



Source Attribution of Poly- and Perfluoroalkyl Substances (PFASs) in Surface Waters from Rhode Island and the New York Metropolitan Area

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1 **Source attribution of poly- and perfluoroalkyl substances (PFASs) in surface**
2 **waters from Rhode Island and the New York Metropolitan Area**

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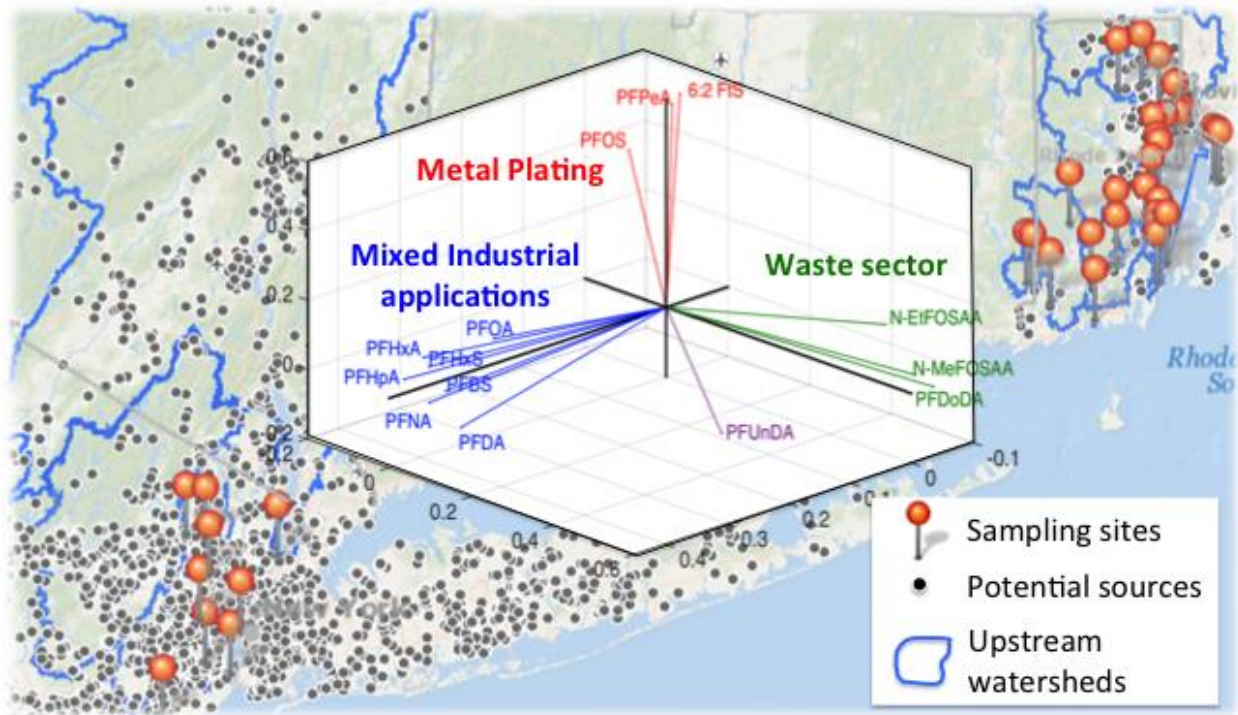
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14 **Abstract**

15 Exposure to poly and perfluoroalkyl substances (PFASs) has been associated with adverse health
16 effects in humans and wildlife. Understanding pollution sources is essential for environmental
17 regulation but source attribution for PFASs has been confounded by limited information on
18 industrial releases and rapid changes in chemical production. Here we use principal component
19 analysis (PCA), hierarchical clustering, and geospatial analysis to understand source
20 contributions to 14 PFASs measured across 37 sites in the Northeastern United States in 2014.
21 PFASs are significantly elevated in urban areas compared to rural sites except for
22 perfluorobutane sulfonate (PFBS), N-methyl perfluorooctanesulfonamidoacetic acid (N-
23 MeFOSAA), perfluoroundecanate (PFUnDA) and perfluorododecanate (PFDoDA). The highest
24 PFAS concentrations across sites were for perfluorooctanate (PFOA, 56 ng L⁻¹) and
25 perfluorohexane sulfonate (PFOS, 43 ng L⁻¹) and PFOS levels are lower than earlier
26 measurements of U.S. surface waters. PCA and cluster analysis indicates three main statistical
27 groupings of PFASs. Geospatial analysis of watersheds reveals the first component/cluster
28 originates from a mixture of contemporary point sources such as airports and textile mills.
29 Atmospheric sources from the waste sector are consistent with the second component, and the
30 metal smelting industry plausibly explains the third component. We find this source-attribution
31 technique is effective for better understanding PFAS sources in urban areas.

32 **Introduction**

33 Exposure to poly- and perfluoroalkyl substances (PFASs) has been associated with many
34 negative health outcomes including compromised immune function, metabolic disruption,
35 obesity, and altered liver function.¹ PFASs in surface waters are an emerging concern for U.S.
36 public water supplies and long-chain compounds bioaccumulate in aquatic food webs, posing
37 health risks to seafood consumers.²⁻⁶ Production of PFASs and their precursors has shifted
38 dramatically over the last two decades toward shorter-chain and polyfluorinated species.⁷
39 Diverse point sources and atmospheric deposition of some PFASs confounds understanding of
40 the dominant contributors to contamination in the aquatic environment. Regulatory databases
41 such as the U.S. EPA's Facility Registry Survey (FRS)⁸ and the Toxic Release Inventory⁹
42 presently contain limited to no information on magnitudes of PFASs released to the environment.

43 Multivariate statistical analyses based on chemical composition profiles can be a
44 powerful tool for diagnosing contamination sources, as illustrated for many other organic
45 contaminants.¹⁰ Principal components analysis (PCA) provides information on
46 interrelationships among various chemicals and is useful for deriving common source
47 profiles. Two-way hierarchical clustering can be used as a confirmatory analysis of PCA by
48 generating a flexible number of subgroups of similar sites (those affected by a common
49 source type) without dictating the number of clusters *a priori*. Clustering of compounds
50 identifies chemicals that co-occur to form a unique signature. These techniques have not
51 been routinely applied to interpret PFAS contamination and show potential for interpreting
52 sources in surface water and seawater.^{4,11}

53 Here we combine PCA and hierarchical clustering of PFAS profiles measured in surface
54 waters from 37 rivers, streams and estuaries in the Northeastern United States with geospatial

55 analysis of potential sources. Few measurements are available for PFASs in U.S. surface waters
56 over the past five years and the importance of different sources is poorly understood. Source
57 regions for air pollution are commonly identified using back trajectories.^{12,13} We apply an
58 analogous approach for identifying sources of aquatic pollution based on hydrological
59 distances within a watershed. The main objective of this study is to identify major sources of
60 surface water PFAS contamination in diverse watersheds using information on chemical
61 composition and geospatial analytical tools that consider surface hydrology.

62 **Methods**

63 *Sample collection and analysis*

64 We collected surface water samples from rivers/creeks and estuaries at approximately 1
65 m depth at 28 sites in the state of Rhode Island (RI) in June, 2014 and 9 sites the New York
66 Metropolitan Area (NY/NJ) in October, 2014 (Figure 1). A complete description of sampling
67 sites is provided in the Supporting Information (SI Table S1). Precipitation and flow rates in
68 rivers tend to be higher in June, potentially resulting in enhanced dilution and a low bias for
69 some PFASs measured in RI rivers compared to NY/NJ.

70 Samples were stored in one-liter pre-rinsed polypropylene bottles at -20 °C and thawed at
71 room temperature. Each sample was shaken vigorously for homogenization before subsampling
72 500 ml for the analysis of 21 PFASs. Each unfiltered sample was spiked with 20 µL of a 0.1 ng
73 µL⁻¹ mass labeled PFAS mixture (Wellington; Guelph, Canada; individual compounds are listed
74 in Table S2) as internal standards for quantification. PFASs were extracted using an Oasis Wax
75 solid phase extraction (SPE) cartridge (6 mL, 150 mg sorbent) following the method of Taniyasu
76 et. al.¹⁴ (see SI Section S1 for details). A nitrogen evaporator (ZIPVAP) was used to concentrate
77 the extract to 1 mL (methanol: water; v:v = 1:1).

78 Sample detection for 21 native PFASs (Tables S2, S3) was performed using an Agilent
79 6460 LC-MS/MS equipped with an online-SPE system (Agilent 1290 Infinity Flex Cube) in
80 dynamic multiple reaction mode (sample chromatogram in Figure S1). At least one negative
81 control (field or procedural blank) and one positive control (spiked with 2 ng of the 21 PFASs in
82 500 ml water) were included in every extraction batch. Whole method recovery tested using the
83 positive controls was 70-120% for all but 4 PFASs that ranged from 60-70%, which is
84 comparable to recoveries reported by previous studies.^{3,14,15} The 4 PFASs are perfluoropentanoate
85 (PFPeA), perfluoroheptanoate (PFHpA), N-methyl perfluorooctanesulfonamidoacetic acid
86 (MeFOSAA) and N-ethyl perfluorooctanesulfon-amidoacetic acid (EtFOSAA). Potential analyte
87 loss during sample preparation was corrected using internal standards spiked prior to sample
88 extraction. The limit of detection (LOD, Figure S2) was defined as equivalent to the blank plus
89 the concentration corresponding to a signal-to-noise ratio of three. Variability between duplicates
90 obtained at two sites was <20%. PFASs in five field blanks (HPLC grade water) prepared
91 following the sample preparation procedure were all below the LOD.

92 We quantified branched isomers for perfluorooctanoate (PFOA), perfluorohexane
93 sulfonate (PFHxS), perfluorooctane sulfonate (PFOS), N-MeFOSAA and N-EtFOSAA using
94 calibration standards for the linear isomers, assuming the same instrumental response factor
95 (Table S3). Seven compounds namely perfluorododecane sulfonate (PFDS), 8:2 fluorotelomer
96 sulfonate (8:2 FTS), perfluorooctane sulfonamide (FOSA), and PFCAs with more than 12 carbon
97 atoms) were detected in less than half of samples and were excluded from additional statistical
98 analysis (see Table S2 for details). For the 14 PFASs that had detection frequencies of greater
99 than 60% (Table S2), we used the Robust Regression on Order Statistics approach for censored

100 log-normally distributed environmental data described by Helsel¹⁶ to assign values to samples
101 with concentrations below the LOD.

102 *Statistical and spatial analysis*

103 We used principal components analysis (PCA) and hierarchical clustering to group sites
104 with statistically distinct PFAS composition profiles. PCA was performed using MATLAB's
105 Statistics Toolbox (MathWorks, Inc.) on normalized (z-score to remove the effect of
106 concentration difference at different sites) PFAS concentration data. The inverse of variances of
107 the data were used as variable weights and varimax rotation was applied to interpret the meaning
108 of extracted principal components. Hierarchical Cluster analysis was conducted using the hclust
109 function in the R statistical computing package (version 3.1.3).

110 We characterized the watershed for each freshwater sampling site using the U.S.
111 Geological Survey's (USGS) National Elevation Dataset (3 arc-second for site 15 and 16 and 1
112 arc-second for others) and the Hydrologic Tool in ArcGIS Pro 1.2 and ArcGIS online. Estuarine
113 sampling sites were excluded from the geospatial analysis due to the confounding influence of
114 tidal waters diluting potential source profiles. Population within each watershed was based on
115 ESRI's U.S. Demographic Database.¹⁷ We used the USGS's StreamStats database (version 4)¹⁸
116 to characterize water flow rates for each location and to compute mass flow (kg/yr) of PFASs at
117 each site and per-capita mass flows (kg/person/yr).

118 For all inland sites (non-estuarine), we acquired a list and geospatial data for plausible
119 PFAS sources from the US EPA Facility Registry Service (FRS) database on facilities and sites
120 subject to environmental regulation (see SI for the search criteria).⁸ These include airports,
121 facilities for metal plating/coating, printing, sewage treatment, waste management (including
122 landfills), and manufacturers of semiconductor, textile, paint/coating/adhesive, ink, paper, and

123 petroleum products. A caveat of this analysis is that not all facilities included in the FRS
124 database necessarily release PFASs and the database may not comprehensively include all
125 possible sources.

126 Hydrological distances of point sources from each sampling site were computed using the
127 ArcGIS Trace Downstream tool. Within each watershed, we defined an indicator for the impact
128 of potential point sources as a function of distance from sampling locations by assuming
129 exponential decay in the source signature¹⁹ (i.e., $impact = 1/e^d$, where d = hydrological distance,
130 km). This approach provides additional information on plausible sources that complements
131 multivariate statistical analysis but cannot be considered a quantitative estimate of contributions
132 to sampling locations since magnitudes of PFAS discharges are not available.

133 **Results and Discussion**

134 ***Concentrations and spatial patterns***

135 Figure 1 shows the compound specific composition and concentrations of PFASs
136 measured in surface water samples as part of this work. Sampling sites in NY/NJ had much
137 greater population density in upstream watersheds (10-43x) compared to RI but the highest
138 concentrations of most PFASs were measured near the city of Providence, RI (Figure 1, Figure
139 S2). The range of measured PFAS concentrations reported here are comparable or lower than
140 U.S. surface waters from other regions collected between 2000-2009 (Table S4).^{2,20-24}

141 All sites had detectable PFOA and PFNA and over 90% contained detectable PFHxS,
142 PFOS, PFDA, and 6:2 FtS (Table S2, S3, Figure S2). The highest individual PFAS
143 concentration across sites was PFOA (56 ng L⁻¹) at Site 31 (Passaic River, NJ). Highest
144 concentrations of PFHxS (43 ng L⁻¹) and PFNA (14 ng L⁻¹) were measured at Site 5 (Mill Cove,
145 RI). The maximum PFOS concentration (27 ng L⁻¹) was measured at Site 2 (Woonasquatucket

146 River, RI) within the City of Providence, RI. This is much lower than maximum levels reported
147 in earlier studies of US surface waters that range between 43-244 ng L⁻¹ (Table S4) and reflects
148 the continued decline in environmental PFOS burdens in North America following elimination of
149 production in 2002.^{25,26}

150
151 Measured PFAS concentrations in urban regions were significantly higher (Wilcoxin rank
152 sum test, $p < 0.017$) than rural sites for all compounds except PFBS, N-MeFOSAA, PFUnDA and
153 PFDoDA (Figure S3). Sites 1-9 in RI and Sites 29-37 in NY/NJ are all urban areas, defined by
154 population densities of greater than 1000 individuals per square mile (2590 km²), and population
155 densities of greater than 500 individuals per square mile in surrounding census blocks.²⁷ We did
156 not find a statistically significant correlation between total population in each upstream
157 watershed and PFAS concentrations measured at each sampling site ($p = 0.12$ to 0.95 across
158 compounds). We derived per-capita discharges (Figure S4) using a similar approach as Pistocchi
159 and Loos.²⁸ Highest median per-capita discharges ($\mu\text{g person}^{-1} \text{ day}^{-1}$) across compounds, in
160 decreasing order, were for PFOA (27), PFHxA (14), PFHpA (10), PFOS (9), PFHxS (7), and
161 PFNA (5) (Figure S4). These are lower than previously reported in Europe ca. 2007 (e.g., PFOA:
162 $82 \mu\text{g person}^{-1} \text{ day}^{-1}$, PFOS: $57 \mu\text{g person}^{-1} \text{ day}^{-1}$).²⁸

163 ***Source identification***

164 Both hierarchical clustering and PCA identified three distinct groupings of PFASs
165 (Figure 2a, b). The first component/cluster explains 46% of variability in the PCA and includes
166 two major end products of the fluorochemical manufacturing industry (PFOA, PFNA), and a mix
167 of other compounds: PFBS, PFHxS, PFHxA, PFDA. Site 5 (Mill Cove, RI) contains the highest
168 summed PFASs across all sites and is dominated by this mixture of PFASs. PCA results suggest
169 Site 5 is statistically similar to the Pawcatuck River, RI sampling locations (Sites 20, 19) and the

170 Passaic River, NJ (Site 31). However, these sites are grouped separately in the hierarchical
171 clustering analysis (Figure 2b), suggesting some differences in source contributions.

172 Geospatial analysis of the watersheds for Sites 5, 19, 20 and 31 reveals a mixture of
173 potential sources (Figure S5). For Site 5, the greatest source impact as a function of distance
174 within the watershed is from T.F. Green Airport, the largest public airport in Rhode Island. Prior
175 work indicates uses of AFFF in modern airports release diverse PFASs to downstream aquatic
176 environments, including the compounds identified as part of the first PCA/cluster.^{4,29-31} For Sites
177 19 and 20, textile mills in the upstream watersheds have the highest impact as a function of
178 distance (Table S5). PFASs are used for water resistant coating in textiles and washing and
179 disposal of wastewater at textile mills provides a vector for their entry to the aquatic
180 environment. For Site 31, PCA scores suggest a mix of components 1-3 (Figure 2 c, d). This site
181 also clusters differently than Sites 19 and 20 (Figure 2b). The FCA database indicates the
182 watershed of Site 31 (Figure S5) contains diverse industrial sources that must account for this
183 profile including metal plating, printing, a landfill, petroleum and coal products manufacturing.
184 Overall, we conclude that the first PCA component and cluster of PFASs (PFOA, PFNA PFBS,
185 PFHxS, PFHxA, PFDA) represents a mixture of contemporary sources including airports and
186 textile mills.

187 The second component/cluster explains 19% of the variability in PFASs and includes two
188 long-chain PFASs (PFUnDA and PFDODA) and two precursors to PFOS (MeFOSAA and
189 EtFOSAA) (Figure 2). PFUnDA and PFDODA mainly originate from fluorotelomer alcohols or
190 other fluotelomer based products.³² Both N-MeFOSAA and N-EtFOSAA are intermediate
191 degradation products from the volatile parent compound N-alkyl perfluorooctane
192 sulfamideoethanol (FOSE) with PFOS as the final degradation product. This profile is most

193 pronounced at Site 3 along the Woonasquatucket River in RI and is also evident at Site 1
194 (Slack's Tributary, RI) and Site 6 (Buckeye Brook, RI). For Site 3, the largest source impact
195 based on distance is from a wastewater treatment plant 1 km upstream. No industrial facilities
196 exist upstream of Sites 1 and 6. Landfill/waste management facilities are located within 2 km of
197 all three sites but are not hydrologically connected to the sampling locations (Figure S5). Both
198 landfills and wastewater treatment plants are known atmospheric sources of fluorotelomer
199 alcohols and FOSE.³³ Concentrations of N-MeFOSAA, PFUnDA and PFDoDA were not
200 spatially variable at most sites and only slightly elevated at Site 3, consistent with an atmospheric
201 input pathway. We thus infer that this component is most likely attributable to sources from the
202 waste sector.

203 The third component explains 15% of the variability in PFASs and includes PFPeA,
204 PFOS, and 6:2 FTS. This component is most pronounced at Site 2 along the Woonasquatucket
205 River, within the City of Providence, RI. GIS analysis of the watershed at this site reveals the
206 presence of 14 metal coating/plating industries upstream (Figure 2d, Table S5, Figure S5). PFOS
207 was historically used as a mist/fume control agent in metal plating, in surface coatings and as the
208 major component in AFFFs for fighting petroleum related fire.^{25,26,34} Some PFOS applications
209 such as metal plating have been replaced by less stable fluorotelomer based chemicals such as
210 6:2 FtS,³⁵ which will eventually degrade into PFPeA and PFHxA (yields of 1.1% and 1.5% in
211 activated sludge).³⁶ It is likely that PFHxA is not included in the cluster because other direct
212 sources can contribute one order of magnitude more PFHxA than PFPeA.^{37,38} We conclude that
213 the distinct PFAS profile at Site 2 is can be explained by the metal plating industry.

214 **Implications**

215 Multivariate statistical tools such as PCA and hierarchical clustering of PFAS profiles
216 combined with data on hydrological proximity of potential sources are useful for identifying
217 sources of surface water contamination. We find aquatic transport pathways (hydrological
218 distance and river flow directions) are critical for source identification. This contrasts many other
219 persistent organic pollutants that are primarily transported atmospherically, allowing sources
220 within a radius surrounding the sampling sites to be linked to concentrations.³⁹ We conclude that
221 the approach demonstrated here for RI and NY/NJ has potential for diagnosing PFAS source
222 contributions in urbanized regions with elevated concentrations and lacking specific information
223 on the magnitude of PFAS discharges from diverse industries. Background PFAS concentrations
224 at most rural sites in this study contain a mix of diverse source signatures that are not statistically
225 distinguishable using these methods. This analysis could be refined in future applications by
226 analyzing additional emerging short-chain PFASs and precursors to develop more unique
227 chemical signatures for specific industries (i.e., those contributing to the first component/cluster).

228 **Supporting Information**

229 Supporting Information Available: Details on analytical methods, data analyses, supporting
230 figures and tables. This material is available free of charge via the Internet at <http://pubs.acs.org>.

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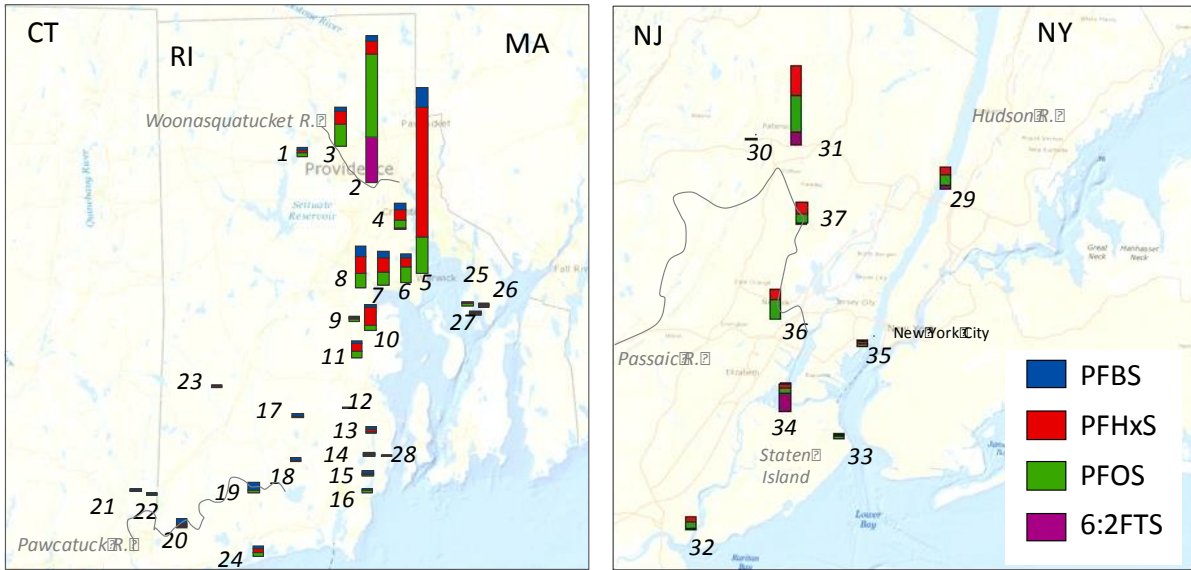
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361 **Figure Captions**

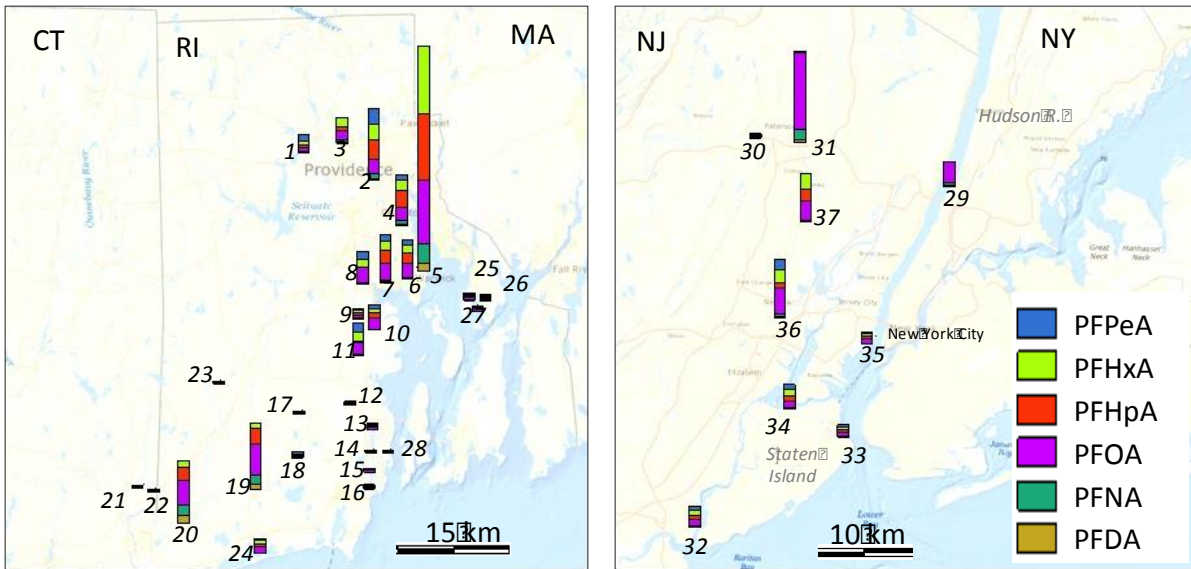
362 **Figure 1.** Concentrations of PFASs measured in surface waters from Rhode Island and the New
363 York Metropolitan Area. Full names of individual compounds are listed in Table S2. N-
364 MeFOSAA and N-EtFOSAA are not shown but were detected in ~70% of the samples at
365 concentrations <1 ng/L.

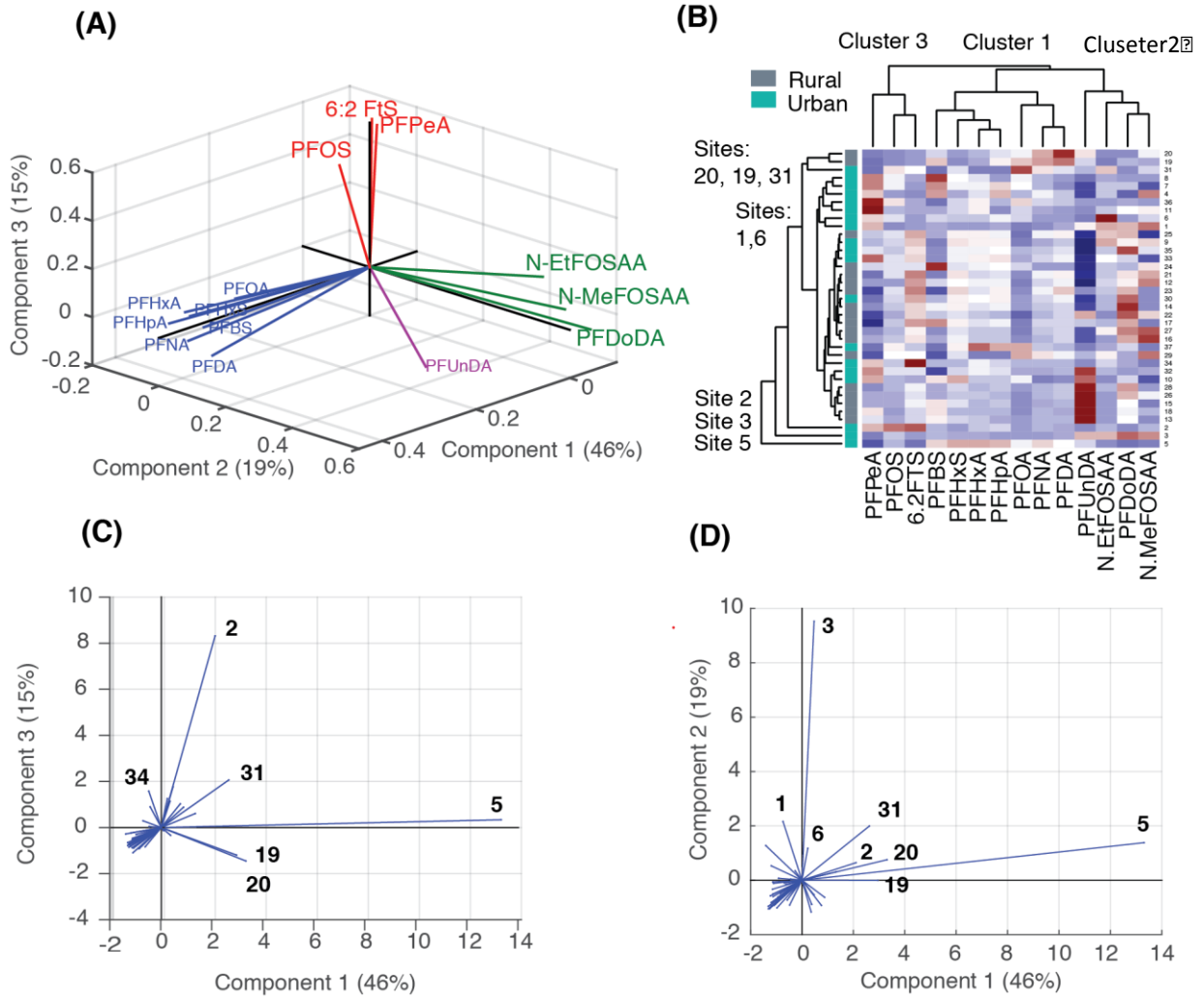
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367 **Figure 2.** Multivariate statistical analysis of surface water data. Panel (A) shows loadings of
368 principal components analysis (PCA) and Panels (C) and (D) show score plots for three
369 components across sampling sites. Panel (B) compares PCA results to hierarchical clustering of
370 compounds and sites. Sites with statistically distinct PFAS profiles are indicated on plots (C)
371 and (D) and highlighted on the hierarchical clustering diagram. The three principal components
372 together explain 80% of the variance in PFAS composition.

Per- and poly-fluoroalkane sulfonates 10 ng/L



Perfluorocarboxylates 30 ng/L





1 **Supporting Information**

2 **Concentrations and source attribution of poly- and perfluoroalkyl substances (PFASs) in surface**
3 **waters from the Northeastern U.S.**

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34	standards of the linear isomers by assuming same response factors between isomers. The limit of	
35	detection (LOD) for each compound is shown as a red bar. Those below detection are assigned	
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37	distributed environmental data as described by Helsel. ²	14
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39	sites (RI sites 1–11 and NY/NJ sites 29–37) and rural sites 12-28 (b) RI sites 1–11 and NY/NJ sites 29–	
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50 **Section S1: Supplemental Information on Methods**

51 **Sample Preparation and Instrumental Analysis**

52 PFASs were extracted from water samples using Oasis Wax (6 ml, 150 mg sorbent) solid phase
53 extraction (SPE) cartridges following the method of Taniyasu et al.¹ Each 500 ml water sample was
54 passed through a preconditioned Oasis Wax (6 ml, 150 mg sorbent) weak ion exchange SPE cartridge
55 mounted on a vacuum manifold at a flow rate of ~1 drop/s. Target analytes were eluted off the
56 cartridges using 6 ml 0.1% NH₄OH in methanol and collected in 15 ml centrifuge tubes (Corning). The
57 extracts were concentrated to 0.5 ml under a gentle stream of high purity nitrogen (5.0 grade),
58 centrifuged at 5000 rpm for 10 minutes, and transferred 1.5 ml polypropylene auto-sampler vials
59 (Microsolv). Before instrumental analysis, 0.5 ml water was added to each sample and vortex mixed.

60 A 300 µL aliquot of each sample was injected and loaded to an Agilent Zorbax SB-Aq
61 (4.6×12.5mm; 5µm) online SPE column with 0.85 ml 0.1% (v:v) formic acid at a flow rate of 1
62 ml/min. Following sample loading, the SPE were eluted and load the analytes to an Agilent Poroshell
63 120 EC-C18 (3.0×50mm; 2.7µm) reverse phase HPLC column. Methanol and water containing 2 mM
64 ammonium acetate were used as mobile phases (flow rate: 0.5 ml/min). Starting from 3% methanol,
65 the elution gradient was linearly increased to 61% in 7 minutes, held for 1 minute, then linearly
66 increased to 100% methanol in 3 min, and was kept until the end of the sample run (14 min).

67 The tandem mass spectrometer equipped with an electrospray ionization source was operated in
68 negative ion mode. Dynamic multiple reaction monitoring (dMRM) mode was used for data
69 acquisition in order to increase sensitivity. The collision gas was 5.0 grade N₂. Optimized MS
70 parameters are as follows: source temperature, 300 °C; capillary voltage, -3.8 kV; nitrogen nebulizer
71 gas, 45 psi and 13 L/min. Methanol was injected and passed through the system to eliminate any
72 potential carry-over after every sample (or calibration standard).

73 Shorter chain PFASs such as PFBA and 4:2 FtS were not analyzed due to their low retention on

74 the C-18 reverse phase HPLC column, which would result in a low accuracy.² A different analytical
75 method (e.g., using a normal phase HPLC column) that can accurately measure those shorter chain
76 PFASs is needed to detect these compounds and represents a limitation of the present analysis.

77 **Data analysis**

78 Helsel² suggests statistical inference bias may occur for data with detection frequencies of less
79 than 30%. PFASs with detection frequencies of 60-70% are included here because they are important
80 for source identification. We tested results of principal component analysis with and without PFASs
81 with low detection frequencies (60-65%: PFPeA, PFHpA, PFDoDA) and find no significant changes in
82 PCA scores (Wilcoxon signed rank tests ($p=0.06-0.5$) and clustering included in the main results of
83 this work.

84 Potential industrial PFAS point sources were retrieved from the US EPA Facility Registry
85 Service (FRS) database and used in the geospatial analysis conducted as part of this research. Filtering
86 of the database was based North American Industry Classification System (NAICS) codes. Facilities
87 and their coordinates were retrieved based on the following NAICS codes: Sewage treatment facilities
88 (22132); textile mills (313); paper manufacturing (322); printing and related support activities (323);
89 petroleum and coal products manufacturing (324); paint, coating, and adhesive manufacturing (3255);
90 printing ink manufacturing (32591); metal coating, engraving, heat treating and allied activities (3328);
91 semiconductor manufacturing (3344); airport operation (48811); waste management and remediation
92 (562)

93 **Section S2: Supporting Tables and Figures**

94 Table S1. Surface water sampling dates, site locations and description.

Site	Location	Coastal	Urban/Rural	Potential sources	Date	Longitude	Latitude
1	Slack's Tributary	N	Urban	No hydrologically connected point sources; a landfill is located 1.9 km to the north	06/19/2014	-71.55	41.85
2	Woonasquattucket River	N	Urban	Metal coating/plating	06/19/2014	-71.44	41.82
3	Woonasquattucket River (Greystone pond)	N	Urban	Wastewater treatment plant, printing activity	06/19/2014	-71.49	41.87
4	Pawtuxet River	N	Urban	Metal coating/plating, semiconductor manufacturing	06/19/2014	-71.40	41.77
5	Brook at Mill Cove	N	Urban	T.F. Green State Airport ~5km upstream	06/19/2014	-71.38	41.71
6	Buckeye Brook	N	Urban	No hydrologically connected point sources; a landfill is located 2.3 km to the west	06/19/2014	-71.39	41.70
7	Southern Creek	N	Urban	No hydrologically connected point sources	06/19/2014	-71.42	41.70
8	Mill Brook	N	Urban	One semiconductor manufacturer making thin film components, networks, and arrays on ceramic and silicon; one company conducting waste management providing service on hazardous waste removal, hazardous waste transportation, oil tank hazardous waste disposal (https://www3.epa.gov/region1/removal-sites/BradfordPrintingFinishing.html)	06/19/2014	-71.46	41.70
9	EG Town Dock	Y	Urban	Estuary of Greenwich Cove; next to an e-waste recycling company	06/19/2014	-71.45	41.65
10	Hunt River	Y	Urban	Two semiconductor manufacturers and one printing company	06/19/2014	-71.44	41.64
11	Sand Hill Brook (Saw Mill Pond Inlet)	N	Urban	A municipal waste transfer station and paint, coating, adhesive manufacturing	06/19/2014	-71.47	41.61
12	Secret Lake-Oak Hill Brook	N	Urban	A legacy landfill site is approximately 2 km to the west of this site	06/19/2014	-71.48	41.55
13	Narrow River Stuart Stream	N	Rural	Outlet of Carr Pond	06/19/2014	-71.44	41.52
14	Narrow	N	Rural	3 km downstream of site 13	06/19/	-71.45	41.49

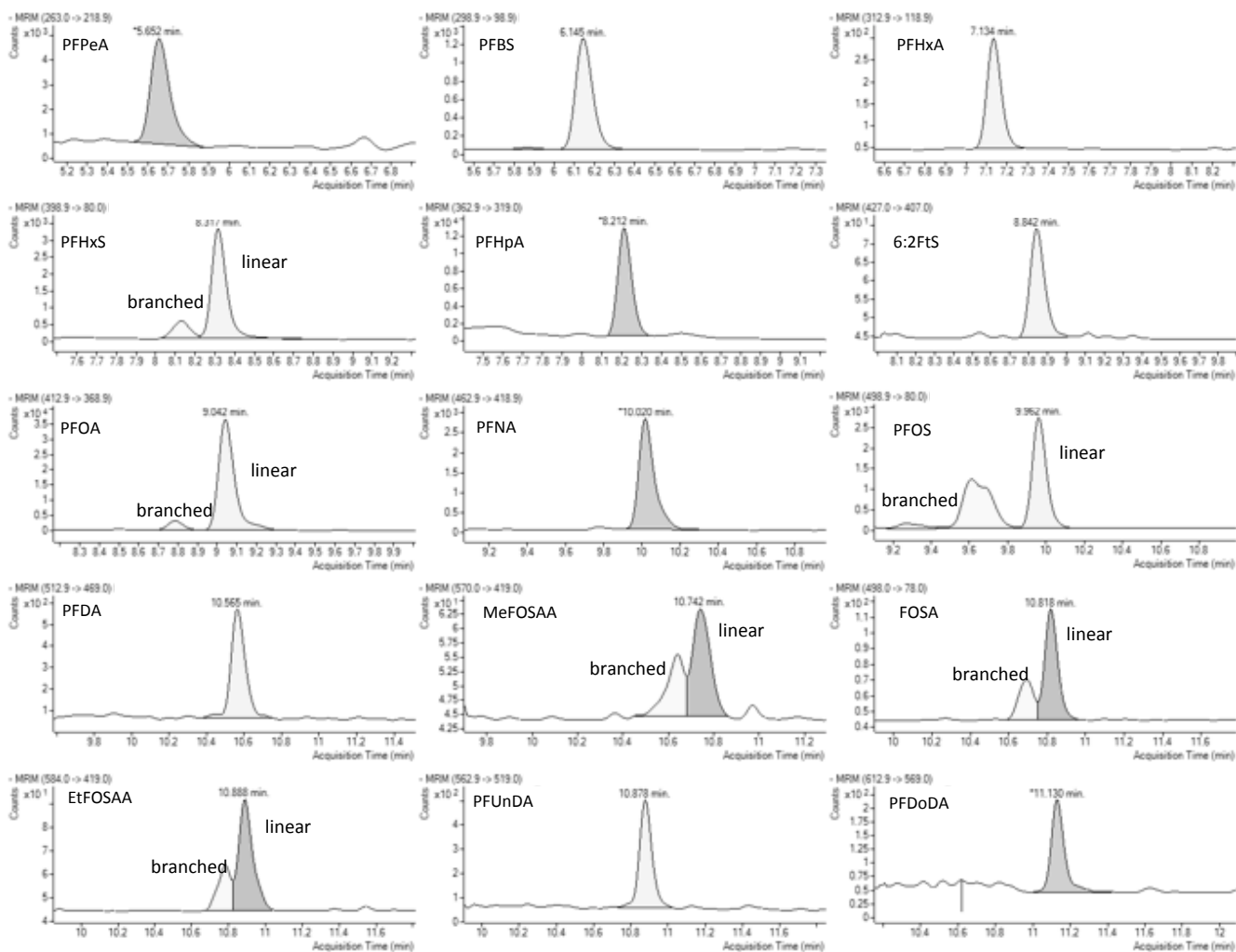
Site	Location	Coastal	Urban/Rural	Potential sources	Date	Longitude	Latitude
	River Lakeside Dr.				2014		
15	Narrow River	N	Rural	2.5 km downstream of site 14	06/19/2014	-71.45	41.47
16	Narrow River	N	Rural	2 km downstream of site 15	06/19/2014	-71.45	41.45
17	Queens River	N	Rural	One river branch upstream of Pawcatuck River (background site)	06/19/2014	-71.56	41.54
18	Chickasheem Brook	N	Rural	River branch upstream of Pawcatuck River; a manufacturer of uninterruptible power supplies, electronics peripherals and data center products is downstream	06/19/2014	-71.56	41.49
19	Pawcatuck River	N	Rural	Where Beaver River merges into Pawcatuck River; a manufacture of military, tactical, and performance synthetic and synthetic blend textiles ~1 km upstream	06/19/2014	-71.63	41.45
20	Pawcatuck River	N	Rural	Adjacent to Bradford Printing & Finishing facility, a textile finishing plant from 1911 until 2012; a large fire occurred in 2007; heavy flooding occurred in 2010; another fire occurred in 2012; Several hundred containers of highly flammable liquid, dyes and unknown compounds were stored next to each another and many containers were visibly leaking in 2012. ³	06/19/2014	-71.75	41.41
21	Green Falls River	N	Rural	Background site; no upstream industrial facilities recorded in FRS database	06/19/2014	-71.82	41.45
22	Green Falls River	N	Rural	~ 2 km downstream of site 21 where Parmenter Brook merges into Green Falls River; no upstream industrial facilities recorded in FRS database	06/19/2014	-71.80	41.44
23	Fall River	N	Rural	Background site; no upstream industrial facilities recorded in FRS database	06/19/2014	-71.69	41.58
24	Allen Cove - Inflow (Green Hill Pond)	N	Rural	Close to Charlestown beach; residential area	06/19/2014	-71.62	41.37
25	Bristol Harbor	Y	Rural	Coastal site; east shore of Bristol Harbor	06/19/2014	-71.29	41.67
26	Bristol Harbor	Y	Rural	Coastal site; east shore of Bristol Harbor	06/19/2014	-71.28	41.67
27	Bristol Harbor	Y	Rural	Coastal site; west shore of Bristol Harbor	06/19/2014	-71.27	41.66
28	South Ferry Rd Pier	Y	Rural	Coastal site; Narragansett Bay	06/19/2014	-71.42	41.49

Site	Location	Coastal	Urban/Rural	Potential sources	Date	Longitude	Latitude
	Dock						
29	Hudson River	N	Urban	There are a sewage treatment plant, a plastic bag manufacturing and printing company, a printing ink manufacture, and a floor coating manufacture within 10 km upstream along the river	10/24/2014	-73.93	40.87
30	Passaic River	N	Urban	West Paterson Recycling Center 2.5 km upstream	10/24/2014	-74.19	40.91
31	Passaic River	N	Urban	Highly industrialized between 30 and 31, including paint, coating, adhesive manufacturing, textile mills, printing ink manufactures, paper manufacturers; semiconductor manufactures and metal coating/plating companies.	10/24/2014	-74.13	40.91
32	Harbortown Rd, NJ	Y	Urban	At the mouth of a tidal strait and a kill separating Staten Island, New York City from mainland New Jersey; some petroleum/coal related industrials within 2 km upstreams	10/25/2014	-74.25	40.52
33	Lower NY Harbor	Y	Urban	A printing ink manufacture 1 km away	10/25/2014	-74.06	40.62
34	Staten Island NY	N	Urban	A company with printing activity; a paint, coating, adhesive manufacture, and a paper manufacture within 1.5 km upstream	10/25/2014	-74.13	40.64
35	Hudson River	Y	Urban	Morris Canal close to Jersey city; two companies on Paint, coating, adhesive manufacturing 1 km away	10/26/2014	-74.04	40.71
36	Passaic River	N	Urban	Close to the city of Newark and the airport; highly industrialized area; Newark wastewater treatment plant is 2.5 km upstream	10/26/2014	-74.15	40.73
37	Passaic River	N	Urban	Upstream of site 36; highly industrial area; within 1 km upstream there is a company related to metal plating and a textile mill.	10/26/2014	-74.12	40.83

96 Table S2. Full names and acronyms of PFASs measured in surface waters, limits of detection (LOD),
 97 concentration ranges measured across sites, and percent of sites with detection. PFASs measured in
 98 >60% of samples analyzed in this study are highlighted in bold

PFAS	Acronym	# of carbons	Internal standard	LOD (ng/L)	Range (ng/L)	Detect. %
<i>Perfluorocarboxylates</i>		<i>PFCAs</i>				
Perfluoropentanate	PFPeA	C5	¹³ C ₂ -PFH _x A	0.38	BD – 10	62%
Perfluorohexanate	PFHxA	C6	¹³ C ₂ -PFH _x A	0.29	BD–48	87%
Perfluoroheptanate	PFHpA	C7	¹³ C ₄ -PFOA	0.62	BD–48	64%
Perfluorooctanate	PFOA	C8	¹³ C ₄ -PFOA	0.07	0.27 – 47	100%
Perfluorononanate	PFNA	C9	¹³ C ₅ -PFNA	0.04	0.07 – 14	100%
Perfluorodecanate	PFDA	C10	¹³ C ₂ -PFDA	0.03	BD – 5.8	92%
Perfluoroundecanate	PFUnDA	C11	¹³ C ₂ -PFUnDA	0.02	BD –1.9	77%
Perfluorododecanate	PFDoDA	C12	¹³ C ₂ -PFDoDA	0.02	BD–2.6	64%
Perfluorotridecanate	PFTrDA	C13	¹³ C ₂ -PFDoDA	0.02	BD–1.2	31%
Perfluorotetradecanate	PFTeDA	C14	¹³ C ₂ -PFDoDA	0.02	BD0.4	18%
Perfluorohexaadecanate	PFHxDA	C16	¹³ C ₂ -PFDoDA	0.01	BD–0.2	26%
Perfluorooctadecanate	PFODA	C18	¹³ C ₂ -PFDoDA	0.08	BD–0.4	8%
<i>Perfluoroalkane sulfonates</i>		<i>PFSAs</i>				
Perfluorobutane sulfonate	PFBS	C4	¹⁸ O ₂ -PFH _x S	0.08	BD–6.2	85%
Perfluorohexane sulfonate	PFHxS	C6	¹⁸ O ₂ -PFH _x S	0.06	BD – 35	90%
Perfluorooctane sulfonate	PFOS	C8	¹³ C ₄ -PFOS	0.05	BD – 23	95%
Perfluorododecane sulfonate	PFDS	C10	¹³ C ₄ -PFOS	0.07	BD–0.6	15%
6:2 fluorotelomer sulfonate	6:2 FtS		¹³ C ₂ -6:2 FtS	0.003	BD – 15	97%
8:2 fluorotelomer sulfonate	8:2 FtS		¹³ C ₂ -6:2 FtS	0.4	BD–0.8	41%
Perfluorooctane sulfonamide	FOSA	C8	¹³ C ₈ -FOSA	0.02	BD–0.2	41%
N-ethyl perfluorooctanesulfonamidoacetic acid	N-EtFOSAA		D ₅ N-EtFOSAA	0.001	BD-9.9	67%
N-methyl perfluorooctanesulfonamidoacetic acid	N-MeFOSAA		D ₅ N-MeFOSAA	0.002	BD-0.6	69%

99 BD = below detection.



100
101
102
103

Figure S1. Chromatograms of PFASs in a sample analyzed using an Agilent 6460 LC-MS/MS equipped with an online-SPE system (Agilent 1290 Infinity Flex Cube) in dynamic multiple reaction mode.

104 Table S3a. Concentrations (pg/L) of poly- and perfluoroalkyl substances with detection frequency
 105 greater than 60%.

Site	PFPeA	PFHxA	PFHpA	PFOA ^a	PFNA	PFDA	PFUnDA	106 107
1	4550	2191	2409	2363	390	405	607	108
2	10357	12137	13577	8832	3134	1133	308	109
3	<LOD	6310	3371	5236	1476	894	1853	110
4	4228	7337	12301	7546	2735	957	114	111
5	<LOD	48414	48159	36806	13986	5625	1286	112
6	4359	5408	7640	8455	733	367	167	113
7	4828	6715	9236	10080	1275	205	46	114
8	5611	5649	<LOD	9237	923	176	48	115
9	927	1562	1597	1972	336	127	97	116
10	3064	2987	3090	6978	308	125	<LOD	117
11	6361	6678	<LOD	6905	799	226	177	118
12	555	565	<LOD	849	165	59	38	119
13	1413	1170	<LOD	1480	253	104	<LOD	120
14	<LOD	665	<LOD	663	104	<LOD	33	121
15	732	556	<LOD	851	136	31	<LOD	122
16	631	543	<LOD	946	174	87	62	123
17	681	550	<LOD	898	155	59	62	124
18	2138	663	<LOD	1006	293	<LOD	<LOD	125
19	<LOD	3740	11793	18974	6182	3808	482	126
20	<LOD	4138	9728	14985	7235	5824	888	127
21	<LOD	<LOD	<LOD	586	232	73	41	128
22	<LOD	493	<LOD	708	206	83	<LOD	129
23	<LOD	<LOD	<LOD	640	200	152	97	130
24	1221	2121	2479	3784	260	52	55	131
25	843	1214	897	1320	400	169	97	132
26	821	964	751	1014	323	134	<LOD	133
27	617	900	800	1170	355	166	78	134
28	<LOD	<LOD	<LOD	267	74	38	<LOD	135
29	<LOD	<LOD	<LOD	11862	2188	685	257	136
30	<LOD	815	947	871	151	59	28	137
31	<LOD	<LOD	<LOD	47254	6658	2154	464	138
32	3032	3529	3226	3738	601	301	<LOD	139
33	1870	1802	1907	2020	363	182	49	140
34	3434	5188	3431	4049	726	347	115	141
35	1111	1710	1852	2805	411	211	59	142
36	7998	9277	3426	15137	2022	719	238	143
37	<LOD	10901	8455	11335	757	152	79	144

^aLinear isomers with calibration standards for quantification

145

146 Table S3b. Concentrations (pg/L) of poly- and perfluoroalkyl substances with detection frequency
 147 greater than 60%.

Site	PFBS	PFHxS ^a	PFOS ^a	6:2 FtS	MeFOSAA ^a	EtFOSAA ^a	PFDoDA	148
1	669	864	777	15	241	348	618	149
2	1652	3758	23226	15292	147	278	89	150
3	1327	3583	5868	55	610	937	2598	151
4	2290	2558	2185	380	227	152	28	152
5	6181	35022	9804	239	113	240	117	153
6	1087	2637	4127	24	90	694	96	154
7	2102	4130	3743	30	<LOD	122	<LOD	155
8	3355	4664	3937	9	23	53	23	156
9	296	695	735	26	38	65	61	157
10	1161	5075	1477	8	<LOD	36	<LOD	158
11	546	2418	1822	5	106	94	313	159
12	278	<LOD	<LOD	<LOD	43	14	25	160
13	889	645	347	6	<LOD	<LOD	<LOD	161
14	368	476	176	<LOD	<LOD	<LOD	<LOD	162
15	705	421	180	10	<LOD	<LOD	<LOD	163
16	226	323	488	3	82	<LOD	131	164
17	466	372	334	7	82	<LOD	131	165
18	973	208	<LOD	10	27	<LOD	<LOD	166
19	2485	<LOD	509	10	60	<LOD	194	167
20	1465	361	612	4	159	24	35	168
21	92	<LOD	290	10	34	<LOD	24	169
22	341	133	292	13	39	<LOD	<LOD	170
23	<LOD	143	238	12	<LOD	<LOD	42	171
24	1185	916	1198	6	55	46	41	172
25	281	343	626	16	<LOD	49	<LOD	173
26	254	282	437	12	47	<LOD	<LOD	174
27	229	320	460	22	80	58	<LOD	175
28	131	<LOD	161	4	<LOD	<LOD	<LOD	176
29	<LOD	2149	2835	1087	160	148	59	177
30	220	224	244	69	<LOD	<LOD	<LOD	178
31	<LOD	8526	9988	4377	166	593	99	179
32	<LOD	1390	1929	464	32	59	25	180
33	226	408	755	58	<LOD	48	31	181
34	467	963	1661	5918	<LOD	92	34	182
35	278	640	790	82	33	31	<LOD	183
36	<LOD	3087	5384	89	40	57	99	184
37	<LOD	3162	2748	43	<LOD	18	128	185

^aLinear isomers with calibration standards for quantification

187

Table S3c. Concentrations (pg/L) of branched isomers^a of poly- and perfluoroalkyl substances

Site	br-PFHxS	br-PFOA	br-PFOS	br-MeFOSAA	br-EtFOSAA
1	201	550	181	56	81
2	695	1635	4298	27	51
3	777	1135	1272	132	203
4	590	1741	504	52	35
5	8228	8647	2303	27	56
6	481	1542	753	<17	127
7	741	1808	671	<17	22
8	896	1775	756	<17	<12
9	<64	114	<51	<17	<12
10	982	1350	286	<17	<12
11	483	1378	364	21	19
12	<64	249	<51	<17	<12
13	76	174	<51	<17	<12
14	76	106	<51	<17	<12
15	<64	125	<51	<17	<12
16	<64	146	75	<17	<12
17	<64	118	<51	<17	<12
18	<64	151	<51	<17	<12
19	<64	3015	81	<17	<12
20	78	3250	133	35	<12
21	<64	74	<51	<17	<12
22	<64	<68	<51	<17	<12
23	<64	<68	<51	<17	<12
24	164	678	215	<17	<12
25	80	306	145	<17	<12
26	93	333	144	<17	<12
27	<64	227	89	<17	<12
28	<64	284	171	<17	<12
29	471	2602	622	35	32
30	<64	193	54	<17	<12
31	1578	8745	1848	31	110
32	294	790	408	<17	12
33	113	557	208	<17	13
34	176	739	303	<17	17
35	158	691	195	<17	<12
36	539	2660	953	<17	<12
37	700	2512	609	<17	<12

^aBranched isomers were quantified based on peak areas assuming the same response factors as the linear isomers.

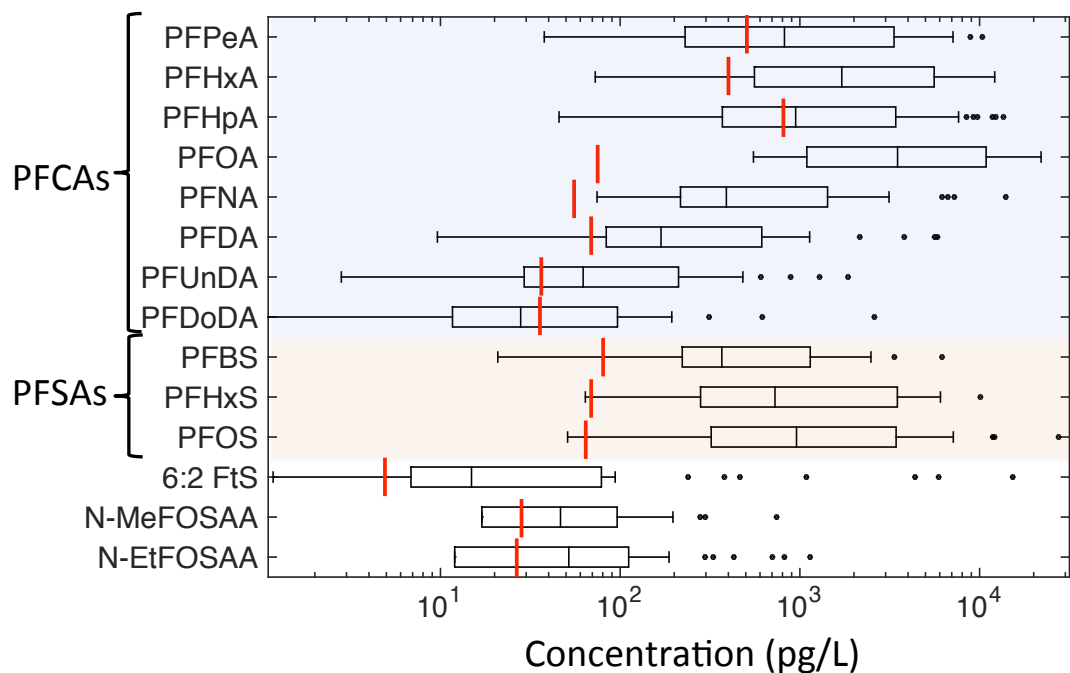
192 Table S4. PFAS concentrations measured in U.S. surface waters in this study and previous work.

PFASs, ng/L (minimum/median/maximum)

Location/ (sites, sampling year)	PFPeA	PFHxA	PFHpA	PFOA	PFNA	PFDA	PFUnDA	PFDoDA	PFBS	PFHxS	PFOS
Tennessee (n=40, 2000)⁴				<25 <25 598							17 52 144
North Carolina (n=11, 2006)⁵		<1 5.14 23	<1 14.8 329	<1 12.6 287	<1 5.7 194	<1 13.2 120	<1 5.67 52.1	<1 1.95 4.46	<1 2.46 9.41	<1 5.66 35.1	<1 28.9 132
Georgia (n=11, 2006)⁶				3 238 1150	<0.6 5.6 369	<0.1 2.1 131	<0.1 <0.1 99				1 6 318
Upper Mississippi River Basin (n=177, 2008)⁷	<LOQ 0.71 31.5	<LOQ 1.59 53.4	<LOQ 2.16 90.2	<LOQ 2.07 125	<LOQ 0.71 72.9	<LOQ 0.71 42	<LOQ 0.71 29.1	<LOQ 0.71 24.7	<LOQ 0.71 84.1	<LOQ 0.71 169	<LOQ 3.01 245
Georgia (n=8, 2008)⁸	<MDL 57 149	<MDL 68 149	<MDL 46 100	<MDL 102 204	<MDL 21 46	<MDL 25 46	<MDL	<MDL	<MDL 124 260	<MDL 13 31	<MDL 150 321
New Jersey (n=12, 2009)⁹	<5 <5 15	<5 <5 17	<5 <5 10	<5 11 100	<5 <5 19	<5 <5 ND			<5 <5 6	<5 <5 46	<5 <5 43
Rhode Island and New York Metropolitan Region (n=37, 2014, this study)*	<0.4 0.8 10.4	<0.3 1.7 48.4	<0.6 0.9 48.2	0.3 3.5 56.0	0.1 0.4 14.0	<0.03 0.2 5.8	<0.03 0.1 1.9	<0.03 0.0 2.6	<0.08 0.4 6.2	<0.12 0.7 43.0	<0.10 0.96 27.5

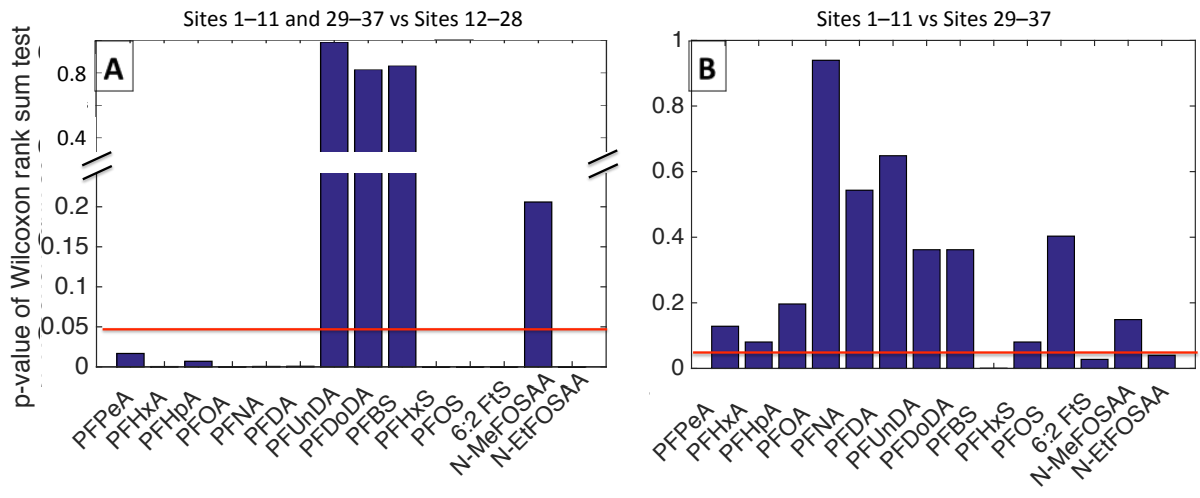
193 *PFOA, PFHxS and PFOS reported here include both linear and branched isomers. The branched
194 isomers were quantified based on peak areas assuming the same response factors as the linear isomers

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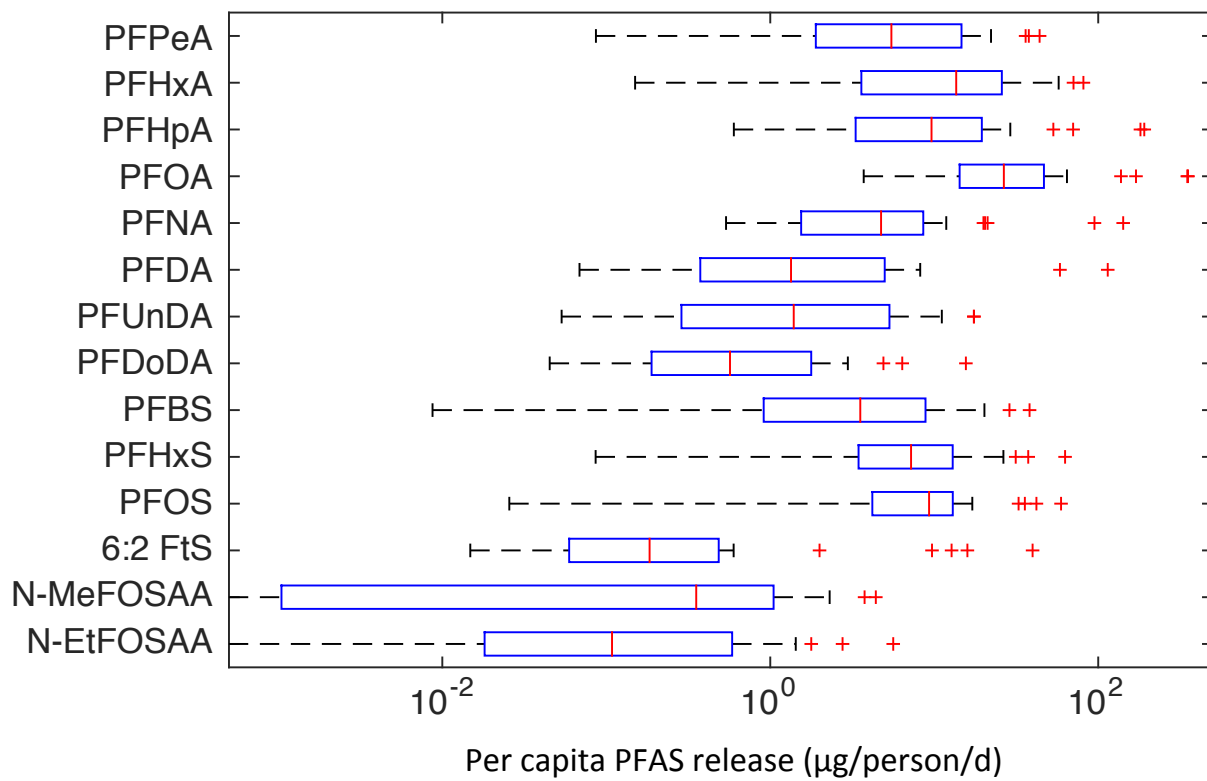
Figure S2. Concentrations of 14-PFASs measured in 37 rivers and estuaries in Rhode Island (RI) and the New York Metropolitan area (NY/NJ). The limit of detection (LOD) for each compound is shown as a red bar. Those below detection are assigned values based on the robust ROS (Regression on Order Statistics) approach for censored log-normally distributed environmental data as described by Helsel.²



203

204 Figure S3. Significance levels for Wilcoxon rank sum tests comparing PFAS concentrations (a)
 205 between urban sites (RI sites 1–11 and NY/NJ sites 29–37) and rural sites 12-28 (b) RI sites 1–11 and
 206 NY/NJ sites 29–37. Red line denotes $p=0.05$, which we use to indicate statistical significance.

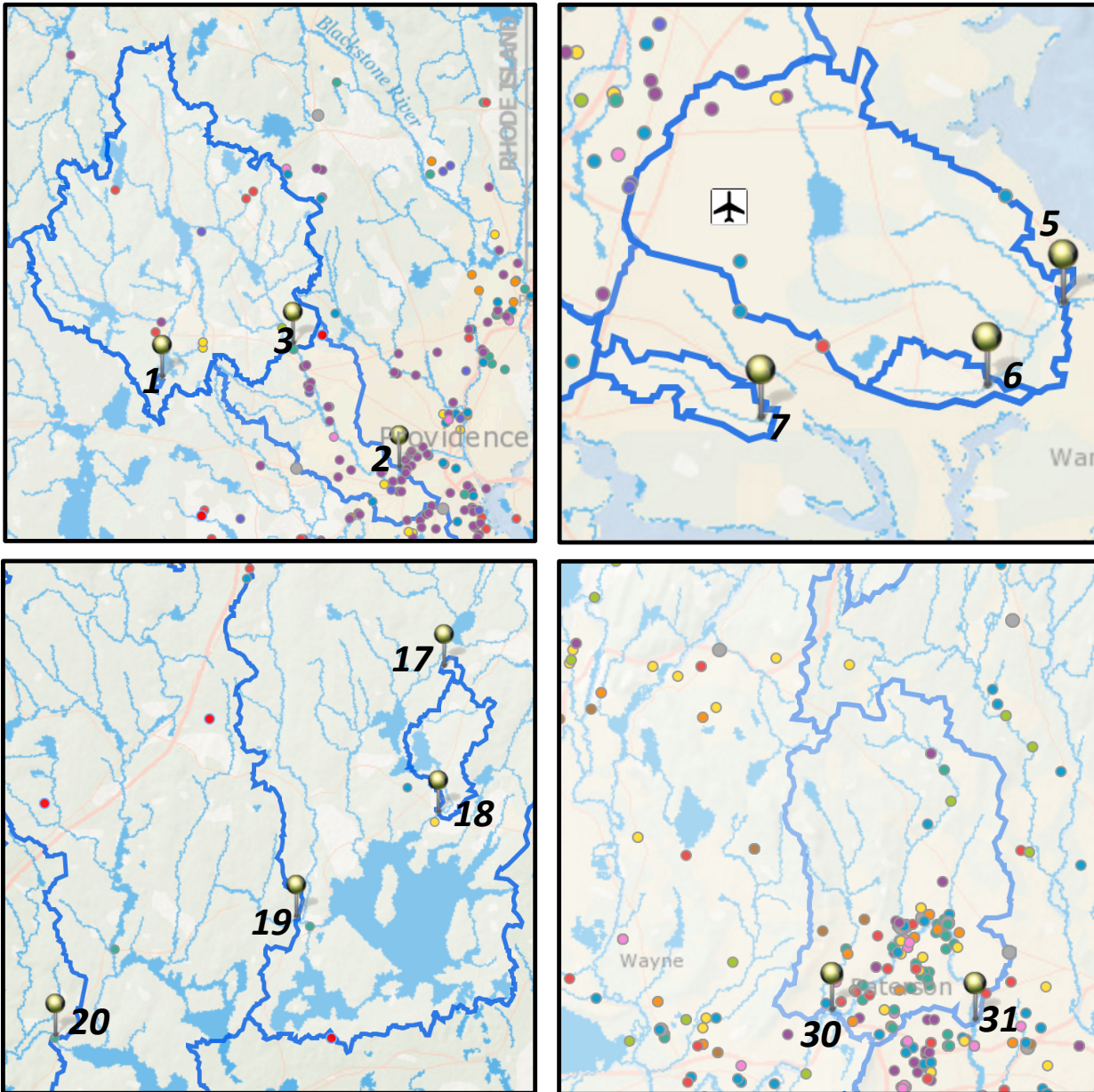
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210 Figure S4. Per-capita release of PFAS ($\mu\text{g}/\text{person}/\text{d}$) estimated based on measured PFAS
211 concentrations, water flow rate and upstream population at each sampling site.

- Waste Management
- Printing Activity
- Sewage Treatment
- Metal Coating Plating
- Paint, Coating, Adhesive Manufacturing
- Semiconductor Manufacturing
- Paper Manufacturing
- Petroleum Coal
- Textile Mills
- Petroleum Coal Products Manufacturing



212

213 Figure S5. Maps showing sampling sites with distinct PFAS composition profiles, the upstream
 214 watersheds and the potential source contributions.

215 Table S5. Impact factors for potential PFAS sources in watersheds upstream of the non-estuarine
 216 sampling sites.

sites	Impact from facilities upstreams													
	Upstream Area (km ²)	Upstream Population	Upstream Population Density (Person/km ²)	Metal Coating Plating	Paint, Coating, Adhesive Manufacturing	Paper Manufacturing	Petroleum Coal Products Manufacturing	Printing Activity	Printing Ink Manufacturing	Semiconductor Manufacturing	Sewage Treatment	Textile Mills	Waste Management (incl. Landfills)	Airport
1	0.1	26	190											
2	124.4	87446	703	1.7E+00	3.7E-01	1.3E-02	2.6E-01	5.5E-05		5.1E-01		1.1E-01	5.5E-08	
3	97.6	24495	251	3.2E-04		1.7E-06				1.2E-06	3.7E-01		3.5E-04	
4	598.2	208255	348	1.3E-01	5.2E-04	1.7E-04	5.8E-04	5.5E-03	1.4E-08	2.8E-02	4.9E-03	1.7E-04	1.6E-01	
5	16.0	16509	1032	9.9E-04				3.0E-03					3.4E-04	4.1E-03
6	1.0	1174	1196											
7	1.4	1792	1306											
8	15.9	12476	783	7.1E-01			2.5E-04	6.1E-04		6.0E-01				
10	59.3	14886	251	4.4E-03	1.3E-03			9.1E-05		3.3E-04			3.0E-04	
11	5.5	1394	254		2.1E-01								6.9E-01	
12	0.6	123	218											
13	12.1	1951	161											
14	21.2	4811	226											
15	24.4	5870	240											
16	33.8	8835	262											
17	0.5	20	42											
18	10.2	746	73											
19	235.3	23112	98					2.0E-04		4.3E-05		3.9E-01		
20	561.1	43081	77			3.8E-15		4.8E-13		1.1E-13		9.1E-01		
21	0.01	1	69											
22	65.7	2647	40											
23	0.3	12	44											
29	12799.8	1994644	156	9.6E-02	1.4E+00	2.2E+00	4.9E-09	5.6E+00	2.0E+00	7.5E-02	3.6E+00	2.8E-01	1.8E+00	
30	2015.8	854842	424	1.2E-02	4.7E-03	6.5E-04	8.0E-04	9.8E-02	1.8E-05	6.3E-02	9.4E-04	3.8E-03	3.8E-04	
31	2090.0	1050694	503	3.2E-02	1.1E-01	2.6E-02	7.6E-04	4.2E-02	1.6E-01	1.2E-01	2.4E-04	1.6E-01	3.1E-02	
34	3345.1	3737691	1117	2.1E-01	5.2E-04	1.8E-01	6.2E-03	2.7E-01	1.5E-04	3.6E-01	2.9E-03	1.7E-01	6.1E-02	2.5E-04
36	2406.9	1903628	791	1.7E+00	3.5E-01	1.8E-01	4.2E-01	7.8E-01	7.6E-03	4.8E-02	6.0E-01	4.2E-02	7.0E-01	
37	2303.7	1494335	649	1.9E+00	5.5E-01	1.9E-01	8.0E-02	1.1E+00	3.7E-02	1.4E-01	1.3E-03	5.8E-01	5.7E-02	

217 *Facilities are based on the U.S. EPA Facility Registry Service database.¹⁰ Impact of potential point
 218 sources as a function of distance from sampling locations by assuming exponential decay in the
 219 concentration (i.e., $Impact = I/e^d$, where d = hydrological distance, km)
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