Aerosol and Air Quality Research, 14: 2051–2066, 2014 Copyright © Taiwan Association for Aerosol Research ISSN: 1680-8584 print / 2071-1409 online doi: 10.4209/aaqr.2014.03.0048



# PM<sub>2.5</sub> Chemical Source Profiles of Emissions Resulting from Industrial and Domestic Burning Activities in India

# JeevanLal Matawle<sup>1</sup>, Shamsh Pervez<sup>1\*</sup>, Shippi Dewangan<sup>1</sup>, Suresh Tiwari<sup>2</sup>, Deewan Singh Bisht<sup>2</sup>, Yasmeen F. Pervez<sup>3</sup>

<sup>1</sup> School of Studies in Chemistry, Pt. Ravishankar Shukla University, Raipur-492010(C.G.), India

<sup>2</sup> Indian Institute of Tropical and Meteorology (IITM), New Delhi, India

<sup>3</sup> Chhatrpati Shivaji Institute of Technology, Durg 491001, India

# ABSTRACT

A study has been performed to develop  $PM_{2.5}$  (particles with aerodynamic diameters  $\leq 2.5$ ) chemically speciated source profiles of different industrial and domestic burning practices in India. A total of fifty-five  $PM_{2.5}$  samples have been collected in emissions resulting from (1) industrial furnaces, (2) household fuels, (3) municipal solid waste burning, and (4) welding workshop burning practices, and categorized for eleven subtypes of sources. The collected samples were subjected to chemical analysis for twenty-one elemental (Al, As, Ca, Cd, Co, Cr, Cu, Fe, Hg, K, Mg, Mn, Mo, Na, Ni, Pb, S, Sb, Se, V, Zn), nine ionic (Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>, Cl<sup>-</sup>, F<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>), OC, and EC source indicator species using atomic absorption spectrometry, ion chromatography and carbon analysis (thermal/optical transmittance method), respectively. The carbonaceous fraction was most abundant in household fuel burning emissions (47.6 ± 7.45% to 65.92 ± 13.13%). The ionic/elemental ratios of major inorganic constituents (Ca<sup>2+</sup>/Ca, Mg<sup>2+</sup>/Mg and Na<sup>+</sup>/Na) have been identified to describe the PM<sub>2.5</sub> emissions from combustion or re-suspension dusts during industrial activities. Brick Kiln processes (BKP) have been identified as the major emitter of the highest number of toxic species (Cd, Co, Mo, Sb and V), followed by steel rerolling mills (Hg and Pb) and steel processing industries (As, Ni). The source marker calculations also confirmed that K<sup>+</sup>, Mn, and As are good markers for biomass burning, metallurgical industrial emission, and coal burning, respectively, similar to the findings in previous studies.

Keywords: Source profiles; PM2.5; Source markers; Industrial source; Household fuel burning emission.

# INTRODUCTION

The source apportionment of PM fractions increases with different trends using receptor models, mostly chemical mass balance (CMB), to develop pollution control and mitigation strategies worldwide (Watson *et al.*, 2002; Samara, 2005; Khan *et al.*, 2010; Kong *et al.*, 2010). Receptor models that derive profiles from ambient/indoor measurements requires systematic chemically speciated emission profiles from prominent sources that were possible to effect pollutant concentration at receptor for verification (Hopke, 1999; Watson *et al.*, 2001, 2002; Brook *et al.*, 2003; Gupta *et al.*, 2007). These source profiles are the fractional mass (abundances  $\pm$  uncertainty) of measured chemical species relative to primary PM mass of source emissions (Watson *et al.*).

Tel.: +91 9425242455

*E-mail address:* shamshpervez@gmail.com; shamshp@yahoo.co.in

*al.*, 2001) and one of the most important parameters (Pant and Harrison, 2012) to: 1) create chemically speciated emission inventories (Cass and McRae, 1983; Kuykendal *et al.*, 1990; Chow *et al.*, 2004), 2) apportion receptor concentrations to source (Watson *et al.*, 1984, 1990, 1991, 2001) and 3) estimate toxic and hazardous pollutant emissions (Chow *et al.*, 2004). Chemical abundance in most of earlier source profiles is accompanied by an uncertainty/standard deviation value that intends to represent the errors/variability of that abundance resulting from differences among separate emitters and between samples taken same/different times from the same emitters; which is essential to CMB runs (Watson *et al.*, 1994, 2001; Chow *et al.*, 2003; Ho *et al.*, 2004; Tsai *et al.*, 2007).

Several source profiles of different PM fractions have been developed for different individual sources and applied widespread (Chow and Watson, 1994; Vega *et al.*, 2001; Watson and Chow, 2001; Watson *et al.*, 2001; USEPA, 2002; Chow *et al.*, 2004; Yatkin and Bayram, 2008). These profiles differ with sources, process operating conditions, geology, and geographic seasonality (Watson *et al.*, 2001; Kong *et al.*, 2011; Pant and Harrison, 2012).Additional profiles are

<sup>&</sup>lt;sup>\*</sup> Corresponding author.

always needed for contemporary inventories and source apportionment studies (Watson *et al.*, 2001).Very limited source emission chemical profiles of PM<sub>2.5</sub> have been developed in Indian context (Gadkari and Pervez, 2007, 2008; CPCB, 2008; Patil *et al.*, 2013).

The current study presented  $PM_{2.5}$  chemical source profiles of emissions resulting from 07 different industrial processes including arc-welding workshops and 04 domestic burning practices involved with household cooking activities and municipal solid waste management in India. These  $PM_{2.5}$ chemical source profiles were developed with the objectives to meet the requirement of location specific and latest source profiles that could be applied for chemical mass balance receptor modelling studies; and to update previous source profiles.

# METHODOLOGY

The development of  $PM_{2.5}$  chemical source profiles of selected burning practices have been carried out as a part of a comprehensive source apportionment study of indoor/ outdoor  $PM_{2.5}$  in a dense urban- industrial zone of India (Balakrishna and Pervez, 2009; Pervez *et al.*, 2012) by following a real-world pooled sampling plan using purposeful

study design (Gilbert, 1987). Sampling of PM<sub>2.5</sub> was conducted at four different types of combustion sources, mainly observed in urban areas of Chhattisgarh, India: (1) Municipal waste burning, (2) Household fuel burning (3 sub-types), (3) Mineral based coal fired industries (6 sub-types) and (4) Fabrication workshops (Table 1) (Balakrishna and Pervez, 2009, 2011; Pervez *et al.*, 2012).

#### **Description of Study Area and Source Characteristics**

Raipur-Bhilai, major industrial cities of Chhattisgarh, India located in global scale of  $21^{\circ}14'22.7''N$ ,  $81^{\circ}38.1''E$ and  $21^{\circ}11'0''N$ ,  $81^{\circ}23'6''E$  respectively, having population 1,635,784 (Census, 2011), is known for most dense heavy industrial zone composed of mainly iron processing, thermal power generation and cement production activities. Identification of PM<sub>2.5</sub> emission sources were based on previous reported air monitoring studies (Dubey and Pervez, 2008; Balakrishna *et al.*, 2011; Pervez *et al.*, 2012), layout map, and a survey of current burning practices involved with industrial and domestic activities. About 1200 tonnes of municipal solid waste (MSW) is generated every day in Raipur-Bhilai, region and about 65% of the MSW (about 650 tonne) was disposed-off using open burning procedure on daily basis. Pervez *et al.* (2012) reported that all populations

**Table 1.** Description of domestic and industrial source characterization, material used in burning practices and sampling method.

S.N.	Profile Code	Source Name	Source frequency	Type & Material burnt	Adopted sampling methodology
Munic	ripal solid w	aste burning source			
1.	MSWB	Municipal solid waste	Over than ~5 major	Synthetic and natural	In-plume
		burning	dumping zone and over than ~200 minor burning places	biomaterials in a 1:8 ratio	
House	chold fuel bi	urning sources			
1.	RSFS	Residential solid fuel stoves	45% households	Conventional mixture of Coal, wood, dung with ratio in 2:1:1	In-plume
2.	RKS	Residential kerosene stoves	30% households	Kerosene	In-plume
3.	RLPGS	Residential LPG stoves	25% households	Liquid petroleum gas	In-plume
Miner	al based cod	al fired industries			
1.	СРІ	Cement production	> 9 major and > 12 minor units	Limestone, Gypsum, Steel slag	Chamber re-suspension of bag-filter dust
2.	FEMNI	Ferroy-Manganese industry	> 90 major and minor units	Iron ore, Coal, Dolomite	Chamber re-suspension of bag-filter dust
3.	SPI	Steel production industry	> 130 major and minor units	Iron ore, Coal, Dolomite	Chamber re-suspension of bag-filter dust
4.	СТРР	Coal based thermal power plant	> 50 major and minor units	Coal	Chamber re-suspension of bag-filter dust
5.	SRM	Steel rolling mills	> 150 major and minor units	Scrap cuttings, Steel ingots, Coal	Chamber re-suspension of bag-filter dust
6.	ВКР	Brick kiln process	> 300 major and minor furnaces	Brick clay, Coal, Wood	In-plume
Works	shops				
1.	EAW	Fabrication and welding	> 1200 major and minor units	Iron, welding material	In-plume

of this area use three categorized stoves (based on fuels) for household burning purposes, namely- Liquid petroleum gas (LPG) stoves, Kerosene stoves and stoves with conventional solid fuels. The census of India (2011) describes the household statistical figures of different type of fuel use for cooking purposes: 25% LPG stoves, 30% kerosene stoves and 45% stoves with conventional solid fuel in Central India. As far as different industrial processes carrying out in the study region are concern, about 1351 industries are currently existing in the study region; out of that heavy, medium and small scale industries numbers are 114, 295 and 942, respectively (DoCI, 2012). Nearly 300 major and medium units of iron processing with consumption of 20 Million Tonne (MT) of iron ore/scrap steels per year, ~50 small, medium and major units of coal-burning power generation with coal consumption of 19.03 MT/yr, 15 units of cement production with lime stones/slag/gypsum consumption of 26 MT/yr occurred in previous years. A total of 32.11 MT/yr coals are consumed in production of power generation, steel processes and cement production in the study area. Accordingly, eleven different types of combustion/burning practices involved with industrial, household and outdoor activities have been chosen for the development of PM<sub>2.5</sub> chemical source profiles. Details of source types, justification of their selections, combustion material used have been described in Table 1 and Fig. 1.

#### Sampling

Two different sampling methodologies have been adopted according to nature and characteristics of different burning practices: real-world in-plume and re-suspension sampling (Chow et al., 2004; Patil et al., 2013). In case of industrial combustion processes, stack emitted bag filter house dust samples were re-suspended tocollectPM<sub>2.5</sub> fractions using standard procedures reported elsewhere (Chow et al., 2004; Gadkari and Pervez, 2007). In case of open burning sources related to household fuel burning (RSFS, RKS, RLPGS), outdoor municipal waste burning practices (MSWB), arcwelding workshops (EAW) and brick kilns (BKP), PM<sub>2.5</sub> was sampled from smoke plume. All these open burning sources does not have stacks and PM<sub>2.5</sub> impactors were positioned in smoke plume. Before in-plume sampling, background PM<sub>2.5</sub> were measured for subtraction from the in-plume concentration (Chakrabarty et al., 2013; Dewangan et al., 2013). In case of PM2.5 sampling using chamber resuspension procedures, sampling duration was optimized according to standard filter loading conditions (Chow et al., 2003; DRI, 2011). In case of in-plume sampling for open burning sources, PM<sub>25</sub> sampling event was conducted in 3-4 episodes to cover whole burning processes (Chow et al., 2003; DRI, 2011; Chakrabarty et al., 2013). PM<sub>2.5</sub> has been collected on quartz fiber filters (QFF) (1851-047, Whatman, UK) using Parallel operation of five PM<sub>2.5</sub> samplers (MINIVOL, Ver. 4.2, Model AirMatrics) in each source site at average flow rate of 5 L/min. Filter selection, preparation, calibration, installation, transportation, preservation, weighing measurements and field blanks were conducted by following the quality control and quality assurance described in air sampling protocol reported elsewhere (USEPA, 1999a, b; CPCB 2008; DRI, 2011; Patil et al., 2013). Details of PM<sub>2.5</sub> source sites, their location frequencies in study region and sampling methodology used has been described in Table 1.



Fig. 1. Location map of study region, an urban-industrial environment, Raipur, India.

#### **Chemical Analysis**

32 Chemical species [Twenty one elemental- Al, As, Ca, Cd, Co, Cr, Cu, Fe, Hg, K, Mg, Mn, Mo, Na, Ni, Pb, S, Sb, Se, V, Zn; nine water soluble ions- Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup>, Ca<sup>2+</sup>, NH<sub>4</sub><sup>+</sup>, Cl<sup>-</sup>, F<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup> and carbonaceous fractions-Organaic (OC) and Elemental (EC) carbon], known for source marker species have been determined in PM<sub>2.5</sub>. In case of inorganic constituents, water extractable ( $Na^+$ ,  $K^+$ ,  $Ca^{2+}$ ,  $Mg^{2+}$ ,  $Cl^-$ ,  $F^-$ ,  $NO_3^-$  and  $SO_4^{2-}$ ) and acid digested (1:3) H<sub>2</sub>O<sub>2</sub>/HNO<sub>3</sub>) (Na, Mg, Al, K, Ca, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, As, Cd, Pb, Mo, Se, Sb and Hg) were quantified separately (Katz, 1977; Chow et al., 2003). Elemental species (Al, K, Ca, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, As, Cd, Pb, Mo, Se, Sb and Hg) and selected water extractable (Na, Mg, K and Ca) were determined graphite furnace based atomic absorption spectrophotometrically (AAS) and cold vapour-AAS (iCE3500 Model, Thermo Fisher) using recommended conditions of operation and reported protocol of analysis (Thermo Fisher, 2008; DRI, 2011). Anions in water extracts ( $F^-$ ,  $Cl^-$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ) and digested samples ( $SO_4^{2-}$ ) were determined ion chromatographically (2000 Model, Dionex) using reported protocol of analysis (Chriswell et al., 1986; Watson et al., 1999; Dionex, 2005; Kulshrestha et al., 2010).  $SO_4^{2-}$  measurements in both water extracts and digested samples of PM2.5 has been carried out to quantify sulphur content of  $PM_{2.5}$  by subtracting the  $SO_4^{2-1}$ (digested) from  $SO_4^{2-}$  (water extracts) (Gurugubelli *et al.*, 2013). NH<sub>4</sub><sup>+</sup>, a marker of biomass burning, is determined spectrophotometrically (1305E Model, Systronics) using recommended procedures (Harrison and Perry, 1986).

As far as OC and EC analysis is concern, thermal optical transmittance (TOT) method presented by National Institute for Occupational Safety and Health- NIOSH-5040 protocol (NIOSH, 1999) using semi-continuous thermal/optical carbon analyzer (Sunset Laboratory, Model 4L, USA) (Birch and Cary, 1996; Schauer, et al., 2003; Pipal et al., 2014) has been adopted. Concentration of selected chemical species measured in PM2 5 were corrected by subtracting them from those found in field blanks; followed by subtraction from those found in background PM2.5. The background corrected concentrations along with their uncertainties and in-plume/ background ratio (for those measured in smoke plume) have been presented in Table 2 and Table 3. Uncertainty of source profile abundances estimated as the standard deviation of the average from five source tests (Watson et al., 2001; Watson and Chow, 2007). Species concentrations in laboratory blank and field blanks have also been presented in Table S1 (supporting information). Chemical species abundances in four different ranges of percent by weight along with comparison with reported values have been presented in Table 4 and Table 5. Components of crustal origin, ionics, trace elements and carbonaceous matter found in selected source profiles have been presented in Fig. 2. Source markers of selected source emitted PM<sub>2.5</sub> has been shown in Table 6.

### **RESULTS AND DISCUSSION**

# **Description of PM<sub>2.5</sub> Chemical Source Profiles**

A source profile comprised offivefractional abundances

of individual chemical species with respect to total PM mass and individual uncertainty or standard deviation value for particular chemical species (Watson et al., 2001). Chemical profiles have shown discernible pattern in relative strengths of selected species in PM<sub>2.5</sub> emissions resulting from industrial and domestic burning practices. Profile-wise measured mass accounted were:  $49.30 \pm 3.35\%$ . (MSWB),  $62.64 \pm 5.26\%$  (RSFS),  $78.05 \pm 5.07\%$  (RKS), RLPGS (74.03 ± 3.85%), CPI (50.65 ± 3.60%), FEMNI  $(46.18 \pm 4.02\%)$ , SPI  $(43.04 \pm 3.57\%)$ , CTPP  $(33.44 \pm$ 1.71%), SRM (58.33 ± 3.62.%), EAW (57.78 ± 3.93.%), and BKP ( $60.31 \pm 4.89.\%$ ); unaccounted content might be due to oxides and water content (Ho et al., 2003; Watson et al., 2012) and silica which were not measured directly by the methods applied to the source emissions. On analysing the association of different chemical components (OC/EC, crustal origin, ionics and trace elements) with PM<sub>25</sub> emission source profiles (Fig. 2), significant variation in OC and EC content of PM<sub>2.5</sub> with variability of 73.75% and 96.88%, respectively, across the selected source sites is observed; similar to earlier reported source profiles(Watson et al., 2001; Chow et al., 2003, 2004; Kong et al., 2011). Total carbonaceous matter (TC) (sum of OC and EC) is accounted for > 50% of PM<sub>2.5</sub> emissions in most of the source sites with OC/EC ratio > 1; attributed to incomplete combustion activities resulting in higher emissions of organic carbon in smouldering phases (Chakrabarty et al., 2013). Species of crustal origin (Al, Mn, Mg, Ca, Fe, K and Na) were also found in different proportions across the source profiles with highest variability of 117.42%; comparable with those reported earlier (Chow et al., 2003; Kong et al., 2011).As far as trace elements (Cu, Zn, As, Pb, Cr, Ni, Co, Cd, Hg, V, Mo, Sb, Se and S) and water soluble ions (Na<sup>+</sup>, NH<sub>4</sub><sup>+</sup>,  $K^{+},\ Mg^{2+},\ Ca^{2+},\ F^{-},\ Cl^{-},\ NO_{3}^{-},\ SO_{4}^{-)}$  are concern, more similarity (65.47% and 35.38%, respectively) compared to carbonaceous fractions and crustal species has been observed (Samara et al., 2003; Chow et al., 2004; Patil et al., 2013). Detailed PM<sub>2.5</sub> chemical profiles have been presented in Table 2 and Table 3.

#### Municipal Solid Waste Burning (MSWB)

Municipal refuse open burning is one of the major contributors to the air pollution in Indian cities (Patil *et al.*, 2013). PM<sub>2.5</sub> emissions from MSWB is prominently loaded with TC (84% of measured mass) having OC/TC ratio of 0.94 and potassium (5.14% of measured mass) having K<sup>+</sup>/K ratio of 0.77; comparable to those reported for vegetative and biomass burning (Watson *et al.*, 1994, 2001; Chow *et al.*, 2004). Abundance of other ions (Na<sup>+</sup>, NH<sub>4</sub><sup>+</sup>, Ca<sup>2+</sup>) were significantly higher (> 1% of measured mass) than species related to crustal origin and trace elements.

### **Residential Fuel Stoves**

Residential cooking stoves are mostly uses three different types of fuel/fuel mixtures for cooking purposes in India: solid fuel mixture (3:1:1 ratio of Coal, wood, dung cakes), kerosene and LPG. Highest abundance of TC (75–87% of measured mass) in PM<sub>2.5</sub> emissions has been observed. In addition, RSFS has shown OC/EC ratio of 6.9; two and

2055

.

	In-plume/ Background	0.01	0.01	12 19	01.04	1 13	C1.1	26.67		1 87	70.1	0.06		0.12		1 60	00.1	0 78	07.0	0.43	Ct.0	0.41	<b>F</b> .0	0.01	0.01	43 64		0.72		0.30	00.0	60.0	0.02	c0.0	10.56	10:00	27.59
actices.	BKP	$0.09032 \pm$	0.00301	$0.10733 \pm$	0.03145	$2.03986 \pm$	0.94582	0.05696±	0.00603	$0.00901 \pm$	0.00634	$0.01489 \pm$	0.00305	$0.09042 \pm$	0.00926	$0.44480 \pm$	0.07704	$0.00874 \pm$	0.00514	$0.42686 \pm$	0.43935	$0.38074 \pm$	0.25303	$0.01540 \pm$	0.00189	$0.02089 \pm$	0.00961	$0.96783 \pm 0.10020$	0.10920	0.04081	$0.02857 \pm$	0.00560	$0.06215 \pm$	0.00158	$0.03926 \pm$	0.03095	$0.02083 \pm 0.01079$
dustrial burning pi	In-plume/ Background		CC.0	L3 CV	10.74	0.78	0.10	6.33		0 52	1	0.26		0.21		11517	11.011	272	04.4	9 70	00	2 10	7.10	A 11	4.11	13 70		0.81		0.09	1 0.1	1.04	0000	00.00	1 80	10.1	37.30
emissions from inc	EAW	$2.95246 \pm$	0.19963	$0.10581 \pm$	0.00948	$1.39425 \pm$	0.10591	$0.01352 \pm$	0.00137	$0.00258 \pm$	0.00187	$0.06146 \pm$	0.00605	$0.15264 \pm$	0.00922	$31.97249 \pm$	2.26139	$0.07021 \pm$	0.00610	$6.60688\pm$	0.53671	$1.96525 \pm$	0.18215	5.04121 ±	0.30103	$0.00656 \pm$	0.00186	$1.09229 \pm 0.20052$	00007-0 + 80600 0	0.00995	$0.60203 \pm$	0.05303	$0.00425 \pm$	0.00215	$0.00705 \pm$	0.00643	$0.02816 \pm 0.00684$
cent by mass) of e	SRM	$0.08400 \pm$	0.00598	$0.00393 \pm$	0.00093	5.55807 ±	0.39318	0.00238 ±	0.00020	$0.00036 \pm$	0.00021	0.00027 ±	0.00008	$0.00782 \pm$	0.00061	$5.08803 \pm$	0.35994	$0.01001 \pm$	0.00085	$0.71173 \pm$	0.05723	$1.08633 \pm$	0.07691	$0.05624 \pm$	0.00400	$0.00032 \pm$	0.00023	$1.34385 \pm 0.00007$	0.00654 +	0.00168	$2.22579 \pm$	0.15830	$0.18249 \pm$	0.01257	$0.00103 \pm$	0.00074	$0.00095 \pm 0.00068$
ofiles (weight per	CTPP	$5.24498 \pm$	0.60980	$0.27300 \pm$	0.07300	$1.62692 \pm$	0.78370	0.00005±	0.00009	$0.00273 \pm$	0.00050	$0.09077 \pm$	0.02470	$0.11047 \pm$	0.08000	$1.81559 \pm$	0.43860	$0.00544 \pm$	0.00040	$1.12695 \pm$	0.13240	$0.92400 \pm$	0.16965	$0.71051 \pm$	0.06500	$0.00004 \pm$	0.00001	$0.57435 \pm$	0.00027	0.00172	$0.04959 \pm$	0.00400	$0.87193 \pm$	0.02094	$0.00014 \pm$	0.00200	$0.05200 \pm 0.00400$
nposite sources pi	IdS	$0.77095 \pm$	0.10408	$0.05500 \pm$	0.03000	$1.12100 \pm$	0.39350	0.00052±	0.00032	$0.00055 \pm$	0.00015	$0.96280 \pm$	0.01600	0.67500 ±	0.01800	$15.27590 \pm$	0.32431	$0.00785 \pm$	0.00340	$1.09798 \pm$	0.35200	$0.99930 \pm$	0.15090	$0.20133 \pm$	0.02443	$0.00489 \pm$	0.00049	$0.67693 \pm$	0.04276 +	0.00111	$0.00461 \pm$	0.00051	$0.00735 \pm$	0.00014	$0.00035 \pm$	0.00012	$0.00158 \pm 0.00080$
able 3. PM <sub>2.5</sub> con	FEMNI	2.46541 ±	0.30952	$0.00710 \pm$	0.00210	$6.54080 \pm$	1.27759	$0.00019 \pm$	0.00015	$0.00182 \pm$	0.00194	$0.71700 \pm$	0.06000	0.82500 ±	0.00600	$6.81153 \pm$	3.79153	$0.00913 \pm$	0.01544	$0.54332 \pm$	0.12590	$1.95458 \pm$	0.42539	$1.97486 \pm$	0.08710	$0.01261 \pm$	0.00363	$3.39504 \pm 0.4710$	0.07519 +	0.00703	$0.00341 \pm$	0.00072	$0.02215 \pm$	0.00135	$0.00095 \pm$	0.00013	$0.00434 \pm 0.00251$
E	CPI	$3.79199 \pm$	0.30528	$0.00010 \pm$	0.00013	$20.0980 \pm$	0.25756	0.00028 ±	0.00127	$0.00202 \pm$	0.00100	$0.03372 \pm$	0.00305	0.94700 ±	0.01000	$3.03099 \pm$	0.24105	$0.00888 \pm$	0.00381	$0.61121 \pm$	0.02858	5.84425 ±	0.53912	$0.05213 \pm$	0.00699	$0.00165 \pm$	0.00022	$3.26488 \pm$	0.00333 +	0.01905	$0.02974 \pm$	0.00267	$0.98579 \pm$	0.04756	$0.00196 \pm$	0.00028	$0.00130 \pm 0.00038$
	Species	1 4	W	<u>۷</u>	CL.	°Ľ	Ca	Cd	5))	C	2)	Cr	5	Cu	5	٩ لا	21	Нα	911 B	Л	4	$M \alpha$	9 TAT	Mn	INTA	Mo		Na		Z	Ju	ГŪ	υ	n	d S	20	Se

Matawle et al., Aerosol and Air Quality Research, 14: 2051–2066, 2014

2056

In-plume/ Background	33.29	0.15	0.24	0.21	0.15	0.13	0.78	0.21	0.47	2.26	0.40	0.96	8.13			
BKP	$0.08092 \pm 0.02066$	$0.08909 \pm 0.00813$	$0.07190 \pm 0.03106$	$0.60190 \pm 0.04750$	$0.70456 \pm 0.07140$	$0.83246 \pm 0.00589$	$0.78128 \pm 0.56128$	$0.62025 \pm 0.30722$	$0.26115 \pm 0.17726$	$1.41742 \pm 0.70371$	$0.11199 \pm 0.01093$	$20.53330 \pm 0.37666$	$31.92240 \pm 1.59610$	$52.45570 \pm 1.97276$	$60.31460 \pm 4.88548$	
In-plume/ Background	8.03	3.46	1.36	0.10	0.00	0.08	0.82	0.11	5.91	1.03	1.03	60.0	0.04			
EAW	$0.01951 \pm 0.00489$	$2.10060 \pm 0.07025$	$0.40982 \pm 0.03096$	$0.27349 \pm 0.05170$	$0.01189 \pm 0.00854$	$0.51454 \pm 0.04012$	$0.81442 \pm 0.06250$	$0.31450 \pm 0.08500$	$3.31681 \pm 0.25985$	$0.64688 \pm 0.20334$	$0.29062 \pm 0.06212$	$1.88401 \pm 0.09420$	$0.16661 \pm 0.00833$	$2.05062 \pm 0.10253$	57.78406 ± 3.92908	
SRM	$0.00670 \pm 0.00078$	$0.78549 \pm 0.05557$	$0.01366 \pm 0.0116$	$0.90901 \pm 0.07097$	$0.39439 \pm 0.07933$	$4.17298 \pm 0.29597$	0.24446 ±	$0.13700 \pm 0.01484$	$0.05234 \pm 0.01421$	$2.33327 \pm 0.08260$	$0.34232 \pm 0.03474$	$24.06154 \pm 1.20308$	$11.47508 \pm 0.57375$	$35.53662 \pm 1.77683$	58.32601 ± 3.61621	
СТРР	$0.01854 \pm 0.00211$	$0.15036 \pm 0.01353$	$1.73480 \pm 0.16500$	0.21566 ±	$0.40000 \pm 0.20000$	$1.16403 \pm 0.06695$	$0.39449 \pm 0.03665$	$0.38322 \pm 0.11700$	$0.82600 \pm 0.05030$	$1.13420 \pm 0.02670$	$0.67000 \pm 0.02500$	$9.16596 \pm 0.19555$	$6.70770 \pm 0.32630$	$15.87366 \pm 0.52185$	$33.44134 \pm 1.70550$	ting in sum%.
SPI	$0.02653 \pm 0.00496$	$0.46236 \pm 0.05455$	$0.71000 \pm 0.0300$	$1.13534 \pm 0.03970$	$0.08527 \pm 0.00621$	$0.53500 \pm 0.00012$	$0.41815 \pm 0.12630$	$0.05175 \pm 0.01115$	$0.70123 \pm 0.08280$	$0.80390 \pm 0.01170$	$0.55240 \pm 0.01420$	$4.37481 \pm 0.45575$	$13.74898 \pm 1.82446$	$18.12379 \pm 2.28021$	43.03669 ± 3.57204	d for double count
FEMNI	$0.01002 \pm 0.00704$	$0.73586 \pm 0.20121$	$0.17410 \pm 0.05500$	$0.47000 \pm 0.17425$	$0.45055 \pm 0.31774$	$0.73854 \pm 0.00447$	$1.98625 \pm 0.17927$	$1.11496 \pm 0.41855$	$0.24413 \pm 0.12670$	$0.76072 \pm 0.17605$	$0.49190 \pm 0.01306$	$11.59053 \pm 0.65233$	$5.58571 \pm 0.43274$	$17.17624 \pm 1.08507$	46.18473 ± 4.01807	C were eliminate
CPI	$0.00187 \pm 0.00394$	$0.31990 \pm 0.03213$	$0.23748 \pm 0.08865$	$0.01654 \pm 0.01059$	$0.00299 \pm 0.00175$	$3.31847 \pm 0.77165$	$1.06891 \pm 0.36055$	$0.00956 \pm 0.00314$	$0.01276 \pm 0.01140$	$2.55696 \pm 0.19926$	$0.00236 \pm 0.00081$	$7.89808 \pm 1.59461$	$0.14000 \pm 0.05097$	$8.03808 \pm 1.64558$	$50.65410 \pm 3.59664$	Na <sup>+</sup> , Mg2 <sup>+</sup> and T
Species	Λ	Zn	اللا بل	Cl-	$NO_{3}^{-}$	$\mathrm{SO_4}^{2-}$	$\mathrm{Na}^+$	$\mathrm{NH_4}^+$	$\mathbf{K}^{+}$	$Ca^{2+}$	${\rm Mg}^{2+}$	00	EC	TC	Sum%	* Ca <sup>2+</sup> , K <sup>+</sup> ,

(continued).	
Table 3.	

Matawle et al., Aerosol and Air Quality Research, 14: 2051–2066, 2014

2057

q	
fe	
ō	
rej	
Ŋ	
ıal	
OL	
ati	
E	
nte	
Ш	
Ň	
all	
on	
Ξţi.	
n	
ith	
R	
n	
isc	
ar	
đ	
ō	
ц	
.jei	
14	
ŭ	
S 3	
ë	
cti	
ra	
요	
Е	
pŋ	
ંગ	
st	
ne	
Ю	
ų	
Q	
Ē	
SUC	
SIC	
iis	
en	
.5	
Σ	
Ы	
.u	
S	
an	
nd	
INC	
al	
ies	
G	
Ś	
al	
.ü	
ЗШ	
h	
<b>(</b> )	
of c	
v of c	
ary of c	
mary of c	
ammary of c	
Summary of c	
4. Summary of c	Š.
ole 4. Summary of c	iles.

<b>Fable 4.</b> Sur profiles.	nmary of chemical speci	es abundance in $PM_{2.5}$ emissions from domesti	c burning practices and their compariso	on with nationally and internation	nally reported
an a	· · · · · · · · · · · · · · · · · · ·		Chemical abundances in percent mass		
Source ID	Various studies	< 0.1%	0.1–1%	1-10%	> 10%
MSWB	Present study <sup>*</sup>	Al, As, Cd, Cr, Cu, Hg, Mg, Mn, Mo, Na, Ni, Pb. S. Sb. Se, V, F <sup>*</sup> , Mg <sup>24</sup>	Ca, Co, Fe, K, Zn, Cl <sup>-</sup> , NO <sub>3</sub> <sup>-</sup> , SO <sub>4</sub> <sup>2-</sup> , Na <sup>+</sup> , NH <sub>4</sub> <sup>+</sup> , Ca <sup>2+</sup>	K <sup>+</sup> , EC	0C
	USEPA Speciate 4.0 PR.No91040**	Sb, As, Cd, Ce, Cr, Cu, Mn, Mo, Ni, Pd, Se, Sn	Cl, V, Zn	Al, $NH_4^+$ , Ca, EC, Fe, Pb, H2. OC, K. Na. $SO_4^{2-}$ . S	
	CPCB, 2008***	<u>Al, Mg, As, Cd, Co, Cr, Cu, Hg, Mn, Mo, Ni,</u> Ph. Sh. Se. V. Mo <sup>2+</sup>	$Fe, Na, K, Zn, F^{-}, NO_{\overline{2}}^{-}, Na^{+}, NH_{\underline{4}}^{+}$ .	Ca, Fe, $SO_4^{2-}$ , $Ca^{2+}$ , EC	OC
RSFS	Present study	Al, As, Cd, Co, Cu, Hg, Mn, Mo, Ni, Pb, Sh. Se, V	<b></b> Ca, Cr, Fe, Mg, Na, S, Zn, F, Na, NH, <sup>+</sup> , Ca <sup>2+</sup> , Mo <sup>2+</sup>	K, Cl <sup>-</sup> , NO <sub>3</sub> <sup>-</sup> , SO <sub>4</sub> <sup>2-</sup> , K <sup>+</sup> , EC,	00
	USEPA Speciate 3.2 PR No -423312 5	Al, As, Ca, Cl, Co, Cu, Fe, Pb, Mg, Mn, Hg, Mo No. S Zn	$NH_4^+$ , $C\Gamma$ , $NO_3^-$ , $K$ , $K^+$ , $SO_4^{2-}$		EC, OC
	<u>CPCB, 2008</u>	Al, Ca, Fe, Mg, As, Cd, Co, Cr, Cu, Hg, Mn, Mo Ni Ph, Sh, So IV NO- NH + Mo <sup>2+</sup>	<u>Na, Zn, F', SO<sub>4</sub><sup>2-</sup>, Na<sup>+</sup>, Ca<sup>2+</sup></u>	$K, CT, K^+, EC$	<u>0C</u>
RKS	Present study	Ca, Cd, Co, Cr, Cu, Hg, Mg, Mn, Mo, Na, Ca, Cd, Ca, Cu, Cu, Hg, Mg, Mn, Mo, Na, Na <sup>+</sup> , Ca <sup>2+</sup> , Mo <sup>2+</sup>	Al, As, Fe, Ni, Pb, Sb, Se, V, Zn, $F^-$	K, S, CT, NO <sup>3</sup> , SO <sup>2-</sup> NH, <sup>4</sup> , K <sup>4</sup>	OC, EC
RLPGS	Present study	As, Ca, Cd, Co, Cr, Cu, Fe, Hg, Mg, Mn, Mo, Ni, Pb, Sb, Se, V, Zn, F <sup>-</sup> , Ca <sup>2+</sup> , Mg <sup>2+</sup>	Al, K, Na, S, CΓ, NO <sub>3</sub> <sup>-</sup> , Na <sup>+</sup> , K <sup>+</sup>	$SO_{4}^{2-}, NH_{4}^{+}$	OC, EC
	USEPA Speciate 4.3 PR. No91156	Cr, Fe, Mn, Ni, Se, V	$Al, NH_4^+, K$	Ca, Na, S, Cl <sup>-</sup> , NO <sup>-</sup> <sub>3</sub> , EC	$SO_4^{2-}, OC$
	<i>CPCB</i> , 2008 <sup>*</sup>	<u>Al, Ca, Mg, Cd, Co, Cr, Cu, Hg, K, Mn, Mo,</u> Ni, Sb, V, Zn	$Fe$ , Na, As, Pb, Se, $F^-$ , NO $\frac{1}{2}$ , NH $\frac{1}{4}$	$\frac{C\Gamma, SO_4^{2-}, Na^+, K^+, Ca^{2+}}{Mg^{2+}, EC}$	<u>0C</u>
* Renrecent r	wofile for PM., fraction.	* Rold font for nresent study. ** Italio font for rer	norted international studies and: *** Italio	r font with underline for reported	Indian ctudy

tor reported indian study. allie Italic tone with under anu, suudes Italic form for reported international tout for present study; DUID \* Represent profile for PM<sub>10</sub> fraction;

-Ct	
ŏ	
Ĕ	
ğ	
-re-	
N	
II	
n	
<u>10</u> .	
at	
Ê	
Ð	
E.	
ğ	
a	
2	
al	
n	
Ĕ	
Ja.	
Ţ	
Ŧ	
12	
-	
5	
.IS	
ar	
đ	
ЯĽ	
3	
Ē.	
ē	
th	
Ч	
ũ	
0	
ğ	
. <u>S</u>	
ਠੁ	
ra	
р.	
ല്	
Ξ	
E	
R	
Ξ	
ia.	
Ħ	
n	
Ċ,	
_	
Ξ.	
n in	
om in	
from in	
is from in	
ons from in	
sions from in	
issions from in	
missions from in	
emissions from in	
2.5 emissions from in	
M <sub>2.5</sub> emissions from in	
PM2.5 emissions from in	
n PM2.5 emissions from in	
in PM <sub>2.5</sub> emissions from in	
ce in PM2.5 emissions from in	
nce in PM2.5 emissions from in	
dance in PM <sub>2.5</sub> emissions from in	
indance in PM <sub>2.5</sub> emissions from in	
oundance in PM2.5 emissions from in	
abundance in PM2.5 emissions from in	
es abundance in PM2.5 emissions from in	
sies abundance in PM2.5 emissions from in	
ecies abundance in PM <sub>2.5</sub> emissions from in	
species abundance in PM <sub>2.5</sub> emissions from in	
A species abundance in PM <sub>2.5</sub> emissions from in	
cal species abundance in PM2.5 emissions from in	
nical species abundance in PM2.5 emissions from in	
emical species abundance in PM2.5 emissions from in	
themical species abundance in PM2.5 emissions from in	
f chemical species abundance in PM <sub>2.5</sub> emissions from in	
of chemical species abundance in PM <sub>2.5</sub> emissions from in	
y of chemical species abundance in PM2.5 emissions from in	
ary of chemical species abundance in PM2.5 emissions from in	
mary of chemical species abundance in PM2.5 emissions from in	
mmary of chemical species abundance in PM2.5 emissions from in	
ummary of chemical species abundance in PM2.5 emissions from in	
Summary of chemical species abundance in PM <sub>2.5</sub> emissions from in	
5. Summary of chemical species abundance in PM <sub>2.5</sub> emissions from in	S.
e 5. Summary of chemical species abundance in PM <sub>2.5</sub> emissions from in	les.
ble 5. Summary of chemical species abundance in PM2.5 emissions from in	ofiles.

U Come	Vorione studios		Chemical abundances in percent	mass	
ource ID	v arious studies	< 0.1%	0.1–1%	1-10%	> 10%
CPI	Present study	As, Cd, Co, Cr, Hg, Mn, Mo, Ni, Pb, Sb, Se, V, Cl, NO <sub>3</sub> , NH <sub>4</sub> <sup>+</sup> , K <sup>+</sup> , Mg <sup>2+</sup>	Cu, K, S, Zn, F <sup>-</sup> , EC	Al, Fe, Mg, Na, SO4 <sup>2-</sup> , Na <sup>+</sup> , Ca <sup>2+</sup> , OC	Са
	Chow et al., 2004	As, Mg, Cd, Co, Cr, Cu, Hg, Mn, Mo, Ni Ph. Sh. So. V. Zn	AI	$Fe, Na, Cl, CT, NO_{3}^{-}, Na^{+}, NH_{1}^{+} FC$	Ca, K, S, $SO_4^{2-K^+}$
EMNI	Present study	As, Cd, Co, Hg, Mo, Ni, Pb, S, Sb, Se, V	$Cr, Cu, K, Zn, F^-, C\Gamma, NO_3^-, SO_4^{2-}, K^+, C_{a^{2+}}M_{a^{2+}}$	Al, Ca, Fe, Mn, Na, Na <sup>+</sup> , NH <sup>+</sup> FC	OC OC
IdS	Present study	As, Cd, Co, Hg, Mo, Ni, Pb, S, Sb, So V NO. <sup>-</sup> NH. <sup>+</sup>	Al, Cr, Cu, Mg, Mn, Na, Zn, F <sup>-</sup> , SO. <sup>2-</sup> Na <sup>+</sup> K <sup>+</sup> Ca <sup>2+</sup> Mo <sup>2+</sup>	Ca, K, Cl <sup>-</sup> , OC	Fe, EC
CTPP	Present study	Cd, Co,Cr, Hg, Mo, Ni, Pb, Sb, Se, V	As, Cu, Mg, Mn, Na, S, Zn, Cl <sup>-</sup> NO. <sup>-</sup> Na <sup>+</sup> NH. <sup>+</sup> K <sup>+</sup> Mo <sup>2+</sup>	Al, Ca, Fe, K, F <sup>-</sup> , SO4 <sup>2-</sup> , Ca <sup>2+</sup> OC FC	
	Watson et al., 2001	Cd, Co, Cr, Cu, Hg, Mn, Mo, Ni, Pb, Sh So V Zn	As, Cl, K, Mg, Na, $NO_3^-$ , $K^+$	$Al, Ca, Fe, S, Cl^{-}, NH_{4}^{+}, OC, FC$	$SO_4^{2-}$
	<u>CPCB, 2008</u>	<u>56, 55, 7, 54</u> <u>Al. Cd. Co. Cu, Hg, K, Mn, Mo, Pb, Sb</u> <u>Se</u> F. Mo <sup>2+</sup>	<u>As, Cr, Mg, Na, V, Zn, Cr, NO<sup>2</sup>.</u> Na <sup>+</sup> K <sup>+</sup> Ca <sup>2+</sup>	Ca, Ni, NH <sub>4</sub> <sup>+</sup> , EC	$\underline{SO}_{\underline{4}}^{2-}, \underline{OC}$
SRM	Present study	Al, As, Cd, Co, Cr, Cu, Hg, Mn, Mo, Ni Sh, So, V, F <sup>+</sup> , K <sup>+</sup>	$K, S, Zn, C\Gamma, NO_3^-, Na^+, NH_4^+, Mo^{2+}$	Ca, Fe, Mg, Na, Pb, SO4 <sup>2–</sup> , Ca <sup>2+</sup>	OC, EC
	<u>CPCB, 2008</u>	Al. As, Cd, Co, Cr, Cu, Hg, Mn, Mo. Ni Sh, So, V F- K <sup>+</sup>	$K, Zn, Cl^{-}, NO_{3}^{-}, Na^{+}, NH_{4}^{+}, Mg^{2+}$	Ca, Fe, Mg, Na, Pb, SO <sub>4</sub> <sup>2-</sup> , Ca <sup>2+</sup> EC	<u>0C</u>
EAW	Present study	Cd, Co, Cr, Hg, Mo, Ni, Pb, Sb, Se, V. NO. <sup>-</sup>	As, Cu, Pb, F <sup>-</sup> , Cl <sup>-</sup> , SO <sup>4-</sup> , Na <sup>+</sup> , NH <sup>4+</sup> , Ca <sup>2+</sup> , Mo <sup>2+</sup> , FC	Al, Ca, K, Mg, Mn, Na, K <sup>+</sup> , OC	Fe
	<u>CPCB, 2008</u>	<u>4s, Cd, Co, Cr, Cu, Hg, Mo, Ni, Sb, Se,</u> <u>V_NO,-</u>	$Na, Pb, F, CI, SO_{4}^{-}, Na^{+}, NH_{4}^{+}, Ca^{2+}, Ma^{2+}, EC$	<u>Al, Ca, Mg, K, Mn, Zn, K<sup>+</sup>, OC</u>	$\overline{Fe}$
BKP	Present study	Al, Cd, Co, Cr, Cu, Hg, Mn, Mo, Ni, Pb, S, Sb, Se, V, Zn, F	As, Fe, K, Mg, Na, CI, NO <sub>3</sub> <sup>-</sup> , SO <sub>4</sub> <sup>2-</sup> , Na <sup>+</sup> , NH <sub>4</sub> <sup>+</sup> , K <sup>+</sup> , Mg <sup>2+</sup>	$\overline{\operatorname{Ca}}$ , $\operatorname{Ca}^{2^+}$	OC, EC
	USEPA Speciate 4.3	As, Cr, Cu, Mn, Ni, Pb, Sb, Se, Zn NH <sup>+</sup> FC	Cl, K, Na, Mg, S, $NO_{3}^{-}$ , $SO_{4}^{2-}$	Al, Fe, OC	Ca

Matawle et al., Aerosol and Air Quality Research, 14: 2051–2066, 2014



**Fig. 2.** Crustal elements (Al, Ca, Fe, K, Mg, Mn and Na), trace element (As, Cd, Co, Cr, Cu, Hg, Mo, Ni, Pb, S, Sb, Se, V, and Zn), ions (F<sup>-</sup>, Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, Na<sup>+</sup>, K<sup>+</sup>, Ca<sup>2+</sup>, Mg<sup>2+</sup>) and organic content in PM<sub>2.5</sub> of selected source emissions.

Table 6. Source signatures of domestic and industrial burning practices in India.

Source type	Source signature	Source type	Source signature
Residential fi	uel burning	Industria	l burning Sources
RSFS	$F^{-}$ , As, $Mg^{2+}$ , $Ca^{2+}$ , $Cr$ , $K^{+}$	CPI	S, Cr, Cu, Al, Mo, Mg
RKS	Pb, Cd, Sb, F <sup>-</sup> , Se, V	FEMNI	Cr, Mo, Mg <sup>2+</sup> , Mn, NO <sub>3</sub> <sup>-</sup> , Cu
RLPGS	Sb, Cd, Pb, S, Mo, Se	SPI	$Cr, As, Mg^{2+}, Cu, EC, Cl^{-}$
Residential fuel combustion <sup>a*</sup>	<u>OC, EC, K<sup>+</sup>, Cl<sup>-</sup></u>	СТРР	As, Cr, S, F <sup>-</sup> , NO <sub>3</sub> <sup>-</sup> , Al
Municipal wa	ste burning	SRM	Pb, Mg <sup>2+</sup> , NO <sub>3</sub> <sup>-</sup> , EC, S, Cl
MSWB	$F-, Co, Cd, Ca^{2+}, Na^+, K^+$	EAW	Mn, $Cr$ , $K^+$ , $Cd$ , Pb, $Mg^{2+}$
Waste burning <sup>a</sup>	$OC$ , $EC$ , $K^+$ , $As$ , $Pb$ , $Zn$	BKP	As, Cd, Mo, EC, $NO_3^-$ , Sb
		Metallurgy <sup>a, b, d</sup>	<u>Mn, Zn, Pb, Cd, Cu, As, Hg</u>
		Industry <sup> a, c</sup> and Coal	<u>V, Ni, <math>SO_4^{2-}</math> and Se, As, Cr, Co,</u>
		<i>Combustion<sup>b</sup></i>	<u>Cu, Al</u>

\* Reported sources and their markers in italic style with underline for comparison.

<sup>a</sup> Watson et al., 2008; <sup>b</sup> Mitra et al., 2002; <sup>c</sup> Viana et al., 2008; <sup>d</sup> Viana et al., 2006.

five-fold higher compared to RKS and RLPGS, respectively. On considering the biogenic markers (K<sup>+</sup> and NH<sub>4</sub><sup>+</sup>), higher values of K<sup>+</sup>/K ratio (0.89–0.97) and NH<sub>4</sub><sup>+</sup> abundance compared to those found in crustal origin and industrial emitted PM<sub>2.5</sub> is attributed to the fact that biogenic sources are dominating in PM<sub>2.5</sub> emissions from household solid fuel burning practices involved with cooking purposes. These values are similar to earlier reports on emissions of biogenic (Watson *et al.*, 1994, 2001; Chow *et al.*, 2004) and solid fossil fuel combustion (Watson *et al.*, 2001; Kong *et al.*, 2011). About 4–13 times higher K<sup>+</sup> emissions has been observed in case of RSFS compared to RKS and RLPGS, respectively. But NH<sub>4</sub><sup>+</sup> was found 2–16 times higher in emissions resulting from kerosene stoves (RKS) compared to RLPGS and RSFS, respectively. As far as trace species

is concern, sulphur constituents (S,  $SO_4^{2-}$ ) along with Se and seven toxic species (As, Cd, Hg, Ni, Pb, V and Zn) were found multi-fold higher in RKS compared to RSFS and RLPGS; similar to earlier reported levels for kerosene stoves (Patil *et al.*, 2013). Other anions (NO<sub>3</sub><sup>-</sup>, F<sup>-</sup> and Cl<sup>-</sup>) have shown higher presence in PM<sub>2.5</sub> emissions resulting from RSFS compared to RKS and RLPGS. Most of markers of crustal origin are found < 0.1% in all types of household fuel burning emissions.

# **Coal Fired Mineral Based Industries**

Six different process based industrial emissions, mostly dominating in study region, have been chosen for the development of  $PM_{2.5}$  chemical source profiles. Out of six chosen industries, only BKP has shown > 50% load of

carbonaceous matter (TC) in their PM<sub>2.5</sub> emissions. Other industrial PM2.5 emissions have shown lesser load of TC in the ranges of 8.03-35.53%. PM<sub>2.5</sub> emissions from two industrial combustion processes (SPI and BKP) have shown higher load of EC compared to OC with EC/OC ratio of 3.25 and 1.55, respectively. Significantly lower measured mass in case of CPI, FEMNI and CTPP is might be due to higher unaccounted silica load in PM2.5 mass. As far as iron particle load in PM<sub>2.5</sub> is concern, sponge iron processing (direct reduction iron processing) (SPI) have shown 2-34 times higher load compared to other industrial processes included in the study. In case of other major constituents like Al is found highest in CTPP (15.5% of measured mass) and Ca is found highest in CPI (39.6% of measured mass). Water soluble/total ratio (Ca<sup>2+</sup>/Ca, Mg<sup>2+</sup>/Mg and Na<sup>+</sup>/Na) describes the pre-dominance of emissions resulting from either dust re-suspension during process handling or combustion activities involved with mineral based coalfired industries (Volkovik, 1983; Watson et al., 1994). The lower values of these ratios in PM25 emissions from CPI (0.13, 0.01 and 0.33, respectively), FEMNI (0.12, 0.25 and 0.58, respectively) and SRM (0.42, 0.31 and 0.18, respectively) attributed the higher abundance from crustal origin, whereas higher values for SPI (0.71, 0.55 and 0.61, respectively), CTPP (0.69, 0.73 and 0.68, respectively) and BKP (0.69, 0.28 and 0.81, respectively) indicated the predominance of emissions from combustion processes. CTPP, SPI and BKP have shown higher enrichment of biogenic potassium ( $K^+$ ) with  $K^+/K$  ratio of 0.73, 0.64 and 0.62, respectively; contrast to CPI and SRM along with earlier reported values for coal fired steel industries (Watson et al., 2001; CPCB, 2008; USEPA, 2013). This might be due to use of poor quality coals in CTPP and SPI and biomass (dung cakes with coals) in brick kilns .Order of variation in association of anions with PM25 emissions from selected source sites is evaluated to be:  $NO^{3-} > F^- > Cl^- > SO_4^{2-}$ .on other hand, uniformity in occurrence of sulphur group species (S,  $SO_4^{2-}$  and Se) across the selected industrial sites is attributed to similar source origin of coal combustion (Volkovik, 1983). Trace elements have also shown significant variation in their relative abundances in PM2.5 emitted from different sources; justified their inclusion in the development of PM<sub>2.5</sub> chemical source profiles for emissions resulting from combustion processes involved with different mineral based coal-fired industries. Brick Kiln processes (BKP) has been identified as the major emitter of highest number of toxic species (Cd, Co, Mo, Sb and V) followed by SRM (Hg and Pb) and SPI (As, Ni). Moderate combustion temperature and poor quality of coals and other combustion materials (dung cakes etc.) are responsible for higher emission of toxic species (Vollkovic, 1983).

In addition to these industrial sources, open fabrication workshops having arc-welding activities (EAW) is also reported to be the significant contributor of outdoor  $PM_{2.5}$  due to their profuse locations within the study region (Pervez *et al.*, 2005). It has been observed that EAW is contributing mainly active iron and manganese particles. Pb, As, Cu and  $F^-$  were observed to be found above 0.1% compared to other toxic species similar as earlier reported profiles of arc-

welding workshops (Swamy et al., 1994; CPCB, 2008).

# Comparison of Developed Source Profiles with Reported Profiles of National and International Origin

To compare the developed PM<sub>2.5</sub> chemical source profiles in this study with previous reported profiles by Chow et al. (2004), Watson et al. (2001), the Central Pollution Control Board, India (2008) and Speciate 4.0 (USEPA). All chemical species of each of selected profiles are grouped in four percentage fractional ranges(Chow et al., 2003, 2004) and summarized in Table 4 and Table 5.The major markers  $(OC, K^+, NH_4^+)$  of emissions resulting from MSWB were observed to be found in different levels in developed and reported profiles. OC was found > 10% in present and CPCB profiles, but that was observed to be within 1-10% in Speciate 4.0 profiles. K<sup>+</sup> is found in different fractional range between present and CPCB profiles with higher abundance (1-10%)in presented profile; confirm the strong variation in biogenic matter content in MSWB. Additionally, the different abundance of K<sup>+</sup> in MSWB emitted PM<sub>2.5</sub> between present and CPCB profiles might be due to different sampling methods; CPCB adopted laboratory scale study, whereas real-world sampling on open air burning was adopted in present study. Apart from major markers, other species of crustal origin were found in similar fractional ranges in MSWB profiles developed in present study and CPCB database as well. In case of PM2.5 profiles of emissions resulting from residential fuel burning (RSFS, RKS and RLPGS), carbonaceous matter was found > 10% across all the comparative profiles, but EC was found > 10% only in presented profiles. As far as anions and cations of fuel combustion markers (K<sup>+</sup>, NH<sub>4</sub><sup>+</sup>, SO<sub>4</sub><sup>2-</sup>, Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>) is concern, most of them found higher (1-10%) in presented profiles compared to databases of CPCB and Speciate 4.0; most of them found higher in emissions of kerosene stoves (RKS). Other species of crustal origin and trace elements have shown more than 50% agreement in their occurrence in similar fractional groups across the selected databases for comparison.

On comparing PM<sub>2.5</sub> chemical profiles of emissions resulting from selected industrial sites with CPCB and Speciate 4.0, different observations were obtained for different industrial sites. Sponge-iron industrial emissions (SPI) and Ferro-manganese industry (FEMNI) were not included in CPCB and Speciate 4.0. In case of CPI and CTPP, major fractional group (> 10%) has shown different inclusion of species across the comparative databases. BKP has shown inverse distribution of OC and Ca between two fractional groups (> 10% and 1-10%) on comparing the presented profile with CPCB profile. On contrary, EAW has shown similar major marker (Fe) in both comparative source profile databases. The element Ca for CPI, OC and EC for SRM & BKP, Fe for EAW, EC for SPI and OC for FEMNI were accounted for > 10% abundances. The water soluble ions and crustal element distributed between 0.1– 1% and 1–10% ranges. The trace elements were contributed < 0.1% relative abundances except As and S for CTPP with range 0.1-1%.

Overall species abundance of source profiles in defined

fractional groups is in  $\sim$ 60–80% agreement with reported source profile databases of CPCB and in  $\sim$ 35–50% agreement with USEPA Speciate 4.0.

#### Mass Closure Analysis

Mass reconstruction of  $PM_{2.5}$  has been carried out using revised IMPROVE18 mass closure (material balance) equation (Ho *et al.*, 2003, Chow *et al.*, 2012); estimate the unmeasured oxides and compared with the total gravimetric measured mass for the quality assurance (Watson *et al.*, 2012):

 $[PM_{2.5}] = 1.375 \text{ SO}_4^{2-} + 1.29 \text{ NO}_3^{-} + 1.8 \text{ OC} + \text{EC} + (2.2 \text{ Al} + 2.49 \text{ Si} + 1.63 \text{ Ca} + 1.94 \text{ Ti} + 2.42 \text{ Fe}) + 1.8 \text{ Cl}^-$ (1)

Two major crustal elements, silica and titanium have not been included in mass closure study; resulting in lower estimated values for crustal fraction in the overall mass closure. In case of RKS and RLPGS, PM2.5 mass reconstruction has been achieved 107.27% and 102.08%, respectively due to negligible abundances of unmeasured silica, titanium and higher abundance value of OC. Domestic solid material burning practices (RSFS and MSWB) have shown relatively lower mass closure values (97.47% and 79.36%, respectively). PM<sub>2.5</sub> Mass closure results for industrial sources have shown close equivalence between the gravimetric mass and reconstructed mass. The highest and lowest mass closure values have been found in case of EAW (98.84%) and CTPP (45.65%), respectively. Mass closure results of PM2.5 emissions, resulting from other industrial source sites, were achieved 77.13%, 62.11%, 66.32%, 68.12% and 85.08% for BKP, FEMNI, SPI, CPI, and SRM, respectively. No overestimation of mass closure result has been occurred in developed profiles.

#### Ion Balance Calculation

The ionic balance calculation has been performed to confirm the acid-base property of PM<sub>2.5</sub> fractions emitted by different burning practices. Conversion of ion mass concentrations into micro equivalents was performed to calculate the cation/anion balance of PM<sub>2.5</sub> (Cao *et al.*, 2005; Zhang *et al.*, 2011; Tao *et al.*, 2013). The cation and anion micro equivalents of particles were calculated as follows:

A (Anion micro equivalents/m<sup>3</sup>) =  $F^{-}/19 + Cl^{-}/35.5 + NO_{3}^{-}/62 + SO_{4}^{-2}/48$  (3)

The well balanced anion/cation (A/C) ratio must be 1. The value higher than 1, indicates the acidic nature of the particle (Kerminen *et al.*, 2001) whereas slightly lower than 1, indicate contribution of unmeasured  $CO_3^{2^-}$  ion, and very low A/C ratio indicates the basic nature of particle (Cao *et al.*, 2005; Shen *et al.*, 2007, 2009). A/C ratios of PM<sub>2.5</sub> emissions resulting from domestic and industrial burning practices have been accounted; ranges from 0.18–0.79 and 0.19–0.74, respectively and confirm that all source emitted

PM<sub>2.5</sub> samples were basic in nature.

#### Source Markers

Source markers of particles are mostly described by specific size distribution, specific suite of elements and specific ratios of compounds, elements or isotopes (Mitraet *al.*, 2002). The relative source indicator species were evaluated for all eleven sources grouped in domestic and industrial burning practices. For the calculation, following formula was applied to define the indicatory species for specific source emitted  $PM_{2.5}$  fraction (Yang *et al.*, 2002; Kong *et al.*, 2011):

$$Ratio_{i,j=} \frac{(X_i / \Sigma X)_j}{(X_i / \Sigma X)_{\min}}$$
(4)

where: X<sub>i</sub> was the i<sup>th</sup> individual species concentration;  $(X_i/\Sigma X)_i$  was the quotient of i<sup>th</sup> individual species ij divided by the summation of 32 species concentrations of emission source j;  $(X_i / \Sigma X)_{min}$  was the quotient of i<sup>th</sup> individual species divided by the summation of 32 species concentrations which were the minimum for all emission sources (Yang et al., 2002; Chen et al., 2003). A normalization procedure has been applied according to Mitra et al. (2002) and Kong et al. (2011) to minimize the effect of physical parameters. Normalized individual species concentration was used by dividing the i<sup>th</sup> individual species ij concentration to the sum of i<sup>th</sup> individual concentration for all the source profiles (Kong et al., 2011). The top six chemical species with highest ratio values for total relative source profiles has been use as relative source indicatory chemical species describe in detail on Table 6 and compare with earlier reports (Kong et al., 2011). Water soluble K<sup>+</sup>, marker of vegetative and biomass burning sources (Watson et al., 2002, 2008), Fand Ca<sup>2+</sup> were evaluated to be the similar source markers of PM<sub>2.5</sub> emissions from burning practices involved with MSWB and RSFS; whereas RKS and RLPGS have shown distinct source markers with major marker groups of (Fand V) and (S and Mo), respectively. Arsenic (As) was found common source marker of PM2.5 emissions from CTPP, SPI, and BKP; one of prominent trace element marker of emissions resulting from coal burning (Mitra et al., 2002). Mn was calculated as common source marker PM<sub>2.5</sub> emissions from FEMNI and EAW. Distinct observations of source markers of PM<sub>2.5</sub> emissions from selected source sites compared to those reported earlier (Mitra et al., 2002; Viana et al., 2006; Watson et al., 2008) is attributed to the importance of development of region specific source profiles to obtained precise results of receptor modeling.

# CONCLUSIONS

The differences with earlier reported/developed similar characteristics profiles created demand of additional and more precise source profiles that represent a study area. In this channel, the present study is an important work in development of source profile database in India. The eleven important stationary sources profiles for PM<sub>2.5</sub> fraction are

reported in this paper. These profiles are comprised of 21 element, 9 water soluble ion, and carbonaceous fractions by following the standard protocol of chemical analysis and data validation. The carbonaceous fractions are most abundant with different OC/TC ratio; ranges from ~0.58-0.94 in PM<sub>2.5</sub> from selected source emissions. Trace metals were found significantly higher in PM2.5 emissions from burning practices involved with household cooking activities and municipal solid waste management practices, compared to crustal origin. Observation of different relative enrichment of defined chemical components (Fig. 2) in PM<sub>2.5</sub> emissions from selected industrial burning sources might be due to use of different raw materials, and combustion temperature and conditions involved with industrial processes. The developed profiles comparatively much similar in > 10%and < 0.1% abundant species with earlier reported profiles for similar sources. The K<sup>+</sup>, Mn, and As were found and source marker for biomass burning, metallurgical industrial emission, and coal burning respectively, shown good agreement with National CPCB, 2008 and global USEPA speciate database also previous reported profiles.

These profiles require update, up gradation and addition of new sources with the sufficient interval of time to better represent changes in characteristics of sources of burning practices in India.

# AUTHOR CONTRIBUTIONS

S.P. designed and led the study, including project coordination; S.P. and J.M. were involved with field measurements, data collection, manuscript preparation and data analysis; J.M., S.D., S.T. and D.B. performed Chemical analysis. All authors discussed the results and contributed to the manuscript.

### ACKNOWLEDGEMENT

This material is based upon work supported by Department of Science and Technology Project SR/S4/AS-61/2010. Authors are grateful to Pt. Ravishankar Shukla University and Indian Institute of Tropical Meteorology, IITM (New Delhi) for providing library and laboratory facilities, respectively.

### SUPPLEMENTARY MATERIALS

Supplementary data associated with this article can be found in the online version at http://www.aaqr.org.

### REFERENCES

- Balakrishna, G. and Pervez, S. (2009). Source Apportionment of Atmospheric Dust Fallout in an Urban-Industrial Environment in India. *Aerosol Air Qual. Res.* 9: 359– 367.
- Balakrishna, G. and Pervez, S. (2011). Soil as a Source Contributor in Mineral Dust Fallout at Urban Industrial Residential Area. *Iran. J. Earth Sci.* 3: 80–88.

Balakrishna, G., Pervez, S. and Bisht, D.S. (2011). Source

Apportionment of Arsenic in Atmospheric Dust fall out in an Urban Residential Area, Raipur, Central India. *Atmos. Chem. Phys.* 11: 5141–5151.

- Birch, M.E. and Cary, R.A. (1996). Elemental Carbon-Based Method for Monitoring Occupational Exposures to Particulate Diesel Exhaust. *Aerosol Sci. Technol.* 25: 221–241.
- Brook, J.R., Vega, E. and Watson, J.G. (2003). In Particulate Matter Science for Policy Makers, A NARSTO Assessment, Part 2, NARSTO, Pasco, WA, p. 7–1.
- Cao, J.J., Lee, S.C., Zhang, X.Y., Chow, J.C., An, Z.S., Ho, K.F., Watson, J.G., Fung, K., Wang, Y.Q. and Shen, Z.X. (2005). Characterization of Airborne Carbonate over a Site near Asian Dust Source Regions during Spring 2002 and Its Climatic and Environmental Significance. J. Geophys. Res. 110: D03203, doi: 10.1029/2004JD005244.
- Cass, G.R. and McRae, G.S. (1983). Source-Receptor Reconciliation of Routine Air Monitoring Data for Trace Metals: An Emission Inventory Assisted Approach. *Environ. Sci. Technol.* 17: 29–139.
- Census (2011). Office of the Registrar General and Census Commissioner, India. Available from: http://www.censusindia.gov.in/2011.
- Chakrabarty, R.K., Pervez, S., Chow, J.C., Watson, J.G., Dewangan, S., Robles, J. and Tian, G. (2013). Funeral Pyres in South Asia: Brown Carbon Aerosol Emissions and Climate Impacts. *Environ. Sci. Technol. Lett.* 1: 44– 48, doi: 10.1021/ez4000669.
- Chen, S.J., Hsieh, L.T. and Chiu, S.C. (2003). Emission of Polycyclic Aromatic Hydrocarbons from Animal Carcass Incinerators. *Sci. Total Environ.* 313: 61–76.
- Chow, J.C. and Watson, J.G. (1994). Contemporary Source Profiles for Geological Material and Motor Vehicle Emissions. Final Report, Cooperative Agreement CR816826-02-01.DRI Document No. 2625.2F, Prepared for Office of Air Quality Planning and Standards, Environmental Protection Agency, Research Triangle Park, NC, by the Desert Research Institute, Reno, NV, 25 February.
- Chow, J.C., Watson, J.G., Ashbaugh, L.L. and Magliano, K.L. (2003). Similarities and Differences in PM<sub>10</sub> Chemical Source Profiles for Geological Dust from the San Joaquin Valley, California. *Atmos. Environ.* 37: 1317– 1340.
- Chow, J.C., Watson, J.G., Kuhns, H., Etyemezian, V., Lowenthal, D.H., Crow, D., Kohl, S.D., Engelbrecht, J.P. and Green, M.C. (2004). Source Profile of Industrial, Mobile, and Area Sources in the Big Bend Regional Aerosol Visibility and Observational Study. *Chemosphere* 54: 185–208.
- Chow, J.C., Watson, J.G. and Lowenthal, D.H. (2012). Evaluation of PM<sub>2.5</sub> and PM<sub>10</sub> Mass Closure Formulae, The Air & Waste Management Association: Aerosol and Atmospheric Optics: Visibility and Pollution Specialty Conference, Whitehead, MT.
- Chriswell, C.D., Mroch, D.R. and Markuszewski, R. (1986). Determination of Total Sulfur by Ion Chromatography Following Peroxide Oxidation in Spent Caustic from the Chemical Cleaning of Coal. *Anal. Chem.* 58: 319–321.

- CPCB (2008). STATIONARY SOURCES EMISSION PROFILES, Central Pollution Control Board, Delhi. Available from http://www.cpcb.nic.in/Stationary\_Sour ces\_Emission\_Profiles.xls.
- CPCB (2010). Air Quality Monitoring, Emission Inventory and Source Apportionment Study for Indian Cities, National Summary Report. Central Pollution Control Board, Delhi. Available from: http://moef.nic.in/downlo ads/public-information/Rpt-airmonitoring-17-01-2011.pdf.
- Dewangan, S., Chakrabarty, R., Zielinska, B. and Pervez, S. (2013). Emission of Volatile Organic Compounds from Religious and Ritual Activities in India. *Environ. Monit. Assess.* 185:9279–9286, doi: 10.1007/s10661-013-3250-z.
- Dionex (2005). Operational Manual: ICS 2000 Ion Chromatography Systems, Dionex Corporation, California, USA.
- DoCI (2012). Annual Report, Department of Commerce and Industry, Government of Chhattisgarh, India.
- DRI (2011). Laboratory Analysis Methods, Desert Research Institute, Reno, NV, USA.
- Dubey, N. and Pervez, S. (2008). Investigation of Variation in Ambient PM<sub>10</sub> Levels within an Urban-Industrial Environment. *Aerosol Air Qual. Res.* 8: 54–64.
- Dubey, N. (2011). Study of Particulate Source Apportionment at Classified Atmospheric Receptors in Selected Defined Urban Areas, Ph. D. Thesis, Pt. Ravishankar Shukla University, Raipur, India.
- Gadkari, N.M. and Pervez, S. (2007). Source Investigation of Personal Particulates in Relation to Identify Major Routes of Exposure among Urban Residentials. *Atmos. Environ.* 41: 7951–7963.
- Gadkari, N. and Pervez, S. (2008). Source Apportionment of Personal Exposure of Fine Particulates among School Communities in India. *Environ. Monit. Assess.* 142: 227– 241.
- Gilbert, R.O. (1987). *Statistical Methods for Environmental Pollution Monitoring*, Van Nostrad Reinhold, New York, USA.
- Gupta, A.K., Karar, K. and Srivastava, A. (2007). Chemical Mass Balance Source Apportionment of PM<sub>10</sub> and TSP in Residential and Industrial Sites of an Urban Region of Kolkata, India. *J. Hazard. Mater*.142: 279– 287.
- Gurugubelli, B., Pervez, S. and Tiwari, S. (2013). Characterization and Spatiotemporal Variation of Urban Ambient Dust Fallout in Central India. *Aerosol Air Qual. Res.* 13: 83–96.
- Harrison, R.M. and Perry, R. (1986). Handbook of Air Pollution Analysis (2nd Eds.), Chapman Hall, New York.
- Ho, K.F., Lee, S.C., Chow J.C. and Watson, J.G. (2003). Characterization of PM<sub>10</sub> and PM<sub>2.5</sub> Source Profile for Fugitive Dust in Hong Kong. *Atmos. Environ.* 37: 1023– 1032.
- Hopke, P.K. (1999). Elemental Analysis of Airborne Particles, Landsberger, S. and Creatchman, M. (Eds.), Gordon and Breach Science, Amsterdam, p. 273–315.
- Katz, M. (1977). *Methods of Air Sampling and Analysis,* American Public Health Association (APHA), Interdisciplinary Book and Periodicals, Washington D.C.

- Kerminen, V.M., Hillamo, R., Teinila, K., Pakkanen, T., Allegrini, I. and Sparapani, R. (2001). Ion Balances of Size-Resolved Tropospheric Aerosol Samples: Implications for the Acidity and Atmospheric Processing of Aerosols. *Atmos. Environ.* 35: 5255–5265.
- Khan, M.F., Hirano, K. and Masunaga, S. (2010). Quantifying the Sources of Hazardous Elements of Suspended Particulate Matter Aerosol Collected in Yokohama, Japan. *Atmos. Environ.* 44: 2646–2657.
- Kong, S., Ji, Y., Lu, B., Chen, L., Han, B., Li, Z. and Bai, Z. (2011). Characterization of PM<sub>10</sub> Source Profiles for Fugitive Dust in Fushun-a City Famous for Coal. *Atmos. Environ.* 45: 5351–5365.
- Kong, S.F., Han, B., Bai, Z.P., Chen, L., Shi, J.W. and Xu, Z. (2010). Receptor Modeling of PM<sub>2.5</sub>, PM<sub>10</sub> and TSP in Different Seasons and Long-Range Transport Analysis at a Coastal Site of Tianjin, China. *Sci. Total Environ.* 408: 4681–4694.
- Kulshreshtha, A. Bisht, D.S., Masih, J., Massey, D.D., Tiwari, S. and Taneja, A. (2010). Chemical Characterization of Water-Soluble Aerosols in Different Residential Environments of Semi Arid Region of India. J. Atmos. Chem. 62: 121–138, doi: 10.1007/s10874-010-9143-4.
- Kuykendal, W.B., Brooks, G., Waddell J.T. and Butler, W.A. (1990). In *Transactions, Visibility and Fine Particles,* Mathai, C.V. (Ed.), Air and Waste Management Association, Pittsburg, PA, p. 447–455.
- Mitra, A.P., Morawska, L., Sharma, C. and Zhang, J. (2002). Chapter Two: Methodologies for Characterization of Combustion Sources and for Quantification of Their Emissions. *Chemosphere* 49: 903–922.
- NIOSH (1999). In Manual of Analytical Methods, National Institute for Occupation Safety and Health, Cincinnati, OH.
- Pant, P. and Harrison, R.M. (2012). Critical Review of Receptor Modeling for Particulate Matter- A Case Study of India. *Atmos. Environ.* 49: 1–12.
- Patil, R.S., Kumar, R., Menon, R., Shah, M.K. and Sethi, V. (2013). Development of Particulate Matter Speciation Profiles for Major Sources in Six Cities in India. *Atmos. Res.* 132–133: 1–11.
- Pervez, S., Mathew, J. and Sharma, R. (2005). Investigation of Personal-Indoor -Outdoor Particulate Relationships in Welding Workshops. *Counc. Sci. Ind. Res.* 64: 454–458.
- Pervez, S., Dubey, N., Watson, J.G., Chow, J.C. and Pervez, Y. (2012). Impact of Different Household Fuel Use on Source Apportionment Results of House-Indoor RPM in Central India. *Aerosol Air Qual. Res.* 12: 49–60.
- Pipal, A.S., Jan, R., Bisht, D.S., Srivastava, A.K., Tiwari, S. and Taneja, A. (2014). Day and Night Variability of Atmospheric Organic and Elemental Carbon during Winter of 2011-12 in Agra, India. *Sustainable Environ. Res.* 24: 107–116.
- Samara, C., Kouimtzis, T., Tsitouridou, R., Kanias, G. and Simeonov, V. (2003). Chemical Mass Balance Source Apportionment of PM<sub>10</sub> in an Industrialized Urban Area of Northern Greece. *Atmos. Environ.* 37: 41–54.
- Samara, C. (2005). Chemical Mass Balance Source Apportionment of TSP in a Lignite Burning Area of

Western Macedonia, Greece. *Atmos. Environ.* 39: 6430–6443.

- Schauer, J.J., Mader, B.T., Deminter, J.T., Heidemann, G., Bae, M.S. and Seinfeld, J.H. (2003). ACE-Asia Intercomparison of a Thermal-Optical Method for the Determination of Particle-Phase Organic and Elemental Carbon. *Environ. Sci. Technol.* 37: 993–1001.
- Shen, Z.X., Cao, J.J., Arimoto, R., Zhang, R.J., Jie, D.M. and Liu, S.X. (2007). Chemical Composition and Source Characterization of Spring Aerosol over Horqin Sand Land in Northeastern China. J. Geophys. Res. 112: D14315, http://dx.doi.org /10.1029/ 2006JD007991.
- Shen, Z.X., Cao, J.J., Arimoto, R., Han, Z.W., Zhang, R.J., Han, Y.M., Liu, S.X., Okuda, T., Nakao, S. and Tanaka, S. (2009). Ionic Composition of TSP and PM<sub>2.5</sub> during Dust Storms and Air Pollution Episodes at Xi'an, China. *Atmos. Environ.* 43: 2911–2918.
- Swamy, K., Kaliaperumal, R. and Swaminathan, G.S. (1994). Wavelength Dispersive X-Ray Fluorescence Spectrometric Technique for Determining Elements in Weld Fumes. *X-Ray Spectrom.* 23: 71–74.
- Tao, J., Zhang, L., Engling,G., Zhang, R., Yang, Y., Cao, J., Zhu, C., Wang, Q. and Luo, L. (2013). Chemical Composition of PM<sub>2.5</sub> in an Urban Environment in Chengdu, China: Importance of Springtime Dust Storms and Biomass Burning. *Atmos. Res.* 122: 270–283.
- Thermo Fisher (2008). *Atomic Absorption Spectrometric Methods Manual: Issue-5*, Thermo Fisher Corporation, Cambridge, UK.
- Tsai, J.H., Lin, K.H., Chen, C.Y., Ding, J.Y., Choa, C.G. and Chiang, H.L. (2007). Chemical Constituents in Particulate Emissions from an Integrated Iron and Steel Facility. *J. Hazard. Mater*.147: 111–119.
- USEPA (1999a). Compendium Method IO-3.4: Determination of Metals in Ambient Particulate Matter Using Inductively Coupled Plasma (ICP) Spectroscopy, Compendium of Methods for the Determination of Inorganic Compounds in Ambient Air.
- USEPA (1999b). Compendium Method IO-3.1: Selection, Preparation and Extraction of Filter Material", Compendium of Methods for the Determination of Inorganic Compounds in Ambient Air.
- USEPA (2002). SPECIATE.EPA's Repository of Total Organic Compound (TOC) and Particulate Matter (PM) Speciated Profiles for a Variety of Sources for Use in Source Apportionment Studies. Website Available from: http://www.epa.gov/ttn/chief/software/speciate.
- USEPA (2013). PM Composite Profiles by Source Category. Available from: http://cfpub.epa.gov/si/speciate/ehpa\_sp eciate\_browse.cfm.
- Vega, E., Mugica, V., Reyes, E., Sanchez, G., Chow, J.C. and Watson, J.G. (2001). Chemical Composition of Fugitive Dust Emitters in Mexico City. *Atmos. Environ*.35: 4033– 4039.
- Viana, M., Querol, X. and Alastuey, A. (2006). Chemical characterisation of PM episodes in NE Spain. *Chemosphere* 62: 947–956.
- Viana, M., Kuhlbusch, T.A.J., Querol, X., Alastuey, A., Harrison, R.M., Hopke, P.K., Winiwarter, W., Vallius,

M., Szidat, S., Prévôt, A.S.H., Hueglin, C., Bloemen, H., Wåhlin, P., Vecchi, R., Miranda, A.I., Kasper-Giebl, A., Maenhaut, W. and Hitzenberge, R. (2008). Source Apportionment of Particulate Matter in Europe: A Review of Methods and Results. *J. Aerosol Sci.* 39: 827–849.

- Volkovic, V. (1983). Trace Elements in Coal, Vol. II, CRC Press, Florida.
- Watson, J.G., Cooper, J.A. and Huntzicker, J.J. (1984). The Effective Variance Weighting for Least Squares Calculations applied to the Mass Balance Receptor Model. *Atmos. Environ.* 18: 1347–1355.
- Watson, J.G., Robinson, N.F., Chow, J.C., Henry, R.C., Kim, B.M., Pace, T.G., Meyer, E.L. and Nguyen, Q. (1990). The USEPA/DRI Chemical Mass Balance Receptor Model, CMB 7.0. *Environ. Software* 5: 38–49.
- Watson, J.G., Chow, J.C. and Pace, T.G. (1991). In Receptor Modeling for Air Quality Management, Hopke, P.K. (Ed.), Elsevier press, New York, NY, p. 83–116.
- Watson, J.G., Chow, J.C., Lu, Z., Fujita, E.M., Lowenthal, D.H. and Lawson, D.R. (1994).Chemical Mass Balance Source Apportionment of PM<sub>10</sub> during the Southern California Air Quality Study. *Aerosol Sci. Technol.* 21: 1–36.
- Watson, J.G., Chow, J.C. and Frazier, C.A. (1999). In *Elemental Analysis of Airborne Particles, Vol. 1,* Landsberger, S. and Creatchman, M. (Eds.), Gordon and Breach Science, Amsterdam, p. 67–96.
- Watson, J.G. and Chow, J.C. (2001). Source Characterization of Major Emission Sources in the Imperial and Mexicali Valleys along the US/Mexico Border. *Sci. Total Environ.* 276: 33–47.
- Watson, J.G., Chow, J.C. and Houck, J.E. (2001). PM<sub>2.5</sub> Chemical Source Profile for Vehicle Exhaust, Vegetative Burning, Geological Material, and Coal Burning in Northwestern Colorado during 1995. *Chemosphere* 43: 1141–1151.
- Watson, J.G., Zhu, T., Chow, J.C., Engelbrecht, J., Fujita, E.M. and Wilson, W.E. (2002). Receptor Modeling Application Framework for Particle Source Apportionment. *Chemosphere*49: 1093–1136.
- Watson, J.G. and Chow, J.C. (2007). In *Introduction to Environmental Forensics*, 2<sup>nd</sup> Edition, Murphy, B. and Morrison, R. (Eds.), Academic Press, New York, NY, p. 279.
- Watson, J.G., Chen, L.W.A., Chow, J.C., Doraiswamy, P. and Lowenthal, D.H. (2008). Source Apportionment: Findings from the U.S. Supersites Program. J. Air Waste Manage. Assoc. 58: 265–288.
- Watson, J.G., Chow, J.C., Lowenthal, D.H., Chen., L.W. and Wang, X. (2012). Reformulation of PM<sub>2.5</sub> Mass Reconstruction Assumptions for the San Joaquin Valley. Report for Peter Biscay, San Joaquin Valley Air Pollution Study Agency 1990 E. Gettysburg Ave. Fresno, CA 93726.
- Yang, H.H., Lai, S.O., Hsieh, L.T., Hsueh, H.J. and Chi, T.W. (2002). Profiles of PAH Emission from Steel and Iron Industries. *Chemosphere* 48: 1061–1074.
- Yatkin, S. and Bayram, A. (2008). Determination of Major Natural and Anthropogenic Source Profiles for

Particulate Matter and Trace Elements in Izmir, Turkey. *Chemosphere* 71: 685–696.

Sources. Atmos. Res. 102: 110–119.

Zhang, T., Cao, J.J., Tie, X.X., Shen, Z.X., Liu, S.X., Ding,
H., Han, Y.M., Wang, G.H., Ho, K.F., Qiang, J. and Li.
W.T. (2011). Water-Soluble Ions in Atmospheric Aerosols
Measured in Xi'an, China: Seasonal Variations and

Received for review, March 28, 2014 Accepted, June 25, 2014

# **Supplementary Materials**

Instrument	Species	Wavelength	Calibration	LOD in	Blank (g/Filter)	Field blank (g/Filter)
AAS-Flame	Са	422.7	0.999	0.0037*	0.000010	0.000010
	Fe	248.3	0.999	0.0043*	0	0.000007
	K	766.5	0.999	0.0009*	0.000002	0.000007
	Mg	285.2	0.999	0.0022*	0.000003	0.000002
	Na	589.0	0.994	0.0037*	0.000024	0.000034
	Zn	213.9	0.999	0.0033*	0.000007	0.000010
AAS-GF/Zeeman	Al	309.3	0.978	0.21	0	0.000099
	Cd	228.8	0.999	0.02	0	0
	Со	240.7	0.998	0.01	0.00000004	0.00000005
	Cr	357.9	0.998	0.025	0.00000009	0.0000003
	Cu	324.8	0.993	0.29	0.0000004	0.0000007
	Mn	279.5	0.962	0.06	0.00000002	0.0000010
	Мо	313.3	0.981	0.31	0.0000002	0.0000004
	Ni	232.0	0.999	0.16	0.0000002	0.0000021
	Pb	283.3	0.982	0.07	0	0
	V	318.5	0.998	2.7	0.00000003	0.0000002
	As	193.7	0.999	0.53	0	0.000000009
	Sb	217.6	0.999	0.4	0.0000000008	0.000000005
	Se	196.0	0.999	0.8	0.00000002	0.00000002
AAS- VP	Hg	253.7	0.999	0.06	0.0000002	0.00000059
Spectrophotometer	$\mathrm{NH_4}^+$	620	0.998	-	0.0000025	0.00000774
Ion chromatograph	F-	-	-	0.01*	0	0.0000185
	Cl	-	-	0.01*	0.0000643	0.000103
	$NO_3^-$	-	-	0.01*	0	0
	$SO_4^{2-}$	-	-	0.01*	0.000027	0.000044
SemiContinuous	OC	-	-	-	0.000109	0.000176
Carbon analyzer	EC	-	-	-	0.000000037	0.00000203

 Table S1. Instrument parameters with filter blank and field blank values.

\*LOD value in ppm unit.