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Article

A Novel Multi-Charged Draw Solute That Removes Organic Arsenicals from Water in a Hybrid Membrane Process

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| 1 | A Novel Multi-Charged Draw Solute That Removes |
|--------|---|
| 2 | Organic Arsenicals from Water in a Hybrid |
| 3 | Membrane Process |
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| 10 | ABSTRACRT: The potential of forward osmosis for water treatment can only be maximized |
| 11 | with suitable draw solutes. Here a three-dimensional, multi-charge draw solute of decasodium |
| 12 | phytate (Na ₁₀ -phytate) is designed and synthesized for removing organic arsenicals from water |
| 13 | using a hybrid forward osmosis (FO) - membrane distillation (MD) process. Efficient water |
| 14 | recovery is achieved using Na ₁₀ -phytate as a draw solute with a water flux of 20.0 LMH and |
| 15 | negligible reverse solute diffusion when 1000 ppm organic arsenicals as the feed and operated |
| 16 | under ambient conditions with FO mode. At 50 °C, the novel draw solute increases water flux by |
| 17 | more than 30 % with water fluxes higher than 26.0 LMH on the FO side, drastically enhancing |

18 water recovery efficiency. By combining the FO and MD processes into a single hybrid process, 19 a 100% recovery of Na₁₀-phytate draw solute was achieved. Crucially, organic arsenicals or 20 Na₁₀-phytate draw solutes are both rejected 100% and not detected in the permeate of the hybrid 21 process. The complete rejection of both organic arsenicals and draw solutes using hybrid 22 membrane processes is unprecedented; creating a new application for membrane separations.

23 TOC Art



24 25

26 INTRODUCTION

Arsenic contamination of water resources arise from their natural geological presence in deep underground water sources or from anthropogenic activities such as mining, farming, and wood preservation.¹⁻³ Arsenic can exist as toxic (inorganic or organic) and nontoxic (organic content in seafood) forms.^{1,3} The toxicity of arsenic depends on valence state, solubility, physical state and purity, rates of absorption and elimination. In general, the toxicity of arsenicals is as follows: Inorganic trivalent > organic trivalent > inorganic pentavalent > organic pentavalent > elemental arsenic.⁴⁻⁶ The toxicity of various arsenic forms guides the design of current arsenic removal technologies. Traditionally, inorganic arsenic compounds were perceived to be more toxic and potent than organic arsenicals.^{4,7} Hence, conventional technologies such as adsorption, ion exchange, coagulation, and membrane separations (nanofiltration, reverse osmosis, and electrodialysis) were investigated for removing inorganic arsenicals.⁶⁻⁸

38 Despite the extensive applications of these technologies for arsenical removal, the following 39 problems are widely present. Large volumes of arsenic-rich sludge and waste can be generated using adsorption and coagulation,^{2,9-11} while membrane separation technologies consume large 40 amounts of energy.^{6,12-14} The removal of organic arsenic has become increasingly important as 41 methyl derivatives of arsenic that are omnipresent in agriculture (monomethylarsenic - MMA, 42 dimethylarsenic – DMA) present a carcinogenic risk.⁴ Moreover, inorganic arsenic compounds 43 can convert into methylated arsenicals under appropriate conditions.⁷ Clearly, there is a need to 44 45 develop a green and cost-effective bespoke technology for removing organic arsenicals.

46 Here we propose to use a hybrid forward osmosis (FO) – membrane distillation (MD) process 47 to remove organic arsenics (Figure 1). FO-MD is an economical and sustainable technology to treat wastewater,¹⁵ concentrate protein solutions,¹⁶ and desalination.¹⁷ This technology separates 48 49 water from dissolved solutes (contaminants or proteins) using the operating principle 50 underpinning FO processes while recovering the draw solute using MD. An osmotic pressure 51 difference is first generated by a concentration gradient using a "draw" solution with higher 52 concentration of dissolved solutes in relation to the feed solution. Water from the feed solution 53 permeates across the membrane, while the contaminants are retained in the FO feed. The FO

permeate now contains a dilute mixture of water and draw solute that requires MD to recover the draw solute while producing high purity water. Clearly, the crux of this hybrid process is the draw solutes.¹⁸ Suitable draw solutes can enhance FO performance, i.e. achieving high water recovery whilst avoiding secondary contamination at lower costs, and recovered. Conventional draw solutes such as NaCl,¹⁹ MgCl₂²⁰ and glucose saccharide²⁰ have been deployed for removing inorganic arsenics, but to minimal effect. This is attributed to either low water permeation flux, or/and severe reverse solute diffusion.



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Figure 1. Using Na₁₀-phytate draw solute to treat water that contains MMA and DMA via a FO process.
Subsequently, the draw solute is dehydrated and regenerated using an in-line MD process.

Excellent draw solutes that optimize FO processes are typically pH-neutral, easily ionized compounds that generate large osmotic pressures.^{21,22} These compounds produce a neutral solution that does not degrade the membrane during FO, whilst providing a large driving force for water transport. Draw solutes with a suitable spatial configuration and high ionization degree 69 are critical for producing a high water recovery and minimizing reverse solute diffusion across the membrane in FO.²² Guided by these requirements, we design and synthesize a novel phytate 70 71 (Na₁₀-phytate) and demonstrate the advantages of this new compound as a draw solute to remove 72 organic arsenicals from water. First, water fluxes of up to 20 LMH (FO mode) are achieved at 73 room temperature when Na₁₀-phytate is deployed in single FO processes. Second, we observed 74 negligible reverse solute diffusion during MMA or DMA removal even at 50 °C. Finally, we 75 show that MMA or DMA can be completely removed from water while the Na_{10} -phytate draw 76 solutes are fully recovered and regenerated.

77 **MATERIALS AND METHODS**

Syntheses of Na₁₀-Phytate. Na₁₀-phytate was synthesized through the neutralization reaction
between phytic acid and NaOH. The experimental details were given in the supporting
information (SI).

81 **Determination of the Sodium ion Numbers in Na**₁₀**-Phytate.** The number of sodium ions in 82 the synthesized sodium phytate was determined via acid-base titration. The detailed information 83 was provided in the SI.

84 **Characterization of Na₁₀-Phytate.** Elemental analyses, the size distribution of Na₁₀-phytate in 85 its aqueous solution, and thermogravimetric analysis (TGA) measurement were used to 86 determine the chemical composition of Na₁₀-phytate and its state in solution prior to the FO 87 applications. The experimental details were included in the SI.

88 **Determination of the Physicochemical Properties of the Na₁₀-Phytate Solution.** The 89 physicochemical properties of the Na₁₀-phytate solution were determined by analyzing their 90 relative viscosity and osmotic pressures. Details were provided in the SI.

FO Processes. The FO experiments were carried out through a bench-scale FO set-up as established elsewhere.²³ Commercial HTI flat sheet membranes and home-made thin-film composite (TFC) FO membranes fabricated on polyethersulfone (PES) hollow fibers (TFC-PES)²⁴ were both used in the FO experiments. The detailed experimental conditions were provided in the SI.

96 **Preparation and Analyses of MMAs and DMAs Solutions**. The MMAs and DMAs 97 solutions were prepared from CH_4AsNaO_3 and $(CH_3)_2AsNaO_2$, respectively. The pH values of 98 these solutions were maintained using HCl and NaOH. The experimental details of the 99 MMAs/DMAs solution preparation and analyses were included in the SI.

100 **Regeneration of Na₁₀-Phytate Solute.** An MD set-up with home-made PVDF hollow fiber 101 membranes²⁵ was employed to concentrate the diluted Na₁₀-phytate solution after FO. The 102 experimental details were disclosed in the SI.

103 RESULTS AND DISCUSSION

104 Synthesis and Characterization of Na₁₀-Phytate. Phytic acid, a hexaphosphate polyacid, was 105 neutralized with a known volume and molar concentration of sodium hydroxide (NaOH) to 106 produce a three-dimensional, multi-charge decasodium phytate (Na₁₀-phytate) neutral draw 107 solute. The leftover NaOH solution was titrated with hydrochloric acid (HCl); revealing a 1:10 ratio of phytic acid to sodium hydroxide i.e. there are 10 Na⁺ ions in the phytate. 108 109 Thermogravimetric analysis (TGA) was used to confirm the elemental composition of the 110 phytate developed here in this work. Thermal decomposition of Na₁₀-phytate occurred in two stages. The first stage corresponds to a release of water molecules between 122 - 155 °C; 111 112 accounting for a weight loss of 9.5 wt. % (Table S1). The second stage at 205 – 398 °C correlates 113 to the decomposition of 63.8 wt. % of dehydrated organic phosphate salt (Figure 2a). The residue

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- 114 (26.7 wt. %) comprises mainly sodium oxides. Weight losses associated with these two stages
- are in excellent comparison with theoretical calculations carrying crystal water molecules.



Figure 2. (a) TGA analysis of Na₁₀-phytate demonstrating its thermal stability up to 205 °C. (b) The osmotic pressures of the precursor compound – phytic acid, conventional draw solute – NaCl, and the novel draw solute developed in this work – Na₁₀-phytate. (c) An osmotic pressure comparison of Na₁₀-phytate with traditional draw solutes. (d) A relative viscosity comparison of Na₁₀-phytate with other draw solutes.

Important material characteristics that govern applications as FO draw solutes include osmotic pressure and relative viscosity. We observe that both osmotic pressure and relative viscosity of Na₁₀-phytate increase non-linearly with higher concentrations (Fig. 2b, Fig. S1). This can be ascribed to the relatively large configuration and chemical composition of Na₁₀-phytate. The

hydrophilic functional groups of Na₁₀-phytate can ionize into multi-charge anions and multiple 126 cations in aqueous solutions. Predicted by the Van't Hoff equation,²⁶ the dissociation of Na_{10} -127 128 phytate into multiple ionic species increases the number of solute particles in the aqueous 129 solution; generating an osmotic pressure (66 atm) which is 65 % larger than that of NaCl at 1.0 M, the conventional draw solute in FO (Figure 2b). At low concentrations, salts dissociate 130 completely into multiple ionic species; generating large increments in osmotic pressures.²⁷ This 131 132 is impeded at higher concentrations. However, the osmotic pressure generated by 1.0 M of Na₁₀-133 phytate is significantly higher than those of synthetic draw solutes such as magnetic nanoparticles,²⁸ polyelectrolytes,²⁹ hydrogels³⁰ and many others^{22,31} (Figure 2c). Meanwhile the 134 relative viscosity (n_r) of Na₁₀-phytate is at least 60 % lower than polyelectrolytes²⁹ (Figure 2d), 135 136 and is comparable to NaCl. Compounds with a reduced relative viscosity are preferred as draw 137 solutes for FO processes as less internal concentration polarization will be caused and water 138 molecules can easily diffuse across the membrane towards the permeate side. The good thermal 139 stability, high osmotic pressure, and low relative viscosity of Na_{10} -phytate are ideal for FO.

140 Na₁₀-phytate as a FO Draw Solute. Initial screening tests (Figure 3) indicated that FO is 141 optimized with a home-made thin film composite (TFC) – polyethersulfone (PES) hollow fiber membrane.²⁴ Regardless of membrane orientation (the active layer facing feed solution i.e. FO 142 143 mode or the active layer facing draw solution containing 0.5 M of Na₁₀-phytate - pressure 144 retarded osmosis (PRO) mode), the water flux of a TFC-PES membrane is 3 times higher than 145 that of a commercial flat sheet membrane from HTI. Coupled with an insignificant salt leakage 146 rate, the TFC-PES membrane in FO mode is ideal for FO water treatment. The drastically higher 147 water flux is ascribed to membrane structural configuration where the thin TFC selective layer supported on a porous PES substrate has lower water transfer resistance relative to the dense and 148

thick HTI membrane.²⁴ Meanwhile the spatial structure and chemical composition of Na₁₀phytate contribute to a negligible reverse solute diffusion. In Na₁₀-phytate, the evenly-distributed phosphates around an aromatic carbon ring develop a three-dimensional structure that form hydrogen-bond with water molecules in the aqueous solution (Figure S2a). This supramolecular structure inhibits reverse diffusion; minimizing leakage rates. The size distribution of the supramolecular network falls in a range much larger than that of an FO membrane pore size which accounts for the negligible leakage of Na₁₀-phytate (Figure S2b).





Figure 3. Effects of draw solute concentration, membrane and membrane orientation on FO performance:
(a) water flux of HTI membrane; (b) water flux of TFC-PES membrane; (c) salt leakage of HTI
membrane; (d) salt leakage of TFC-PES membrane. DI water as the feed solution.

161 We investigated the impact of Na_{10} -phytate as a draw solute to remove organic arsenicals in a 162 FO process using TFC-PES hollow fiber membranes at FO operation times between 20 to 120 163 mins, MMA or DMA concentrations ranging from 0 - 1000 ppm, pH between 3 to 11, and 164 temperatures from 25 to 60 °C. These parameters are key for creating an osmotic pressure 165 differential across the membrane to drive FO water treatment. Longer operation times (20 to 120 166 min) reduced water flux by 20 % with the DI water feed, possibly due to water transfer 167 concentration polarization and/or membrane fouling (Figure 4a). Water flux is reduced by 35 and 168 45 % with 1000 ppm of DMA or MMA, respectively. Organic arsenicals in the feed solution 169 enhance feed osmotic pressure; reducing the net driving force across the membrane, hence 170 further decreasing water flux. The higher concentration of anions in a MMA solution (relative to DMA solution) can dissociate into more ionic solute particles;⁹ generating a slightly higher 171 172 osmotic pressure. Thus, a lower water flux was observed with MMA present in the feed solution. The high anion concentration in MMA solutions creates a stronger charge repulsion with the 173 174 surface of the electronegative TFC-PES membrane; rejecting more than 99.5 % of MMA.



175

Figure 4. FO performances (water flux and organic As rejection) for removing MMA or DMA from
water using TFC-PES hollow fiber membranes at various operating conditions: (a) 20 – 120 mins of

operation time; (b) 0 – 1000 ppm MMA or DMA in deionized water; (c) pH 3 – 11; and (d) temperature
between 25 to 60 °C. Experimental conditions for each parameter are shown in the Supporting
Information.

181 Different MMA concentrations did not impact on both water flux and rejection rates (Figure 4b). 182 However, higher DMA content in the feed solution reduced the rejection rate of organic 183 arsenicals to less than 96 %. Neutral DMA solution consists 14 % neutral species, and 86 % monovalent anions.⁹ Higher DMA content will increase the number of neutral DMA; hence 184 185 reducing the rejection rate. The ratio of neutral species to anions in MMA/DMA solutions is also 186 sensitive to pH values (Figure 4c). As pH of a MMA feed solution changes from acidity (3) to alkalinity (11), the neutral species are converted into various anions;⁹ hence reducing water flux 187 188 due to a lower driving force attributed to enhanced osmotic feed pressures but increasing MMA 189 rejection rates. Since DMA solution consists more neutral DMA species than MMA solution, the 190 increase in pH from 3 to 11 will only slightly reduce water flux. Higher temperatures will also 191 enhance the osmotic pressure differentials that consequently increase water flux (Figure 4d). 192 Ideally, to remove 1000 ppm of MMA or DMA more effectively from water, the FO process 193 must be conducted under neutral conditions with a higher temperature. To fully optimize the 194 benefits of FO for removing organic arsenicals, the draw solutes must be recovered/regenerated.

195 **Regeneration of Na₁₀-phytate Draw Solute.** Membrane distillation (MD) is a proven 196 technology for regenerating draw solutes such as hydroacid complexes,³² polyelectrolytes,¹⁵ and 197 thermosensitive polymers¹⁷ from dilute FO permeate solutions. Different from FO processes 198 where draw solution concentration significantly impacts on water flux, temperature dominates 199 water transport in MD.³³ This is also observed here when aqueous solutions containing Na₁₀-200 phytate at different concentrations were deployed in a single MD process using home-made 201 polyvinylidene fluoride (PVDF) hollow fiber membranes.²⁵ The increase in water flux due to 202 increasing temperatures can be attributed to the exponential generation of higher water vapor 203 pressures. The impacts of both Na_{10} -phytate concentration and operation time on water flux were 204 negligible (Figure S3).

205 By combining the FO and MD processes into a single hybrid process, we report complete 206 recovery of Na₁₀-phytate draw solute whilst producing 100 % water (Figure 5). The water flux of 207 this hybrid process at 50 °C is higher than 26.0 LMH on the FO side; sufficient to handle large quantities of wastewater that are typically associated with industrial-scale operations.³⁴⁻³⁶ This 208 209 FO-MD hybrid technology is more efficient that single FO processes for dehydrating dilute 210 MMA/DMA feed solutions. This is because the MD process can immediately concentrate the FO 211 permeate consisting of Na₁₀-phytate draw solute. The combination of our novel Na₁₀-phytate 212 draw solute with a hybrid FO-MD process demonstrates a simple and effective technology for removing organic arsenicals. Additional complex processes such as pre-oxidation³⁷ and 213 operation in acidic conditions³⁸ are not required in this FO-MD process, while side reactions of 214 MMA conversion into more toxic inorganic species³⁹ is inhibited. The combination of our novel 215 Na₁₀-phytate draw solute with FO-MD hybrid process surpasses the water flux and organic 216 arsenical rejection of current technologies.^{1,6,8} 217



218

Figure 5. (a) The water flux and water transfer rate, (b) concentrations and rejection rates of organic arsenicals from water of a FO-MD hybrid process using 0.5 M Na₁₀-phytate as draw solute at 50 °C. 20 °C deionized water was deployed on the MD permeate side.

222 Using a plant-derived compound, phytic acid, we designed and synthesized a novel draw 223 solute that is suitable for removing organic arsenicals from water in a FO-MD hybrid process. 224 The presence of hydrophilic functional groups arranged in a three-dimensional configuration 225 around an aromatic carbon ring can form supramolecular structures through hydrogen bonds with 226 water molecules that prevent reverse diffusion during FO processes. This draws water across a 227 TFC-PES hollow fiber membrane at rates that can handle industrial-scale quantities of 228 wastewater. The regeneration of the novel Na₁₀-phytate draw solute is optimized using an in-line 229 MD process; streamlining the draw solute recovery process with minimal energy consumption.

| 230 | Phytate-based draw solutes can be potentially deployed for the removing other types of |
|-----|---|
| 231 | contaminants during water treatment and are especially suitable for protein enrichment. The |
| 232 | negligible reverse diffusion of Na ₁₀ -phyate in FO can avoid denaturation of impurity-sensitive |
| 233 | proteins and produce high-quality products. |
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| 242 | ■ ASSOCIATED CONTENT |
| 243 | Supporting Information Available |
| 244 | Materials; syntheses and characterizations of Na ₁₀ -phytate; the physicochemical properties of |
| 245 | Na ₁₀ -phytate solutions; FO processes; the preparation and analyses of MMAs and DMAs |
| 246 | solutions; the regeneration of Na ₁₀ -phytate; molecular structures of MMAs and DMAs; the TGA |
| 247 | spectrum; weight losses of Na10-phytate in TGA measurements; relative viscosity of Na10- |
| | |

- 248 phytate and phytic acid; osmotic pressure of Na₁₀-phytate, phytic acid and NaCl; Supermolecular
- 249 network and size distribution of Na₁₀-phytate in the aqueous solution.

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