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12	Lethal trap created by adaptive evolutionary response to exotic resource
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24 Global transport of organisms by humans provides novel resources to wild species that often respond maladaptively. Native herbivorous insects have been killed feeding on toxic exotics, 25 which acted as "ecological traps<sup>1-4</sup>." We document a **novel** trap stemming from the opposite 26 27 effect, high fitness on an exotic resource despite lack of adaptation to it. Plantago lanceolata 28 was introduced to Western North America by cattle-ranching. Feeding on this exotic plant 29 released a large, isolated population of the native butterfly *Euphydrvas editha* from a 30 longstanding tradeoff between maternal fecundity and offspring mortality. Because of this 31 release, and despite reduced insect developmental rate on the exotic, *Plantago* immediately supported higher larval survival than the insects' traditional host. Collinsia parviflora<sup>5</sup>. Prior 32 work in the 1980's documented evolving preference for *Plantago* by ovipositing adults<sup>6</sup>. We 33 34 predicted that, if this trend continued, the insects could endanger themselves, since availability of *Plantago* to butterflies is controlled by humans, and humans change 35 management faster than butterflies evolve<sup>6</sup>. Here we report fulfillment of this prediction. The 36 37 butterflies abandoned Collinsia and evolved total dependence on Plantago. The trap was set. 38 In 2005 humans withdrew their cattle, springing the trap. Grasses grew around the *Plantago*, 39 cooling the thermophilic insects, which went extinct. This extinction could have been 40 prevented if the population had retained partial use of *Collinsia*, which occupied drier microhabitats unaffected by cattle removal. The flush of grasses abated quickly, rendering the 41 42 meadow once again suitable for *Euphydryas* feeding on either host, but no butterflies were 43 observed from 2008-2012. In 2013-4 the site was naturally recolonized by Euphydryas 44 feeding exclusively on *Collinsia*, returning the system to its starting-point and setting the 45 stage for a repeat of the anthropogenic evolutionary cycle. 46

47 Subject terms: *Euphydryas*, eco-evolutionary dynamics, anthropogenic evolution,
48 conservation, exotic host, ecological trap, insect diet, host shift, oviposition preference,
49 evolutionary trap.

50

Main The late Gary Polis described *Homo sapiens* as a "ubiquitous keystone pest<sup>7</sup>." And so we are<sup>8</sup>. Although adaptation to human activities allows some wild species to coexist with us<sup>9</sup>, many fail to adapt to human land management and suffer in consequence. An example is the setting by humans of "ecological traps<sup>10,11,"</sup> defined as follows: "in an environment altered suddenly by human activities, an organism makes maladaptive habitat choices based on formerly reliable environmental cues, despite availability of higher quality habitat<sup>11</sup>."

Most ecological traps result from preference by wild organisms for novel resources that are unsuitable or toxic. Australian monitor lizards suffered population crashes after feeding on toxic exotic cane toads<sup>12</sup>. Bees maladaptively preferred crops grown from seed treated with neonicotinoid insecticides<sup>13</sup>. Several examples involved insect herbivores feeding on exotic hosts that reduced insect fitness<sup>1-4</sup> but fell short of causing population extinctions, because traditional hosts were still used alongside the exotics.

65

In the "ecological trap" scenario the novel resource is accepted as food, but is initially detrimental<sup>11</sup>, with the expectation that evolution should lead either to behavioural avoidance of the resource or physiological ability to use it<sup>3</sup>. Here, we describe a different type of trap formed by a novel resource that immediately supported such high fitness that a butterfly population evolved complete dependence on it, causing local extinction when humans withdrew our apparent gift. This paradoxical phenomenon is previously undescribed and doesn't fit published definitions, either of "ecological trap" or "evolutionary trap<sup>1,11</sup>."

73

74 Our study insects, the thermophilic<sup>14</sup> Edith's checkerspot butterflies, *Euphydryas editha*,

formed a sedentary, isolated population around the margins of a spring-fed meadow,

76 Schneider's Meadow, at 1800m elevation in Carson City, Nevada. They have one generation

per year. Their novel host was the exotic perennial, *Plantago lanceolata*, and their traditional

78 host the ephemeral native annual *Collinsia parviflora* (Extended data Fig. 1).

79

The geographically closest *E. editha* populations of the same ecotype as Schneider, but where *Plantago* had not arrived, used *Collinsia* as their sole host<sup>5</sup>. When neonate larvae from one of these "ancestral" populations were transplanted to Schneider, their survival on *Plantago* was identical to that of the local Schneider insects<sup>5</sup>. The "ancestrals" were ready to use *Plantago* from the moment of its introduction. It is not surprising, then, that this exotic has been colonized twice by other North American *Euphydryas*<sup>15,16</sup>. Oregon *E. editha taylori* are now dependent on the exotic, though it is not clear if this is due to evolution of the butterfly, since
the original host(s) have disappeared<sup>16</sup>.

88

89 During the 1980's, survival of E. editha at Schneider was consistently higher on the exotic than on the traditional host (Table 1), despite larval growth being c.18% slower on *Plantago<sup>5</sup>*. 90 Where the principal host is ephemeral, as is Collinsia, female E. editha face a trade-off 91 92 between maternal fecundity and offspring survival. Prolonging larval development can 93 increase fecundity, but the resulting delay in adult emergence augments the risk of offspring mortality from host senescence<sup>17</sup>. The evolutionary response to this trade-off has been to 94 delay emergence to the point where many offspring routinely starve from phenological 95 asynchrony with their hosts<sup>17</sup>. The stage is set for the time constraint to be released and 96 97 fitness increased by host-switching to the longer-lived *Plantago*, despite slower larval growth 98 on it. Indeed, the majority of larval mortality observed on Collinsia was from host 99 senescence, while *Plantago* did not senesce during the seasons of larval activity. 100 No adult females sampled from "ancestral" populations preferred *Plantago* over *Collinsia* for 101 102 oviposition, but around 20% accepted both hosts equally<sup>5</sup>. In contrast, by 1982 *Plantago* was already preferred for oviposition by a minority (c.7%) of preference-tested adults at 103 104 Schneider<sup>6</sup>. Evolution of *Plantago* preference had begun. 105 106 Given natural selection for oviposition on *Plantago*, and given that oviposition preferences at 107 Schneider were both heritable (estimated heritability 0.9) and correlated with offspring performance<sup>18</sup>, we expected to see rapid evolution of preference. We did: the proportion of 108 insects preferring *Plantago* increased to around 50% by 1990<sup>6</sup>. This change was heritable. 109 110 Laboratory-raised, Collinsia-fed offspring of field-caught 1990 butterflies were significantly more *Plantago*-preferring than similarly-raised offspring of field-caught 1983 insects<sup>6</sup>. 111 112 113 Here, we report that the bout of anthropogenic evolution continued until monophagy on 114 *Plantago* was achieved. In 2005 and 2007, all tested females preferred to oviposit on the 115 exotic (Fig. 1a) and in 2007 all larvae found in the field were on *Plantago* (Fig. 1b). The 116 insects had abandoned both their traditional host, Collinsia, and the minor host, Penstemon 117 rydbergii, that had been incorporated into their diet during the host shift (Extended data 118 Table 1, Extended data Fig. 2). 119

120 In 1993 we wrote that this episode of anthropogenic evolution was "foreshadowing a new

121 problem in Conservation Biology. By adapting genetically to human-induced changes, the

122 insects risk becoming dependent on continuation of the same practices. This is a serious risk,

123 because human cultural evolution can be even faster than the rapid genetic adaptation that the

124 insects can evidently achieve."<sup>6</sup>

125

126 This prediction was fulfilled. In late 2005, following the death of "Uncle Harry" Schneider,

127 the meadow was sold and cattle-grazing ceased. Grasses grew freely. By March 2007, 96%

128 of *Plantagos* had become embedded in grass (Table 2, Extended data Figs. 3,4a) and *E*.

129 editha larvae wandered among dense vegetation, no longer able to bask in sunlight on bare

130 ground adjacent to their hosts.

131

132 The conspicuous communal webs spun by gregarious young larvae render *E. editha* easy to

133 census (Extended data Fig. 4b) and population extinction feasible to establish. Searches for

134 larval webs, eggs and adults found no *E. editha* in 2008, 2009, 2010 or 2012 (Fig. 1c,

135 Extended data Table 1). The population was extinct.

136

By analogy with known cause-effect relationships involving other thermophilic butterflies, 137 we attribute this extinction to the flush of lush vegetation caused by cattle removal. Lushness 138 is associated with high rates of predation on butterfly larvae<sup>19</sup> and increased lushness caused 139 by abandoning of traditional management in Europe has caused ground-level cooling 140 resulting in butterfly declines and local extinctions<sup>20-22</sup>. The UK extinction of the Large Blue 141 butterfly was attributed to lushness-caused microclimatic cooling, following myxomatosis-142 related reduction of grazing by rabbits<sup>22</sup>. Restoration of grazing and re-warming of ground-143 level microclimate were essential for this butterfly's successful re-introduction<sup>22</sup>. 144 145

Particularly in sparse vegetation, sunshine creates thermal stratification with microclimate hotter close to the ground<sup>23</sup>, speeding insect development<sup>24</sup>. We used prior observations of natural ovipositon sites at Schneider to measure "eggspace" temperatures. Eggspaces on exposed *Plantagos* were augmented by 13.4°C above ambient, compared to 6.0°C on embedded *Plantagos*, leaving "eggs" on embedded plants >7°C cooler, on average, than those on exposed plants (Extended data Table 2). In light of the high proportion of plants embedded in 2007 (Table 2), and of the known adverse effects of host embedding<sup>16,19</sup> and microclimatic cooling<sup>20-22</sup> on other butterflies, it is not surprising that this cooling was
followed by extinction.

155

As anthropogenic nutrients were used up, the flush of grasses abated naturally in 2008, since
when *Plantagos* exposed to full sunlight have again been available (Table 2, Extended data
Fig. 5). In 2013-4 butterflies recolonized: an exhaustive search in 2014 revealed nine *E*. *editha* larval webs, all on the ancestral host, *Collinsia* (Fig.1b,c, Extended data Fig. 4b).
Recolonization had occurred from a population resembling the starting condition, prior to the
anthopogenic evolution.

162

163 The distance from Schneider's Meadow to the nearest known population of the subspecies, *E.e. monoensis*, at Simee Dimeh summit, is 37.7km. How far is this to a *Euphydryas*? 164 Harrison<sup>25</sup> measured colonizations of empty habitat patches by the Bay Checkerspot (E. 165 166 editha bayensis) and found that the greatest cumulative distance travelled in ten years was 167 4.5km. Given this sedentary nature of the butterfly and the physical isolation of the meadow, we did not consider the possibility of rapid recolonization. Our mindset in 2014 was simply 168 169 to reconfirm the extinction and we were astonished to find larvae. After the event, we discovered that the "Carter Springs" fire in September 2012 had positively impacted the 170 171 Simee Dimeh butterflies, dramatically extending size and lifespan of *Collinsia* (Extended data Fig. 6), resulting in a population boom of *E. editha* similar to previously-documented 172 response to fire<sup>26</sup> and providing a plausible source of *Collinsia*-feeding *E. editha* despite the 173 174 distance.

175

176 Could the original population have survived if it had retained its traditional diet of *Collinsia* 

177 alongside *Plantago*, as it did from 1982 to 2002? *Collinsia* was both most abundant and most

used by the insects in dry sagebrush around the meadow edge (Extended data Figs. 1,2,3),

179 where removal of cattle did not result in embedding of *Collinsias*, even at peak lushness in

180 2007 (Table 2, Extended data Fig. 3). If the butterflies had adopted the exotic less

181 completely, they would likely have survived the change of land use. Conversely, if they had

remained monophagous on *Collinsia* they may not have survived the bottleneck in 1988-9<sup>6</sup>,

183 when, after record-breaking frost without insulating snow (-25°C at Minden on January 1,

184 1988), the population was spatially restricted to a small, sheltered, south-facing area from

185 which *Collinsia* was coincidentally absent (Fig. 1b,c; discussion with extended data Fig. 2).

187 Evolution of *E. editha* at Schneider illustrates the process by which, long ago, European grassland butterflies evolved widespread dependence on human haymaking and grazing, 188 rendering themselves vulnerable to abandonment of traditional management techniques<sup>20-22</sup>. 189 190 However, E. editha as a species is not threatened by the trap we document. Its ecotypic 191 variation and rapid evolution augur well for resilience to environmental fluctuations, whether natural or anthropogenic<sup>23</sup>. In contrast, substantial perturbation occurred at the subspecies 192 level, since *E. e. monoensis* is currently restricted to four known sites: two isolated 193 194 populations and two metapopulations distributed along 235km of the eastern Sierra Nevada 195 from McGee Creek at latitude 37.29 to Schneider at latitude 39.11. 196 197 Unless the Schneider population sent out successful propagules, which we judge unlikely, the lineage we observed from 1982-2007 is extinct. At the population level, the changes we 198 199 observed exemplify dramatic, oscillating anthropogenic evolution of a species not directly 200 targeted by humans. This example of small-scale oscillating diet evolution oddly mimics 201 repeated recolonizations of abandoned hosts detected across millions of years by phylogenetic analysis of the butterfly family Nymphalidae, to which E. editha belongs<sup>27</sup>. 202 203 Attempts at "evolutionary rescue" of wild species<sup>28,29</sup> may be compromised when 204 anthropogenic traps such as that documented here remain undetected. These traps may be 205 cryptic to humans, but understanding them may become increasingly important to species 206

207 conservation in the Anthropocene.

209	Data availability. Raw data are included in the Figure, Tables and Extended data.
210	
211 212	References
213	1. Schlaepfer, M. A., Sherman, P. W., Blossey B. & Rynge M. C. Introduced species as
214	evolutionary traps. Ecol. Lett. 8, 241-246 (2005).
215	
216	2. Nakajima, M., Boggs, C. L., Bailey, S., Reithel, J. & Paape, T. Fitness costs of butterfly
217	oviposition on a lethal non-native plant in a mixed native and non-native plant community.
218	<i>Oecologia</i> <b>172,</b> 823-832 (2013).
219	
220	3. Morton, T. A. L., Thorn, A., Reed, J. M., van Driesche R. G., Casagrande R. A. & Chew,
221	F. S. Modeling the decline and potential recovery of a native butterfly following serial
222	invasions by exotic species. Biological Invasions 17, 1683-1695 (2015).
223	
224	4. Yoon, S. & Read, Q. Consequences of exotic host use: impacts on Lepidoptera and a test
225	of the ecological trap hypothesis. Oecologia 181, 985-996 (2016).
226	
227	5. Thomas, C. D., Ng, D., Singer, M. C., Mallet J. L. B., Parmesan, C. & Billington, H. L.
228	Incorporation of a European weed into the diet of a North American herbivore. <i>Evolution</i> <b>41</b> ,
229	892-901 (1987).
230	
231	6. Singer, M. C., Thomas, C. D. & Parmesan, C. Rapid human-induced evolution of insect-
232	host associations. Nature 366, 681-683 (1993).
233	
234	7. Polis, G. A. Address to the Ecological Society of America (1999).
235	
236	8. Gotanda K. M., Hendry, A. P. & Svensson, E. I. Human influences on evolution, and the
237	ecological and societal consequences. Philos Trans Roy Soc B 372, 20160028 (2017).
238	
239	9. Thomas, C. D. Inheritors of the Earth: How Nature Is Thriving in an Age of Extinction.
240	Penguin/Allen Lane Books (2017).
241	
242	

243	10. Robertson, B. A., Rehage, J. S. & Sih, A. Ecological novelty and the emergence of
244	evolutionary traps. Trends Ecol Evol 28, 552-559 (2013).
245	
246	11. Schlaepfer, M. A., Runge, M. C. & Sherman, P. W. Ecological and evolutionary traps.
247	Trends Ecol Evol 17, 474-480 (2002).
248	
249	12 Jolly, C. J., Shine R., & Greenlees, M. J. The impacts of a toxic invasive prey species
250	(The cane toad, <i>Rhinella marina</i> ) on a vulnerable predator (the lace monitor, <i>Varanus varius</i> )
251	Biological Invasions 18, 1499-1509 (2016).
252	
253	13. Kessler S. C. et al. Bees prefer foods containing neonicotinoid pesticides. Nature 521, 74-
254	76 (2015).
255	
256	14. Weiss, S. B., Murphy, D. D., & White, R. R. Sun, slope, and butterflies: topographic
257	determinants of habitat quality for Euphydryas editha. Ecology 69, 1486-1496 (1988).
258	
259	15. Brown, L. M., Breed, G. A., Severns P. M. & Crone, E. E. Losing a battle but winning
260	the war: moving past preference-performance to understand native herbivore-novel hostplant
261	interactions. Oecologia 183, 441-453 (2017).
262	
263	16. Severns, P. M. & Warren, A. D. Saving an imperiled butterfly, Euphydryas editha taylori
264	(Taylor's checkerspot), by selectively conserving and eliminating exotic plants. Animal
265	Conserv 11, 476–483 (2008).
266	
267	17. Singer, M. C. & Parmesan, C. Phenological asynchrony between herbivorous insects and
268	their hosts: signal of climate change or pre-existing adaptive strategy? Philos Trans Roy Soc
269	<i>B</i> <b>365</b> , 3161-3176 (2010).
270	
271	18. Singer, M. C., Ng, D. & Thomas, C. D. Heritability of oviposition preference and its
272	relationship to offspring performance within a single insect population. Evolution 42, 977-
273	985 (1988).
274	
275	19. Wiklund, C. & Friberg, M. Enemy-free space and habitat-specific host specialization in a
276	butterfly. Oecologia 157, 287-294 (2008).

278	20. WallisDeVries, M. F. & van Swaay, C. Global warming and excess nitrogen may induce
279	butterfly decline by microclimatic cooling. Glob Chang Biol 12, 1620–1626 (2006).
280	
281	21. Warren, M. S. & Bourn N. A. D. Ten challenges for 2010 and beyond to conserve
282	Lepidoptera in Europe. J Insect Conserv 15, 321-326 (2010).
283	
284	22. Thomas, J. A., Simcox, D. J. & Clarke, R. T. Successful conservation of a threatened
285	Maculinea butterfly. Science 325, 80-83 (2009).
286	
287	23. Bennett, N. L., Severns, P. M., Parmesan, C. & Singer, M. C. Geographic mosaics of
288	phenology, host preference, adult size and microhabitat choice predict butterfly resilience to
289	climate warming. <i>Oikos</i> <b>124</b> , 41-53 (2015).
290	
291	24. McBride, C. S. & Singer, M. C. Field studies reveal strong postmating isolation between
292	ecologically divergent butterfly populations. <i>PLoS Biology</i> <b>8</b> , Issue 10 Article Number:
293	e1000529 DOI: 10.1371/journal.pbio.1000529 (2010).
294	
295	25. Harrison, S. Long-distance dispersal and colonization in the Bay Checkerspot butterfly,
296	Euphydryas editha bayensis. Ecology 70, 1236-1243 (1989).
297	
298	26. Thomas, C. D., Singer, M. C., & Boughton D. A. Catastrophic extinction of population
299	sources in a butterfly metapopulation. Amer Natur 148, 957-975 (1996).
300	
301	27. Nylin, S., Slove, J. & Janz, N. Host plant utilization, host range oscillations and
302	diversification in Nymphalid butterflies. Evolution 68, 105-124 (2014).
303	
304	28. Gonzalez, A., Ronce, O., Ferriere, R. & Hochberg, M. E. Evolutionary rescue: an
305	emerging focus at the intersection between ecology and evolution. <i>Philos Trans Roy Soc B</i>
306	368, 20120404 (2012)
307	20 Comell S. D. et al. Angleine conduction on high contract address shall all shallowers. Science
3U8 200	29. Carron, S. P. et al. Applying evolutionary biology to address global challenges. Science <b>346</b> , 212+ DOI: 10.1126/science.1245002 (2014)
210	<b>340</b> , 515 $\pm$ DOI. 10.1120/SCIENCE.1243995 (2014).
510	

311	
312	30. Singer, M. C. Butterfly-host relationships: host quality, adult choice and larval success.
313	Pp 81-88 in The Biology of Butterflies, Symp. Roy. Ent. Soc. 13, (1984).
314	
315	31. Newcombe, R. G. Two-sided confidence intervals for the single proportion: a
316	comparison of seven methods. Statistics in Medicine 17, 857-872 (1998).
317	
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331	Author Contributions Both authors participated in field censuses and writing. MCS
332	performed oviposition preference tests and statistical analyses.
333	
334	Competing Financial interests The authors declare absence of financial interests, whether
335	competing or not.

### 336 Figure 1. Changes of *E. editha* oviposition preference (a), larval diet (b) and population

337 **density (c).** The two preference graphs (a) are not mirror-images because insects without

- 338 preference are not shown. Changes in strength of preference are shown in Extended data Fig.
- 339 7. (b) absence of larvae on *Collinsia* in 1988-89 caused by temporary spatial restriction
- 340 during and after bottleneck; in those two years larvae were restricted to *Plantago* and
- 341 *Penstemon* (extended data Table 1 and extended data Fig. 2). Numbers within graphs
- 342 represent biologically independent sample sizes; individual adult insects (a) or larval groups
- 343 (b,c). Error bars = 95% c.i. calculated according to Newcombe<sup>31</sup>, with continuity correction;
- 344 source data in Extended data Table 3. No error bars for larval diet in years of inadequate
- 345 sampling (1982, 1983, 2015) or total census (1988, 1989, 2014).

## 346 Table 1. Survival of *E. editha* placed in the field on *Collinsia* and *Plantago* at

347 Schneider's Meadow in the 1980's. Proportions of groups surviving analyzed as 2x2

Year	Life Stages	Group	Group	Individual	Individual	Statistical
And reference	measured	Survival	Survival	survival	survival	significance
		Collinsia	Plantago	Collinsia	Plantago	(group survival)
1980	Oviposition	33%	70%	Not recorded	Not recorded	P = 0.044
(previously	to second	(n = 15)	(n = 20)			
unpublished)	instar					
1982 <sup>30</sup>	Oviposition	17%	55%	4%	24%	P = 0.000011
	to second	(n = 58)	(n = 86)	(n = 1810)	(n = 2764)	
	instar					
1985 <sup>5,18</sup>	Larval	62%	84%	19.9%	27.1%	P = 0.0085
	survival for	(n = 63)	(n = 62)	(n = 1260)	(n= 1240)	(combining data
	ten days					from both
	from hatch					references)
1986	Oviposition	29%	80%	Not recorded	Not recorded	P = 0.00028
(previously	to second	(n = 28)	(n = 25)			
unpublished)	instar					

348 contingency tables by Fisher's exact test, two-tailed.

	1			
Year	<i>Plantago</i> density,	Collinsia density,	Percent	Percent
	plants/m <sup>2</sup>	plants/m <sup>2</sup>	Plantago	Collinsia
			embedded	embedded
1984	2.3	103	<10%	<5%
	(50 in 244x 30cm quadrats	(2266 in 244x30cm quadrats		
2005	9.4	27.5	<10%	<5%
	(317 in 15 x 2.25 m <sup>2</sup> quadrats)	(330 in 12 x 1m <sup>2</sup> quadrats)		
2007	0.21	35.2	96.4%	10.5%
	(18 in 85 x 1m <sup>2</sup> quadrats)	(528 in 15 1m <sup>2</sup> quadrats)	(189/196)	(69/657)
2008	2.8	26.1	18.6%	13.8%
	(44 in 16 x 1m <sup>2</sup> quadrats)	(418 in 16 x1m <sup>2</sup> quadrats)	(43/231)	(77/559)
2009	2.5	45.7	11.8%	11.7%
	(48 in 19 x 1m <sup>2</sup> quadrats)	(869 in 19x 1m <sup>2</sup> quadrats)	(251/2115)	(114/975)
2014	3.4	52.4	8.5%	18.0%
	(173 in 50 x 1m <sup>2</sup> quadrats)	(472 in 9 x 1m <sup>2</sup> quadrats)	(28/328)	(150/833)

350	Table 2.	Estimated host densities and percentages of each host embedded and	exposed.

- 352 Methods
- 353

354 Statistics. All tests are two-tailed. Error bars on Figures, given only where justified by sampling techniques, are 95% c.i., calculated by vassarstats.net, using methodology of 355 Newcombe<sup>31</sup> for c.i of proportions, with continuity correction. Figure 1a does not explicitly 356 depict insects without preference but they can be deduced by subtracting the depicted 357 358 percentages from 100%. Otherwise no relevant data have been omitted from any experiment 359 or set of observations. Sample sizes were largely limited by feasibility. 360 361 Data availability. Raw data are included in the Figure, Tables and extended data. 362 Blind preference testing. Oviposition preference tests to estimate heritability<sup>18</sup> were 363

performed blind, in the sense that the tester did not know which insects were sibs or offspring of particular parents. The high heritability estimate from this blind testing (0.9) gave confidence that the tests were not subject to severe observer bias; repeated blind testing of the same butterfly by different observers has also given confidence. However, in the present MS the appropriate blind test would require the tester to not know which year it was, and we were not able to achieve this without compromising the quality of the data.

370

371 Figure 1a. Butterflies were captured in the field and their oviposition preferences tested by a 372 standardized technique, in which encounters are staged between the tested insect and each 373 plant in alternation. Plants were undisturbed in their natural habitats or freshly transplanted into pots in their own soil. Acceptance of plant taste was judged from full abdominal curling 374 and extrusion of the ovipositor for  $3 \sec^{32}$ . Acceptance and rejection were recorded at each 375 encounter, but oviposition was not allowed<sup>32</sup>. Videos showing acceptance in such staged 376 377 encounters are in reference 24. During each test the range of plants that would be accepted, if 378 encountered, expands over time with increasing motivation to oviposit. Therefore, acceptance 379 of plant A followed by rejection of plant B is recorded as preference for A over B. Testing of 380 assumptions underlying this technique described in reference 32. Because insects without 381 preference are not shown in the Figure, percentages do not sum to 100% except in 2005 & 382 2007, when preference for *Plantago* was unanimous among tested butterflies. Raw data are in 383 Extended data Table 3. A more detailed comparison between early and late periods, showing 384 strength as well as direction of preferences, is given in Extended data Fig. 7. The assumption

that these insects' preferences are not influenced by prior experience, either as larvae or as
 adults, is supported by prior observation and experiment<sup>18,32,33</sup>.

387

388 Figure 1b. Percentages of egg clutches/larval groups found on *Plantago*. Raw data in 389 Extended data Tables 1 and 3. Low sample sizes in 1988 and 1989 were total counts, 390 reflecting bottleneck. High proportion of larvae on *Plantago* in 1989 reflects lag in 391 recolonization of Collinsia after spatial restriction in bottleneck, shown in Extended data Fig. 392 2. In 2002-2007 we searched a larger area for larvae on Collinsia than for larvae on Plantago. 393 For those years the Figure, showing the proportion of larvae found on the two hosts, 394 overestimates the overall proportion on Collinsia; the areas searched are in Extended data 395 Table 1. To be conservative we indicated data for 2002 and 2005 as "published" although neither sample sizes nor confidence limits were given previously<sup>33</sup>. The graph shows that in 396 397 2015 we found a single group of hatching eggs on Collinsia. However, we performed no 398 census in 2015, our visit was too early.

399

Figure 1c. Estimates of density of larval webs, on all hosts combined, per 10,000m<sup>2</sup>. Raw 400 401 data in Extended data Table 1. In each census, individual host plants of all species were 402 searched for eggs, larvae, larval webs and typical damage. Very different scales of census 403 were conducted in different years. For small patches, every individual host plant could be 404 searched, for larger patches, stratified line transects were used. Confidence limits are not 405 given, but the log scale of the Y-axis makes clear the dramatic scale of population changes; for example, the raw data (Extended data Table 1) include the finding of 34 groups in only 406  $70m^2$  in 1982 and 4 groups in 20,000m<sup>2</sup> in 1988. No density estimate is given for 2007, 407 because, although 38 groups were found (Fig. 1b), at the time of the last census some were 408 409 still eggs, so harder to find than larvae. Sample calculation for 2002, (see methods for 410 Extended data Table 1): estimated number of webs on *Plantago* =  $67 \times 4,000/170 = 1576$ . Total number on *Collinsia* = 3 (all were counted). Estimated total number webs in entire area 411 of  $20,000m^2 = 1579$ ; density per  $10,000m^2 = 789$ . 412 413

Table 1. Effects of oviposition host on fitness: eggs were placed out on randomly-chosen
hosts in the field, by manipulating butterflies to lay (videos in ref. 24). Gregarious neonate
larvae were placed out with a sable brush in groups of 20-35. After ten days of larval life, just
before diapause, each group was gathered in. It is in pre-diapause life that the principal

418 effects of oviposition on fitness are manifest, since post-diapause larvae are mobile and can419 switch between host species.

420

421 Table 2. Host densities estimated from quadrats along line transects placed randomly within 422 the strata of the ecotone where each plant was concentrated (Extended data Fig. 1). Early 423 rough estimates of percent embedded from photographs, memory and anecdotal observation, 424 since embeddedness was uncommon and its value as a trait was not anticipated prior to the 425 sudden embedding of Plantago in 2007. A plant was classed as "embedded" if surrounded for 426 > 50% of circumference by vegetation taller than itself. In most cases the differences between 427 "embedded" and "exposed" plants were striking (Extended data Fig. 3); in the March census 428 in 2007 some embedded plants were hidden and not found, pressed under thatch from winter 429 snow, inaccessible to post-diapause larvae feeding in March but reappearing and, if no longer 430 completely embedded, accessible to ovipositing butterflies in May. Extended data Fig. 4a has 431 a photo of eggs naturally laid on such a reappearing *Plantago*, emerging from winter thatch, 432 in May, 2007.

433

434 Extended data Table 1. Table contains counts over wider areas including lower host densities than the "core areas" censused in Table 2. Total area where suitable Plantago might 435 be found was around 4,000m<sup>2</sup>; total area that might contain suitable *Collinsia* was larger, 436 maximally about 17,000m<sup>2</sup> (1982 map in Extended data Fig. 2) but less in dry years. About 437  $1,000 \text{ m}^2$  overlapped between the two distributions. Wider areas were searched in 1988 and 438 439 1989 to check whether we had missed part of the population in prior work. We had not, so 440 since we found no habitat in wider search, subsequent searches were restricted to meadow margins and adjacent sagebrush, approximately 20,000m<sup>2</sup> (Extended data Figs.1 & 2). 441 442

In 2002, 2005 and 2007 the entire area where larvae might have been found on *Collinsia* was searched, but *Plantago* areas were not searched in entirety, merely sufficiently to get an estimate of plant density and occupancy by the butterflies. After the extinction in 2007-8, the entire habitat was searched in each census. In most years, more plants were searched for *E. editha* than were included in censuses to estimate plant density reported in Table 2.

The Table shows data from areas censused in which plants with and without larvae werecounted. Where maps in Extended data Fig. 2 show more insects than the Table, as in 1982,

the insect distribution in the map is derived from rapid assessment in which insects wereobserved and counted but plants were not.

453

Extended data Table 2. Eggspace temperatures were measured with a fine thermocouple
(MT-29/1B insect probe, type T, copper-constantan). After measuring each exposed *Plantago*we measured one or two adjacent (within 1m) embedded plants growing in the same position
within the ecotone; we took care that embedded plants were not systematically growing in
more humid microsites. When two embedded plants were measured, we used the mean value
of the two for analysis.

460

461 We found three errors in our prior publications: (1) number of groups found on *Collinsia* in

462 1990 is here corrected to 18 from  $6^{6}$ ; (2) number of butterflies preference- tested in 1986 is

463 corrected from 31 to  $36^{5,6}$ ; (3) misleading early reference to population as feeding on

464 *Plantago* in 1969<sup>34</sup> stemmed from initial visit in June, after *Collinsia* had senesced and

465 disappeared. Faster-growing larvae on *Collinsia* had entered diapause and were not found,

466 while some still remained feeding on *Plantago*. Next visit in 1971 made clear that, although

467 *Plantago* was already being used, *Collinsia* was the principal host. No censuses were468 performed in 1969-71.

469

470 References

471

472 32. Singer, M.C., Vasco, D. A., Parmesan, C., Thomas, C. D. & Ng, D. Distinguishing

473 between preference and motivation in food choice: an example from insect oviposition.

474 *Animal Behaviour* **44**, 463-471 (1992).

475

33. Singer, M. C., Wee, B., Hawkins, S. & Butcher, M. Rapid natural and anthropogenic diet
evolution: three examples from checkerspot butterflies. Pp. 311–324 *in* K. J. Tilmon, ed. *The*

478 evolutionary ecology of herbivorous insects: speciation, specialization and radiation. Univ.

479 of California Press, Berkeley, CA (2008).

480

481 34. Singer M. C. Evolution of food-plant preference in the butterfly *Euphydryas editha*.

482 *Evolution* **25**, 383-389 (1971).

483 Extended data Figure 1. Habitat and host distributions at Schneider's Meadow. a:

- 484 Butterfly habitat is a single, isolated, spring-fed wet meadow in the centre of the picture,
- 485 surrounded by non-habitat for the butterflies: dry sagebrush scrub and coniferous forest. b:
- 486 Distribution in the meadow-edge ecotone of the principal hosts, Plantago lanceolata and
- 487 *Collinsia parviflora*, plus the minor host *Penstemon rydbergii*. **c:** Typical difference in
- 488 phenology between Plantago and Collinsia in May 2014. In the foreground are red, senescent
- 489 Collinsia plants that are edible to the insects but will die within a few days; behind them is a
- 490 single green, budding *Plantago* that will remain edible until after all *E. editha* larvae have
- 491 entered diapause. d: Hatching egg clutch on *C.parviflora* cotyledon in hot, dry microhabitat. 492
- 493 Extended data Figure 2. Changes in distribution of early stages of *E. editha* (eggs or

#### 494 larvae) from 1982-2007.

Data added by hand to GoogleMaps image. Most stars represent several groups. For example, 496 in 1989, 23 groups were found on *Plantago* and one on *Penstemon*. The restricted 497 distribution in that year followed a bottleneck in 1988 after record-breaking cold in January, 498 without the usual insulating snow cover. Schneider's Meadow is at 1700m elevation. Nearby 499 towns at lower elevations recorded -25C on January 1 1988 (Minden, 1444m elevation) and -500 20C on January 18 (Carson City, 1424m). Note recolonization of Collinsia as the insects 501 expanded back into the distribution of Collinsia in 1990 and 1993. Larval groups recorded in

- 502 1988 and 1989 were clustered around an attractive nectar source (Wyethia sp.); it is possible
- 503 that adults attracted to this nectar in 1988 had survived as larvae on *Collinsia* in 1987-8, then
- 504 laid eggs in 1988 on *Plantago* adjacent to nectar. This possibility prevents us from making a
- definite conclusion that the population would have become extinct if eggs in 1987, prior to 505
- 506 the bottleneck, had been laid only on *Collinsia*. Data for 2005 exist, and closely resemble
- 507 those for 2002.
- 508

- 509 Extended data Figure 3. Effects of cessation of grazing: Plantagos embedded, Collinsias
- 510 unaffected. Data in Table 2. Panel a shows a *Plantago* at Schneider in 1984, exposed to full
- 511 sunlight and physically acceptable to ovipositing *E. editha*. Panel **b** shows the meadow edge
- 512 in May 2007, after cattle removal. In the foreground is *Plantago* habitat with thick grasses; in
- 513 the background is *Collinsia* habitat not grassed-in, with barren spaces between the sagebrush.
- 514 Panel c shows Collinsia parviflora in May 2007, unaffec ted by the embedding that
- 515 simultaneously affected the Plantagos shown in panels d and e. Embedding in grasses not

- 516 only cooled the *Plantagos* (Extended data Table 2) but rendered them hard to find, both by
- 517 butterflies seeking oviposition sites and by larvae seeking food.
- 518
- 519 Extended data Figure 4. a. Natural egg clutch laid in May 2007 on *Plantago*
- 520 The plant is pushing through winter thatch, and would have been unlikely to be acceptable to
- 521 ovipositing butterflies prior to cattle removal, when plants like the one in Extended data Fig
- 522 3a were available. **b. Communal web spun after recolonization**. Second-instar larvae on
- 523 *Collinsia* at Schneider in May 2014. This is a single group of larvae probably stemming from
- a single oviposition event; there were 9 such groups, all on *Collinsia*. Oddly, this group is not
- 525 on the most exposed *Collinsia* available.
- 526
- 527 Extended data Figure 5. The return of mostly-exposed *Plantagos* after anthropogenic
- 528 lushness abated. Photos taken in 2014, but Table 2 shows that they could have been taken in529 2008 or subsequently.
- 530
- 531 Extended data Figure 6. Effect of fire on size and longevity of *Collinsia parviflora*.
- 532 A single fire-enhanced Collinsia parviflora individual at McGee Creek (East of Bishop,
- 533 California) is still blooming. There is a small web of *E. editha* larvae at its base from a
- 534 naturally-laid egg clutch. The fifteen senescent individual *C. parviflora* lying on the ground
- are a haphazard sample gathered from unburned microsites within 2m of the enhancedindividual.
- 537

## 538 Extended data Figure 7. Strength and direction of oviposition preferences of butterflies 539 sampled at Schneider in 1983 and (2005+2007).

The number over each bar is the sample size of biologically independent samples: individual butterflies captured in the field. The "discrimination phase" is the length of time for which the insect would search, during which it would consistently accept the preferred host and consistently reject the second-ranked host. At the end of this phase, if it does not succeed in ovipositing, the insect enters an "acceptance phase," after reaching the level of oviposition motivation at which either host would be accepted, whichever were next encountered.

Insects in the blue 1-4 column on the left of the Figure would search for 1-4 hours during
which only *Collinsia* would be accepted. If they failed to find *Collinsia* within 4 hours, they
would subsequently accept either host, until actual oviposition occurred. Green central bar

- shows butterflies without preference. Sample size for 2005-7 is smaller than in Fig.1a
- because we include on Fig.1a (and here omit) 5 butterflies for which we determined the
- 552 direction of preference, but not the strength.
- 553

Extended data Table 1. Census results: areas searched and numbers of egg clutches or larval webs found on each host. Data for 1982-1993 from reference 6. These are not the same data shown in Extended data Fig. 2; see Methods. Extension of survey to 50,000m<sup>2</sup> in 1989 did not reveal additional habitat, so density in Fig.1c for that year is calculated using the estimated maximum habitat area of 20,000m<sup>2</sup>, giving a higher estimate than reported in reference 6.

560

# 561 Extended data Table 2. Measurements of temperature excess over ambient air at 1m height at 562 three potential types of oviposition site at Schneider's Meadow: embedded *Plantago*, exposed

- 563 *Plantago* and exposed *Collinsia*.
- 564 Measures taken at eggspace height (2-4cm) between noon and 15:00 in May 2015. Each
- 565 measure of exposed *Collinsia* or *Plantago* came from a different, haphazardly chosen habitat
- 566 patch. On occasion we measured two embedded *Plantagos* in the same patch; when this
- 567 occurred we show the mean of the two values and used those means in analysis, treating the
- two plants as a single sample. Consequently, each data point in the Table represents a
- 569 biologically independent sample, and the independent sample sizes are 12, 18 and 18 as
- 570 indicated at the base of each cell.
- 571
- 572 Comparison between exposed *Plantago* (mean excess 13.4) and embedded *Plantago* (mean
- 573 excess 6.0) by two-sided t-test: t = 7.55, df = 28, p < 0.0001
- 574 Comparison between exposed *Collinsia* (mean excess 11.46) and exposed *Plantago* by two-
- 575 sided t-test: t = 1.77, df = 28, p = 0.09
- 576
- 577 Extended data table 3. Source data for means and confidence limits of adult preference 578 and larval diet, shown in Fig. 1a,b. Statistical test described in caption to Fig.1.
- 579