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Measurements of low-level anthropogenic radionuclides from soils around Maralinga

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Abstract. The isotopes ²³⁹Pu and ²⁴⁰Pu are present in surface soils as a result of global fallout from nuclear weapons tests carried out in the 1950's and 1960's. These isotopes constitute artificial tracers of recent soil erosion and sediment movement. In practice the high throughput capabilities and high sensitivity of the AMS technique makes the study of Australia's geographically large areas viable using Pu isotopes. As part of its weapons development program the United Kingdom carried out a series of atmospheric and surface nuclear weapons tests at Maralinga, South Australia in 1956 and 1957. The contribution from the Maralinga tests to the Pu isotopic abundances present in the region around Maralinga is largely unknown. In global fallout, for example, the ²⁴⁰Pu/²³⁹Pu ratio is typically in the range 0.17 - 0.19, but the influence of the regional tests could lead to values outside this range. This would impact on the assessment techniques used in the soil and sediment tracer studies. We report recent measurements on soil samples collected from across the Maralinga Test site.

1 Introduction

From the point of view of using plutonium as a tracer of soil movement nuclear weapons tests can be divided into two categories: (i) high yield explosions that inject Pu into the stratosphere, where it remains with a residence half-life of 8 - 24 months [1] and becomes relatively well mixed before returning to the surface as "global fallout", and (ii) low yield detonations ($< \sim 300$ kT) for which the bulk of the nuclear debris are confined to the troposphere, for a residence half-life estimated at ~70 days [2] and which leads to "regional fallout" over an area determined largely by the fission yield of the device and by the meteorological conditions at the time of the test and in the period thereafter. Fallout in the immediate vicinity of tests carried out close to ground level can also have relatively large particles derived from bomb fragments or contaminated ground material that are deposited from the atmosphere as "local fallout" within hours or days of the test [3]. The final relative amounts of the Pu isotopes from each test depend on the individual weapon design and yield, hence local and regional fallout can have plutonium isotopic signatures that differ significantly from that of global fallout.

Between 1952 and 1957 twelve nuclear weapons were detonated in Australia, with an average yield of \sim 15 kT equivalent TNT. These tests were carried out at three sites: at the Montebello Islands, off the coast of Western

Australia (3 tests), and in South Australia at Emu field (2 tests) and Maralinga (7 tests). In addition to the nuclear detonations at Maralinga there were also numerous safety trials that involved the burning and/or dispersal by conventional explosives of plutonium, uranium, and other radionuclides [4]. In particular the 12 Vixen B trials, each of which involved approximately 4 TBq of ²³⁹Pu, left Taranaki the most severely contaminated site at Maralinga, with approximately 22 kg of ²³⁹Pu being distributed across the local environment. This is comparable with the ~30 kg of ²³⁹Pu deposited across the whole of mainland Australia as a result of global fallout derived from all atmospheric tests world-wide.

The Maralinga region sits on a low calcrete rise at \sim 200 m altitude north of the Nullarbor Plain, on the southern edge of the Great Victoria Desert. Relief in the area is typically less than 6 m and average annual rainfall is 200 mm.a⁻¹. Vegetation is largely open woodland with a low shrub understory, or hummock grassland. Weapons tests at Maralinga were carried out within an area of a few km², approximately 35 km to the north of Maralinga village (figure 1). The safety trials were conducted at several locations to the south and south-east of the test site, and up to ~30 km to the north-east of the village.

In order to use global fallout Pu as a tracer of soil movement in Australia the regional fallout component attributable to the Maralinga tests needs to be characterized. This is best achieved using samples collected from close to the test site area, where the signal

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is greatest, but outside the region affected by local fallout from the safety trials. While the vast bulk of the local fallout dispersed during the trials was removed during clean-up and rehabilitation projects, the extent of the area where local Pu fallout is a significant fraction of the total Pu fallout has not been established. The aim of the present work was to determine optimal sampling locations to avoid local fallout from the trials while maximizing the regional fallout signal from the tests in Australia.

2 Methods

Soil cores for the study reported here were collected at approximately 5 km intervals alongside the access road between Maralinga village and the test site area. Sampling sites were roughly 20-30 m from the access road, and at the time of collection showed no evidence of anthropogenic disturbance. Eight cores were collected into lengths of tube, each ~100 mm long and 38 mm in diameter, sealed at each end and transported intact to the Australian National University (ANU) laboratories for Pu analysis using Accelerator Mass Spectrometry (AMS).



Figure 1. The Maralinga test site area. Crosses indicate approximate sampling locations used in the present work. Sites where weapons tests were carried out are indicated with solid circles, and safety trial locations by open circles. Note that the Taranaki site was also used for the 25 kT Antler round 3 weapons test.

Sample preparation and measurement details were based on those described in [5]. Briefly, each core was first sectioned into ~20 mm depth increments, and the material from each increment thoroughly mixed to homogenise the sample. The surface sample from each core was then analysed for ¹³⁷Cs content at the CSIRO Land and Water laboratories in Canberra using HPGe detectors to count the 662 keV γ -rays that arise from ¹³⁷Cs decay [6]. Typically, a 4 pg (i.e. 10¹⁰ atoms) ²⁴²Pu spike was then added to a 10 g aliquot of each homogenised sample and the plutonium leached with hot nitric acid for ~24 hours. The acid solution was then purified using ion exchange columns and the ²³⁹Pu and ²⁴⁰Pu concentrations and ²⁴⁰Pu/²³⁹Pu ratio determined with AMS using the 14UD pelletron accelerator at the ANU [7, 8]. The purified Pu fractions of a few selected replicate samples were separated off for electro-deposition [9] and analysis by α -spectroscopy at the CSIRO laboratories.

3 Results

The ²³⁹Pu concentrations, as determined by the AMS measurements, are presented in figure 2 and table A.1as a function of soil depth for all eight cores. The profiles for samples T001 - T004 show maximum concentrations typical of those seen elsewhere in Australia [10-13], but which peak closer to the surface and only penetrate to around a quarter of the usual depth. This latter behaviour can most likely be attributed to the dry, desert-like environment of the area. Samples T006 - T008, on the other hand, show progressively larger surface maxima, reaching over 400 mBq ²³⁹Pu/g at site T008. These profiles also peak very close to the surface and have relatively small penetration depths. The depth profile for sample T005 however, shows very little Pu, and suggests that the soil at this site has either been removed or buried by the addition of >10cm of topsoil: given the extensive activities but poorly documented early history of the area it is quite plausible that this was the case (see below).



Figure 2. Soil depth profiles from alongside the Maralinga test site access road. Note the log scale for sites T007 and T008.

The ²⁴⁰Pu/²³⁹Pu atom ratio for each sample location, determined from the weighted average of all depth increments, is listed in table 1 and plotted in figure 3 as a function of distance from the Taranaki test site. Our results for sites T007 and T008 are in good agreement

Sample	²³⁹ Pu (mBq/g) ^a	²⁴⁰ Pu/ ²³⁹ Pu atom ratio ^b	²³⁹⁺²⁴⁰ Pu (mBq/g) ^a		137 Cs (mBq/g) ^{a,c}	¹³⁷ C/ ²³⁹⁺²⁴⁰ Pu activity
			AMS	a-spec.		ratio ^{a,c}
T001	0.071 ± 0.005	0.194 <u>+</u> 0.018	0.123 ± 0.007		3.15 ± 0.16	25.7 ± 1.9
T002	0.153 ± 0.008	0.144 ± 0.008	0.236 ± 0.010	0.244 ± 0.008	6.81 ± 0.21	28.9 ± 1.4
T003	0.109 ± 0.007	0.136 ± 0.008	0.154 ± 0.009		3.15 ± 0.15	20.4 ± 1.5
T004	0.103 ± 0.005	0.133 ± 0.009	0.147 ± 0.006		2.59 ± 0.16	17.6 ± 1.3
T005	0.007 ± 0.001	0.107 ± 0.031	0.010 ± 0.001		b.d.l. ^d	
T006	0.328 ± 0.014	0.098 ± 0.005	0.444 ± 0.016	69 ± 10^{e}	8.18 ± 0.24	18.4 ± 0.8
T007	1.734 ± 0.073	0.0412 ± 0.002	1.981 <u>+</u> 0.074		0.89 <u>+</u> 0.11	0.45 ± 0.06
T008	408 <u>+</u> 17	0.0532 ± 0.002	488 <u>+</u> 17		3.89 ± 0.20	0.0080 ± 0.0004
$T008^{f}$	$0.136 \pm 0.006^{\rm f}$	$0.066 \pm 0.005^{\rm f}$	$0.169 \pm 0.007^{\rm f}$	$0.956 \pm 0.026^{\rm f}$		

 Table 1. Plutonium and caesium activities and ratios in surface soils at Maralinga. Uncertainties include systematic and statistical errors and correspond to one standard deviation.

^a Uppermost depth increment

^b Average ratio for entire core

^c As measured on 31 Mar 2011

^d Below detection limit

^e The large error here is due to tailing of the very large ²³⁹⁺²⁴⁰Pu peak propagating into the peak of the ²⁴²Pu tracer

^f Lowermost (77 – 95 mm) depth increment only

with high resolution γ -ray spectrometry measurements (atom ratio: 0.027-0.046) reported for Taranaki [14, 15]. These earlier atom ratios were determined prior to the 1995–2000 (MARTAC) clean up when soil activities were considerably higher than is presently the case. Replicate samples also measured by AMS agreed within error for sites T001 – T006, as does the result for ²³⁹⁺²⁴⁰Pu from a replicate sample at site T002 measured using α -spectroscopy. The AMS replicates and α -spectroscopy results for sites T006-T008 however, show variability that is greater than statistically expected. This is discussed in more detail below.

Table 1 also lists the ²³⁹Pu and ¹³⁷Cs concentrations and the ¹³⁷Cs/²³⁹⁺²⁴⁰Pu activity ratios for the uppermost depth increment from each site. Notably, the¹³⁷Cs concentration is relatively constant between sites T001 and T008 and shows no obvious trend with distance from the test site, while that for ²³⁹Pu increases by a factor of more than 5000, with the maximum close to the Taranaki test and trial site.



Figure 3. The 240 Pu/ 239 Pu atom ratio for sites T001 – T008.

4 Discussion

Figure 2 shows the surface ²³⁹Pu concentration increases by more than 3 orders of magnitude between sites T001– T004 and site T008, which is close to the Taranaki test site. Figure 3 shows the ²⁴⁰Pu/²³⁹Pu ratios for sites T006– T008 are well below the global fallout average of ~0.18 [16, 17], and also below the Australian average of ~0.14 [13]. The high concentrations and low ratios, which at sites T007 and T008 yield ²⁴⁰Pu/²³⁹Pu ratios consistent with undetonated weapons material [18], indicate that the Pu at these two sites is overwhelmingly derived from the safety trials which dispersed the weapons material using conventional explosives. The data for site T006 also suggests the presence of trials material, with ratios and concentrations between the two end members.

Although site T007 is the closest to the Taranaki test site, the Pu surface concentrations and inventory are significantly lower than at site T008. Furthermore, the ¹³⁷Cs concentration is also low at site T007, and the data therefore imply that material may also have been removed from this site. Site T007 is ~480 m south-southwest of the detonation point, and while it appeared undisturbed at the time of sampling, is only ~10 m from an area where earthworks have clearly taken place. Site T008, on the other hand is over 250 m from the nearest earthworks. Excluding sites T005 and T007, where the ¹³⁷Cs concentration shows similar behaviour to that of Pu; i.e. is unusually low, the relatively constant ¹³⁷Cs concentration across the remaining sites (1-10 mBq¹³⁷Cs per gram of soil) is similar to that reported elsewhere in Australia [5,19]. This indicates that local and regional fallout from the Australian weapons tests are not major contributors to the total Cs fallout at the sampling locations.

The ²³⁹Pu and ²³⁹⁺²⁴⁰Pu inventories for the eight sites, based on the AMS data, are listed in Table 2. Sites T001-T006 have inventories that are at the lower end of

those typically observed in Australia, however there is not yet enough data to make meaningful comparisons. The inventories of sites T001-T004 show quite substantial variability from site to site, and at present we have no definitive explanation for this observation. The inventory at site T005 is particularly low, however there is some evidence from aerial photographs that indicate this site is within a few meters of past road works and the possibility that the soil surface has been disturbed cannot be ruled out, as is suggested above. Sites T007 and T008 have inventories that are far larger than observed elsewhere in Australia. Taken together the ¹³⁷Cs concentrations, ²⁴⁰Pu/²³⁹Pu ratios and Pu inventories indicate the dominant source of Pu at these sites is derived from local fallout derived from the safety trials of the 1960's. Further supportive evidence for this conclusion is provided by the ²⁴⁰Pu/²³⁹Pu and ¹³⁷Cs/²³⁹⁺²⁴⁰Pu ratios, which are at a minimum at sites T007 (Taranaki) and T008, respectively. Both ratios return to values consistent with global fallout within 25

km of the test site. The ${}^{137}Cs/{}^{239+240}Pu$ activity ratios at sites T001 and T002 are in good agreement with the (decay corrected) fallout ratios of 23.5 reported for Antarctica [20], our value of 24.7 ± 1.4 for the Herbert River, Australia [5], and with an estimated ratio of ~28.6, based on ${}^{90}Sr$ deposition and the total annual world-wide activity releases for ${}^{90}Sr$, ${}^{137}Cs$ and ${}^{239+240}Pu$ [21]. Sites T007 and T008 give ${}^{137}Cs/{}^{239+240}Pu$ activity ratios that are much lower but are in good agreement with, and show similar variability to, those reported for Maralinga [14, 15]. The other sites show ${}^{137}Cs/{}^{239+240}Pu$ activity ratios intermediate between the global fallout and weapons test site values, reflecting the diminishing influence of the trials as the distance from the testing area increases.

Table 2. Plutonium inventories from Maralinga, determined using the measured densities ($\sim 1.6 \text{ g.cm}^3$) for each increment.

Sample	²³⁹ Pu Inventory	²³⁹⁺²⁴⁰ Pu Inventory			
Sample	(mBq/cm^2)	(mBq/cm^2)			
T001	0.421 ± 0.005	0.729 ± 0.007			
T002	1.43 ± 0.01	2.17 ± 0.01			
T003	1.08 ± 0.01	1.62 ± 0.01			
T004	0.496 ± 0.007	0.750 ± 0.008			
T005	0.040 ± 0.001	0.056 ± 0.001			
T006	1.90 ± 0.02	2.60 ± 0.02			
T007	21.2 ± 0.1	24.5 ± 0.1			
T008	1241 ± 17	1483 ± 17			

The variability in the replicate samples at sites T006 – T008 suggests the sampled material is not homogenous. However, at site T006 the disagreement between the AMS and α -spectroscopy replicates is greater than a factor of 200; and at site T008 replicates differ by more than a factor of 5. It is unlikely that such differences could be the result of a simple failure to mix the sample material adequately prior to separation of each sample aliquot. The presence of sub-millimetre plutoniumbearing particles is however, well documented at Maralinga [4], and given the proximity of these sampling locations to the Vixen B safety trials, it is suspected that

such "hot particles" are the likely source of the discrepancies observed in the present data.

5 Conclusion

The ¹³⁷Cs, ²³⁹Pu and ²⁴⁰Pu data from sites T001 and T002 indicate that tropospheric fallout deposition at distances greater than ~ 25 km to the south of the test area has been negligible. Our data are consistent with that of global fallout in Australia, and are also consistent with fallout distribution maps from the Australian tests [22, 23] which indicate the bulk of the regional fallout from the Maralinga weapons detonations was to the north and east of the test site. The data from sites T006-T008 however, show clear signatures of local fallout Pu derived from the safety trials, with indications of the presence of "hot particles" of Pu bearing material. The 137Cs/239+240Pu ratio, which changes by over 3 orders of magnitude across the present sampling sites, appears to provide the most sensitive probe of the presence of local or regional fallout, however only the ²⁴⁰Pu/²³⁹Pu ratio provides definitive information on the source of the Pu. Excluding sites T005-T008, the ²³⁹Pu inventories are at the low end of the range typically observed in Australia, and show greater variability than expected. Importantly from the point of view of using Pu as a tracer of soil movement, the contribution to the Pu inventories from the weapons tests and safety trials does not appear to be significant at distances greater than 10 km to the south of the test site area. The variability in the inventories however, could be important and needs to be understood, as does the extent of the impacted area in other directions. Further work is in planning to address these issues.

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Table A.1. Plutonium concentrations, as determined by AMS, for the samples from the present work. Replicate samples are indicated with a *.

Sample	Depth (cm)	²³⁹ Pu	ı (n	nBq/g)	²⁴⁰ Pu	(m	Bq/g)
T001_A	0 - 2.7	0.071	±	0.005	0.051	±	0.005
T001_B	2.7 - 4.7	0.016	\pm	0.002	0.012	\pm	0.002
$T001_B^*$	2.7 - 4.7	0.015	±	0.001	0.014	\pm	0.002
T001_C	4.7 - 6.7	0.004	±	0.001	0.003	±	0.002
T001_D	6.7 - 8.7	0.002	\pm	0.001	0.0013	\pm	0.0006
T001_E	8.7 - 10.7	0.0014	±	0.0004	0.0005	\pm	0.0004
T002_A	0 - 2.9	0.153	±	0.008	0.083	±	0.006
T002_B	2.9 - 4.8	0.105	\pm	0.005	0.055	\pm	0.004
$T002_B^*$	2.9 - 4.8	0.096	±	0.005	0.043	±	0.003
T002_C	4.8 - 7.1	0.050	±	0.004	0.025	±	0.004
T002_D	7.1 - 9.0	0.023	±	0.002	0.012	±	0.002
T002_E	9.0 - 11.0	0.021	±	0.002	0.012	±	0.002
T003 A	0 - 2.8	0.109	±	0.007	0.045	±	0.005
T003_B	2.8 - 4.8	0.158	±	0.008	0.081	±	0.005
$T003_B^*$	2.8 - 4.8	0.106	±	0.005	0.060	±	0.004
T003_C	4.8 - 6.9	0.046	±	0.004	0.025	±	0.004
T003_D	6.9 – 8.9	0.028	±	0.002	0.015	±	0.002
T003_E	8.9 - 10.9	b	.d.1		b	.d.1	
T004_A	0 - 0.6	0.094	±	0.005	0.042	±	0.003
$T004_A^*$	0 - 0.6	0.115	\pm	0.006	0.045	\pm	0.004
T004_B	0.6 - 2.6	0.080	±	0.004	0.040	±	0.003
$T004_B^*$	0.6 - 2.6	0.069	±	0.005	0.034	±	0.004
T004_C	2.6 - 4.6	0.014	\pm	0.001	0.010	\pm	0.002
$T004_C^*$	2.6 - 4.6	0.015	±	0.001	0.008	±	0.001
T004_D	4.6 - 6.6	0.0040	±	0.0005	0.0036	\pm	0.0008
$T004_D^*$	4.6 - 6.6	0.0028	±	0.0007	0.0026	±	0.0009
T004_E	6.6 - 8.6	0.0022	±	0.0005	0.0013	±	0.0006
$T004 E^*$	6.6 - 8.6	0.0034	±	0.0006	0.0020	±	0.0007
T004_F	8.6 - 10.6	0.0020	±	0.0005	0.0014	±	0.0006
$T004_F^*$	8.6 - 10.6	0.0023	±	0.0007	0.0006	±	0.0006

Sample	Depth (cm)	²³⁹ Pu (mBq/g)		²⁴⁰ Pu (mBq/g)			
T005_A	0 - 2.5	0.0062	\pm	0.0009	0.0021	±	0.0007
$T005_A^*$	0 - 2.5	0.0081	±	0.0009	0.0030	±	0.0007
T005_B	2.5 - 4.4	0.0013	±	0.0003	0.0006	±	0.0004
$T005_B^*$	2.5 - 4.4	0.0017	±	0.0004	0.0005	±	0.0003
T005_C	4.4 - 6.4	0.0005	±	0.0002	b.d.1		
$T005_C^*$	4.4 - 6.4	0.0001	\pm	0.0001	b.d.1		
T005_D	6.4 - 8.4	0.0003	±	0.0001	b.d.l		
T005_E	8.4 - 10.4	0.0015	\pm	0.0004	0.0012	\pm	0.0006
T006_A	0 - 2.6	0.328	±	0.014	0.116	±	0.006
T006_B	2.6 - 4.2	0.152	±	0.007	0.057	±	0.003
$T006_B^*$	2.6 - 4.2	0.138	±	0.007	0.060	±	0.004
T006_C	4.2 - 6.2	0.068	±	0.004	0.024	±	0.002
T006_D	6.2 - 8.2	0.0117	\pm	0.0009	0.0054	±	0.0008
T006_E	8.2 - 10.3	0.017	±	0.001	0.006	±	0.001
T007_A	0 - 2.7	1.734	±	0.073	0.247	±	0.012
T007_B	2.7 - 4.3	1.424	±	0.060	0.218	±	0.009
$T007_B^*$	2.7 - 4.3	1.055	±	0.045	0.153	±	0.009
T007_C	4.3 - 6.2	1.446	±	0.062	0.217	±	0.011
T007_D	6.2 - 8.2	1.001	±	0.042	0.165	±	0.009
T007_E	8.2 - 10.3	0.335	±	0.016	0.053	±	0.006
T008_A	0 - 2.0	408	±	17	80	\pm	3
T008_B	2.0 - 3.9	4.76	\pm	0.20	0.92	±	0.04
$T008_B^*$	2.0 - 3.9	18.1	\pm	0.8	2.80	±	0.12
T008_C	3.9 - 5.7	1.739	\pm	0.073	0.329	±	0.018
T008_D	5.7 – 7.7	0.386	±	0.017	0.082	±	0.005
T008_E	7.7 – 9.5	0.136	±	0.006	0.033	±	0.003