| Mn | 0.743 | 0.151 | 0.289 | 0.617 | 0.794 | 0.573 | 0.443 |  |  |  |  |  |  |  |
| Zn | 0.671 | 0.255 | -0.048 | 0.211 | 0.732 | 0.385 | 0.217 | 0.786 |  |  |  |  |  |  |
| Na | 0.011 | 0.521 | -0.466 | -0.394 | 0.011 | 0.488 | 0.351 | -0.184 | -0.063 |  |  |  |  |  |
| C:N | -0.613 | -0.966 | 0.557 | 0.199 | -0.514 | -0.493 | -0.574 | -0.160 | -0.240 | -0.454 |  |  |  |  |
| $\mathrm{P}: \mathrm{N}$ | -0.329 | -0.755 | 0.962 | 0.716 | -0.251 | -0.157 | -0.057 | 0.219 | -0.105 | -0.516 | 0.730 |  |  |  |
| CEC | 0.614 | 0.716 | -0.409 | -0.015 | 0.580 | 0.606 | 0.332 | 0.394 | 0.491 | 0.438 | -0.680 | -0.514 |  |  |
| BS | 0.838 | 0.586 | -0.104 | 0.336 | 0.775 | 0.656 | 0.419 | 0.643 | 0.571 | 0.012 | -0.596 | -0.224 | 0.668 |  |
| Fine particles | 0.458 | 0.208 | 0.187 | 0.498 | 0.608 | 0.624 | 0.492 | 0.699 | 0.382 | 0.080 | -0.191 | 0.119 | 0.419 | 0.440 |

the northern plots (effective temperature sum $700-1100$ d.d.) were analysed separately, there was no notable correlation between the N concentration and $H_{100}$. The N concentrations on the northernmost plots (Sevettijärvi No. 1 and Kevo No. 22) were slightly higher than those in central Lapland, and this had a clear effect on the results. The effective temperature sum, as well as the annual precipitation, correlated positively with the $H_{100}$ values of both species (Table 6).

The correlations between the number of species and the N concentration in the organic layer varied in the different plant groups (Fig. 5f and Table 7). The number of herb species correlated the best with the N concentration on both the pine and spruce plots, as well as in the combined data. The number of grass species also showed a positive trend with the N concentration, especially on the spruce plots. On the other hand, the species richness of dwarf shrubs and lichens on the spruce plots, and of hepatics on the pine plots, correlated negatively with the N concentration in the organic layer. In general, the relationship between the species numbers and $H_{100}$ (Fig. 6b) was relatively similar to that for the N concentration (Fig. 5f). On the pine plots the number of herb species correlated even better with $H_{100}(r=$ $0.735)$ than with the N concentration ( $r=0.531$ ) (Table 7). The cover curves of the species or species groups along the $H_{100}$ gradient were, in most cases, wide and did not give any clear indication of the productivity level (Fig. 6c and d).

Table 6. Pearson correlations between the site index $\left(H_{100}\right)$ values of Scots pine and Norway spruce plots and the variables of organic layer and climate giving the highest correlation with the vegetation pattern. See Table 3 for explanation of the variable names.

|  | Scots pine <br> $(n=13)$ | Norway spruce <br> $(n=14)$ |
| :--- | :---: | :---: |
| Total N | $0.523^{\circ}$ | $0.898^{* * *}$ |
| pH | 0.352 | 0.314 |
| Exch. Ca | $0.648^{*}$ | 0.211 |
| Exch. Mg | $-0.543^{\circ}$ | 0.213 |
| BS | 0.310 | 0.227 |
| Org. layer thickness | 0.386 | -0.144 |
| Eff. temperature sum | $0.896^{* * *}$ | $0.909^{* * *}$ |
| Annual precipitation | $0.850^{* * *}$ | $0.478^{\circ}$ |
| Statistical significance: ${ }^{* * *}=p<0.001,{ }^{*}=p<0.05,{ }^{\circ}=$ |  |  |
| $p<0.10 . n=$ number of plots. |  |  |



Fig. 5. (a) The cover of the selected bryophyte species and the sum of lichens in relation to the N gradient of the organic layer. The cover of the selected vascular plant species in relation to (b) N , (c) $\mathrm{C} / \mathrm{N}$ ratio, (d) pH, and (e) Ca gradient of the organic layer. (f) The number of species in different functional plant groups in relation to the N gradient of the organic layer. om = organic matter content. All the curves are the predictions given by GAM models. The values for N were In -transformed in the models.


Fig. 6. - a: The site's $H_{100}$ index in relation to the N gradient in the organic layer (linear regression lines marked for the Scots pine and Norway spruce plots). - $\mathbf{b}$ : The number of species. - $\mathbf{c}$ : The cover of selected vascular plant species. - d: Sum of mosses, lichens and liverworts in relation to the $H_{100}$ index. om = organic matter content. The curves in $\mathbf{b}-\mathbf{d}$ are the predictions given by GAM models.

Table 7. Pearson correlations between the species numbers in different plant groups and the total N concentration in the organic layer and the site's $H_{100}$ index.

|  | Scots pine ( $n=13$ ) |  | Norway spruce ( $n=14$ ) |  | All plots ( $n=27$ ) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total N | $H_{100}$ | Total N | $H_{100}$ | Total N | $H_{100}$ |
| Dwarf shrubs | -0.329 | -0.401 | $-0.673^{* *}$ | $-0.632^{*}$ | $-0.656^{* * *}$ | $-0.586{ }^{* *}$ |
| Herbs | $0.531^{\circ}$ | 0.735** | 0.717** | 0.617* | $0.793^{* * *}$ | $0.663^{* * *}$ |
| Grasses | 0.454 | $0.492^{\circ}$ | 0.687** | 0.587* | $0.764^{* * *}$ | $0.619^{* * *}$ |
| Hepatics | $-0.533^{\circ}$ | -0.886*** | -0.391 | -0.394 | -0.216 | -0.493** |
| Mosses | $-0.474^{\circ}$ | -0.601* | 0.357 | 0.420 | 0.457* | 0.268 |
| Lichens | -0.182 | -0.791** | -0.648* | -0.550* | -0.549** | -0.659*** |

Statistical significance: ${ }^{* * *}=p<0.001,{ }^{* *}=p<0.01,^{*}=p<0.05,^{\circ}=p<0.10 . n=$ number of plots.

## Discussion

The strong climatic shift from south to north, combined with the range in properties of the organic layer, formed a complex environmental gradient that affected both the understorey vegetation and the tree stands. In many cases it was impossible to separate the independent effects of the inter-correlated environmental factors on the vegetation. The age of the stands also increased towards the north, which added variation to the results. The difference between the unmanaged and managed plots appeared to be obscured by the effect of stand age. Our results support earlier findings concerning the vegetation-soil relationships in boreal forests (e.g. Cajander 1949, Dahl et al. 1967, Økland and Eilertsen 1993, Giesler et al. 1998), that the most important chemical variables in the soil causing broad-scale differentiation of the boreal understorey vegetation are $\mathrm{N}, \mathrm{pH}$ and BS. Our results also indicate that the relationships between these variables and the dominant vegetation gradient were close to linear, although one or two plots (e.g. Uusikaarlepyy No. 23 , which was located relatively close to a fur farm) caused irregularities in the overall pattern. However, the importance of N for the vegetation pattern was not so clear when only part of the climatic gradient was investigated. In northern Finland for instance, the vegetation was also related to the P and K concentrations in the organic layer.

Many studies have shown that N deficiency is one of the primary ecological factors that limit the biological production of boreal forests (e.g. Tamm 1991). In our study, only the total N concentration of the organic layer was determined. Generally, only about $1 \%$ of the total N is available to plants in mineral form $\left(\mathrm{NH}_{4}\right.$ and $\left.\mathrm{NO}_{3}\right)$ in boreal forests (Smolander et al. 2005). Current knowledge, however, emphasizes that plants can also take up some N in organic forms (e.g. amino acids) via mycorrhizal associations (Näsholm et al. 1998). Read (1991) suggested that, in regions with a cool, humid climate, low pH and slow rate of decomposition, organic N and ammonium are the main sources of N for plants. Species possessing ectomycorrhizal (ECM) and ericoid mycorrhizal (ERM) associations should predominate in these regions. Experimental evidence has shown
that ECM and ERM fungi can mobilise N and P from organic polymers (Read and Perez-Moreno 2003). In contrast, in warmer, drier regions with a higher pH and more rapid N mineralization and nitrification, nitrate is the prevailing N source for plants. In these regions, plants with arbuscular mycorrhizae (AM) form dominant communities and these plant species can assimilate amino acid N (Read and Perez-Moreno 2003).

Our results on the distribution of vascular plant species along the total N gradient are consistent with Read's (1991) hypothesis. The two most abundant dwarf shrubs (Vaccinium myrtillus and V. vitis-idaea), which possess ERMs, were found on all the plots, but their abundance increased towards the N -poor plots at high latitudes. Mäkipää (1999) did not find differences in the ecological N optima between these two species in the material from southern Finland, but in our data $V$. myrtillus had the maximum cover at higher N level than V. vitis-idaea. The evergreen dwarf shrubs, Calluna vulgaris and Empetrum nigrum, reached their maximum cover on more acidic and nutrient-poor soils than the Vaccinium species. The dominance of evergreen dwarf shrubs in nutrient-poor environments can be explained by their low rates of nutrient loss (Aerts 1995) and their polyphenolic-rich leaves and litter (Jalal et al. 1982), which accumulate in the organic layer and can be decomposed by their own fungal partner (ERM) (Read and PerezMoreno 2002). On the other hand, the herbs (e.g. Maianthemum bifolium, Rubus saxatilis and Oxalis acetosella) and grasses (e.g. Calamagrostis arundinacea), which possess AMs, were the most abundant on southern plots with higher N concentrations and somewhat higher pH . The most common grass, Deschampsia flexuosa, which is known to be able to take up the amino acid glycine (Näsholm et al. 1998), was also present on the northern plots.

The moss and lichen layer of boreal forests is known to respond to a complex of factors, including moisture and light conditions, as well as the structural properties of the organic layer (Økland and Eilertsen 1993, Rydgren 1996). Bryophytes obtain almost all of their water and nutrients from precipitation and throughfall, but some of them also obtain nutrients from water that has been in contact with the organic layer (Bates
1992). The N retention ability of different boreal bryophyte species varies, and this can control the between-species interactions. For instance, the dominance of Pleuroziumin schreberi may be due to its high biomass production per tissue N concentration (Salemaa et al. 2008). Bryophytes (DeLuca et al. 2002), and leather lichens (Kallio et al. 1972) in the northern N -poor regions have epiphytic cyanobacteria, which are able to biologically fix N . The vegetation-soil dynamics on the northern plots, where the cover of feathermosses often exceeds $50 \%$ and the occurrence of leather lichens (Nephroma sp. and Peltigera sp.) is not rare, may be influenced by the presence of N -fixing organisms.

In addition to N , the BS of the organic layer also correlated positively with the broad-scale vegetation differentiation in our data. BS (i.e. the proportion of base cations occupying the exchange sites on organic soil particles), which is mainly used to depict the acid-buffering capacity of soils, is also an indicator of the availability of $\mathrm{Ca}, \mathrm{Mg}$ and K to plants. BS had a higher positive correlation with pH than with the N concentration. On some of the plots classified as herb-rich heath forests (e.g. Punkaharju No. 17 and Uusikaarlepyy No. 23), the N concentration in the organic layer was relatively high ( 20 and $24 \mathrm{~g} \mathrm{~kg}^{-1}$, respectively), but the BS low ( $54 \%$ and $47 \%$ ). Specific local conditions, e.g. the Uusikaarlepyy plot is located on an acid sulphate soil, and the Punkaharju plot on a partly paludified site earlier subjected to slash-and-burn agriculture, undoubtedly explain this unusual relationship.

On the plots located in the northern boreal zone, the P and K concentrations in the organic layer were better explanatory variables of the vegetation pattern than the total N concentration. This was probably due to the fact that herb-rich heaths were absent from the plot network in the northern boreal zone, and the range of N concentrations was narrower than in the south. On the other hand, local geological conditions affect the high P and K concentrations on plot Nos. 7, 8 and 21 in the Oulanka area, which is characterized by phosphate-rich alkaline bedrock.

Although the broad-scale differentiation of the vegetation was clearly associated with the chemical composition of the organic layer, in some cases there was greater variation in the
structure of the vegetation than in the chemical properties of the organic layer between the site types. In these cases, factors other than the chemical composition of the organic layer, such as the moisture conditions, shading by the tree stand etc., play an important role in determining the composition of the understorey vegetation.

In addition to soil nutrients, climatic factors also had a considerable influence on the vegetation pattern. Solantie (2005) suggested that the effective temperature sum, duration of the vegetation period and the maximum soil frost penetration, are the most important factors determining the zonation of forest vegetation in Finland, and also presented evidence that these factors predict forest productivity. One typical feature in the boreal climatic gradient is that, although precipitation decreases from south to north, the soil moisture content is higher in the north owing to the lower rate of evapo-transpiration caused by the cooler conditions (Solantie 1987). Thus many moisture-demanding plant species that grow on peatlands in the south (e.g. Vaccinium uliginosum) have found suitable habitats also on heaths in the north. In general, fine-textured soils retain moisture and are relatively fertile, whereas coarse-textured soils are highly permeable to water, and hence drier, and have much lower nutrient concentrations (Urvas and Erviö 1974, Sepponen et al. 1979). In our study, however, the main vegetation gradients did not correlate well with the proportion of fine particles in the soil.

Tree species was a primary attribute for the vegetation structure in our material, because the more fertile plots were dominated by Norway spruce and the less fertile ones by Scots pine. The prevailing tree species has regenerated naturally on the majority of the plots, and thinnings have directed the succession towards single treespecies stands on the managed plots. Tree species affects the vegetation through light transmittance, chemical composition of the throughfall, soil nutrient concentrations and properties of the litter (Økland and Eilertsen 1993, Barbier et al. 2008). In general, Norway spruce shades the understorey more than Scots pine. Furthermore, the chemodiversity of trees, e.g. in needle monoterpene levels, may affect the occurrence of specific plant species in the local environment (Iason et al. 2005).

The $H_{100}$ values varied within a very wide range, indicating considerable differences in production potential between the stands. The $H_{100}$ values of different tree species growing on the same site are seldom fully comparable. In Finnish conditions, however, Norway spruce and Scots pine have relatively similar site indices under similar conditions (Vuokila and Väliaho 1980). The $H_{100}$ value of the stands correlated positively with the N concentration in the organic layer for both species, but the correlation was higher for Norway spruce. Tamminen (1993) reported a corresponding relationship in a study that included over 1200 stands in southern Finland. However, other factors can also explain the variation in the $H_{100}$ index (Tamminen 1993); in our data the considerable variation in $H_{100}$ on the northern, nutrient-poor sites was obviously related to the cold climate. According to Nieppola (1993), the presence of individual plant species, and especially the number of herbs, is a good indicator of the site index in Scots pine stands in southern Finland. In contrast, the species cover was poorly related to site index. The same conclusion can be drawn from both the Scots pine and Norway spruce plots and the extensive latitudinal gradient in our material.

In this paper, we presented quantitative evidence that the most important factors behind the extensive "fertility gradient" of boreal forest vegetation, were the total N , exchangeable Ca and Mg concentrations, and the BS and pH of the organic layer. We conclude, that our results support the hypothesis that the site index $H_{100}$ correlates with the N concentration of the organic layer, although the correlation was weaker for Scots pine than for Norway spruce and not so clear in the north. Furthermore, the plant species composition, especially the number of herb and grass species, well indicated the site index, despite the small size of the material. The description of the vegetation continuum along an extensive climatic gradient using multivariate methods will also enable us to make hypotheses e.g. on how climate change or eutrophication may alter the understorey vegetation in the future.

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