

# 1 **Accumulation of marine microplastics along a trophic gradient as determined by an Agent-** 2 **Based Model**

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## 11 **Abstract**

12 Microplastics are ubiquitous in the marine environment and are now consistently found in almost all  
13 marine animals. This study examines the rate of accumulation in a modelled filter feeder (mussels)  
14 both from direct uptake of microplastics and from direct uptake in addition to trophic uptake (via  
15 consuming plankton which have consumed microplastic themselves). We show that trophic uptake  
16 plays an important role in increasing plastic present in filter feeders, especially when consumption  
17 of the plastic does not reduce its overall abundance in the water column (e.g. in areas with high  
18 water flow such as estuaries). However, we also show that trophic transfer increases microplastic  
19 uptake, even if the amount of plastic is limited and depleted, as long as plankton are able to  
20 reproduce (for example, as would happen during a plankton bloom). If both plankton and plastic  
21 are limited and reduced in concentration by filter feeding, then no increase in microplastic by  
22 trophic transfer occurs, but microplastic still enters the filter feeders. The results have important  
23 implications for large filter feeders such as baleen whales, basking and whale sharks, as these  
24 animals concentrate their feeding on zooplankton blooms and as a result are likely to consume  
25 more plastic than previous studies have predicted.

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27 Key words: microplastic; plankton; mussel; filter feeder; trophic ecology; trophic transfer

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## 1. Introduction

Plastic, especially microplastics, have become ubiquitous in the marine environment (Eriksen et al. 2014), with recent studies showing their presence in almost all marine animals including those from the deep sea (Taylor et al. 2016). Microplastic ingestion by marine organisms can cause a range of negative effects including endocrine disruption, mutagenicity and carcinogenicity (Rios et al. 2007), which can have repercussions for growth, sexual development, fecundity, morbidity and mortality (reviewed by Cole et al. 2013).

Trophic transfer of microplastics has been demonstrated in laboratory studies, from zooplankton to mysid shrimp (Setälä et al. 2014) and from mussels to crabs (Farrell and Nelson 2013). However, little is known about the accumulation of microplastics through trophic transfer outside of laboratory studies, partially due to the difficulties of tracking microplastics and small organisms such as plankton through space and time.

In this study we present an agent-based modelling approach to investigate the role of trophic transfer of microplastics. We modelled plastic microbeads, plastic thread, zooplankton (three 'species' with three different feeding preferences for microbeads and other zooplankton) and mussels as agents in the model. As much research has previously been conducted on zooplankton uptake of microbeads, we assumed in the models that microbeads could be consumed by zooplankton and mussels, where as thread could only be consumed directly by mussels; hence comparing thread to microbead concentration in mussels allowed us to assess the effects of trophic transfer (we are subsequently aware of some research indicating thread can be consumed by zooplankton e.g. Dedman, 2014, but in the model, this was not permitted as it allows for comparisons of trophic transfer on uptake). We examined scenarios where filter feeding by mussels would: 1) not affect the concentration of microplastic and zooplankton in the water (i.e. both were highly abundant, or there was continuous movement of water); 2) not effect the concentration of zooplankton, but would reduce the abundance of microplastics (i.e. 'clean' water with little microplastic, but with rapid growth in zooplankton, such as a plankton bloom) and; 3) reduce both the concentration of plastic and of zooplankton as they were consumed.

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## 60 **2. Methods**

61 Agent-based models were built in R (R Core Team 2015; see  
62 [www.rickstafford.com/plastic\\_models.html](http://www.rickstafford.com/plastic_models.html) for source code) to simulate the actions and interactions  
63 of the following six agents; mussels, selective feeding zooplankton (e.g. nauplii and cirripede  
64 nauplii), non-selective zooplankton (e.g. gastropods) and predatory feeding zooplankton (e.g.  
65 copepod, decapod and worms), and microplastic (both bead and thread) in order to assess the  
66 uptake of microplastics by mussels either directly (by examining thread uptake, which did not pass  
67 through zooplankton in the model, see introduction), or by direct and trophic transfer uptake (by  
68 examining beads, which were consumed by zooplankton as well as directly by mussels). By  
69 modelling thread and beads in this manner, it was possible to examine the differences in uptake  
70 between only direct uptake, and uptake through trophic transfer.

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72 The model was run in a 100 x 100 grid arena and lasted 100 time-steps. Mussels were non-moving  
73 and always present (but positions of mussels were randomly generated on the grid), whereas the  
74 zooplankton and microplastic moved around and once ingested, in some simulations, were  
75 replaced by new agents in random locations (regeneration). Mussels were programmed to uptake  
76 beads, threads, and all 3 types of zooplankton, if in the same grid cell or one of the neighbouring  
77 nine grid cells to the mussel. Uptake was stochastic with a certain probability defined for likelihood  
78 of consumption if the agent to be consumed was in the specified cells. Selective and non-selective  
79 feeding zooplankton were programmed to uptake beads only, if both were in the same grid square,  
80 and predatory feeding zooplankton were programmed to uptake beads and both selective and non-  
81 selective feeding zooplankton. In all cases, uptake was not guaranteed, but stochastic and based  
82 on probability estimates of uptake of zooplankton and microplastic as defined in Cole et al. (2013),  
83 see Table 1 for the probability values used in this study.

84

85 Zooplankton, beads and threads moved by one grid square per time-step (including diagonal  
86 movement), with a heading generated from that of the heading of the previous time-step.

87 Following directionality rules used in previous ecological ABM models (Stafford and Davies 2005)

88 plastic particles could adjust their bearing by up to 90 degrees per time step and plankton by up to  
89 45 degrees per time step. These changes in heading were generated from random numbers drawn  
90 from a uniform distribution.

91

92 Three plastic scenarios were simulated based on the empirical data results; 1 = equal amounts of  
93 thread and beads, 2 = more thread than beads and 3 = more beads than threads. Three different  
94 ratios of plastic to zooplankton were also conducted based on the empirical data results;  
95 Plastic:Plankton ratio 1 = 75:25, 2 = 50:50 and 3 = 25:75. Four zooplankton community structures  
96 were used: 1 = medium to high numbers of most species, 2 = medium to high numbers of copepod  
97 and cirripede, 3 = low to medium numbers of most species, and 4 = low to medium numbers of  
98 copepod, decapod and gastropod. In all cases, the numbers of mussels remained fixed (see Table  
99 2 for exact numbers used in each simulation).

100

101 In total 36 scenarios were run, each scenario was run 3 times and a mean taken (total n = 108  
102 model runs). Model 1 regenerated both microplastic and zooplankton, so once a plastic bead,  
103 thread or plankton agent was consumed, and another reappeared in a random location. Model 2  
104 was run to regenerate zooplankton only (hence microplastic in the water column was depleted over  
105 time) and Model 3 was run with no regeneration of either zooplankton or microplastic.

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### 107 **3. Results**

108 A number of factors influence microplastic uptake in the models. For model 1, the different input  
109 parameters and plastic uptake in each biological agent type are shown in Table 2. Not only does  
110 the amount of plastic increase in plankton and mussels with increasing amounts of plastic in the  
111 water, but more plankton also result in more plastic accumulating in the mussels.

112

113 The three Trophic Interaction Agent-based Models, showed different results in total microplastic  
114 uptake based on the different regeneration scenarios (Figure 1). When both microplastic and  
115 zooplankton were regenerated there was a large increase in the uptake of microplastic in the  
116 presence of zooplankton, with three times as much microplastic ingested at some levels of

117 microplastic concentration compared to no regeneration of either plastic or plankton (Figure 1a).  
118 This difference was reduced when there was no regeneration of microplastic. However, there was  
119 still a higher uptake of microplastic in the presence of zooplankton, with ~ 50% more microplastic  
120 ingested if passing through zooplankton as an additional uptake route (Figure 1b). If there is no  
121 regeneration of either microplastic or zooplankton then the amount of uptake is similar between  
122 beads (which are consumed by zooplankton) and threads (not consumed by zooplankton)  
123 indicating no significant increase in microplastic uptake in mussels was occurring through trophic  
124 transfer (Figure 1c). The variability of plastic bead concentration in mussels increased with plastic  
125 bead concentration in the water due to the changes in the amount of plankton in the model in  
126 different scenarios (as seen in Table 2), so while the overall trend was for increases in plastic  
127 beads in mussels as their concentration in the water increased, this was modified by plankton  
128 density. This created heteroscedasticity of data making it unsuitable for parametric statistical  
129 analysis. However, the difference in gradients between beads and threads in models 1 and 2 are  
130 clear and do not require statistical verification.

131

#### 132 **4. Discussion**

133 The results demonstrate that under two of the three studied scenarios, the ingestion of microbeads  
134 by zooplankton, and subsequent consumption of zooplankton by mussels increased the amount of  
135 plastic found in mussels as compared to routes with no trophic intermediate stage present (as  
136 determined by thread uptake in the mussels).

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138 These scenarios where plastic and/or plankton are 'regenerated' after consumption are not  
139 ecologically unrealistic. Coral reefs, for example, exist in nutrient poor areas, and the basis of the  
140 plankton-based food chain is through plankton continuously drifting over the reef (Odum and Odum  
141 1955; Atkinson and Grigg, 1984). Such currents and condition which bring plankton are also likely  
142 to carry microplastics. The same is likely to be true of many coastal environments, especially tidal  
143 areas such as estuaries, where again, much material is imported with each tidal cycle (Peterson et  
144 al. 1985). Both estuaries and coral reefs are also important grounds for commercial fishing and  
145 shellfish stocks, meaning that further transfer into humans is then possible.

146

147 Equally, 'regeneration' of zooplankton would be likely to occur during plankton blooms, as  
148 reproduction and growth is normally rapid and opportunistic based on phytoplankton abundance.  
149 Hence, even where the amount of plastic in the water may be limited, high numbers of  
150 zooplankton can result in faster rates of uptake than may have been previously thought. This may  
151 have implications for plastic uptake in large filter feeders, such as baleen whales and basking or  
152 whale sharks, as they are known to selectively target these high abundance patches of  
153 zooplankton when feeding (e.g. Sims and Quayle, 1998).

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155 Microplastics are another increasingly important stressor on marine ecosystems, already under  
156 stress inflicted by factors such as climate change, overfishing and other pollutants (Halpern et al.  
157 2008). While there are policies and procedures designed to protect against further plastic pollution,  
158 e.g. the EU's Good Environmental Status (Galgani et al. 2013; Wright et al. 2013), these policies  
159 only consider the effects of plastics directly in the water column. While further work is necessary to  
160 fully quantify the magnitude of trophic transfer in situ, this current study demonstrates the potential  
161 increase in uptake that could occur in higher trophic level species. Consequently, the role of trophic  
162 transfer needs to be given substantial consideration when developing appropriate limits for  
163 microplastic in the ocean.

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231 Table 1. Uptake probabilities (%) used for all scenarios in Model 1, 2 and 3. If random number  
232 was  $\leq$  probability when in the same grid cell (or additional 9 neighbouring grid cells for mussels)  
233 then the object would be consumed. Zooplankton feeding rate probabilities taken from Cole et al.  
234 (2013)

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<b>Scenario</b>	<b>Probability</b>
Selective plankton feeding on bead	0.8
Non-selective plankton feeding on bead	0.9
Predatory plankton feeding on bead	0.8
Predatory plankton feeding on selective plankton	0.7
Predatory plankton feeding on non-selective plankton	0.7
Mussel feeding on bead	0.9
Mussel feeding on selective plankton	0.9
Mussel feeding on non-selective plankton	0.9
Mussel feeding on predatory plankton	0.9
Mussel feeding on thread	0.9

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251 Table 2. The 36 scenarios of different plastic and plankton concentrations used in each model  
 252 and the mean outputs from three replicate runs for each scenario for model 1.

Plastic thread	Plastic bead	Selective plankton	Non-selective plankton	Predatory plankton	Mussels	Plastic in mussels	Plastic thread in mussels	Plastic in Selective plankton	Plastic in non-selective plankton	Plastic in predatory plankton
200	200	200	100	200	10	169	56	131	80	378
200	400	200	100	200	10	357	99	277	154	763
100	100	200	100	200	10	98	23	74	31	200
100	300	200	100	200	10	294	25	213	108	570
200	600	200	100	200	10	546	49	404	247	1145
50	150	200	100	200	10	131	9	101	61	277
300	100	200	100	200	10	81	92	65	36	192
600	200	200	100	200	10	184	169	134	79	396
150	50	200	100	200	10	39	36	45	19	101
200	200	50	50	400	10	172	48	29	27	539
400	400	50	50	400	10	358	102	55	49	1119
100	100	50	50	400	10	88	26	11	12	269
100	300	50	50	400	10	251	21	35	40	861
200	600	50	50	400	10	509	65	72	80	1645
50	150	50	50	400	10	122	20	20	20	425
300	100	50	50	400	10	83	73	9	9	271
600	200	50	50	400	10	171	143	30	30	536
150	50	50	50	400	10	42	46	6	8	132
150	150	100	100	100	10	101	40	66	81	128
400	400	100	100	100	10	263	97	194	205	331
50	50	100	100	100	10	28	14	21	33	42
100	200	100	100	100	10	129	27	86	100	159
200	600	100	100	100	10	359	54	289	301	536
25	50	100	100	100	10	32	7	21	26	47
250	100	100	100	100	10	58	59	47	49	84
600	200	100	100	100	10	116	140	90	108	167
140	50	100	100	100	10	27	43	22	44	31
200	200	50	150	200	10	150	50	32	117	345
400	400	50	150	200	10	327	111	69	232	689
100	100	50	150	200	10	72	26	19	58	171
100	300	50	150	200	10	212	22	53	188	551
200	600	50	150	200	10	421	43	90	362	1049
50	150	50	150	200	10	121	13	18	82	244
300	100	50	150	200	10	86	74	23	53	167
600	200	50	150	200	10	177	169	36	116	338
150	50	50	150	200	10	33	40	7	25	80

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259 Figure 1. Relationship between amount of plastic thread in the water and uptake by mussels  
260 (grey line) compared to the relationship between amount of plastic beads in the water and uptake  
261 by mussels (direct uptake and via plankton, black line). (a) Model 1 – regeneration of consumed  
262 beads and plankton, (b) Model 2 – regeneration of plankton only, (c) Model 3 – no regeneration

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