

1 **This manuscript contextually corresponds with the following paper:**

2 Ónodi, G., Botta-Dukát, Z., Kröel-Dulay, Gy., Lellei-Kovács, E. and Kertész, M. (2018) Reduction in primary
3 production followed by rapid recovery of plant biomass in response to repeated mid-season droughts in a
4 semiarid shrubland. *Plant Ecology*, 219(5), 517-526. doi: 10.1007/s11258-018-0814-6

5 **Availability of the original paper and the electronic supplementary material:**

6 <http://link.springer.com/article/10.1007/s11258-018-0814-6>

7

8 **Reduction in primary production followed by rapid recovery of plant biomass in response to repeated mid-**
9 **season droughts in a semiarid shrubland**

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22 **Acknowledgements**

23 This study was funded by the VULCAN project (EU FP5 grant EVK2-CT-2000-00094)), the INCREASE project
24 (EU FP7 grant 227628), the Hungarian Scientific Research Fund (OTKA K112576), and the National Research,
25 Development and Innovation Office (GINOP 2.3.3-15-2016-00019). We are grateful to the Kiskunság National Park
26 (Hungary) for the support of our field work. The authors thank the anonymous reviewers of this manuscript for their
27 valuable comments which have helped us to improve the quality of the paper.

28 Abstract

29 The frequency and severity of extreme weather events, including droughts, are expected to increase due to the
30 climate change. Climate manipulation field experiments are widely used tools to study the response of key
31 parameters like primary production to the treatments. Our study aimed to detect the effect of drought on the
32 aboveground biomass and primary production both during the treatments as well as during the whole growing
33 seasons in semiarid vegetation.

34 We estimated aboveground green biomass of vascular plants in a Pannonian sand forest-steppe ecosystem in
35 Hungary. We applied non-destructive field remote sensing method in control and drought treatments. Drought
36 treatment was carried out by precipitation exclusion in May and June, and was repeated in each year from 2002. We
37 measured NDVI before the drought treatment, right after the treatment, and at the end of the summer in 2011 and
38 2013.

39 We found that the yearly biomass peaks, measured in control plots after the treatment periods, were decreased or
40 absent in drought treatment plots, and consequently, the aboveground net primary production was smaller than in the
41 control plots. At the same time, we did not find general drought effects on all biomass data. The studied ecosystem
42 proved resilient, as the biomass in the drought treated plots recovered by the next drought treatment. We conclude
43 that the effect of drought treatment can be overestimated with only one measurement at the time of the peak biomass,
44 while multiple within-year measurements better describe the response of biomass.

45

46 Keywords

47 Aboveground Net Primary Production; Climate change experiment; Drought; Multi-seasonal biomass estimation;

48 NDVI; Semiarid shrubland

49 Introduction

50

51 The ongoing climate change increases the frequency and severity of extreme weather events (IPCC 2013). One of the
52 key ecological parameters affected by changing climate is primary production. Extreme drought events have been
53 shown to reduce primary production in Europe (Ciais et al. 2005) and the increase in drought frequency negatively
54 affected grassland production worldwide (Zhao and Running 2010). At plot scale, climate manipulation experiments
55 are particularly effective way to study ecological consequences of climate change (Wu et al. 2011), especially long-
56 term multi-site field experiments (Kröel-Dulay et al. 2015).

57 In climate manipulation experiments, effects of precipitation treatments on vegetation performance are often
58 estimated once per year, during or right after seasonal treatments (Köchy and Wilson 2004; Brancaleoni et al. 2007;
59 Mänd et al. 2010; Byrne et al. 2013; Tielbörger et al. 2014). However, the effects emerging in the treatment period
60 may change during the rest of the growing season. Thus, additional within-year measurements may provide further
61 information about the changes in plant biomass, either targeting legacy effects in long-term experiments before the
62 yearly treatments, or aiming to follow the relaxation period after. Aboveground plant biomass may recover after the
63 drought period in semiarid grassland and shrubland communities adapted to high variability in precipitation
64 (Miranda et al. 2011). Even if drought strongly reduces aboveground biomass, it can recover quickly during the late
65 summer, because belowground parts are less affected by the drought (Shinoda et al. 2010). Therefore, treatment
66 effects should be checked multiple times during the growing season.

67 Conducting multiple within-year measurements on aboveground plant biomass is one of the major challenges in
68 long-term field experiments. For this purpose, application of non-destructive sampling methods is suggested (Gamon
69 et al. 1995). When effects on primary production is in focus, field spectroscopy is one of the feasible solutions for
70 estimating aboveground biomass, or leaf area index (Goodin and Henebry 1997; Pontauiller et al. 2003; Mulla 2013;
71 Nestola et al. 2016; Ónodi et al. 2017a). The Normalized Differential Vegetation Index (NDVI) obtained by field
72 spectroscopy is an accurate proxy for aboveground biomass estimations (Gamon et al. 1995; Ónodi et al. 2017b).
73 However, it is rarely applied in multi-seasonal measurements in long-term ecological experiments, but see (Goodin
74 and Henebry 1997; Filella et al. 2004; Boelman et al. 2005; Wang et al. 2016; Nestola et al. 2016).

75 Our goal was to study the effect of two-month drought treatments (i.e. rain exclusions) on the aboveground biomass,
76 and primary production via proxies (NDVI, and the sum of positive NDVI increments accordingly) both during the
77 treatments as well as during the whole growing seasons. We applied field spectroscopy to observe within-year
78 changes in aboveground green biomass of vascular plants in the semiarid Kiskunság forest-steppe vegetation (Lellei-
79 Kovács et al. 2008a; Kröel-Dulay et al. 2015). According to climate change scenarios for Hungary, the frequency of
80 extreme dry and wet years is expected to increase in the study region (Bartholy et al. 2003; Bartholy and Pongrácz
81 2007).

82 Our specific questions were as follows: What is the effect of drought treatment on the aboveground biomass estimate
83 (NDVI) in different seasons in a long-term climate manipulation experiment? What is the effect of drought treatment
84 on annual primary production and production of different seasons, i.e. treatment and post-treatment changes of
85 biomass calculated as the sum of positive NDVI increments?

86

87

88 Methods

89

90 *Study site and experimental design*

91 Our study site is part of the EU FP5 VULCAN and the EU FP7 INCREASE projects (Beier et al. 2004; Peñuelas et
92 al. 2007; Kröel-Dulay et al. 2015) representing the continental semi-arid forest-steppe vegetation of Central Europe
93 in the multi-site surveys. The site is in the Kiskunság National Park (N46°52', E19°25'), in a Pannonian sand forest-
94 steppe vegetation mosaic (Lellei-Kovács et al. 2008b) of high plant diversity and nature protection value (Fekete et
95 al. 2002; Molnár et al. 2012). In our study plots, we sampled open grassland patches where also shrubby root suckers
96 of white poplar (*Populus alba*) occurred. The soil is calcareic arenosol which enhances the semi-desert character of
97 the vegetation. Climate of the study area is temperate continental. The vegetation period starts in April and finishes
98 in October. Based on regional 30-years average values (1961-1990), mean annual temperature is 10.4°C, mean
99 monthly temperature ranges from -1.9 °C in January to 21.1 °C in July, while mean annual precipitation is 505 mm
100 with a peak in June (Kovács-Láng et al. 2000).

101 The climate manipulation experiment described by Lellei-Kovács et al. (2008) were conducted in three replications
102 of controls and drought treatments. The vegetation of the replicates differed from each other in the abundance of
103 poplar shoots, but within each control - drought treatment pair, the plots were similar in this respect. Plot size was 4
104 m x 5 m, and the experiment started in 2002. Automatically controlled rain exclusion during May and June was
105 applied as drought treatment.

106

107 *Sampling design and data collection*

108 In our study, we estimated aboveground green biomass of vascular plants (referred as plant biomass hereinafter) by
109 means of NDVI in the control and drought treatment plots in 2011 and 2013. The planned 2012 measurements had to
110 be cancelled for technical reasons. We applied a multi-seasonal non-destructive plant biomass sampling method (Fig.
111 1). As the first step, we measured a baseline of the plant biomass at the beginning of the vegetation period of the first
112 year (M0 in Fig. 1), when plant activity is still very low as the soil surface is covered by litter. Afterwards, we
113 estimated the plant biomass three times a year: at the turn of April and May (before treatment measurement, M1), at
114 the turn of June and July (after treatment measurement, M2) and after a relaxation period at the turn of August and
115 September (end-of-summer measurement, M3). Precipitation was monitored in all plots separately. Rain exclusion
116 data were calculated as differences between average values collected in the three control and the three drought
117 treatment plots. Annual and monthly precipitation data were calculated as average values of the control plots (Online
118 Resource 1).

119 In 2011, drought treatment started at 30 April and ended at 07 July. During this period we excluded 88.8 mm out of
120 the 112.8 mm precipitation (78.7%). Annual precipitation in 2011 was 408.0 mm. Dates of the biomass estimation
121 measurements were: (M0) 01 April, (M1) 02 May, (M2) 28 June, and (M3) 30 August.

122 In 2013 drought treatment started later, it was conducted between 15 May, and 30 June. During this period we
123 excluded 111.7 mm out of the 118.4 mm precipitation (94.4%). However, 30.6 mm rain was not excluded during the
124 first two weeks of May. Annual precipitation in 2013 was 597.8 mm. Dates of the biomass estimation measurements
125 were: (M1) 29 April, (M2) 10 July, (M3) 04 September. Thus, 111.7 mm precipitation out of 149.0 mm (75.0%) was
126 excluded between the M1 and M2 measurements.

127 We estimated the amount of plant biomass by non-destructive field spectroscopy techniques in each measurement
128 event (Online Resource 2). We applied a portable Cropscan MSR87 multispectral radiometer (Cropscan, Inc.,
129 Rochester, MN) for measuring incoming and reflected light intensity. We used an aluminium frame for moving the
130 sensor above the plots at a height of 1.5 meter. In each of the six plots (three control and three drought treated plots)
131 we sampled twelve subplots arranged in a 3 x 4 grid. The area of the circular subplots were 0.44 m² (diameter: 0.75
132 m), and the distance between centre points of the neighbouring subplots were 1 meter. The frame allowed us to
133 repeat the sampling of each subplot at the same position during the different measurement events. We calculated
134 NDVI (Rouse et al. 1974) values based on the measured light intensity data at red (660 nm) and near infrared (810
135 nm) wavelengths. According to our previous investigation, NDVI provides an accurate proxy for plant aboveground
136 green biomass estimation in the studied vegetation complex (Ónodi et al. 2017b). Thus, differences in NDVI values
137 are interpreted as differences in aboveground green biomass henceforth. Baseline NDVI data collected at the first
138 (M0) measurement event ($NDVI_{AVG \pm SE} = 0.205 \pm 0.003$) provides an empirical zero point for calculation of
139 increments of yearly plant biomass. The 0.205 average is in agreement with our long term experience (Ónodi et al.
140 2017a) and the low standard error value we got shows that the baseline is not sensitive to the differences in litter
141 cover and composition. We consider the increase in NDVI as proxy for aboveground primary production. Thus, we
142 count the sum of the positive increments as proxy for the annual aboveground net primary production (ANPP),
143 according to Sala and Austin (2000).

144

145 *Statistical analyses*

146 In the first analysis, dependence of the measured NDVI values on treatments, years and measurement events and
147 their interactions (including three-way interaction) were analysed by fitting linear mixed models (Zuur et al. 2009).
148 In this analysis, subplots nested in plots were random factors in the model, since simplification of the random part
149 would result in higher AIC values. In order to avoid this, while not losing the inside-plot variation information, we
150 applied the nested design, in line with Colegrave and Ruxton (2018). Significance of fixed factors was tested by
151 maximum likelihood ratio tests (Zuur et al. 2009).

152 The following null-hypotheses were tested using contrasts. The hypothesis 1 refers to the measured NDVI values in
153 each sampling date. The hypotheses 2-4 refer to the changes of the NDVI values in time, and they are arranged into

154 pairs where (a) probe whether there is significant increase or decrease in the given time span at the level of a certain
155 treatment, and (b) compare the changes between control and drought.

- 156 1 NDVI values in control and drought treatments do not differ (tested in each measurement event);
- 157 2 (a) changes in NDVI between M1 and M2 (hereafter called treatment change) do not differ from zero;
- 158 2 (b) treatment changes do not differ between control and treatment plots;
- 159 3 (a) changes in NDVI between M2 and M3 (hereafter called post-treatment change) do not differ from zero;
- 160 3 (b) post-treatment changes do not differ between control and treatment plots;
- 161 4 (a) changes in NDVI between M1 and M3 (hereafter called whole-season change) do not differ from zero;
- 162 4 (b) whole-season changes do not differ between control and treatment plots.

163 P-values were corrected by single-step procedure (Hothorn et al. 2008) to avoid their inflation due to multiple
164 testing.

165 In the second analysis, the sum of positive NDVI increments for each subplot, as a proxy for ANPP was the
166 dependent variable, while year and treatment were fixed factors in the model. The random part was the same as in
167 the previous analysis. Significance of fixed factors was tested by series of maximum likelihood ratio tests (Zuur et al.
168 2009).

169 All calculations were done in R statistical environment (R Core Team 2017) using nlme (Pinheiro et al. 2017),
170 multcomp (Hothorn et al. 2008) and lsmeans (Lenth 2016) add-on packages for fitting models, doing post-hoc tests
171 and drawing figures, respectively.

172

173

174 Results

175

176 We found significant three-way (treatment \times year \times measurement event) interaction (likelihood ratio = 12.875, d.f. =
177 2, $p = 0.002$) on the NDVI data, thus treatment effects had to be tested in each sampling time by post-hoc test using

178 contrasts. Post-hoc tests showed that drought treatments significantly affected NDVI values after treatment
179 measurements (M2), in both years and also at the end-of-summer measurement (M3) in 2013 (Table 1, upper six
180 rows, NDVI with C vs. D comparisons). However, the differences are not significant in the other sampling times,
181 even if NDVI values were higher in control plots in all six measurement events (Fig. 2, positive estimates in Table
182 1).

183 Regarding the increase or decrease of plant biomass between the measurement events, we found significant increase
184 of NDVI values during the treatment change (M2-M1 in Fig. 3; M1 vs. M2 in Table 1) and its significant decrease
185 during the post-treatment change (M3-M2 in Fig. 3; M2 vs. M3 in Table 1) except the drought treatment in 2011.
186 There were no significant changes in the NDVI values in the whole-seasons (M3-M1 in Fig. 3; M1 vs. M3 in Table
187 1).

188 Regarding the treatment effects on the changes of plant biomass during the treatment periods, we found significantly
189 higher biomass increase in the control than in the drought treatment in both years (M2-M1 in Fig. 3; C vs. D
190 comparisons of Δ NDVI (M2-M1) in Table 1). In the post-treatment periods the plant biomass decrease was
191 significantly greater in the control than in the drought treatment in 2011 (M3-M2 in Fig. 3; C vs. D comparisons of
192 Δ NDVI (M3-M2) in Table 1).

193 In the analysis of the sums of the positive increments as proxy variables for ANPP (Fig. 4), the two-way interaction
194 between treatment and year proved to be significant (likelihood ratio = 8.809, $df = 1$, $p = 0.003$). Effect of treatment
195 (likelihood ratio = 16.046, $df = 1$, $p < 0.001$) was significant in both years (2011: $z = 2.224$, $p = 0.040$; 2013: $z =$
196 3.823 , $p < 0.001$), however it was stronger in 2013 ($t = 3.018$, $df = 70$, $p = 0.004$).

197

198

199 Discussion

200

201 Based on multiple NDVI measurements, we found consistent negative effects of drought treatment both on yearly
202 peak plant biomass and on the ANPP, in line with Estiarte et al. (2016) and Reinsch et al. (2017). Drought treatment

203 decreased the biomass in both years in June (M2 in Fig. 2), and at the end of summer in 2013 (M3 in Fig. 2).
204 However, NDVI values showed no overall significant treatment effect. Treatment and measurement event had
205 interactive effects on biomass, similarly to Hoover et al. (2014) who also found both significant year effect and
206 significant year \times drought treatment interactions in their two-year extreme drought and heatwave experiment in
207 central U.S. grassland. In our study, we showed that besides the significant treatment effect in June, the plant
208 biomass did not differ in the treated and the control plots at the beginning of the studied vegetation periods (M1 in
209 Fig. 2) and at the end of summer in 2011 (M3 in Fig 2). The treatment and post-treatment changes of NDVI values
210 (Fig. 3) show also strong effects of drought. While the biomass increased markedly in the control plots, we did not
211 find increment in drought plots in both years (M2-M1 in Fig. 3), only in 2013, when it was significantly less than in
212 the control. Furthermore, the post-treatment biomass decrease was also less in 2011 compared to the control (M3-M2
213 in Fig. 3). Consequently, the estimated ANPP decreased in the case of our drought treatment (Fig. 4), similarly to the
214 findings of most studies in arid or semiarid ecosystems (Beier et al. 2012).

215 The NDVI values responded sensitively to the treatments. The detected treatment effects depended on the relative
216 timing of treatments and measurement events. Delay of starting the treatment resulted in detection of significant
217 biomass increase during the treatment period also in the drought treatment plots in 2013 (M2-M1 in Fig. 3), even if
218 this increase was significantly smaller compared to that in the control plots. We assume that the reason for the
219 biomass increase is that the study site had 30.6 mm precipitation during the two weeks long delay period, which
220 promoted significant vegetation growth also in the drought treated plots.

221 We applied multiple sampling of biomass in a year in order to gain deeper knowledge on the pattern of plant biomass
222 changes in grasslands. First of all, multiple biomass estimates are required for monitoring the amount of biomass in
223 the course of the vegetation season, revealing which periods of the growth season were affected by the treatment. We
224 found that drought eliminated peak biomass in June (M2 in Fig. 2, as well as M2-M1 in Fig. 3), characteristic for the
225 open sand grasslands (Kovács-Láng 1974), while it has slight or no effect at early and late season stages. On the
226 other hand, multiple estimates are required for assessing the primary production following the method of the sum of
227 positive increments in plant biomass (Sala and Austin 2000). This allows a more reliable comparison of ANPP than
228 estimation only using a measurement of peak biomass (Scurlock et al. 2002). The method we applied is based on the
229 calculation of the positive increments between repeated measurement events and it needs an estimate for the base-

230 line. For this purpose, we executed a sampling plan which covers the starting point (M0) and three measurement
231 events (M1, M2, M3) during the vegetation period. Application of the sum of positive increment method allowed us
232 to take the biomass at the time of all the three measurement events into account. Besides the already mentioned
233 drought effect (Fig. 4), we detected that difference between ANPPs in control and drought plots was higher in 2013.
234 This is in accordance with the fact that spring precipitation in 2013 was much higher than in 2011 (Fig. 1), which
235 resulted in higher peak biomass in the control. Furthermore, the late summer drought in 2013 (Fig. 1) prevented the
236 regeneration in the drought treated plots. However, our ANPP estimate, being mostly governed by M2-M1
237 difference, is not sensitive to the regeneration of the plant biomass by the time of the next treatment.

238 We emphasize the importance of biomass measurements multiple times in the growing season in an experiment
239 where yearly drought treatments are applied, in contrast to most of the studies from which only annual data are
240 published. Our results supplement the findings of Estiarte et al. (2016) and Reinsch et al. (2017) who got consistent
241 drought effect applying one annual biomass estimation by point-intercept method right after the treatment period.
242 Our study reveals that in late-successional grassland-shrubland ecosystems, like ours (Kröel-Dulay et al. 2015),
243 compensation may occur before the next drought. With one measurement per year we could only detect the effect of
244 drought treatment on the peak biomass. Although our investigation started in the 10th year of the climate change field
245 experiment, we could not observe general treatment effect on the biomass taking three annual measurements into
246 consideration. While both summer drought treatments caused significant differences in NDVI by the end of the
247 treatment periods, among four before-treatment and end-of-season measurements only one showed significant
248 treatment effect. Furthermore, we found no whole-season (M3-M1) differences in NDVI between the control and
249 treated plots. Thus, the studied ecosystem proved drought resistant both in terms of Vicente-Serrano et al. (2013),
250 reacting to the drought only at a short time scale, and according Hoover et al. (2014), recovering by the end of the
251 season. This resistance is in agreement with the findings of Tielbörger et al. (2014) in long-term experiments in
252 Mediterranean shrublands. We suppose that the main reason for rapid recovery of biomass in the studied vegetation
253 mosaic is that the drought treatment did not lead to regime shift which occurs after strong disturbance events (Kröel-
254 Dulay et al. 2015). The presence of poplar shoots might contribute to the late season recovery of the grass layer after
255 drought through shading, in line with the findings of Erdős et al. (2014). In contrast with our results, in the post-fire
256 successional vegetation of the Catalanian VULCAN site, Filella et al. (2004) found long-term around-the-year

257 divergence in biomass (also estimated by NDVI) due to drought treatment from the first year after the start of the
258 experiment, which is in accordance to the findings of Kröel-Dulay et al. (2015) in early-successional ecosystems.

259 According to Ónodi et al. (2017a), drought can temporarily change the NDVI - biomass relationship. Several
260 structural and physiological changes may result in lower NDVI readings, such as decreased specific leaf area, light
261 absorbance, and green biomass to standing biomass ratio because of drought treatment (Cornic and Massacci 1996).
262 However, Filella et al. (2004) and Mänd et al. (2010) found NDVI a reliable proxy for biomass estimation across
263 treatments, seasons, and sites in the same experimental design. As there were not remarkable long-term
264 compositional changes in the vegetation due to the drought treatment at our site (Kröel-Dulay et al. 2015), we
265 conclude that the lower NDVI value after drought treatments indicated less aboveground green biomass because of
266 increased drying and reduced sprouting.

267 The loss of biomass peak in consecutive years due to drought could lead to severe changes in the carbon budget of
268 the ecosystem. Nagy et al. (2007) found that net ecosystem exchange (NEE) in semi-arid grasslands of the same
269 ecosystem can turn to positive (i.e. carbon releasing) in dry years. However, according to Pintér et al. (2008), the
270 NEE in the same vegetation type is negative (i.e. carbon accumulating) in years of normal or above normal
271 precipitation. Our finding that plant biomass recovers by the next drought treatment show the resilience of this
272 drought-adapted vegetation. Considering the long term climate prediction of increasing frequency of both extreme
273 dry and extreme wet years (Bartholy and Pongrácz 2007), there is no direct danger of desertification in the studied
274 community, as the carbon loss in dry years can be compensated by carbon accumulation in wet years.

275 In conclusion, we want to underline two of our findings. First, by means of application of field remote sensing, we
276 demonstrated the negative effect of drought treatment on the aboveground plant biomass and the ANPP in a diverse
277 semi-arid shrubland-grassland community. At the same time, we showed that only one yearly measurement right
278 after the treatment may overestimate the effect of drought, disregarding the compensation processes of late-
279 successional ecosystems, which can be detected using multiple within-year measurements.

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419 **Table 1** Comparisons tested using contrasts in the mixed effect linear model fitted to NDVI. M2-M1: treatment
 420 change, M3-M2: post-treatment change, M3-M1: whole-season change

Response variables	Subset	Comparison	Estimate	Std. Error	Z-value	p-value
NDVI	2011 M1	C vs. D	0.026	0.022	1.151	0.964
NDVI	2013 M1	C vs. D	0.065	0.022	2.911	0.063
NDVI	2011 M2	C vs. D	0.102	0.022	4.598	<0.001
NDVI	2013 M2	C vs. D	0.104	0.022	4.666	<0.001
NDVI	2011 M3	C vs. D	0.021	0.022	0.928	0.992
NDVI	2013 M3	C vs. D	0.085	0.022	3.831	0.003
NDVI	2011 C	M1 vs. M2	0.054	0.009	6.045	< 0.001
NDVI	2011 D	M1 vs. M2	0.023	0.009	-2.581	0.150
NDVI	2013 C	M1 vs. M2	0.075	0.009	8.404	< 0.001
NDVI	2013 D	M1 vs. M2	0.036	0.009	4.011	0.001
NDVI	2011 C	M2 vs. M3	-0.074	0.009	-8.341	< 0.001
NDVI	2011 D	M2 vs. M3	0.007	0.009	0.845	0.996
NDVI	2013 C	M2 vs. M3	0.062	0.009	-6.961	< 0.001
NDVI	2013 D	M2 vs. M3	-0.043	0.009	-4.871	< 0.001
NDVI	2011 C	M1 vs. M3	-0.020	0.009	-2.296	0.284
NDVI	2011 D	M1 vs. M3	0.015	0.009	-1.736	0.680
NDVI	2013 C	M1 vs. M3	0.013	0.009	1.442	0.863
NDVI	2013 D	M1 vs. M3	-0.008	0.009	-0.860	0.995
ΔNDVI (M2-M1)	2011	C vs. D	0.077	0.013	6.099	< 0.001
ΔNDVI (M2-M1)	2013	C vs. D	0.039	0.013	3.106	0.034
ΔNDVI (M3-M2)	2011	C vs. D	-0.082	0.013	-6.495	< 0.001
ΔNDVI (M3-M2)	2013	C vs. D	-0.019	0.013	-1.478	0.845
ΔNDVI (M3-M1)	2011	C vs. D	-0.005	0.013	-0.396	1.000
ΔNDVI (M3-M1)	2013	C vs. D	0.020	0.013	1.628	0.755

422 List of Figure Captions

423 **Fig. 1** The timing of measurement events (M0 to M3). Horizontal bars stand for time intervals of drought treatments.

424 Vertical bars are proportional to the monthly precipitation, while the unfilled parts of the vertical bars show the

425 amounts of excluded precipitation during the drought treatments. The heights of the bars range from 0 mm

426 (November 2011) to 126.3 mm (March 2013). See dates and more values in the text, as well as in Online Resource 1.

427 **Fig. 2** NDVI values (least-square means and 95% confidence intervals estimated by the fitted mixed-effect model)

428 for the measurement events in 2011 and 2013 in the control (C) and drought treatment (D) plots (see also Online

429 Resource 2); before treatment: M1, after treatment: M2, end-of-summer: M3.

430 * denotes significant drought effect

431 **Fig. 3** Estimated changes of NDVI values between measurement events (least-square means and 95% confidence

432 intervals): treatment change, i.e. M2-M1; post-treatment change, i.e. M3-M2; whole-season change, i.e. M3-M1; in

433 2011 and 2013 in the control (C) and drought (D) treatments;

434 * denotes significant drought effect

435 **Fig. 4** Sum of positive NDVI increments from the 0.205 start-of-season baseline (M0) to the end-of-summer

436 measurements as a proxy for annual aboveground net primary production, in 2011 and 2013 in the control (C) and

437 drought treatment (D) plots (least-squares means and 95% confidence intervals). The difference between control and

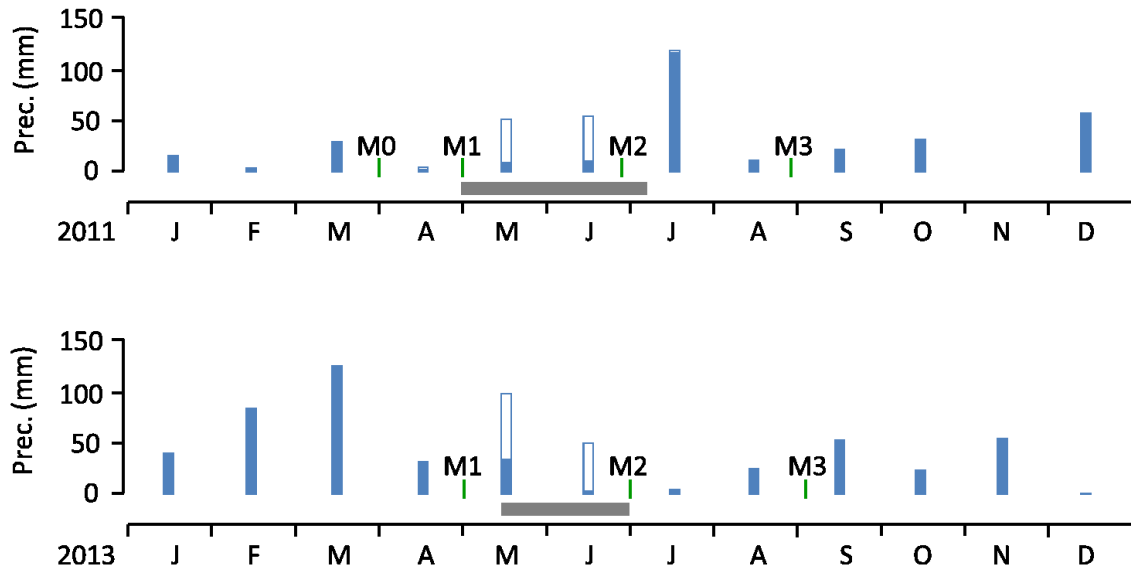
438 drought treatment is significant in both years and its value is significantly greater in 2013.

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441 Figures

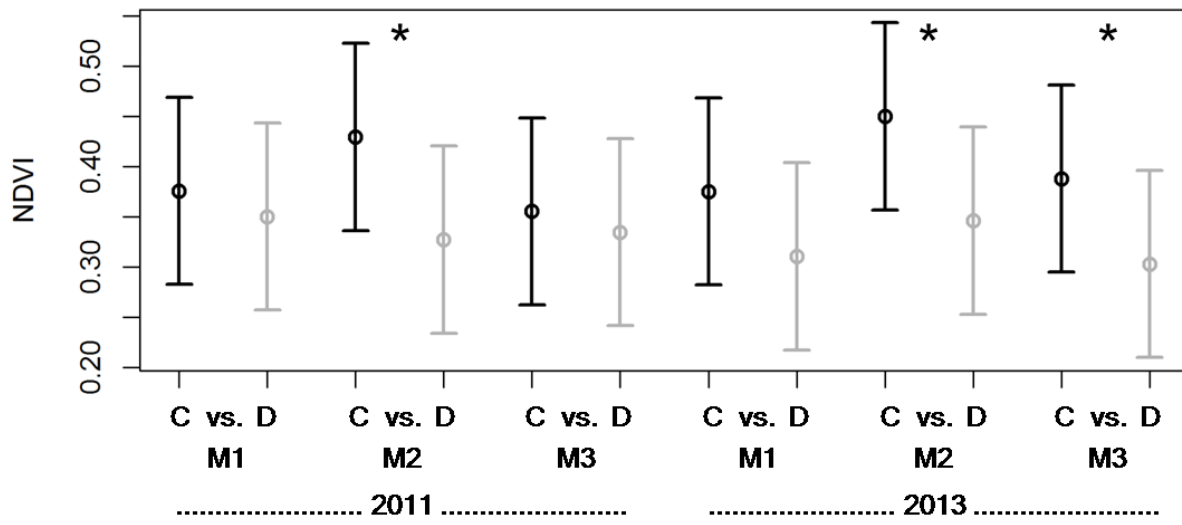
442 Fig. 1



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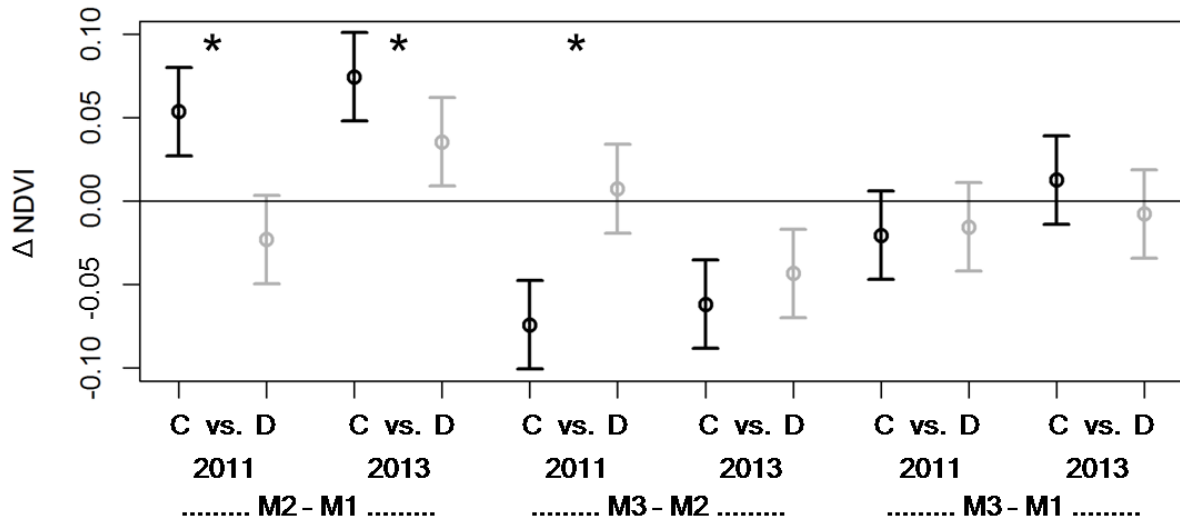
445 Fig. 2



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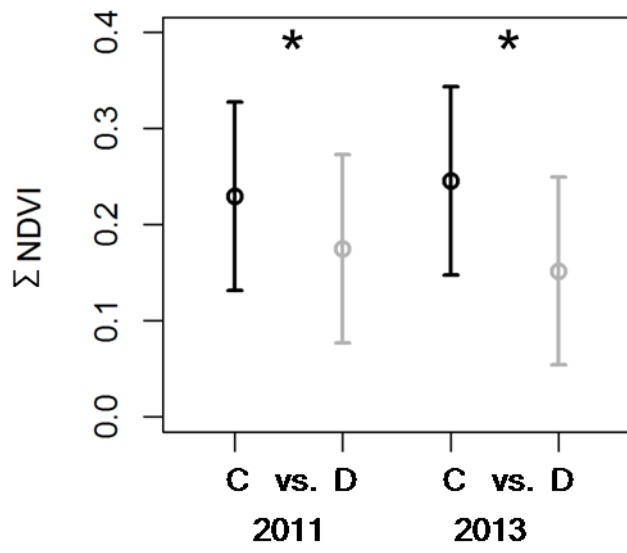
448 Fig. 3



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451 Fig. 4



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