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1 **The Ornthodolite as a tool to quantify animal space use and habitat selection; a case**  
2 **study with birds diving in tidal waters**

3

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6

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13

14 **Abstract**

15 Animal-attached technologies can be powerful means to quantify space-use and  
16 behaviour, however, there are also ethical implications associated with capturing and  
17 instrumenting animals. Furthermore, tagging approaches are not necessarily well-  
18 suited for examining the movements of multiple individuals within specific, local areas  
19 of interest. Here, we assess a method of quantifying animal space use based on a  
20 modified theodolite with an inbuilt laser rangefinder. Using a database of > 4,200  
21 tracks of migrating birds, we show that detection distance increases with bird body  
22 mass (range 5 g - >10 kg). The maximum distance recorded to a bird was 5500 m and

23 measurement error was  $\leq 5$  m for targets within this distance range; a level  
24 comparable to methods such as GPS tagging. We go on to present a case study where  
25 this method was used to assess habitat selection in seabirds operating in dynamic  
26 coastal waters close to a tidal turbine. Combining positional data with outputs from a  
27 hydrographic model revealed that great cormorants (*Phalacrocorax carbo*) appeared to  
28 be highly selective of current characteristics in space and time; exploiting areas where  
29 mean current speeds were  $< 0.8$  m s<sup>-1</sup>, and diving at times when turbulent energy  
30 levels were low. These birds also orientated into tidal currents during dives. Taken  
31 together, this suggests that collision risks are low for cormorants at this site, as the two  
32 conditions avoided by cormorants (high mean current speeds and turbulence levels),  
33 are associated with operational tidal turbines. Overall, we suggest that this modified  
34 theodolite system is well-suited to the quantification of movement in small areas  
35 associated with particular development strategies, including sustainable energy  
36 devices.

37

38 **Keywords:** GPS, movement ecology, seabird, tidal turbine, habitat use

39

40

41 **Competing interests:** We have no competing interests.

42

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## 63 **Introduction**

64 Electronic tagging can now be used to provide data on the spatial movements of  
65 animals with sub-second temporal resolution (Ropert-Coudert & Wilson 2005).  
66 Nonetheless, the size of loggers providing reasonably high frequency data for  
67 significant durations (i.e. substantial battery life) means that the use of this technology  
68 is still limited to relatively large animals (Chittenden *et al.* 2009; Ropert-Coudert &  
69 Wilson 2005; Wilson *et al.* 1986). The present recommendation followed by many  
70 scientists, that the weight of the logger should not exceed 3 % of the weight of the  
71 bird, is a contentious issue. Indeed, the only way to know that there are no deleterious  
72 impacts would be to compare and evaluate the behaviour of “control” birds without  
73 any device attached (Nicolaus *et al.*, 2008). There are also ethical considerations  
74 associated with the capture, handling and instrumentation of individuals (Wilson &  
75 McMahon 2006). In some cases, tags can affect the behavior of the individual and  
76 hence influence the very measurements such devices are designed to make (Elliot  
77 2016; Saraux *et al.* 2011; Stothart *et al.* 2016).

78

79 Beyond the ethical implications of instrumenting animals, biotelemetry may not be the  
80 best approach for addressing particular study questions. For instance, questions such  
81 as how animals operate with respect to specific developments may be concerned with  
82 the movements of large numbers of individuals, or even different taxa, in a relatively  
83 small area. In these cases, tagging may not be ideal, as tagged individuals may not

84 necessarily use the site of interest, or if they do, patterns of resource selection may be  
85 based on a low number of individuals, relative to the number using the site.

86

87 Eulerian, or static measurements, have also been important in quantifying animal  
88 locations (Turchin 1998). Like tagging, the range and resolution of resulting data varies  
89 among the different techniques. Aerial surveys can provide accurate information on  
90 the distribution of individuals over large areas (Camphuysen *et al.* 2004); however, the  
91 costs of this technique mean that few surveys tend to be run per study, limiting the  
92 ability to monitor changes in space-use through time. This method also provides point  
93 counts rather than movement trajectories. Radar can provide vast, high resolution  
94 datasets on space-use relative to a particular location, or series of connected locations  
95 (Alerstam 1990; Chapman & Graber 1997; Eastwood 1967; Gauthreaux & Belser 2003;  
96 Gürbüz *et al.*, 2015) and it can also be used to derive movement trajectories. However,  
97 it is rarely possible to automate the identification of targets or even achieve  
98 identification at all (McCann & Bell 2017). The data processing requirements (e.g. to  
99 remove signal backscatter, from the movement of non-target objects) are also  
100 substantial, although international initiatives such as the European Network for the  
101 Radar surveillance of Animal Movement may lead to advances here (Alves *et al.* 2014).

102

103 Theodolites are instruments originally used for land surveying, and have also been  
104 used for animal tracking (Bailey & Thompson 2006, Piersma *et al.* 1990). This approach  
105 to animal tracking combines aspects of both Eulerian and Lagrangian methods, as

106 whilst it is place-based, individuals can be identified and, in some cases, selected  
107 according to species or behavior. Individuals can also be followed, allowing users to  
108 reconstruct movement tracks (Bailey & Thompson 2006). Theodolites are relatively  
109 straightforward when it comes to the collection and processing of data (relative to  
110 radar data, for instance). They can also provide locations with high accuracy and  
111 precision when compared to land-based or seagoing surveys that use grids to allocate  
112 observations to geographic areas. Traditional theodolites measure azimuth and  
113 elevation angles and do not measure distance directly. If the height of the observer is  
114 known, relative to the target, a single theodolite can be used to derive the 2D position  
115 of an object on a flat substrate (McCormack 1991). Otherwise, a dual theodolite  
116 system is needed to derive the target's 3D position (Tucker & Schmidt-Koenig 1971).  
117 Other non-invasive static methods like 3D video tracking can yield similar precision  
118 with finer temporal resolution than theodolites (positional error of 3D video tracking  
119 can be a few centimetres at closer ranges) but these operate across ranges of up to a  
120 few hundred metres (Cavagna *et al.* 2008; De Margerie *et al.* 2015; Evangelista *et al.*  
121 2017).

122

123 Theodolites that incorporate a distance measure can be used to estimate a target's  
124 position accurately whether the animal is on the substrate or in flight (Wilson & Wilson  
125 1988; Hedenstrom & Alerstam 1994; Piersma *et al.* 1990). Various researchers have  
126 developed the system further in order to estimate the airspeeds of flying animals.  
127 Double theodolite systems were first used to quantify airspeeds using triangulation of

128 horizontal and vertical angles to resolve distance and subsequently combining  
129 positional data with measurements of wind speed (Tucker & Schmidt-Koenig 1971).  
130 This superseded previous methods where birds were followed by vehicles to estimate  
131 ground speed (Michener & Walcott 1967). A further modification was proposed by  
132 Pennycuick (1982), who combined an anemometer and a coincidence rangefinder to  
133 produce a single, portable system that could track objects in flight and estimate their  
134 airspeed. This system is now based on a laser rangefinder incorporated in a pair of  
135 Vector 21 binoculars (Pennycuick *et al.* 2013), which measure distances directly and  
136 provide improved accuracy and precision. As this system was specifically developed to  
137 quantify airspeed in birds, it is only very recently that it has been used to examine  
138 animal distributions (Hedenström & Åkesson 2016; Shepard *et al.* 2016). We suggest  
139 that this technique has potentially broad ecological applications, which have yet to be  
140 fully realised. We note, however, that the incorporation of the laser range-finder  
141 means that the system cannot get a return from the water surface or the smooth,  
142 water-covered surfaces of most cetaceans.

143

144 In this study, we use the Vector Ornithodolite (hereafter, VOD) to examine the factors  
145 affecting the fine-scale space-use of seabirds operating in a highly dynamic tidal  
146 environment. Data were collected in Ramsey Sound, Pembrokeshire, UK, where a tidal  
147 turbine is currently installed but non-operational (Evans *et al.*, 2015). We use  
148 hydrodynamic numerical model simulations of current flows in the Sound to  
149 investigate the conditions that birds select during foraging. The utility and limitations



150 of the equipment for the wider community of movement ecologists are also examined,  
151 specifically through the assessment of the measurement error and whether maximal  
152 detection distances vary according to body size. The latter was investigated using a  
153 large database of 4,284 positional fixes taken from birds during migration.

154

## 155 **Methods**

### 156 System performance

157 The workings of the VOD have been described in detail elsewhere in terms of the use  
158 of this equipment for the measurement of animal location and airspeed (Pennycuick *et*  
159 *al.* 2013), hence only a summary will be given here. The Vectronix USMC Vector 21 is a  
160 pair of binoculars with an inbuilt laser rangefinder, digital compass (giving azimuth  
161 angle), and inclinometer, providing both inclination and azimuth angles (Vectronix™  
162 2004). The user obtains co-ordinates of a target by pressing and releasing two buttons  
163 when the target is between the cross-hairs in the view finder and positions are sent to  
164 a laptop via a cable. In this study, a simple programme was written in Visual Basic  
165 (Microsoft) to enable users to append information including species and behavior to  
166 each set of co-ordinates.

167

168 The Vector measures distances from 5 m to over 10 km (Vectronix™ 2004). The error  
169 associated with distance measurement must be ascertained by the user. We therefore  
170 used the following protocol to quantify this: Locations were taken to a fixed target, in  
171 this instance an area next to a prominent ledge, approximately 1 m<sup>2</sup> situated on

172 Mumbles boat house (51°34'12.0"N 3°58'32.4"W) in Swansea Bay. Fixes were taken at  
173 increasing distances from 50 m to 5 km, with 10 fixes being taken at each of 12  
174 distance intervals. Intervals were selected based on the ability to have a clear view of  
175 the target.

176  
177 The ability to get returns from the laser (and hence record the target's co-ordinates) in  
178 some cases, may be related to the target characteristics (i.e. size, color etc.) and the  
179 experience of the observer. In order to examine how maximum distance varied in  
180 relation to body size, multiple, sequential, locations of birds migrating past Ottenby  
181 observatory, southern Sweden, were collected from 2012 to 2017. The methods are  
182 detailed in full by Hedenström and Åkesson (2016). Each series of locations from an  
183 individual bird is hereafter referred to as a 'run'. The furthest distance measurement  
184 per run was selected for further analysis. We note that observers were not aiming to  
185 get returns from the furthest targets they could observe and the resulting distances  
186 are therefore only an indication of those that could be attained. Data were collected by  
187 experienced ornithologists, with one observer operating the VOD and the other  
188 identifying birds using a telescope, although it is possible for a single person to operate  
189 the system using a telescope to identify distant targets where necessary. This approach  
190 thus provides an insight into the distances that can be obtained where experience in  
191 bird identification is not a limiting factor.

192

193 Data analysis

194 Generalised Linear Models were used to assess whether the maximum distances were  
195 affected by the mass, wingspan and flock size of the target, with the global model  
196 including these terms and an interaction between body mass and flock size. As mass  
197 and wingspan are related, the residual variation from the allometric prediction of  
198 wingspan was used in the model, with the predicted wingspan being taken as  $\text{mass}^{0.39}$   
199 for each of the 151 study species (Pennycuik 2008). Distance and body mass were  
200 log<sub>10</sub> transformed and regressions were run in base R (R core group 2017). Models  
201 were compared using their AIC scores.

202

203 Space use within Ramsey Sound

204 Data collection took place in Ramsey Sound, Pembrokeshire, from a vantage point  
205 based near St Justinian 51°52'42.4"N 5°18'38.4"W, which provided views of the entire  
206 Sound. Data collection began on the 24<sup>th</sup> April 2017 and included a total of 35 visits.  
207 Surveys were conducted in periods of calm and dry weather with good visibility (i.e.  
208 where the horizon remained visible), and for sea states of  $\leq 2$  on the Beaufort scale  
209 (corresponding to wind speeds of  $\leq 3 \text{ m s}^{-1}$ ). The locations of seabirds within the  
210 Sound were recorded across the entire tidal cycle using the VOD. A full scan of the area  
211 was completed every 15 minutes for a minimum session length of 4 hours and the tidal  
212 state was noted (flood, ebb or slack water which occurred 2.5 hours after high and low  
213 water respectively). Locations were recorded for all birds observed within a scan, with  
214 birds being identified to species level (distance permitting). Group size and behavior

215 were also recorded. If foraging behavior was observed, individuals were followed after  
216 the main scan in order to take positional fixes at the start and end points of individual  
217 dives. Care was taken to ensure the entire Sound was searched systematically during  
218 each 15-minute scan to reduce any spatial bias in sightings.

219

220 Azimuth, elevation angle and distance data for bird observations were subsequently  
221 converted to latitude and longitude, using the observer's known GPS position. These  
222 polar coordinates were then used to identify areas of high general use within the  
223 Sound and areas specifically associated with foraging. Distributions were plotted using  
224 fixed kernel density estimation (KDE) in the statistical analysis software R using the  
225 packages 'ggmap' (Khale and Wickham 2016) and 'MASS' (Ripley *et al.* 2017). An  
226 estimate of all-encompassing foraging range of great cormorants (*Phalacrocorax*  
227 *carbo*), was provided by the 90% KDE contour (as the most frequent diving species).

228

229 To investigate how cormorants dived in relation to current vectors, the horizontal  
230 distance covered between the start and end points of a dive was calculated using the  
231 Haversine formula (Jenness 2011). The dive bearing was also calculated, assuming the  
232 bird followed a straight line from its start to end position (Wilson & Wilson 1988). The  
233 convention with axial data, such as those collected here, is to transform the bearings  
234 so they lie between 0 and 180°, calculate the mean, and finally back-transform to plot  
235 the data as a circle diagram (Cox 2001). These data were visualised using Oriana, which

236 was also used to perform a Rayleigh's Z test to assess whether bearings conformed to a  
237 uniform distribution (Kovach 2011).

238  
239 The Telemac-2D (v7r2) open-source hydrodynamic ocean modelling software suite was  
240 used to quantify spatial and temporal variation in current speed ( $\text{m s}^{-1}$ ), turbulent  
241 energy ( $\text{J kg}^{-1}$ ) and water depth (m) within Ramsay Sound for the entire study period.  
242 This model solves the depth integrated Saint-Venant free surface flow equations,  
243 derived from the full Reynolds Averaged Navier Stokes (RANS) equations for  
244 momentum and continuity (Hervouet 2007). The finite element unstructured mesh  
245 varies from coarse (approximately 10 km at model boundaries) to fine (approximately  
246 50 m around the North Wales coast) for a domain encompassing the Irish Sea ( $50^{\circ}\text{N}$  to  
247  $56^{\circ}\text{N}$ ,  $8^{\circ}\text{W}$  to  $3^{\circ}\text{W}$ ). Values of hydrodynamic conditions were provided at  
248 approximately 300 m and 10-minute resolution in Ramsay Sound. Model simulations  
249 are forced at domain boundaries with tidal harmonic constituents only and no other  
250 influences to dynamics are considered. However, in shallow coastal regions where the  
251 water column remains well mixed, vertically homogenous velocities can be expected  
252 above the bottom boundary layer. Therefore depth-averaged approximations provide  
253 good estimation of flow characteristics. Full details of numerical model set up,  
254 calibration and validation are detailed elsewhere (Piano *et al.* 2017; Piano *et al.* 2015).

255

256 To facilitate comparisons with the spatial and temporal distributions of dives, values of  
257 hydrodynamic conditions were transposed onto an orthogonal grid of 100 m resolution  
258 using kriging interpolation. Kriging was performed using the 'automap' package in R

259 (Hiemstra *et al.* 2009). The spatial distribution of dives was compared to that of mean  
260 current speeds in Ramsay Sound. As tidal environments are broadly divisible into areas  
261 of comparatively fast and slow mean current speeds (Benjamins *et al.* 2015, Waggitt *et*  
262 *al.* 2017), such comparisons provide useful insights into general habitat-use.  
263 Furthermore, as tidal stream turbines generally occupy areas of faster mean current  
264 speeds (Fraenkel 2006), these comparisons would also identify the likelihood of  
265 interactions between diving birds and installations (Waggitt & Scott 2014). The  
266 temporal distribution of dives across tidal states (ebb-flood) within persistently used  
267 areas was also examined in relation to current speed, turbulent energy and depth.  
268 These comparisons would identify the hydrodynamic conditions experienced by  
269 individuals during dives. Estimates of hydrodynamic conditions were extracted using  
270 the mean coordinates of dives, which were highly aggregated.

271

## 272 **Results**

### 273 System performance

274 Variance in distance measurements increased with distance (Figure 1). The standard  
275 deviation was around 1-2 m for distances < 2 km and was close to 0.1% of distance  
276 measurements overall. Note that the error measured here reflects random deviation  
277 only. The overall accuracy given in the user manual is  $\pm 1$  m (Vectronix™ 2004).

278

279 Over 4,200 runs were recorded for migrant birds in Sweden. These were filtered to  
280 obtain a “maximum distance” for each of the 151 species in the dataset. The smallest

281 species recorded was a goldcrest (*Regulus regulus*) weighing ~ 6 g, and the maximum  
282 distance achieved for this species was 913 m. A whooper swan (*Cygnus cygnus*) with a  
283 weight of ~ 9 kg, was recorded 2,742 m from the observer, and the largest overall  
284 distance, obtained from a migrating flock of barnacle geese (*Branta leucopsis*), was  
285 5,498 m. The majority of observations were from single birds, with 56 observations  
286 being from flocks of between 2 and an estimated 450 individuals.

287

288 Maximum distance was best explained by a model with bird body mass as the sole  
289 explanatory variable (beta = 0.12, F=120.9, df=149, p<0.001, adj R<sup>2</sup> = 0.44) (Figure 2). A  
290 model including both mass and the residual wingspan received equivalent support  
291 ( $\Delta$ AIC < 2), although wingspan had a low effect size and was non-significant (beta <  
292 0.001, p = 0.4). The global model that also included flock size and the interaction  
293 between flock size and body mass, showed that this interaction did not significantly  
294 influence maximum distance (z =1.178, df =150, p>0.1) and neither did flock size in  
295 isolation (z =1.531, df=150, p>0.1).

296

#### 297 Current selection within Ramsey Sound

298 Seven seabird species were recorded in 140 hours of survey effort: common guillemot  
299 (*Uria aalge*), razorbill (*Alca torda*), European shag (*Phalacrocorax aristotelis*), Northern  
300 gannet (*Morus bassanus*), great black-backed gull (*Larus marinus*), lesser black-backed  
301 gull (*Larus argentatus*) and the great cormorant (Table 1). The majority of all bird  
302 locations were of individuals rafting or flying, these were not included in the analysis

303 (see supplementary information, S1). The cormorant was the only species with > 10  
304 dives recorded across all surveys (n = 56). Birds avoided the main channel where mean  
305 current speeds were > 1.5 m s<sup>-1</sup>, preferring to both loaf and forage in relatively slack  
306 waters, where mean current speeds were < 0.8 m s<sup>-1</sup> (Figure 3). Cormorants foraged  
307 close to the mainland (0.1 - 0.7 km from the vantage point) in a highly restricted area  
308 which is characterised by low current speeds (min = 0.29 m s<sup>-1</sup>, max = 0.81 ms<sup>-1</sup>, mean  
309 = 0.48 ms<sup>-1</sup>) and water depths of 24.5 - 26 m.

310

311 When it comes to the particular times that cormorants dived, over 80% of cormorant  
312 dives occurred 4 hours after high water or later, when tidal height was rising (Figure 4).

313 There was no clear pattern when it came to the selection of current speeds, which  
314 varied from ~ 0.2 - 1.0 m s<sup>-1</sup> in this area across the tidal cycle (Figure 4). However, dive  
315 times did coincide with periods of falling turbulence, with over 80 % of dives occurring  
316 when the turbulence was < 0.02 J kg<sup>-1</sup> (with turbulence increasing up to a mean of 0.04  
317 J kg<sup>-1</sup>).

318

319 Dive bearings were not uniformly distributed (n= 40, Rayleigh's Z<sub>5.503</sub>, p< 0.005) and  
320 the mean orientation (mean=168.8 ± 8.3°) was into the current (Figure 5). Birds also  
321 covered short distances during dives (mean = 44.5 m, median = 17.8 m, max = 261 m,  
322 min = 0.6 m) supporting the notion that birds are orientating into the flow. However,  
323 birds can be drifted backwards where swim speed is less than the current strength,  
324 effectively producing a bearing that is coincident with the current vector.

325



## 326 **Discussion**

### 327 System performance

328 Our results show that the standard deviation of distance measurements is 1- 2 m  
329 within a 2 km range. The real 3D positional error for moving birds may be increased by  
330 (i) systematic error of the laser distance measurement, and (ii) possible influences of  
331 target size and color (the latter would be difficult to test as this may vary depending on  
332 whether the upper or lower surface of the wing is visible, which varies within the  
333 wingbeat cycle). Errors are also likely in (iii) azimuth and inclination angles, in fact,  
334 azimuth error is probably the main source of positioning error within the VOD.  
335 Measuring these effects is beyond the scope of the present study, but we assume that  
336 these additional sources of position error are of the same order of magnitude as the  
337 random error we measured for distance. Therefore, VOD positioning error is probably  
338 comparable to what is generally accepted for GPS data, which is estimated to be in the  
339 range of 3-28 m (Frair et al. 2010). However, while spatial error in tagging technology  
340 can lead to the misrepresentation of behaviors in a scale-dependent manner  
341 (Browning *et al.* 2017; Costa *et al.* 2010), animal locations can be coded according to  
342 behavior (as well as species, age, and other factors that may be of interest) with the  
343 VOD. The downside of the VOD is that it has relatively intensive requirements when it  
344 comes to survey effort.

345

346 The “maximum” distance recorded to birds migrating past the Swedish coast increased  
347 with bird body mass. This suggests that larger birds are detected more readily at

348 greater distances (the same may be true of larger flocks), which could lead to some  
349 sampling bias in studies recording locations of smaller species. Although large flocks of  
350 birds may be detected by an observer earlier than individuals, this did not influence  
351 the ability to obtain a fix using the VOD in our study. However, flocking may have more  
352 complex effects on the ability to detect targets, for instance the type of flock formation  
353 may influence detection ability: echelon formations may be easier for observers to  
354 spot at distance as opposed to clustered flocks, and these flocking principles could also  
355 be affected by body mass. Larger species, such as geese and swans (*Anatidae* & *Cygnus*  
356 *sp.*) tend to form echelon formations whilst smaller birds, like doves (*Columbidae*),  
357 form clusters. Our experiences during data collection also suggest that it can be  
358 difficult to obtain a fix from species at the smaller end of the size spectrum, even when  
359 they have been detected with optics and are within range. Nonetheless, a location was  
360 obtained from the smallest species (5.5 g) when it was ~ 1 km from the observer and  
361 there were several instances where birds weighing 50 - 100 g were recorded ~ 2 km  
362 away, demonstrating that small birds (including those too small to be tagged) can be  
363 detected and recorded at substantial distances. When it comes to the model  
364 predictions of how the VOD generally performed, birds of 10, 100 and 1,000 g were  
365 readily recorded at distances of 500, 1,000 and 2,000 m, respectively.

366

367 Spatial bias is well documented for land-based surveys, which use distance bands or  
368 grid systems for assessing the locations of birds foraging in near-shore tidal habitats  
369 (Waggitt *et al.* 2014; Waggitt *et al.* 2016a). Here, birds are less likely to be detected if

370 they forage further from the shore. It seems unlikely that this affected the results in  
371 the present study, given that the full length of Ramsey Sound (1.9 km) is less than the  
372 distance over which large birds such as seabirds can be detected and recorded (see S1  
373 for a map of the raw data), and that surveys were conducted in periods of low swell  
374 height. Therefore, while some of the limitations of shore-based surveys still apply to  
375 the use of the VOD in a general sense, with both being based on the use of a telescope  
376 and binoculars to scan for birds, we consider it unlikely that we have underestimated  
377 the usage of fast flowing currents that lie further from the coastline.

378

379 Like GPS tagging and land-based surveys, the VOD can be affected by environmental  
380 conditions. The probability of detecting a target or getting a return with the VOD may  
381 be influenced by sea state and surface conditions (although these factors were not  
382 investigated directly here), and false returns can be given from fog or cloud, although  
383 spurious returns are easy to identify and remove. The system can also be affected by  
384 high winds that make the equipment unsteady to hold and difficult to obtain a fix on  
385 the target bird.

386

387 Many studies have discussed the potential impacts of bird capture and recapture and  
388 the deleterious effects of tags (Bennisson *et al.* 2017; Calvo & Furness 1992; Götmark  
389 1992; Phillips *et al.* 2003; Vandenabeele *et al.* 2011; Wilson & Vandenabeele 2012).  
390 The VOD has advantages here, as it does not involve marking animals and in fact  
391 observers can be placed at a vantage point away from breeding colonies, thereby

392 reducing disturbance. The operational range of the system also far exceeds predicted  
393 flushing distances, which can be a factor in other surveys, including boat-based work  
394 (Schwemmer *et al.* 2011). Finally, the VOD uses a laser tachometer to measure  
395 distance. It seems unlikely that this could have adverse effects on target animals, as  
396 medical literature citing retinal injuries from handheld laser devices indicates that risk  
397 of injury is high if the primary light source is in the 'green' end of the light spectrum  
398 and if pointed directly at the eye from less than one metre away (Luttrull & Hallisey  
399 1999; Mainster *et al.* 2004; Wyrsh & Baenninger 2010).

400

#### 401 Habitat Selection

402 Relatively few studies have quantified habitat use at very fine scales in seabirds (Holm  
403 & Burger 2002; Waggitt *et al.* 2016a; Waggitt *et al.* 2016b; Zamon 2003). Here we  
404 show that cormorants were highly selective in terms of both the area and the time of  
405 the tidal cycle they chose to dive. Ramsey Sound experiences extreme tidal variation  
406 with current speeds  $> 3.5 \text{ m s}^{-1}$  and strong eddy formation over the rocky reefs.  
407 Cormorants dived in a highly localised area of the Sound, showing a general avoidance  
408 of high current speeds in the main channel and areas of high turbulence caused by  
409 rocky reefs at 'the Bitches' and 'Bishop's and Clerks'. While we did not test whether  
410 current speed was the ultimate driver of space-use, it seems likely that birds were  
411 responding to low current speed, prey availability, or a combination of both these  
412 factors. This follows from the observation that birds were orientating into the current  
413 during their dives, as has been hypothesised by previous studies (Gremillet *et al.*, 1998;

414 Wilson & Wilson 1988), and travelling short distances. This pattern of diving repeatedly  
415 in the same place, suggests that cormorants are more likely to be foraging mid-water,  
416 as the rate at which benthic prey would be replenished by the changing tide would be  
417 negligible (Rahel 1988; Schneider & Piatt 1986). Furthermore, the sea bed in Ramsey  
418 Sound consists of gravel and hard rock which is less suitable for benthic fish species  
419 (Fischer 2000). It therefore seems likely that cormorants were targeting shoaling fish in  
420 highly specific areas of the Sound.

421

422 Pelagic fish tend to shoal in areas of minimal water turbulence where current speeds  
423 are relatively low (Cury & Roy, 1989; Fréon & Misund 1999), which generally accords  
424 with the conditions that cormorants selected. As stated above, the area where birds  
425 were diving was characterised by a relatively low current speed compared to the main  
426 channel. Within this, and over the changing conditions of the tidal cycle, birds showed  
427 less selectivity of current speed, diving over a reasonably wide range of available  
428 speeds, up to  $\sim 0.75 \text{ m s}^{-1}$ , and only appearing to avoid the strongest currents of  $\sim 1 \text{ m}$   
429  $\text{s}^{-1}$ . What was striking, however, was the tendency to dive on the flood tide, which  
430 appeared to be strongly related to turbulence levels, with birds selecting times of low  
431 turbulence.

432

433 Overall therefore, when and where cormorants dive appears to be influenced by a  
434 hierarchy of factors operating in space and time. In contrast to these findings, Holm  
435 and Burger (2002) showed that pelagic cormorants (*Phalacrocorax pelagicus*) showed

436 no significant response to tidal height or current strength. In fact, individuals were  
437 more likely to dive in areas of high turbulence within eddies (although values of  
438 current speed and turbulence coefficients are not known) (Sealy 1975). Whilst Waggitt  
439 *et al.* (2017) found that European shags *Phalacrocorax aristotellus* were generally  
440 associated with areas of low mean current strengths among five locations in Scotland,  
441 there were exceptions to this rule. These differences may well be driven by patterns of  
442 prey availability, which can vary between sites and predator species. Nevertheless, this  
443 study agrees with a growing consensus that associations with areas of fast mean  
444 currents are comparatively rare among UK cormorant species (Waggitt *et al.* 2017).

445

446 The need for detailed information on foraging patterns, including how tidal stream  
447 features contribute to foraging success and the direction of travel in relation to  
448 currents, has been highlighted in a recent review by Benjamins *et al.* (2015), as these  
449 factors will ultimately influence the likelihood of animals exploiting hydrodynamic  
450 features. Such associations are important in advancing our understanding of the  
451 species and sites where collisions between seabirds and tidal turbines are most likely.  
452 The risk of diving seabirds being pulled into the path of moving components of tidal  
453 stream devices persists through either being i) passively dragged by strong currents or  
454 by ii) birds actively foraging *with* the direction of the coincident current vector  
455 (Benjamins *et al.* 2015; Waggitt & Scott 2014). However, there are several indicators  
456 that the tidal turbine may represent a relatively low risk to seabirds in Ramsey Sound.  
457 The Sound appears to be used by relatively few seabirds, at least in the conditions

458 sampled during this study, despite the fact that around 2,000 pairs of auks breed on  
459 Ramsey Island (Mitchell *et al.* 2004). Furthermore, seabirds tended not to use the main  
460 channel, which has the greatest current speeds, making it most suitable for marine  
461 energy (ME) installations (Mueller & Wallace 2008; Pelc & Fujita 2002; Piano *et al.*  
462 2017).

463

464 Cormorants have previously been identified as one of the species most at risk from  
465 tidal turbine developments (Furness *et al.* 2012; Langton *et al.* 2011) due to their high  
466 usage of tidal races for foraging and their propensity to forage on benthic prey  
467 (Furness *et al.* 2012; Garthe & Hüppop 2004). Cormorants were the birds most  
468 commonly diving in Ramsey Sound, even though no cormorants are recorded as  
469 breeding on Ramsey and the nearest sizeable colony (Thorn Island, 32 pairs) is located  
470 25 km away (Mitchell *et al.* 2004). Therefore, the cormorants observed in our study  
471 were likely to be non-breeding individuals or those choosing to forage some distance  
472 from the main colony. These individuals may be exposed to lower risk of collision with  
473 tidal turbines than would have been predicted based on previous studies, due to their  
474 tendency not only to forage in areas of low current strength, but also to cover far less  
475 distance than is typical of cormorants diving in other areas (Holm & Burger, 2002;  
476 Schneider & Piatt 1986). If these birds are avoiding areas of high turbulence, then this  
477 would also tend to keep them away from the downstream end of operational turbines,  
478 due to turbulence in the wake (Chen *et al.* 2015). However, further research is required

479 to ascertain whether cormorants show a general avoidance of turbulence, or whether  
480 this represents a site, or individual-specific phenomenon.

481

482 In conclusion, we suggest that the VOD is a potentially valuable addition to the  
483 armoury of tools being used to quantify animal responses to specific, small-scale  
484 anthropogenic impacts, such as renewable energy devices. The system provides 3-d  
485 coordinates within a radius of several kilometres with a measurement error that is  
486 commensurate with GPS tags. Though the initial start-up costs for the VOD are  
487 relatively high (\$18,900 at the time of this study), there is no requirement to pay data  
488 subscriptions over the lifetime of the product or recover any technology from animals  
489 to access data. The variety and quantity of data that can be collected mean that it is  
490 likely to prove cost effective in the longer term, particularly when compared to anima-  
491 borne tags, with each GPS tag costing \$70 - \$800 depending on the method of data  
492 transmission and the hardware itself (Hebblewhite *et al.* 2007). The VOD system has  
493 relatively low training requirements and simple post-processing of the resulting data,  
494 but above all, it represents a method of tracking animals that has little to no ethical  
495 implications for the target animals. Finally, the ability to track even the smallest  
496 passerines means that opportunities arise to assess how a wide range of animals may  
497 respond to developments on land, as well as at sea, from patterns of land use to the  
498 installation of wind farms (Hedenström & Alerstam 1994; Piersma *et al.* 1990).



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719

720 Table 1. The number of individual locations recorded with the VOD in Ramsey Sound  
 721 for animals that were performing behaviour on the water surface (n=301 positional  
 722 fixes).

723

724

| Species                  | Rafting | Diving | Flying |
|--------------------------|---------|--------|--------|
| Guillemot                | 48      | 6      | 9      |
| Shag                     | 2       | 1      | 0      |
| Gannet                   | 1       | 7      | 38     |
| Great black-backed gull  | 2       | 0      | 0      |
| Lesser black-backed gull | 66      | 0      | 0      |
| Cormorant                | 43      | 56     | 19     |
| Razorbill                | 4       | 0      | 2      |

725

726

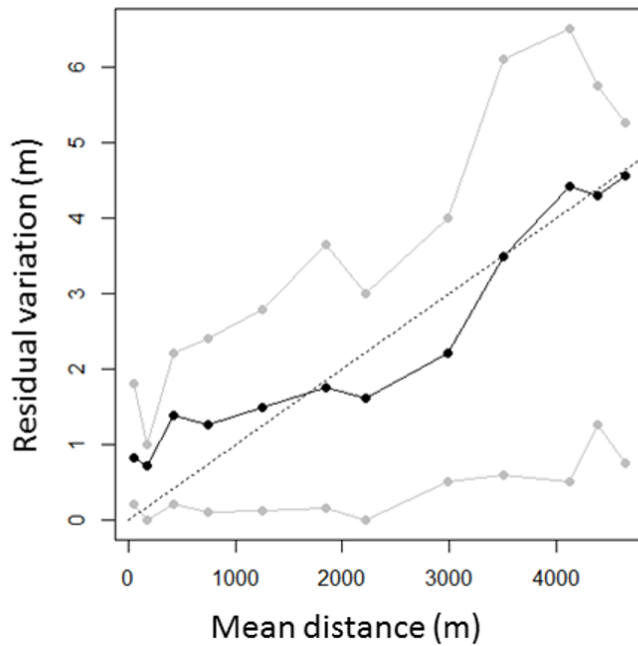
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732 **Figures**

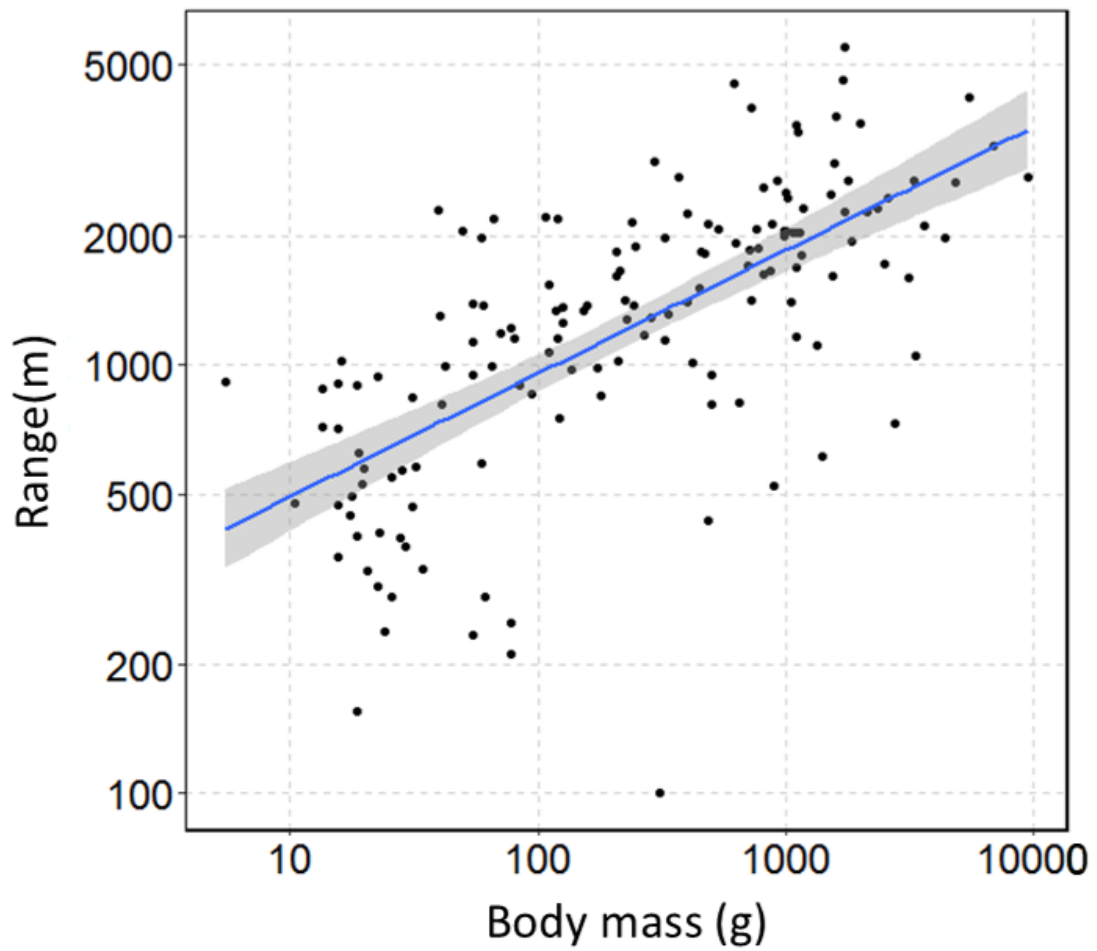
733

734 Figure 1. The residual variation of range measurements as a function of distance to a  
 735 fixed, 1 m<sup>2</sup> target, as measured by the VOD (n = 120 fixes, 10 fixes per distance  
 736 interval). The standard deviations are given in black, while the minimum and maximum  
 737 deviations are given in grey for each distance. The dashed line indicates the variance  
 738 that would be equivalent to 0.1% of the distance value.

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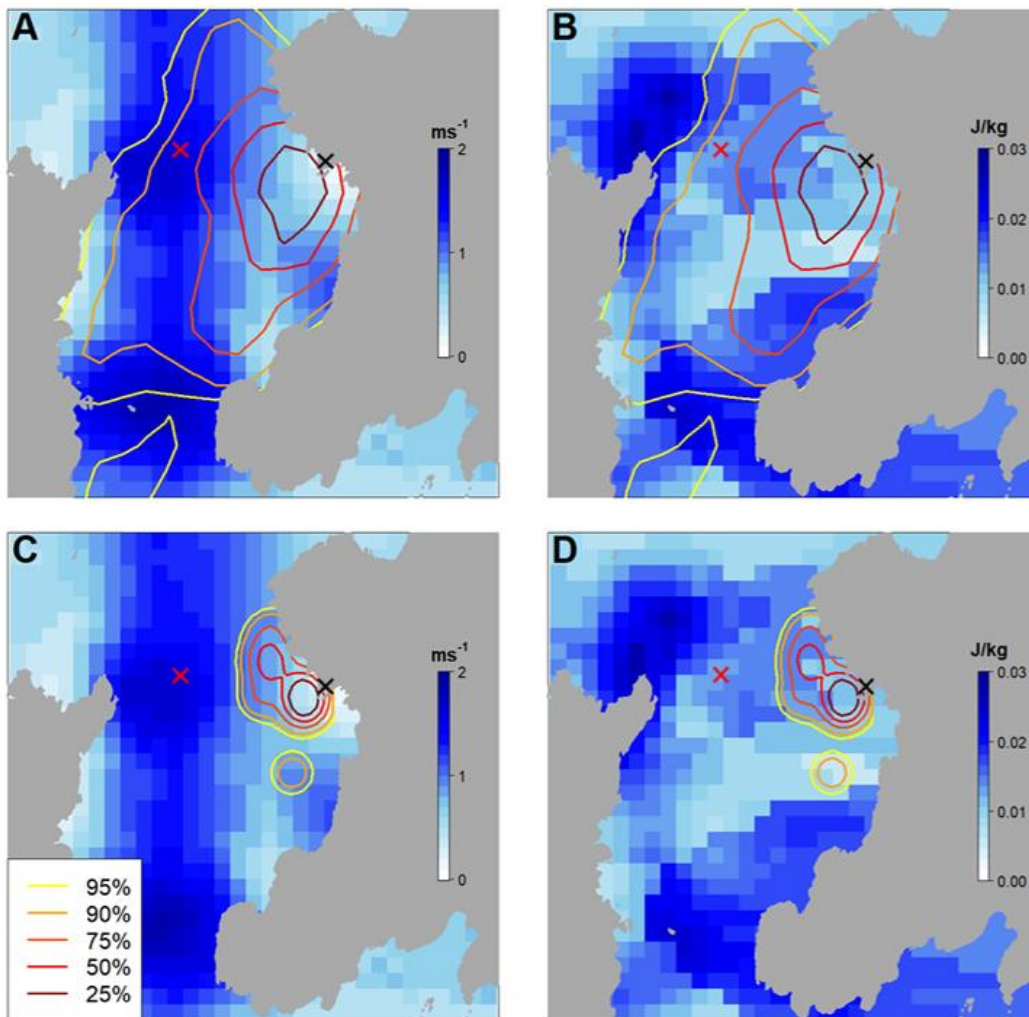


742

743 Figure 2. The maximum range of avian targets from the VOD, in relation to body mass.

744 The blue line equates to the model prediction.

745

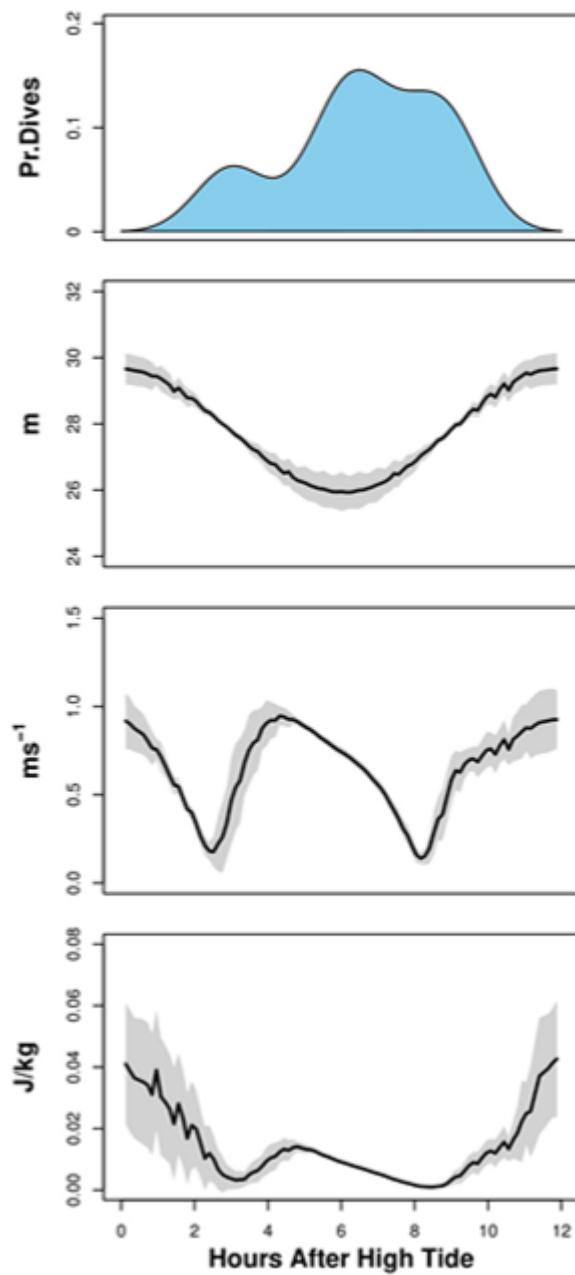


746

747

748 Figure 3. Kernel density contours showing distributions of A) all seabirds and all  
 749 behaviours plotted in relation to mean current speed, B) all species and behaviours  
 750 over mean turbulence, C) all cormorant dives plotted against mean horizontal current  
 751 speed and D) all cormorant dives plotted against mean turbulence in Ramsey Sound.

752 The black cross represents the vantage point at St Justinian's (51°52'42.4"N  
 753 5°18'38.4"W) whilst the red cross marks the location of the DeltaStream tidal turbine  
 754 device.



755

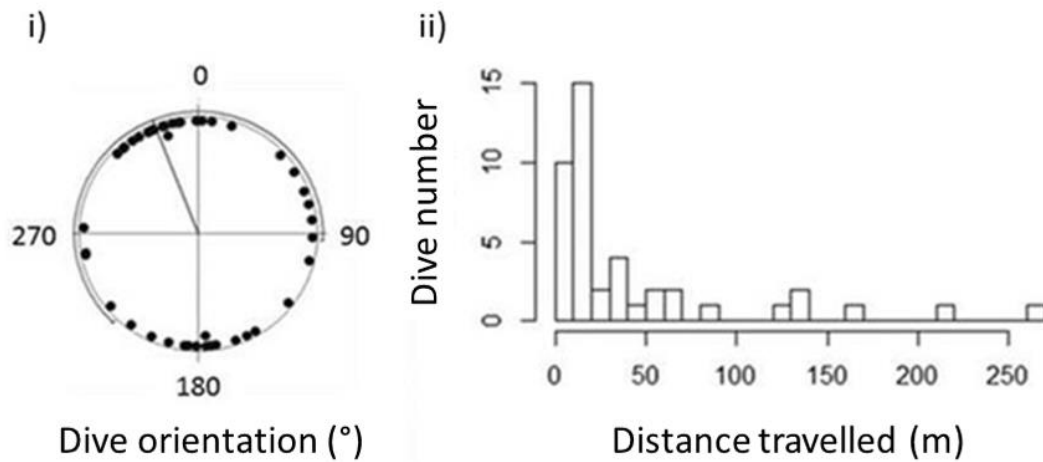
756 Figure 4. The times when cormorants were diving are given in relation to i) tidal height,

757 ii) current strength and iii) turbulence, as modelled using the hydrodynamic model. A

758 density plot is used to show the proportion of dives in relation to time.

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 761 Figure 5. (i) The dive bearings for cormorants foraging in Ramsey Sound illustrate that  
 762 birds forage into the current (mean bearing is given by the line from the centre of the  
 763 circle and the line around the outside indicate the inter-quartile range), which was  
 764 flowing in a Southerly direction from 0 to 180 degrees. The strategy of orientating into  
 765 the flow resulted in birds travelling relatively low horizontal distances during dives, as  
 766 displayed in (ii).