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Cometabolic enzymatic transformation of organic micropollutants under methanogenic conditions

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15 ABSTRACT

16 Anaerobic digestion (AD) has been shown to have the biological potential to decrease the 17 concentrations of several organic micropollutants (OMPs) from sewage sludge. However, the 18 mechanisms and factors behind these biotransformations, which are essential for elucidating the 19 possible transformation products and to foster the complete removal of OMPs via operational 20 strategies, remain unclear. Therefore, this study investigated the transformation mechanisms of 21 20 OMPs during the methanogenic step of AD with a focus on the role of acetate kinase (AK), 22 which is a key enzyme in methane production. The results from lab-scale methanogenic reactors 23 showed that this step accounts for much of the reported OMPs biotransformation in AD. 24 Furthermore, enzymatic assays confirmed that AK transforms galaxolide, naproxen, 25 nonylphenol, octylphenol, ibuprofen, diclofenac, bisphenol A and triclosan. Except for 26 galaxolide, for which further studies are required to refine conclusions, the OMP's chemical 27 structure was a determinant for the AK action because only compounds that contain a carboxyl 28 or a hydroxyl group and that have a moderate steric hindrance were enzymatically transformed, 29 likely by phosphorylation. For these 7 compounds, this enzymatic mechanism accounts for 10-30 90% of the measured methanogenic biotransformation, suggesting that other active enzymes of 31 the AD process are also involved in OMPs biotransformation.

32 Abstract Art



33 INTRODUCTION

The increasing use of pharmaceuticals, personal care products, hormones and many other organic compounds in our daily life has resulted in the release of these organic micropollutants (OMPs) into the environment via diverse pathways.¹ There is evidence of negative effects of OMPs on human and ecosystem health,² which suggests that reduction measures should be applied, especially for one of the main sources of OMP discharge: effluents of sewage treatment plants (STPs).

40 Anaerobic digestion (AD) is widely used in STPs for sludge stabilization prior to its 41 application in agricultural soils as a biosolid. Significant quantities of OMPs arrive at the sludge 42 treatment line sorbed onto solids³ but also solubilized in the water phase of the sludge, achieving 43 concentrations up to 90-140 μ g L⁻¹ for musk fragrances (galaxolide and tonalide), 40 μ g L⁻¹ for 44 triclosan, 25 μ g L⁻¹ for ibuprofen, and 1-10 μ g L⁻¹ for hormones (estrone and 17 β -estradiol).⁴ 45 Most of these OMPs remain in the digested sludge,⁴ thus its use as a fertilizer might transfer 46 OMPs to the soil.⁵

Despite the environmental risks associated with biosolid-amended soils,⁵ few studies have 47 investigated the fate of OMPs during AD.^{4,6-12} They conclude that AD is able to biologically 48 49 transform OMPs, but the degree of removal of some compounds is still controversial, and few 50 transformation products (TPs) have been identified, which indicates the poor understanding of the microbial mechanisms and factors behind these biotransformations.³ Therefore, to develop 51 52 strategies that promote the complete elimination of OMPs and to predict the environmental risks 53 of the TPs that are generated, it is essential to understand the biotransformation pathways 54 involved in AD.

55 To date, only rare and inconclusive information is available about the relationship between anaerobic populations and the removal of OMPs.¹³ This lack of knowledge is justified by the 56 57 complexity of the biological and chemical processes involved in the four steps of AD (Figure 1): hydrolysis, acidogenesis, acetogenesis and methanogenesis.¹⁴ According to the taxonomic 58 analysis of Guo et al.,¹⁵ Bacteria (~93%) was more abundant than Archaea (methanogens) (~6%) 59 60 on sewage sludge AD. Among bacteria, Proteobacteria (41%), Firmicutes (13%) and 61 Bacteroidetes (10%) are the most abundant populations, while Methanosaeta (26%), 62 Methanospirillum (13%), Methanosarcina (13%), Methanoculleus (11%) and Methanoregula (8%) are the dominant methanogenic genera. During the four AD steps, these microorganisms 63 64 produce a broad diversity of enzymes to degrade the organic substrates. Depending on their 65 specificity, these enzymes can also modify the structures of OMPs despite being non-growth 66 substrates. This biochemical process is known as cometabolism, and it is considered by many 67 authors to be a major removal mechanism of OMPs during the biological treatment of wastewater.^{16–21} The action of enzymes on OMPs is poorly investigated, and most studies have 68 focused on oxygenases.^{21–23} A recent study by Krah et al.²⁴ tested the activity of a cocktail of 69 70 extracted enzymes from activated sludge towards OMPs. By measuring several TPs and using 71 several enzymatic inhibitors, they concluded that amide hydrolases could be involved in the 72 biotransformation of acetaminophen, acetyl sulfamethoxazole, atenolol and bezafibrate, that 73 oxidoreductases transform 10-OH-carbamazepine and that erythromycin is affected by 74 hydrolases acting on ester bonds. However, the suspected participation of particular enzymes in the cometabolic biotransformation of OMPs was not definitively confirmed because the 75 76 identification of enzymes by indirect measurements is unreliable, and the isolation of target enzymes has not yet been reported.²¹ Although no studies about the enzymatic transformation 77 78 routes of OMPs during AD have been found, it could be hypothesized that hydrolases from the two first AD steps, which perform relatively simple reactions, are involved in the biotransformation of some OMPs, as was demonstrated for activated sludge systems.²⁴ For the more specific acetogenic and methanogenic steps, the removal of polycyclic aromatic hydrocarbons (PAH) has been linked to methanogenesis,²⁵ but the type of enzymatic biotransformation that could occur is not clear.

84 The aceticlastic methanogenesis (Figure 1) is performed by the archaea Methanosaeta and Methanosarcina and it is considered the main pathway in the synthesis of methane in anaerobic 85 digesters.^{15,26} The first step in the methanization of acetate by *Methanosarcina* species is the 86 phosphorylation of acetate to acetyl phosphate (Figure 1 and Figure S1), which is performed by 87 88 the key intracellular enzyme acetate kinase (AK). Then, the acetyl phosphate is further converted to acetyl-CoA by phosphotransacetylase (PTA)²⁷ and the aceticlastic methanization continues 89 90 with the action of other enzymes (Figure S1). In addition, AK associated with PTA catalyzes the 91 reverse reaction (conversion of acetyl-CoA to acetate; Figure 1) in fermentative prokaryote bacteria during acetogenesis.²⁸ AK is a relatively specific enzyme but, in addition to acetate, it is 92 able to phosphorylate other substrates to a lower extent;^{29,30} thus, AK might also have effects on 93 94 some OMPs.



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Figure 1. Schematic representation of the steps of anaerobic digestion, including the roles of the
enzymes acetate kinase (AK) and phosphotransacetylase (PTA).

The major aim of this work was to gain insight into the biotransformation pathways involved in the removal of OMPs during methanogenesis. In particular, the enzymatic transformation of OMPs was assessed via in vitro assays with AK. Furthermore, the cometabolic impact of AK on the biotransformation of OMPs in methanogenic reactors (MRs) was quantified. To the best of our knowledge, this is the first attempt to clarify the role of AD enzymes on the biotransformation of OMPs.

104 MATERIALS AND METHODS

Organic micropollutants. This study focuses on 20 compounds that are commonly detected in sewage sludge^{3,4,8,9,31–33} and whose chemical structures (Table S1), applications and physicochemical properties (Table S2) are representative of a huge range of OMPs. The selected pollutants are: the musk fragrances galaxolide (HHCB), tonalide (AHTN) and celestolide (ADBI); the anti-inflammatories ibuprofen (IBP), naproxen (NPX) and diclofenac (DCF); the antibiotics sulfamethoxazole (SMX), trimethoprim (TMP), erythromycin (ERY) and roxithromycin (ROX); the neurodrugs fluoxetine (FLX), carbamazepine (CBZ) and diazepam (DZP); and the endocrine-disrupting compounds triclosan (TCS), bisphenol A (BPA), 4octylphenol (OP), 4-nonylphenol (NP), estrone (E1), 17 β -estradiol (E2) and 17 α -ethinylestradiol (EE2). These substances were purchased from Sigma-Aldrich (Steinheim, Germany) except for the fragrances, which were provided by Ventos (Spain). Stock solutions were prepared in HPLC grade methanol or acetone, depending on the compound, and stored at -18 °C.

Methanogenic reactor. Two lab-scale (14 L) continuously stirred (IKA RW20, 150 rpm) 117 MRs were operated under mesophilic (37 °C) conditions. Both reactors were inoculated with 118 119 sludge from a mesophilic STP anaerobic digester. The feeding consisted of a synthetic mixture of 120 volatile fatty acids (VFA) (acetic:butyric:propionic 50:25:25, %COD) and other trace nutrients 121 (section S4), in order to promote the growth of acetogenic bacteria and aceticlastic methanogenic archaea (i.e. Methanosaeta and Methanosarcina).^{15,34,35} After a start-up period, both MRs 122 123 reached steady-state operation at a hydraulic retention time (HRT) of 10 d, an organic loading rate (OLR) of 1 g COD $L^{-1} d^{-1}$ and a methanization efficiency above 70%. After 1-2 months 124 under these conditions (Table S4), a pulse of the selected OMPs (100 μ g L⁻¹ except for the 125 hormones, which were 10 μ g L⁻¹, section S4) was added to each MRs, and their concentrations 126 127 were followed in the liquid (17 samples) and solid (10 samples) phases for 10 d.

Acetate kinase activity. Acetate kinase (AK, EC 2.7.2.1) is a homodimer with two active sites that catalyze the reversible Mg-dependent transfer of the Y-phosphoryl group from adenosine triphosphate (ATP) to acetate.²⁷ AK from *Methanosarcina thermophila* was chosen for this study because it is a well characterized and investigated enzyme²⁹ and because *Methanosarcina* species appear to be key organisms in AD.³⁶ AK from *M. thermophila*

recombinant, expressed in *E. coli*, was purchased from Sigma-Aldrich (USA) as lyophilized powder (885 U mg⁻¹ protein (Bradford) and 6.94 U mg⁻¹ solid). One unit (U) phosphorylates 1.0 μ mol of acetate to acetyl phosphate per min at pH 7.6 and 25 °C. The AK powder was stored at -20 °C. Immediately before use, a solution of AK (288 μ g mL⁻¹, 2.0 U mL⁻¹) in a potassium phosphate buffer (0.1 M, pH 7.4) was prepared.

Hydroxamate assay. The hydroxamate assay^{37,38} is the simplest and most convenient method to determine the AK activity in the direction of acetyl phosphate synthesis.²⁹ This standard assay measures the rate of the forward Reaction 1 in the presence of hydroxylamine, which reacts with acetyl phosphate (Reaction 2) to form a colored complex in the presence of trivalent iron (Reaction 3). Under standard conditions, dephosphorylate acetyl phosphate is thermodynamically more favorable (Reaction 1); therefore, hydroxylamine is required to shift the reaction equilibrium to the right through the removal of acetyl phosphate.³⁷

145 Acetate + ATP
$$\stackrel{AK}{\longleftrightarrow}$$
 Acetyl-P + ADP (Reaction 1)

146 Acetyl-P + Hydroxylamine
$$\rightarrow$$
 Acetyl hydroxamate + HPO₄²⁻ (Reaction 2)

147 Acetyl hydroxamate + FeCl₃
$$\rightarrow$$
 Ferric acetylhydroxamate (Reaction 3)

An adaptation of the hydroxamate $assay^{30}$ was used in this study. A stock solution was prepared with the following components: 290 mM of Tris-HCl (pH 7.4, neutralized with KOH), 400 mM of potassium acetate, 20 mM of MgCl₂·6H₂O, 20 mM of ATP and 1410 mM of hydroxylamine hydrochloride (pH 7.4, neutralized with KOH). This reaction mixture was prepared immediately before use, since the aqueous solution of ATP at room temperature and the neutralized hydroxylamine solution have a limited stability. The reaction started by adding 1.0 mL of the enzyme solution over 1.0 mL of the pre-warmed (25 °C) reaction mixture. After

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incubation for 12 min at 25 °C, the reaction was stopped by the addition of 2.0 mL of 10% trichloroacetic acid. Then, the color reaction was initiated by adding 2.0 mL of FeCl₃ (2.5% in 2 N HCl). After 5-30 min of incubation to allow for the formation of the colored complex,³⁸ the absorbance was recorded spectrophotometrically (Cecil CE-7200, UK) at 540 nm.

A standard curve was prepared by adding $0-1.7 \text{ UmL}^{-1}$ of AK to the reaction mixture following the previously described methodology (Figure S2). Furthermore, because 1 U of enzyme dephosphorylates 1 µmol of ATP per minute, the consumption of ATP (µmol mL⁻¹) can also be correlated with the measured absorbance (Equation S1).

AK activity in the methanogenic reactor. The AK activity inside the reactor was 163 164 determined without special precautions to avoid the presence of air (AK is not sensitive to oxygen)^{30,39} following a methodology adapted from Mu et al.⁴⁰ Three samples 25 mL were 165 166 withdrawn at different time points after the OMPs spike (10 min, 3 d and 10 d) and then 167 centrifuged at 3107 g for 15 min. The supernatant was discarded, and the biomass was washed 168 and resuspended in 25 mL of 0.1 M sodium phosphate buffer (pH 7.4). This procedure was 169 repeated three times. The last resuspended mixture was sonicated at 20 kHz and 4 °C for 10 min 170 to break down the cell walls and release the intracellular AK. The sample was then centrifuged at 15344 g and 4 °C for 30 min to remove the waste debris. When needed, the extracts were stored 171 172 at -20 °C before measuring the enzyme activity assay via the hydroxamate assay.

Experiments with commercial AK. *Preliminary assays*. The relevance of several key parameters (temperature, pH and OMPs concentration) on the AK activity was first evaluated because they could impair the possible action of AK over the OMPs. Studies with AK from *M. thermophila* have typically been performed at 37 °C,^{27,30,41,42} but the product information from SIGMA recommends a temperature of 25 °C for the enzymatic assay. Therefore, several

theoretical AK activities (0.01, 0.05, 0.2 U mL⁻¹) were evaluated via the hydroxamate assay at both temperatures. The results did not show significant differences (Table S5); thus, the lower temperature (25 °C) was chosen for the next assays to minimize possible activity losses by enzyme denaturation in long-term experiments.⁴³

A similar procedure was followed for the pH selection. According to Aceti and Ferry³⁰ and the product information from SIGMA, the maximum activity of AK appears between pH 7.0-7.6. Both extreme pHs were tested and, as expected, the effect on the AK activity (0.02 U mL⁻¹) was negligible (Table S5). To prevent the pH from decreasing below 7.0 during the OMP assays due to the consumption of acetate (weak base), the initial pH was adjusted with KOH to 7.6.

187 Finally, a possible inhibition of the AK activity by the OMPs was examined at 3 OMPs/AK ratios: 0.35 μ g mg⁻¹ (50 μ g OMPs L⁻¹ and 144 mg AK L⁻¹), 3.5 μ g mg⁻¹ (100 μ g OMPs L⁻¹ and 188 28.8 mg AK L^{-1}) and 69 µg mg⁻¹ (100 µg OMPs L^{-1} and 1.44 mg AK L^{-1}). In all cases, a negative 189 190 control without OMPs was included. The results (Table S5) indicated that AK activity was not 191 altered by any OMPs/AK ratio tested, and thus neither by the corresponding solvents added with the OMPs pulse. The OMPs/AK ratio employed in the OMPs assays was $35 \ \mu g \ mg^{-1}$ (based on 192 the theoretical AK activity of a mesophilic AD operating at an OLR of 2 g COD $L^{-1} d^{-1}$ with a 193 194 70% methanization efficiency).

195 *Enzymatic transformation of OMPs.* Once the operating parameters were selected (25 °C, 196 pH 7.6, 100 μ g L⁻¹ of OMPs and 2.9 mg L⁻¹ of AK), the enzymatic transformation of the OMPs 197 was evaluated at two reaction times (1 d and 5 d) with and without the primary substrate (acetate) 198 in duplicate in 100 mL Erlenmeyer flasks. The final reaction media contained 145 mM of Tris-199 HCl, 10 mM of MgCl₂·6H₂O, 705 mM of hydroxylamine hydrochloride, 200 mM of potassium 200 acetate (in excess) and 50 mM of ATP. The acetate concentration, ATP consumption (Equation

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S1), AK activity, pH and temperature were monitored (n=2) at different reaction times (5-11 time points) and the concentration of OMPs was measured in replicate at the beginning (n=4) and end (n=4) of each experiment.

204 Abiotic disappearance of OMPs. The disappearance of the parent micropollutants by adsorption was evaluated by comparing the concentration of OMPs (100 µg L⁻¹ in distilled 205 water) after 1 d at 25 °C in contact with AK (2.9 mg L⁻¹) and without the enzyme. No 206 207 representative differences were found (data not shown), so the adsorption of OMPs during the 208 enzymatic assays with commercial AK was dismissed. Furthermore, to discard possible losses of 209 the parent compound by evaporation, analytical difficulties or chemical reaction with the media 210 (145 mM of Tris-HCl, 10 mM of MgCl₂, 705 mM of hydroxylamine hydrochloride, 50 mM of 211 ATP, 200 mM of potassium acetate), duplicated negative controls (without AK) were incubated 212 under the same conditions as the AK assays specified in the previous section. Trichloroacetic 213 acid was not used to stop the enzymatic activity at the end of the assays with OMPs because reductions of the concentrations of ERY, ROX and SMX were observed when it was added. The 214 215 enzymatic reaction was assumed to stop once the solid phase extraction (SPE) was performed.

Analytical methods. *Conventional parameters*. The operation of the methanogenic digesters was monitored in terms of the temperature, pH, total suspended solids (TSS), volatile suspended solids (VSS), alkalinity, ammonium, and total and soluble COD.⁴⁴ Biogas production was recorded using Ritter milligas counters (Dr. Ing. Ritter Apparatebau GmbH, Bochum, Germany), and its composition was determined through gas chromatography (HP 5890 Series II). Volatile fatty acids (VFA) were measured individually in a gas chromatograph (HP 5890A) equipped with a flame ionization detector (HP 7637A).

223 Organic micropollutants analysis. Samples from the MRs were centrifuged at 1880 g for 224 15 min. The supernatant was pre-filtered (AP4004705, Millipore) and filtered at 0.45 µm 225 (HAWP04700, Millipore) before performing the SPE with 100 mL samples and 60 mg OASIS HLB cartridges (Waters, Milford, MA, USA).^{4,16} Samples from the enzymatic assays (50 mL) 226 227 did not require any pre-treatment prior to SPE. To quantify the OMPs sorbed onto the 228 methanogenic sludge, ultrasonic solvent extraction (USE) was conducted before SPE, as described by Gonzalez-Gil et al.⁴ The limits of quantification (LOQ) and recoveries applied to 229 230 the MRs and the AK assays are shown in Table S3.

231 Statistical analysis. All of the enzymatic assays with OMPs were conducted in duplicate, 232 and each micropollutant concentration was measured twice (n=4). The results are expressed as 233 mean ± standard deviation. The significant differences between the enzymatic results were 234 statistically tested by analysis of variance (ANOVA) followed by the Dunnett T3 test for 235 multiple comparisons. The normal data distribution was analyzed with the Shapiro-Wilk test, and 236 the variance homogeneity was analyzed with the Levene test. When the variances were not 237 homogeneous, Brown-Forysthe analysis was used to assess the significant differences. All of the 238 statistical tests were performed at a 5% significance level using the IBM SPSS statistics® 239 software 20.0.

240 RESULTS AND DISCUSSION

Biotransformation of OMPs during methanogenesis. Figure 2 aims to highlight, in a semi-quantitative representation, the relevance of the methanogenic step on the overall removal efficiencies reported for OMPs during AD of sewage sludge. The methanogenic biomass showed the capacity to biotransform all of the tested OMPs, although the efficiencies varied depending on the compound. For example, SMX, NPX, TMP, OP, NP, FLX, EE2, TCS and the musk
fragrances were significantly removed during methanogenesis, while the other compounds had
biotransformations of less than 50%.

248 It was not easy to get a single value for the removal of OMPs during AD process due to the 249 divergences on the reported values. Trying to narrow these differences, the average AD removals 250 depicted on Figure 2 correspond to studies that used mainly continuous mesophilic digesters, 251 treating sewage sludge at a HRT between 10 and 30 d (Table S6). Even though, IBP, DCF, BPA, 252 NP, FLX, the hormones and the musk fragrances still presented high deviations, so it was not 253 possible to accurately determine the influence of the methanogenesis on their disappearance. 254 Most of the OMPs appear near the diagonal (between the two dashed lines) of Figure 2, 255 suggesting that the methanogenic biomass is the main responsible for their biotransformation 256 during AD. Therefore, the action of key enzymes participating in the methanogenic route, 257 particularly AK, were further investigated in the next sections. ROX and TMP are the only 258 compounds that clearly showed a higher removal during the overall AD, which indicates that 259 other anaerobic communities (i.e., hydrolytic and acidogenic) widely participate on their 260 removal.



261

Figure 2. Semi-quantitative representation of the removal efficiencies of OMPs in the methanogenic reactors (MR, x-axis) versus the average removal during AD of sewage sludge^{4,8–} $^{12,45-47}$ (y-axis). Compounds with large divergences on the reported AD removals (standard deviations above 25%) are depicted by open ellipses. No data about the removal of OP in AD was found (60-70% in the MRs). The two dashed lines delimit the region where methanogenesis explains to a high extent the overall AD removal of the OMPs.

Transformation of OMPs in the AK assays. The experiments were performed at two reaction times (1 and 5 days) and, with the exception of the negative control in the absence of AK, consisted of two assays with and without acetate, which were intended to assess the role of the main substrate on the possible AK catalytic transformation of OMPs. No significant differences were found between the negative control and the two AK assays in the 1-day experiment (data not shown), likely because the reaction time was not sufficient to illustrate the action of AK on OMPs. Hence, this section focuses on the 5-day assays.

Performance of assays. The ATP and acetate concentrations during the 5-day assay are 275 shown in Figure 3. The AK activity during the first 2 d was $4.8 \pm 0.2 \text{ U L}^{-1}$, which resulted in a 276 moderate conversion of acetate and ATP (approximately 30%). To increase the enzymatic 277 activity, a second pulse of AK (8.6 mg L⁻¹) and ATP (40 mM) was added to the reaction media 278 279 after 52 h. As a result, more rapid acetate and ATP consumption was achieved (Figure 3), which led to an average AK activity of 13.4 ± 1.5 U L⁻¹. After approximately 4 d, the depletion of both 280 281 reagents almost stopped, which suggests that the phosphorylation of acetate (Reaction 1) reached 282 an equilibrium. This is likely due to the decomposition of hydroxylamine at room temperature in neutralized solutions,^{38,48} which is required to perform Reaction 2 and thus to avoid equilibrium 283 284 of Reaction 1. This hypothesis is also supported by the fact that the acetate concentration on day 285 5 was higher than the value that was estimated stoichiometrically from the ATP consumption 286 rates (Figure 3). The average AK activity during the 5-day assay with acetate was much higher (8.0 U L^{-1}) than in the assay without acetate (0.1 U L^{-1}) , where a considerably lower conversion 287 288 of ATP (2% after 5 d) was also achieved because the only potential substrates for AK are OMPs 289 at a much lower concentration than the acetate.



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291 Figure 3. ATP (limiting reagent; squares) and acetate (dots) concentrations in the 5-day AK 292 assay with acetate. The arrow indicates the second addition of ATP and AK. The dashed line 293 represents the stoichiometric acetate concentration considering the consumption rates of ATP. 294 Chemical OMP transformation. The initial concentration of OMPs in distilled water is 295 compared to the concentration of OMPs after 5 d in the negative control (reaction media without 296 AK) in Figure 4. DZP, AHTN, ADBI, ERY, E1, EE2 and CBZ nearly completely disappear, 297 likely due to the reaction between the ketone group of these compounds (except EE2) with the 298 hydroxylamine required to perform the enzymatic assays. This hypothesis was confirmed 299 experimentally because the OMPs concentration did not decrease in a reaction media without this 300 amine. However, hydroxylamine was included in the reaction media to shift the action of AK 301 towards acetate phosphorylation (Reaction 1). Obviously, this decision masks any possible effect 302 of AK on the biotransformation of these seven OMPs.



303

Figure 4. Chemical transformation of OMPs expressed as the ratio between the concentration in the negative control (reaction media without enzyme) after 5 d (C) to the initial concentration in distilled water (C_0).

Enzymatic transformation. The concentrations of the 13 OMPs that were not chemically transformed in the negative control after the 5-day assays with AK are shown in Figure 5. Based on their chemical structures (Table S1), these compounds are classified into three groups. The first group includes carboxylic compounds (IBP, NPX, DCF), the second group are OMPs with hydroxyl groups (NP, OP, BPA, TCS, ROX, E2), and the third group comprises compounds with

312 other functional groups (HHCB, SMX, TMP, FLX). All data were normally distributed and, 313 except for NPX, OP and HHCB, their variances were homogeneous. FLX, TMP, SMX, E2 and ROX were not affected by AK, since no significant differences were found between the negative 314 315 control and both enzymatic assays (with and without acetate). In contrast, this difference was 316 statistically significant (p<0.05) for IBP, NPX, DCF, NP, OP, BPA, TCS and HHCB proving 317 that these 8 compounds are transformed by the action of the enzyme. No statistical differences 318 were found between the concentrations of these 8 OMPs in the AK assays performed in the 319 presence and absence of acetate, which suggests that no competitive inhibition occurred between the primary substrate (acetate) and the cometabolic substrates (OMPs).^{19,20} Moreover, these 320 321 results suggest that once AK is available it can directly transform OMPs in the absence of 322 acetate. However, the presence of the primary substrate is needed to trigger the synthesis of AK during the AD process; therefore, according to the definition of cometabolism,^{17,20,21} the 323 324 transformation of OMPs (non-growth substrates) is fortuitous and inherently linked to acetate.

325 To understand how AK can transform these compounds, it is necessary to look into the AK 326 specificity. The action of AK from *M. thermophila* is quite restrictive because the size of the 327 hydrophobic pocket and its affinity to the methyl group of acetate are determinant for substrate specificity.⁴² However, this enzyme still shows slight activity over larger substrates with 328 329 carboxyl groups (propionate, butyrate), alcohols (ethanol) and even over compounds without hydrophobic groups (formate, glycerol, glycine, glycolic acid).^{29,30} For that reason, IBP, NPX 330 331 and DCF, which have a carboxyl group (Group 1, Table S1) and are relatively small (1-2 332 benzene rings), were significantly (p < 0.05) biotransformed by AK (10-15%). In addition, 333 compounds with hydroxyl groups in their chemical structure (Group 2, Table S1) could also be 334 appropriate substrates for AK. Nevertheless, the molecular size appears to be determinant

335 because only the smaller compounds (NP, OP, BPA and TCS) were biotransformed (15-32%). 336 The steric hindrance of ROX and E2 could hinder the formation of the enzyme-substrate 337 complex and their further reaction. Similarly, ERY, EE2 and E1, which reacted chemically with 338 the hydroxylamine, would not be affected by AK due to their large molecular size. As expected, 339 the concentrations of the rest of the OMPs (Group 3, Table S1) did not decrease during the 340 enzymatic assays, excluding HHCB (45-50%), for which a clear explanation was not found. It is 341 hypothesized that AK could attack the ether group of this compound, but no references were 342 found to support this enzymatic transformation.



Figure 5. Transformation of OMPs by AK after 5 days of reaction time. The dark bars represent the concentrations of OMPs in the negative control (reaction media without enzyme), the light bars refer to the assay with acetate, and the white bars refer to the assay without acetate. The asterisks indicate the statistical differences (p<0.05) of both AK assays with respect to the control. The compounds are sorted according to their chemical structures in three groups: (G1) OMPs with a carboxyl group, (G2) OMPs with a hydroxyl group and (G3) OMPs with other functional groups.

352 The results of these experiments demonstrate the importance of combining the specificity of 353 the enzymes and the chemical structure of OMPs to understand the biotransformation 354 mechanisms. Accordingly, we hypothesized that some of the evaluated OMPs (Figure 5) are 355 enzymatically transformed by the cometabolic action of AK when ATP is available in the media 356 and disregarding the presence of the primary substrate (acetate). The proposed pathway predicts 357 the formation of phosphorylated TPs. For example, the carboxyl group of NPX could act as the 358 acceptor of a phosphoryl group, as occurs with acetate (Figure 6). Because the detection and 359 identification of TPs is very challenging, a deep understanding of the cometabolic enzymatic 360 biotransformation of OMPs allows TPs to be predicted, as was revealed in this study.



361

Figure 6. Proposed cometabolic pathway for the biotransformation of some OMPs, such as
NPX, by the enzyme AK during methanogenesis.

Relevance of AK to the biotransformation of OMPs during methanogenesis. The previous section suggests that when sufficient ATP is present in the media, AK could cometabolically phosphorylate compounds with carboxyl or hydroxyl groups and relatively low steric hindrance, such as IBP, NPX, DCF, NP, OP, BPA and TCS. Therefore, the next step is to

determine the relevance of this mechanism to the overall biotransformation that was observed for the aforementioned OMPs in the MRs. Figure 7 compares the biotransformation rates (μ g L⁻¹ d⁻¹) that were obtained in the AK assays with acetate and in the MRs at the same reaction time (5 d). It is important to note that the AK activities in the MRs and the enzymatic assay are quite similar (7.0 and 8.0 U L⁻¹, respectively). Assuming that 2/3 of the total CH₄ is produced from acetate by aceticlastic archaea,²⁶ the theoretical AK activity in the MRs (OLR of 1 g COD L⁻¹ d⁻¹ and 70% methanization would be 10 U L⁻¹, which is similar to the measured value.

Depending on the compound, the biotransformation rates in the MRs varied from 3.0 to 375 14 ug OMP L⁻¹ d⁻¹, while this range is narrower in the AK assays (1.5–4.0 ug OMP L⁻¹ d⁻¹). IBP 376 and DCF are hardly biotransformed during methanogenesis (3.0–3.5 μ g OMP L⁻¹ d⁻¹), although 377 378 an important removal mechanism (>45%) is related to AK action. The AK cometabolic 379 mechanism is also relevant (61%) in the methanogenic biotransformation of BPA (6 µg OMP L⁻ 1 d⁻¹). In contrast, the transformation of NPX by AK (12%) does not explain its high depletion 380 during methanogenesis (14 μ g OMP L⁻¹ d⁻¹). Likewise, AK accounts for only 28-38% of the 381 overall biotransformation observed for OP, NP and TCS in the MRs (10–13 μ g OMP L⁻¹ d⁻¹). On 382 the other hand, AK would explain the whole biotransformation rate of HHCB in the MR 383 (9 μ g OMP L⁻¹ d⁻¹); what is a quite surprising fact, especially because it cannot be ascertain that 384 this transformation occurs via the proposed cometabolic phosphorylation pathway. 385

Based on these results, no direct relationship was found between the degree of biotransformation during methanogenesis and the contribution of the AK cometabolic mechanism. Moreover, the biotransformation rate was always higher in the MRs, which suggests that other active enzymes in addition to AK could further transform them. Some candidate enzymes could be those that continue the aceticlastic methanogenesis (Figure S1), those involved

in the hydrogenotrophic methanogenesis⁴⁹ or other kinases, such as propionate and butyrate 391 kinase. These two kinases have a larger hydrophobic pocket than AK to directly phosphorylate 392 propionate and butyrate, respectively;^{50,51} therefore, they could be involved in the enzymatic 393 394 transformation of the same OMPs as AK, but they could also promote the transformation of 395 larger OMPs with hydroxyl groups, such as ERY, ROX, E1, E2 and EE2, which were not 396 affected by AK. Otherwise, the high methanogenic biotransformations of TMP and SMX (Figure 2) should be caused by other types of enzymes because kinases will not modify their chemical 397 398 structure. In summary, by combining enzyme action and the OMP's chemical structure, new 399 insights are provided to understand the transformation mechanisms of OMPs during AD, which 400 is useful for developing new strategies to maximize the elimination of OMPs from sewage 401 sludge and to predict the TPs that form and their potential risks.



Figure 7. Biotransformation rates of the OMPs prone to be cometabolically phosphorylated by
AK in the MRs (empty green bars) and in the AK assay with acetate (blue-pointed bars) after 5
days. The percentages indicate the contribution of AK to the methanogenic biotransformation.
G1 refers to OMPs with carboxylic compounds, and G2 refers to OMPs with hydroxyl groups.
HHCB was excluded from this figure due to the lack of information regarding the
phosphorylation action of AK on ether groups.

409 ASSOCIATED CONTENT

410 Supporting Information

- 411 The Supporting Information is available free of charge on the ACS Publications website.
- 412 Enzymatic pathways of the aceticlastic methanogenesis (Figure S1); chemical structures (Table
- 413 S1) and physicochemical properties (Table S2) of the OMPs; limits of quantification and
- 414 recoveries of the analytical methods (Table S3); details of the MRs performance (section S4,
- 415 Table S4); details of the acetate kinase activity determination (section S5, Figure S2); results
- 416 from the preliminary assays with AK (Table S5) and summary of the removal of OMPs during
- 417 AD (Table S6).

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422 Notes

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429 **REFERENCES**

- 430 (1) Kümmerer, K. Pharmaceuticals in the Environment. *Annu. Rev. Environ. Resour.* 2010, *35*431 (1), 57–75.
- 432 (2) Santos, L. H. M. L. M.; Araújo, A. N.; Fachini, A.; Pena, A.; Delerue-Matos, C.;
 433 Montenegro, M. C. B. S. M. Ecotoxicological aspects related to the presence of
 434 pharmaceuticals in the aquatic environment. *J. Hazard. Mater.* 2010, *175*, 45–95.
- 435 (3) Stasinakis, A. S. Review on the fate of emerging contaminants during sludge anaerobic digestion. *Bioresour. Technol.* 2012, *121*, 432–440.
- 437 (4) Gonzalez-Gil, L.; Papa, M.; Feretti, D.; Ceretti, E.; Mazzoleni, G.; Steimberg, N.;
 438 Pedrazzani, R.; Bertanza, G.; Lema, J. M.; Carballa, M. Is anaerobic digestion effective
 439 for the removal of organic micropollutants and biological activities from sewage sludge?
 440 *Water Res.* 2016, *102*, 211–220.
- 441 (5) Chen, F.; Ying, G.-G.; Ma, Y.-B.; Chen, Z.-F.; Lai, H.-J.; Peng, F.-J. Field dissipation and risk assessment of typical personal care products TCC, TCS, AHTN and HHCB in biosolid-amended soils. *Sci. Total Environ.* 2014, *470–471*, 1078–1086.
- des Mes, T. Z. D.; Kujawa-Roeleveld, K.; Zeeman, G.; Lettinga, G. Anaerobic
 biodegradation of estrogens Hard to digest. *Water Sci. Technol.* 2008, 57 (8), 1177–
 1182.
- 447 (7) Alvarino, T.; Suarez, S.; Lema, J. M.; Omil, F. Understanding the removal mechanisms of
 448 PPCPs and the influence of main technological parameters in anaerobic UASB and
 449 aerobic CAS reactors. *J. Hazard. Mater.* 2014, *278*, 506–513.
- (8) Carballa, M.; Omil, F.; Ternes, T.; Lema, J. M. Fate of pharmaceutical and personal care products (PPCPs) during anaerobic digestion of sewage sludge. *Water Res.* 2007, *41* (10), 2139–2150.
- (9) Narumiya, M.; Nakada, N.; Yamashita, N.; Tanaka, H. Phase distribution and removal of pharmaceuticals and personal care products during anaerobic sludge digestion. *J. Hazard. Mater.* 2013, *260* (2013), 305–312.
- (10) Samaras, V. G.; Stasinakis, A. S.; Thomaidis, N. S.; Mamais, D.; Lekkas, T. D. Fate of
 selected emerging micropollutants during mesophilic, thermophilic and temperature co phased anaerobic digestion of sewage sludge. *Bioresour. Technol.* 2014, *162*, 365–372.
- 459 (11) Malmborg, J.; Magnér, J. Pharmaceutical residues in sewage sludge: Effect of sanitization and anaerobic digestion. *J. Environ. Manage.* 2015, *153*, 1–10.
- 461 (12) Yang, S.; Hai, F. I.; Price, W. E.; McDonald, J.; Khan, S. J.; Nghiem, L. D. Occurrence of
 462 trace organic contaminants in wastewater sludge and their removals by anaerobic
 463 digestion. *Bioresour. Technol.* 2016, *210*, 153–159.
- 464 (13) Braun, F.; Hamelin, J.; Bonnafous, A.; Delgenès, N.; Steyer, J.-P.; Patureau, D. Similar
 465 PAH fate in anaerobic digesters inoculated with three microbial communities
 466 accumulating either volatile fatty acids or methane. *PLoS One* 2015, *10* (4), 1–20.
- 467 (14) Christy, P. M.; Gopinath, L. R.; Divya, D. A review on anaerobic decomposition and

468 enhancement of biogas production through enzymes and microorganisms. *Renew. Sustain.*469 *Energy Rev.* 2014, *34*, 167–173.

- 470 (15) Guo, J.; Peng, Y.; Ni, B.-J.; Han, X.; Fan, L.; Yuan, Z. Dissecting microbial community
 471 structure and methane-producing pathways of a full-scale anaerobic reactor digesting
 472 activated sludge from wastewater treatment by metagenomic sequencing. *Microb. Cell*473 *Fact.* 2015, *14* (1), 33.
- 474 (16) Fernandez-Fontaina, E.; Pinho, I.; Carballa, M.; Omil, F.; Lema, J. M. Biodegradation
 475 kinetic constants and sorption coefficients of micropollutants in membrane bioreactors.
 476 *Biodegradation* 2013, 24 (2), 165–177.
- 477 (17) Delgadillo-Mirquez, L.; Lardon, L.; Steyer, J.-P.; Patureau, D. A new dynamic model for
 478 bioavailability and cometabolism of micropollutants during anaerobic digestion. *Water* 479 *Res.* 2011, 45 (15), 4511–4521.
- 480 (18) Pomiès, M.; Choubert, J.-M.; Wisniewski, C.; Coquery, M. Modelling of micropollutant
 481 removal in biological wastewater treatments: a review. *Sci. Total Environ.* 2013, 443,
 482 733–748.
- 483 (19) Plósz, B. G.; Leknes, H.; Thomas, K. V. Impacts of competitive inhibition, parent compound formation and partitioning behavior on the removal of antibiotics in municipal wastewater treatment. *Environ. Sci. Technol.* 2010, *44* (2), 734–742.
- 486 (20) Criddle, C. S. The kinetics of cometabolism. *Biotechnol. Bioeng.* **1993**, *41* (11), 1048– 487 1056.
- 488 (21) Fischer, K.; Majewsky, M. Cometabolic degradation of organic wastewater
 489 micropollutants by activated sludge and sludge-inherent microorganisms. *Appl. Microbiol.* 490 *Biotechnol.* 2014, 98 (15), 6583–6597.
- 491 (22) Fernandez-Fontaina, E.; Gomes, I. B.; Aga, D. S.; Omil, F.; Lema, J. M.; Carballa, M.
 492 Biotransformation of pharmaceuticals under nitrification, nitratation and heterotrophic
 493 conditions. *Sci. Total Environ.* 2015, *541*, 1439–1447.
- 494 (23) Kassotaki, E.; Buttiglieri, G.; Ferrando-Climent, L.; Rodriguez-Roda, I.; Pijuan, M.
 495 Enhanced sulfamethoxazole degradation through ammonia oxidizing bacteria co496 metabolism and fate of transformation products. *Water Res.* 2016, 94, 111–119.
- 497 (24) Krah, D.; Ghattas, A.-K.; Wick, A.; Bröder, K.; Ternes, T. a. Micropollutant degradation
 498 via extracted native enzymes from activated sludge. *Water Res.* 2016, *95*, 348–360.
- 499 (25) Trably, E.; Patureau, D.; Delgenes, J. P. Enhancement of polycyclic aromatic
 500 hydrocarbons removal during anaerobic treatment of urban sludge. *Water Sci. Technol.*501 2003, 48 (4), 53–60.
- 502 (26) Conrad, R. Contribution of hydrogen to methane production and control of hydrogen
 503 concentrations in methanogenic soils and sediments. *FEMS Microbiol. Ecol.* 1999, 28 (3),
 504 193–202.
- 505 (27) Gorrell, A.; Ferry, J. G. Investigation of the Methanosarcina thermophila Acetate Kinase
 506 Mechanism by Fluorescence Quenching. *Biochemistry* 2007, *46*, 14170–14176.

- 507 (28) Ferry, J. G. Acetate Kinase and Phosphotransacetylase. In *Methods in Methane* 508 *Metabolism, Part A*; Elsevier Inc., 2011; Vol. 494, pp 219–231.
- 509 (29) Iyer, P.; Ferry, J. G. Acetate Kinase from Methanosarcina thermophila, a Key Enzyme for
 510 Methanogenesis. In *Methods in Biotechnology-Microbial Enzymes and*511 *Biotransformations*; José Luis, B., Ed.; Humana Press Inc: Totowa, New Jersey, 2005;
 512 Vol. 17, pp 239–246.
- 513 (30) Aceti, D. J.; Ferry, G. Purification and characterization of Acetate Kinase from Acetate-514 grown Methanosarcina thermophila. *J. Biol. Chem.* **1988**, *263* (30), 15444–15448.
- (31) Langford, K. H.; Reid, M.; Thomas, K. V. Multi-residue screening of prioritised human pharmaceuticals, illicit drugs and bactericides in sediments and sludge. *J. Environ. Monit.*2011, 13 (8), 2284–2291.
- (32) Radjenović, J.; Petrović, M.; Barceló, D. Fate and distribution of pharmaceuticals in
 wastewater and sewage sludge of the conventional activated sludge (CAS) and advanced
 membrane bioreactor (MBR) treatment. *Water Res.* 2009, 43 (3), 831–841.
- (33) Bolz, U.; Hagenmaier, H.; Körner, W. Phenolic xenoestrogens in surface water,
 sediments, and sewage sludge from Baden-Württemberg, south-west Germany. *Environ. Pollut.* 2001, *115* (2), 291–301.
- 524 (34) Demirel, B.; Scherer, P. The roles of acetotrophic and hydrogenotrophic methanogens
 525 during anaerobic conversion of biomass to methane: A review. *Rev. Environ. Sci.*526 *Biotechnol.* 2008, 7 (2), 173–190.
- (35) Regueiro, L.; Lema, J. M.; Carballa, M. Key microbial communities steering the
 functioning of anaerobic digesters during hydraulic and organic overloading shocks.
 Bioresour. Technol. 2015, 197, 208–216.
- 530 (36) De Vrieze, J.; Hennebel, T.; Boon, N.; Verstraete, W. Methanosarcina: the rediscovered 531 methanogen for heavy duty biomethanation. *Bioresour. Technol.* **2012**, *112*, 1–9.
- (37) Rose, I. A.; Grunberg-Manago, M.; Korey, S. R.; Ochoa, S. Enzymatic phosphorylation of
 acetate. *J. Biol ogical Chem.* 1954.
- (38) Lipmann, F.; Tuttle, L. C. A specific micromethod for the determination of acyl
 phopsphates. J. Biol. Chem. 1945.
- (39) Bock, A. K.; Glasemacher, J.; Schmidt, R.; Schönheit, P. Purification and characterization
 of two extremely thermostable enzymes, phosphate acetyltransferase and acetate kinase,
 from the hyperthermophilic eubacterium Thermotoga maritima. *J. Bacteriol.* 1999, *181*(6), 1861–1867.
- 540 (40) Mu, H.; Chen, Y.; Xiao, N. Effects of metal oxide nanoparticles (TiO2, Al2O3, SiO2 and
 541 ZnO) on waste activated sludge anaerobic digestion. *Bioresour. Technol.* 2011, *102* (22),
 542 10305–10311.
- 543 (41) Gorrell, A.; Lawrence, S. H.; Ferry, J. G.; Bacteriol, J. Structural and Kinetic Analyses of
 544 Arginine Residues in the Active Site of the Acetate Kinase from Methanosarcina
 545 thermophila. J. Biol. Chem. 2005, 280 (11), 10731–10742.

- 546 (42) Ingram-Smith, C.; Gorrell, A.; Lawrence, S. H.; Iyer, P.; Smith, K.; Ferry, J. G.
 547 Characterization of the Acetate Binding Pocket in the Methanosarcina thermophila
 548 Acetate Kinase. J. Bacteriol. 2005, 187 (7), 2386–2394.
- 549 (43) Bisswanger, H. Enzyme assays. Perspect. Sci. 2014, 1 (1-6), 41-55.
- APHA. Standard Methods for the Examination of Water and Wastewater, 21st ed.;
 American Public Health Association/American Water Works Association/Water
 Environment Federation: Washington DC, USA, 2005.
- (45) Clara, M.; Gans, O.; Windhofer, G.; Krenn, U.; Hartl, W.; Braun, K.; Scharf, S.;
 Scheffknecht, C. Occurrence of polycyclic musks in wastewater and receiving water
 bodies and fate during wastewater treatment. *Chemosphere* 2011, 82 (8), 1116–1123.
- Paterakis, N.; Chiu, T. Y.; Koh, Y. K. K.; Lester, J. N.; McAdam, E. J.; Scrimshaw, M.
 D.; Soares, A.; Cartmell, E. The effectiveness of anaerobic digestion in removing estrogens and nonylphenol ethoxylates. *J. Hazard. Mater.* 2012, *199–200*, 88–95.
- 559 (47) Bergersen, O.; Østnes, K.; Vasskog, T. Bioresource Technology Anaerobic treatment of
 560 sewage sludge containing selective serotonin reuptake inhibitors. *Bioresour. Technol.*561 2012, 117, 325–332.
- 562 (48) Wang, Q.; Wei, C.; Pérez, L. M.; Rogers, W. J.; Hall, M. B.; Mannan, M. S. Thermal decomposition pathways of hydroxylamine: theoretical investigation on the initial steps. *J. Phys. Chem. A* 2010, *114* (34), 9262–9269.
- 565 (49) Ferry, J. G. The chemical biology of methanogenesis. *Planet. Space Sci.* 2010, 58 (14–15),
 566 1775–1783.
- 567 (50) Stams, A. J. M. Metabolic interactions between anaerobic bacteria in methanogenic 568 environments. *Antonie Van Leeuwenhoek* **1994**, *66*, 271–294.
- 569 (51) Allen, S. H.; Kellermeyer, R. W.; Stjernholm, R. L.; Wood, H. G. Purification and
 570 Properties of Enzymes Involved in the Propionic Acid Fermentation. J. Bacteriol. 1964,
 571 87, 171–187.