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# Comparing RADseq and microsatellites to infer complex phylogeographic patterns, an empirical perspective in the Crucian carp, Carassius carassius, L. 

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## Keywords

Approximate Bayesian Computation, Postglacial recolonisation, Landscape Genetics, Conservation Biology, Study design, Population structure

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#### Abstract

The conservation of threatened species must be underpinned by phylogeographic knowledge. This need is epitomised by the freshwater fish Carassius carassius, which is in decline across much of its European range. Restriction site associated DNA sequencing (RADseq) is increasingly used for such applications, however RADseq is expensive, and limitations on sample number must be weighed against the benefit of large numbers of markers. This trade-off has previously been examined using simulation studies, however, empirical comparisons between these markers, especially in a phylogeographic context, are lacking. Here, we compare the results from microsatellites and RADseq for the phylogeography of C. carassius to test whether it is more advantageous to genotype fewer markers (microsatellites) in many samples, or many markers (SNPs) in fewer samples. These datasets, along with data from the mitochondrial cytochrome b gene, agree on broad phylogeographic patterns; showing the existence of two previously unidentified C. carassius lineages in Europe; one found throughout northern and central-eastern European drainages, and a second almost exclusively confined to the Danubian catchment. These lineages have been isolated for approximately 2.15 M years, and should be considered separate conservation units. RADseq recovered finer population structure and stronger patterns of IBD than microsatellites, despite including only $17.6 \%$ of samples ( $38 \%$ of populations and $52 \%$ of samples per population). RADseq was also used along with Approximate Bayesian Computation to show that the postglacial colonisation routes of C. carassius differ from the general patterns of freshwater fish in Europe, likely as a result of their distinctive ecology.


## Introduction

Phylogeographic studies have revealed that the contemporary distributions of European taxa and their genetic diversity have been largely shaped by the glacial cycles of the Pleistocene epoch, and in particular by range shifts during recolonisation from glacial refugia (Hewitt 1999). In freshwater fishes, the dynamics of recolonisation are tightly linked to the history of river drainage systems (Bianco 1990; Bănărescu 1990, 1992; Bernatchez \& Wilson 1998; Reyjol et al. 2006). For example, watersheds pose a significant barrier to fish dispersal, often resulting in strong genetic structuring across separate drainage systems (Durand et al. 1999; Hänfling et al. 2002). However, during glacial melt periods, ephemeral rivers and periglacial lakes can arise, providing opportunities for colonisation (Gibbard et al. 1988) of otherwise isolated drain basins (Grosswald 1980; Arkhipov et al. 1995). These processes have resulted in complicated recolonisation scenarios in Europe, which, in contrast to North America (Bernatchez \& Wilson 1998), appear to possess few general patterns of population structure. Furthermore, previous phylogeographic studies have predominantly focused on highly mobile, obligatory or facultatively lotic species, with more sedentary, lentic species being largely overlooked.

The crucian carp, Carassius carassius (Linnaeus 1758), is native to parts of central, eastern and northern Europe and almost exclusively restricted to lentic ecosystems, including lakes, ponds and river floodplains (Copp 1991; Copp et al. 2008). C. carassius, has recently experienced sharp declines in the number and sizes of populations throughout its native range, leading to some local population extinctions. The reasons for these declines include habitat loss through drought and terrestrialisation in England (Copp 1991; Wheeler 2000; Sayer et al. 2011), acidification (Holopainen \& Oikari 1992), poor water quality in the Danube river catchment (Navodaru et al. 2002), and hybridisation with several non-native species (Copp et al. 2010; Savini et al. 2010; Mezhzherin et al. 2012; Wouters et al. 2012; Rylková et al. 2013). The susceptibility of $C$.
carrassius to genetic isolation and bottlenecks is compounded by small population sizes (Hänfling et al. 2005) and low dispersal (Holopainen et al. 1997). Strong geographic structure is therefore likely in this species. Although the threats to C. carassius populations are recognised on a regional level (Lusk et al. 2004; Mrakovčić et al. 2007; Wolfram \& Mikschi 2007; Simic, V et al. 2009; Copp \& Sayer 2010), a global conservation strategy is missing. Broad scale phylogeographic data and definition of evolutionary significant units are essential for informing unified conservation efforts for this species (Frankham et al. 2002).

Phylogeographic data have traditionally been collected using mitochondrial gene regions and/or nuclear markers such as AFLPs and microsatellites. However, cost and time often limits the number of these nuclear markers used, which can result in low power for addressing phylogeographic questions (Cornuet \& Luikart 1996; Luikart \& Cornuet 2008; Landguth et al. 2012; Peery et al. 2012; Hoban et al. 2013). Single nucleotide polymorphisms (SNPs) are increasingly used in phylogeography for assessments of population structure (for example see Morin et al. 2010; Emerson et al. 2010; Hess et al. 2011; Hauser et al. 2011). However, being bi-allelic, SNP loci contain less information than highly polymorphic microsatellites (Coates et al. 2009) and therefore large numbers of SNPs are needed to provide adequate statistical power. SNP discovery and assay development, which has been costly and slow in the past, has recently been greatly facilitated by the invention of restriction site associated DNA sequencing (RADseq, (Miller et al. 2006)), which enables the fast identification of thousands of orthologous SNP markers in non-model organisms. Nevertheless, although next generation sequencing costs are falling, RADseq remains a relatively expensive approach, which often constrains the number of biological samples that can be included in a given study. Researchers are, therefore, faced with a trade-off between the number of samples and loci during study design. The optimal balance between the two is likely to be based on several important but often unknown properties of the study system in question, for example the strength of population structure (i.e. $F_{\mathrm{ST}}$ ). Identifying these properties and comparing the relative strengths and
weaknesses of different molecular markers have recently been highlighted as priority topics in landscape genetics and phylogeography (Epperson et al. 2010; Balkenhol \& Landguth 2011). Recent simulation studies have provided some important insights into this trade-off, for example, Schwartz \& McKelvey (2009) find that patchy geographic sampling along an IBD gradient could result in falsely identified distinct lineages, whereas Landguth et al. (2012) find that increasing the number of loci can strengthen the correlation between genetic and geographic distance for a given sample set. To date, comprehensive empirical comparisons between microsatellite and SNP markers in a phylogeographic context are lacking (but see Bradbury et al. 2015).

In the present study, we use a combination of mitochondrial DNA (mtDNA), microsatellites and genome-wide SNPs obtained from RADseq in order to: 1) produce a comprehensive phylogeography for C. carassius as a basis for Europe-wide conservation strategies, 2) test competing scenarios of postglacial recolonisation that have potentially contributed to the contemporary distribution of the species, and 3) compare the power of microsatellites and RADseq based population structure analyses, in the context of the first two objectives. In this third aim, we specifically ask, whether the benefits gained by the high numbers of markers obtained from RADseq outweigh the potential loss of power associated by the reduction in the number of samples in our system.

## Materials and Methods

## Sample collection and DNA extraction

C. carassius is a Cyprinid native to much of continental Europe; latitudinally from the North Sea and Baltic Sea basins, through central Europe north of the Alps down to the Ponto-Caspian region and longitudinally from Belgium and perhaps northern France into Siberia (Lelek 1980). However, the true extent of this native range is unknown, largely due to difficulties in morphologically distinguishing it from three closely related, introduced and widespread species: Carassius auratus,

Carassius gibelio, and Cyprinus carpio (Wheeler 2000; Hickley \& Chare 2004). We initially collected 1354 samples from 72 populations across 13 European countries, but due to frequent hybridisation between the C. carassius and the three species mentioned above, it was necessary to identify and remove hybrids from this sample set. To this end, all samples were first genotyped at 6 species diagnostic microsatellite loci. We removed all samples identified as hybrids from the dataset and, to safeguard against cryptic hybridisation, we also removed all C. carassius that were sympatric with hybrids (see SI text for full details of species identification and hybrid detection). This left 867 C. carassius samples from 57 populations across the species' distribution in central and northern Europe (Table 1, Fig. 1). Sample sizes ranged from $n=4$ to $n=37$, with a mean of $n=17$ (Table 1). Fish were anaesthetised by a UK Home Office (UKHO) personal license holder (GHC) in a $1 \mathrm{~mL} \mathrm{~L}^{-1}$ bath of 2-phenoxyethanol prior to collection of a $1 \mathrm{~cm}^{2}$ tissue sample from the lowercaudal fin, and wounds were treated with a mixture of adhesive powder (Orahesive) and antibiotic (Cicatrin) (Moore et al. 1990). Tissue samples were immediately placed in $\geq 95 \%$ ethanol, and stored at $-20^{\circ} \mathrm{C}$. DNA was extracted from $2-4 \mathrm{~mm}^{2}$ of each tissue sample using either the Gentra Puregene DNA isolation kit or the DNeasy DNA purification kit (both Qiagen, Hilden, Germany). For the RADseq library, DNA was quantified using the Quant-iT ${ }^{\text {TM }}$ PicoGreen ${ }^{\circledR}$ dsDNA Assay kit (Invitrogen) and normalised to concentrations $\geq 50 \mathrm{ng} \mathrm{ml}^{-1}$. Gel electrophoresis was then used to check that DNA extractions contained high molecular weight DNA.

## Molecular markers and methods

Three types of molecular markers were used in this study. Mitochondrial DNA sequencing was used to identify highly distinct lineages and to date the divergence between them through phylogenetic analysis. Two sets of nuclear markers; microsatellites and RADseq-derived SNPs, were used to investigate more recent and complex structure in a population genetics framework and to compare the relative power of each marker to do so.

## Mitochondrial DNA amplification

A total of 83 C. carassius individuals, randomly chosen from a subset of 30 populations, which were chosen to represent all major catchment areas and the widest possible geographic range (min. $\mathrm{n}=1$, max. $\mathrm{n}=4$, mean $\mathrm{n}=2.7$ ), were sequenced at the cytochrome b (cytb) gene (Table 1). PCR reactions were carried out following the protocol in Takada et al. (2010) using the forward and reverse primers L14736-Glu and H15923-Thru on an Applied Biosciences® Veriti Thermal Cycler. PCR products were sequenced in both directions on an ABI3700 by Macrogen Europe. The forward and reverse cytb sequence reads were aligned using a GenBank sequence from the UK (accession no. JN412539, Table 1) as a reference and ambiguous nucleotides were manually edited using CodonCode aligner v.2.0.6 (CodonCode Corporation).

## Microsatellite amplification

Of the 867 samples identified as pure C. carassius, 19 samples were in populations with sample numbers which were too low to be useful for population genetics analyses ( $<4$ ). The remaining 848 samples, from 49 populations, were genotyped at 13 microsatellite loci, including the six species diagnostic loci used for hybrid identification (Supporting Information (SI) Table 1). Microsatellites were amplified in three multiplex PCR reactions, using the Qiagen multiplex PCR mix with manufacturer's recommended reagent concentrations, including $Q$ solution and $1 \mu \mathrm{l}$ of template DNA. Primer concentrations for each locus are provided in SI Table 1 and PCRs were performed on an Applied Biosciences ${ }^{\circledR}$ Veriti Thermal Cycler. The annealing temperature used was $54^{\circ} \mathrm{C}$ for all reactions, and all other PCR cycling parameters were set to Qiagen multiplex kit recommended values. PCR products were run on a Beckman Coulter CEQ 8000 genome analyser
using a 400 bp size standard and microsatellite alleles scored using the Beckman Coulter CEQ8000 software.

## RADseq

A total of 160 individuals ( 18 populations, min. $n=8$, max. $n=10$, mean $n=8.9$ ), identified as pure C. carassius with the diagnostic microsatellites, were used in the RADseq (Table 1). These samples were chosen to represent a wide geographic range and all major phylogeographic clusters identified using the microsatellite data. These samples were split across 13 libraries prepared at Edinburgh Genomics (University of Edinburgh, UK) according to the protocol in Davey et al. (2012) using the enzyme $S b f 1$. Libraries were then sequenced using paired-end sequencing across five lanes of two Illumina HiSeq 2000 flowcells (Edinburgh Genomics).

## Data analyses

## Phylogenetic analysis of mtDNA

In addition to the 83 sequenced samples (SI Table 2), we retrieved 19 published C. carassius and three C. carpio cytb sequences from GenBank to be used as an outgroup. The C. carpio samples were chosen to include samples from multiple, distant lineages of C. carpio located in Japan, Greece and India. All sequences used were validated through cross checking with their original publications (Table 1). Sequence alignment was performed in MEGA6 (Tamura et al. 2013) using default settings, and DNAsp v.5.0 (Librado \& Rozas 2009) was used to calculate sequence divergence and to identify haplotypes.

Haplotypes of all C. carassius samples and the three C. carpio outgroup individuals were exported to BEAST v.1.7.5 (Drummond et al. 2012) for phylogenetic analyses in order to identify the major
phylogenetic lineages within European C. carassius. Phylogenetic model testing with jModeltest2 v.2.1.7 (Guindon et al. 2003; Darriba et al. 2012) using Akaike information criterion (AIC), Bayesian Information Criteria (BIC) and the decision-theoretic performance-based (DT) approach showed that HKY (Hasegawa et al. 1985) was the most appropriate substitution model for our dataset. Using this model, the splits between the major phylogenetic clades were then dated using a relaxed molecular clock method in BEAST. The widely-used Dowling et al. (2002) cyprinid cytb divergence rate of $1.05 \%$ pairwise sequence divergence / MY was used after converting to a per lineage value of 0.0053 mutations/site/MY for use in BEAST. We used a 'coalescent: constant size' tree prior, which assumes an unknown but constant population size backwards in time, as recommended for intraspecific phylogenies (Drummond et al. 2012) . MCMC chain lengths were 1 x $10^{7}$ with samples taken every 1000 iterations. A gamma site heterogeneity model was used, with the default of four categories. Substitution rates, rate heterogeneity and base frequencies were unlinked between each codon position to allow substitution rate to vary between them. Default values were used for all other parameters and priors.

## Population structure and diversity analyses using microsatellites

Allele dropout and null alleles in the microsatellite data were tested using Microchecker (Van Oosterhout et al. 2004). FSTAT v. 2.9.3.2 (Goudet 1995) was then used to check for linkage disequilibrium (LD) between loci (using 10,000 permutations), deviations from Hardy-Weinberg equilibrium (HWE) within populations (126500 permutations) and for all population genetic summary statistics. Genetic diversity within populations was estimated using Nei's estimator of gene diversity $\left(H_{\mathrm{o}}\right)\left(\right.$ Nei 1987) and Allelic richness $\left(A_{\mathrm{r}}\right)$, which was standardised to the smallest sample size $(\mathrm{n}=4)$ using rarefaction (Petit et al. 1998). Pairwise $F_{\text {ST }}$ values were calculated according to (Weir \& Cockerham 1984) and 23520 permutations and sequential Bonferroni correction were used to test for significance of $F_{\mathrm{ST}}$.

IBD was investigated using a Mantel test in the adegenet v1.6 (Jombart \& Ahmed 2011) package in R v3.0.1 (R Core Team 2013). We then tested for an association between $A_{\mathrm{r}}$ and longitude and latitude, which is predicted under a stepping-stone colonisation model (Ramachandran et al. 2005; Simon et al. 2014), using linear regression analysis in R.

Population structure was then further examined using Discriminant Analyses of Principal Components (DAPC) also in adegenet (DAPC, see SI text and Jombart et al. 2010 for more details). DAPC has been shown to perform as well or better than the commonly used program, STRUCTURE (Pritchard et al. 2000) for both simple and complex models of population structure (Jombart et al. 2010). Furthermore, unlike STRUCTURE, DAPC is free of underlying assumptions of Hardy-Weinberg equilibrium, which are likely to be violated when effective population sizes are small, as is often the case in C. carassius (Hänfling et al. 2005).

In preliminary DAPC analysis using all 49 C. carassius populations, Sweden (SWE9) was found to be so genetically distinct from the rest of the data set that it masked the variation between the other populations. This population was therefore omitted from further DAPC analyses. To infer the appropriate number of genetic clusters in the data, we used BIC scores (SI Fig. 5a), in all cases choosing lowest number of genetic clusters from the range suggested. Spline interpolation (Hazewinkel 1994) was then used to identify the appropriate number of principal components to use in the subsequent discriminant analysis (SI Fig. 5a).

## RADseq data filtering and population structure analysis

The quality of the RADseq raw read data was examined using FastQC (Andrews 2010), the dataset was then cleaned, processed and SNPs were called using the Stacks pipeline v 1.19 (Catchen et al. 2011). Preliminary tests were carried out in order to identify optimal Stacks parameters (See SI text). Final parameter values for the respective Stacks module were as follows; ustacks: $\mathrm{M}=2, \mathrm{~m}=8$,
removal (-r) and deleveraging (-d) algorithms were also used; cstacks: $\mathrm{N}=2$ ( n populations $=18, \mathrm{n}$ individuals = 160); populations module: one SNP per RAD locus was used (--write_single_snp) and SNPs were only retained if they were present in $70 \%$ of individuals ( $\mathrm{r}=0.7$ ) in at least 17 out of the 18 populations in the study $(\mathrm{p}=17)$, which allows for mutations in restriction sites that may cause loci to dropout in certain lineages. All other parameters were kept at default values. Finally, we filtered out loci which had a heterozygosity of $>0.5$ and $F_{\text {IS }}<0.0$ in one or more populations in order to control for the possibility of erroneously merging ohnologs resulting from the multiple genome duplications that have occurred the in Cyprinus and Carassius genera (Henkel et al. 2012; Xu et al. 2014). The resulting refined SNP set was then used in subsequent phylogeographic analyses. The R package Adegenet v. 1.42 was used to calculate $H_{\mathrm{o}}$ and pairwise $F_{\mathrm{ST}}$, test for IBD and genetic clusters were inferred using DAPC.

## Reconstructing postglacial colonisation routes in Europe

DIYABC v. 2.0 (Windows, Cornuet et al. 2014) was used to reconstruct the most likely $C$. carassius recolonisation routes through Europe after the last glacial maximum. We used the RADseq data set for this analysis as it showed a much clearer pattern of population structure than the microsatellite data in DAPC analyses (see Results). Furthermore, preliminary DIYABC analyses using microsatellites failed to identify a scenario which was significantly more likely than its counterparts, suggesting low power in this dataset for reconstructing complex phylogeographic patterns over long timescales.

As DIYABC is a computationally intensive method, it was necessary to perform analyses on a subset of 1000 randomly-selected SNP loci from the full RADseq dataset to reduce computation time. This SNP subset was first analysed with DAPC to confirm that it produced the same population structure as the full dataset and was then used to compare the likelihood of a number of user defined colonisation scenarios (i.e. a specific population tree topology, together with the
parameter prior distributions that are associated with it). First, 1 million datasets were simulated for each scenario. These simulated summary statistic datasets represented the theoretical expectation under each scenario, and were compared to the same summary statistics calculated from the observed data, in order to identify the most likely of the tested scenarios. In DIYABC, two methods of comparison between simulated and observed datasets are used; logistic regression and "direct approach", the latter method identifies the scenario that produces the largest proportion of the $n$ number of closest scenarios to the observed, where $n$ is specified by the user. The goodness-of-fit of scenarios was also assessed using the model checking function implemented in DIYABC (Cornuet et al. 2014). In all analyses, the single-sample summary statistics used were the mean and variance of gene diversity across all polymorphic loci and the mean gene diversity across all loci. The twosample summary statistics used were the mean and variance of $F_{\mathrm{ST}}$ and Nei's distance for loci with $F_{\mathrm{ST}}$ greater than zero between two samples and the mean $F_{\mathrm{ST}}$ and Nei's distance for all loci. Finally, for scenarios including admixture events, the maximum likelihood estimates of admixture proportions were also used. See Cornuet et al. (2014) for the exact equations used and their implementation in DIYABC.

To reduce the number and complexity of possible scenarios, we split DIYABC analysis into three stages (Table 2). In stage 1, we tested 11 broad scale scenarios (Scenarios $1-11$, SI Fig. 1).

Populations were grouped into three pools in order to reduce the number and complexity of possible scenarios (Table 2); Pool 1 - all northern European populations (npops $=17, n=155$ ), Pool $2-$ Don population (npops $=1, n=9$ ), Pool $3-$ Danubian population (npops $=1, n=6$ ). In six scenarios ( 1 , 2, 8-11), northern European and the Don population diverged from each other more recently than from Danubian populations. These scenarios differ in the patterns of effective population size change and the presence or absence of a bottleneck. In scenarios 3 and 4, northern European and Danubian populations are more closely related to each other than to the Don population. And in the remaining three scenarios, one pool of populations is the product of an admixture event between the
other two. Population poolings and scenarios were both chosen on the basis of the broad phylogeographic structure identified in the mtDNA and RADseq population structure analysis (see Results).

In the second and third stages, we performed a finer scale analysis, focussing on the 17 northern European populations alone. Populations were again pooled, this time into six groups, on the basis of both population structure and geography (Table 2). In stage 2 we tested five scenarios (Scenarios 12-16, see SI Fig. 2a for graphical description of each scenario), with no bottlenecks included, which represented the major topological variants that were most likely, given population structure results from DAPC. We then identified the most likely of these scenarios in DIYABC and took this forward into the final stage of the analysis where we tested 6 multiple bottleneck combinations (SI Fig. 2b) around this scenario. This three stage approach allowed us to systematically build a complex scenario for the European colonisation of C. carassius. Finally, we used the posterior distributions of the time parameters, simulated using the scenario identified as most likely in stages one and three, to estimate the times of the major lineage splits in European C. carassius. These parameters, calculated by DIYABC in generations, were converted to years using an average generation time of 2 years (Tarkan et al. 2010).

## Comparison of microsatellite and RADseq data

Finally, we compared the results derived from population structure analyses on microsatellite and RADseq data to assess their suitability for addressing our phylogeographic question. It is important to note that differences between the full microsatellite and RADseq datasets could be attributable to one or a combination of the following; the number of populations, the geographic distribution of populations, the number of samples per population, the number of markers, or the information content of the marker type. To disentangle these sources of variation, we created two microsatellite
data subsets; M2, which included only individuals used in RADseq, (excluding three individuals for which microsatellite data was incomplete, $n=146$, npops $=19$ ), and M3, which contained all individuals for which microsatellite data was available in populations that were used in RADseq ( $n$ $=313$, npops $=19$;

Table 3). This gave us three pairs of datasets for comparison: 1) RADseq Vs. M2: same individuals but different marker types, 2) M1 vs M2: full microsatellite dataset versus a subset of the populations, and 3) M2 vs M3: same populations but different number of individuals per population. This strategy enabled us to test for the influence of marker, sampling of populations and individuals per population respectively. Comparisons were performed between datasets on heterozygosities and pairwise $F_{\mathrm{ST}}$ using both Pearson's product-moment correlation coefficient and paired Student's t-tests in R. IBD results were compared using Mantel tests (Jombart \& Ahmed 2011), and DAPC results were compared on the basis of similarity of number of inferred clusters and cluster sharing between populations.

## Results

## Phylogenetic analyses of mitochondrial data

The combined 1090 bp alignment of 100 cytb C. carassius mtDNA sequences yielded 22 haplotypes, which were split across two well supported and highly differentiated phylogenetic lineages (Fig. 2, SI Table 3). Lineage 1 was found in all northern European river catchments sampled, as well as eastern European (Dnieper) and southeastern European (Don and Volga) catchments, whereas Lineage 2 was almost exclusively confined to the River Danube catchment. There were, however, a few exceptions to this clear geographical split; two individuals, one from the Elbe and one from the Rhine in northern Germany, belonged to mtDNA Lineage 2, as did one individual from the River Lahn river catchment in western Germany. Also one population in the Czech Republic, located on the border between the Danube and Rhine river catchments, was found to contain individuals belonging to lineages 1 and 2 .

The mean number of nucleotide differences within lineages 1 and 2 was 2.25 and 2.00 , respectively, which equated to a sequence divergence $0.2 \%$ and $0.18 \%$, respectively. Between the two lineages
there was an average of 22.5 nucleotide differences ( $2.06 \%$ mean sequence divergence), with 19 of these being fixed. BEAST molecular clock analysis dated the split between lineages 1 and 2 to be 1.30-3.22 million years ago (MYA), with a median estimate of 2.15 MYA (Fig. 2).

## Nuclear marker datasets and quality checking

Microchecker showed no consistent signs of null alleles or allele dropout in microsatellite loci and no significant LD was found between any pairs of loci. No populations showed significant deviation from Hardy-Weinberg proportions (adjusted nominal level 0.0009).

After filtering raw RADseq data, de novo construction of loci across the 19 populations produced 35709 RADseq loci that were present in at least $70 \%$ of individuals in at least 17 populations. These loci contained a total of 29927 polymorphic SNPs (approx. 0.84 SNPs per locus). Only the first SNP in each RADseq locus was retained, to avoid confounding signals of LD. This yielded a total of 18908 loci with a mean coverage of 29.07 reads (SI Fig. 3b). Finally 5719 of these SNP loci were filtered out due to high ( $>0.5$ ) heterozygosity and/or $\mathrm{F}_{\text {IS }}$ of $<0.0$ in at least one population. In doing so, we removed many high coverage tags (SI Fig. 3a), which was consistent with over-merged ohnologs having higher coverage (i.e. reads from more than two alleles) than correctly assembled loci. The final dataset therefore contained 13189 SNP loci, with a mean coverage of 27.72 reads.

## Within population diversity at nuclear loci

Observed heterozygosity $\left(H_{0}\right)$, averaged across all microsatellite loci within a population, ranged from 0.06 (SWE9) to 0.44 (BLS), with a mean of 0.25 across all populations ( $\mathrm{SD}=0.105$ ), and was highly correlated with $A_{\mathrm{r}}(t=19.67, P<0.001, \mathrm{df}=40)$, which ranged from 1.26 (FIN1) to 2.96
(POL3) with a mean of $1.92(\mathrm{SD}=0.51)$. Mean $H_{\mathrm{o}}$ averaged across all RADseq loci for all populations was $0.013(\mathrm{SD}=0.013)$, ranged from 0.001 to 0.057 and was significantly correlated with $H_{\mathrm{o}}$ from microsatellite loci at populations shared between both datasets $(\mathrm{r}=0.69, t=3.74, P=$ $0.002, \mathrm{df}=15)$. Microsatellite $A_{\mathrm{r}}$ significantly decreased along an east to west longitudinal gradient (adj. $R^{2}=0.289, P<0.001$, SI Fig. 4b) consistent with decreasing diversity along colonisation routes. However, $A_{\mathrm{r}}$ did not decrease with increasing latitude (adj $\mathrm{R}^{2}=-0.007, \mathrm{P}=0.414$, SI Fig. 4a). We also repeated this analysis after removing samples from mtDNA Lineage 2 in the Danube catchment. Again there was no relationship between $A_{\mathrm{r}}$ and latitude $\left(R^{2}=-0.023, P=0.254\right.$, SI Fig. 4c), but the relationship between $A_{\mathrm{r}}$ and longitude was strengthened (adj. $R^{2}=0.316, P<0.001$, SI Fig. 4d).

## Population Structure in Europe based on nuclear markers

Population structure was strong, as predicted. Using the full (M1) microsatellite dataset, mean pairwise $F_{\mathrm{ST}}$ was $0.413(\min =0.0 ; \mathrm{BEL} 2$ and BEL3), $\max =0.864(\mathrm{NOR} 2$ vs GBR2), with 861 of the 1128 pairwise population comparisons being significant $F_{\text {ST }}\left(P<0.05\right.$, SI Table 4). Pairwise $F_{\text {ST }}$ calculated from the RADseq dataset also showed strong structure (SI Table 5), ranging from 0.067 (DEN1, DEN2) to 0.699 (NOR2, GBR4), and these values were highly correlated with the same population comparisons in the M3 microsatellite dataset $(\mathrm{r}=0.66, t=9.01, P<0.01, \mathrm{df}=104)$.

BIC scores obtained from initial DAPC analyses of the microsatellite dataset, using all 49 populations, indicated that between 11 and 19 genetic clusters (SI Fig. 5a) would be an appropriate model of the variation in the data. As a conservative estimate of population structure, we chose 11 clusters for use in the discriminant analysis, retaining eight principal components as recommended by the spline interpolation a-scores (SI Fig. 5a). This initial analysis showed that populations belonging to Cluster 10 (RUS1, Don river catchment) and Cluster 11 (GER3, GER4, CZE1,

Danubian catchment) were highly distinct from clusters found in northern Europe (Fig. 1b). Since the marked genetic differentiation between these three main clusters masked the more subtle population structure among northern European populations (see Fig. 1b), we repeated the DAPC analysis without the populations from the Danube and Don (RUS1, GER3, GER4, CZE1, Fig. 1b). The results of this second DAPC analysis revealed an IBD pattern of population structure, across Europe (Fig. 1). Mantel tests excluding the Danubian and Don populations corroborated these results; showing significant correlation with geographic distance in northern Europe (adjusted $R^{2}=$ $0.287, P<0.001$, SI Fig. 6a), with Danubian populations shown to be more diverged than their geography would predict (data not shown).

In the RADseq DAPC analysis, BIC scores suggested between four and ten genetic clusters, a lower number than that inferred from the microsatellite data set. Again we chose the lowest number of suggested clusters (four) clusters to take forward in the analysis (SI Fig. 5b). Following spline interpolation, we retained six principal components and kept two of the linear discriminants from the subsequent discriminant analysis (SI Fig. 5b). The inferred population structure showed that the Danubian population (HUN2) and the Don population (RUS1) were highly diverged from the northern European clusters. Unfortunately, HUN2 is not present in the microsatellite dataset for direct comparison, however both datasets, and the mtDNA data show the same pattern of high divergence between northern Europe and Danubian populations. DAPC analyses of RADseq data again showed an IBD pattern in northern European populations, which was confirmed with Mantel tests when the Danubian population HUN2 was excluded (adjusted $R^{2}=0.722, P<0.001$; SI Fig. 6 b ).

## Postglacial recolonisation of C. carassius in Europe

DAPC results of the 1000 SNP RADseq dataset used in DIYABC showed that it produced the same population structure as the full RADseq dataset (SI Fig. 7). For the broad-scale scenario tests in stage one of the DIYABC analysis, both logistic regression and direct approach identified Scenario 9 as being most likely to describe the true broad-scale demographic history (SI Fig. 8). Model checking showed that the observed summary statistics for our data fell well within those of the posterior parameter distributions for scenario 9 (SI Fig. 8c). Scenario 9 agrees with the mtDNA results, suggesting that the Danubian populations have made no major contribution to the colonisation of northern Europe. The median posterior distribution estimate of the divergence time between Danubian and northern European populations is 2.18 MYA ( $95 \% \mathrm{CI}=1.03-5.12 \mathrm{MYA}$ ), assuming a two-year generation time (Tarkan et al. 2010)), which is strikingly similar to that of mtDNA dating analysis. Scenario 9 also suggests that the northern European populations experienced a population size decline after the split of Pool 1 from the population in the Don river catchment, which lasted approximately 8920 years $(95 \% \mathrm{CI}=616-13700$ years $)$ and reduced $N_{e}$ by $32 \%$.

In stage two of the DIYABC analysis, we tested the major variant scenarios for the colonisation of northern Europe. In assessing the relative probabilities of scenarios, there was some discrepancy between the direct approach, which revealed Scenario 14 to be most likely, and the logistic regression, which favoured Scenario 13 (with Scenario 14 being the second most likely). However, the goodness-of-fit model checking showed that the observed dataset fell well within the posterior parameter distributions for Scenario 14 (SI Fig. 9a), but not for Scenario 13 (not shown). Therefore, Scenario 14 was carried forward into stage three in which we tested six more scenarios (SI Fig. 2b) to compare combinations of bottlenecks using the same population tree topology as in Scenario 14. Direct approach, logistic regression and model checking all found scenario 14 d to be the most likely (SI Fig. 9b), we therefore accepted this as the scenario for the colonisation of C. carassius in
northern Europe (SI Fig. 9b). This scenario infers an initial split between two sub-lineages in northern Europe approximately 33600 YBP (Fig. 4), one of which re-colonised northwest Europe and one that re-colonised Finland through the Ukraine and Belarus. Scenario 14d also inferred a secondary contact between these sub-lineages approximately 15940 YBP , resulting in the populations currently present in Poland; these admixed populations provided the source of one colonisation across the Baltic into Sweden, and a second route was inferred into southern Sweden from Denmark (Table 3, SI Fig. 9b).

## Comparing microsatellite datasets and $R A D$ sequencing data

The results from the RADseq $(n=149$, npops $=16)$ dataset and the full microsatellite dataset $(\mathrm{M} 1$, $n=848$, npops $=49)$ largely agreed on the inferred structure and cluster identity of populations. However, there were some important differences between them. Firstly, the IBD pattern of population structure in northern Europe was much stronger in the RADseq data $\left(\mathrm{R}^{2}=0.722, P<\right.$ 0.001 , SI Fig. 6) compared to the M1 dataset ( $\mathrm{R}^{2}=0.287, P<0.001$, excluding Danubian populations and SWE9 from both datasets, SI Fig. 6). Secondly, clusters inferred by the RADseq DAPC analysis are much more distinct, i.e. there is much lower within-cluster, and higher betweencluster variation in the RADseq results than in the M1 dataset results (Fig. 3).

As the properties of the RADseq and M1 datasets differ in four respects, namely marker type, number of populations, number of samples per population (Table 3) and uniformity of sampling locations, (SI Fig. 10) it was not possible to identify the cause of discrepancies in their results. Therefore, below we report the results from the pair-wise dataset comparisons, which isolate the effects of these parameter differences.

1) M1 Vs. M3: the effect that the number of populations and the uniformity of sampling locations might have on inferred population structure. The geographic distribution of sampling locations was more clustered in M1 (full microsatellite dataset) than in M3 (containing microsatellite for samples in populations used in RADseq (SI Fig. 10), and IBD patterns were considerably stronger in the M3 subset (adj. $R^{2}=0.447, P<0.001$ ) than in the full M1 dataset (adj. $R^{2}=0.287, P<0.001$ ). In contrast DAPC results were very similar between datasets, with inferred cluster number, structure and population identity of clusters generally agreeing well (Fig. 1, Fig. 3c).
2) $M 2 V s$. $M 3$ : the effect of reducing the number of samples per population on the inferred population structure. The number of samples per population in the M2 subset (microsatellite data only for the samples used in RADseq, mean $=9.125 \pm 0.8$ ) was significantly lower than that of the M3 subset (mean, $19.6 \pm 9.0, t=-4.66, \mathrm{df}=15, P<0.001$ ), as was the number of alleles per population $(\mathrm{M} 2$ mean $=24.4 \pm 7.3, \mathrm{M} 3$ mean $=27.4 \pm 8.1, t=-5.72, \mathrm{df}=15, P<0.001)$. Population heterozygosities were significantly different between M 2 and $\mathrm{M} 3(\mathrm{M} 2$ mean $=0.21$, M 3 mean $=$ $0.23, t=-2.4, \mathrm{df}=15, P=0.012)$, but highly correlated $(\mathrm{r}=0.94, t=-11.13, P<0.001, \mathrm{df}=15)$. Pairwise $F_{\mathrm{ST}}$ were very strongly correlated $(\mathrm{r}=0.97, t=46.26, P<0.001, \mathrm{df}=105)$, but again, still significantly different between the two datasets (M2 mean $=0.46$, M3 mean $=0.49, t=-6.21, P<$ $0.001, \mathrm{df}=15$, Table 4). The patterns of IBD were almost identical for $\mathrm{M} 2\left(\mathrm{R}^{2}=0.455, P<0.001\right)$ and M3 ( $\mathrm{R}^{2}=0.447, P<0.001$, SI Fig. 6 ) and population structure inferred by DAPC was again similar. BIC scores suggested a similar range of cluster number for M2 and M3, the smallest of which was nine in both cases.
3) RADseq Vs. M3: The effect of the number and the type of markers used on the phylogeographic results. We compared the results from the RADseq and M2 datasets, which contain exactly the same samples (with the exception of three individuals missing in M2). Significant correlations were again found between heterozygosities estimated for the two datasets $(\mathrm{r}=0.69, t=3.73, P=0.002$,
$\mathrm{df}=15)$ and pair-wise $F_{\mathrm{STS}}(\mathrm{r}=0.70, t=10.09, P<0.001, \mathrm{df}=105)$, but RADseq data yielded much lower pairwise $F_{\mathrm{ST}}$ (mean $\mathrm{RAD}=0.29$, mean $\mathrm{M} 2=0.46, t=13.74, P<0.001, \mathrm{df}=15$ ). DAPC analysis of RADseq data resolved populations into much more distinct clusters (Figs. 3a, $3 b)$, and the IBD pattern found was considerably stronger in the RADseq $\left(\mathrm{R}^{2}=0.722, P<0.001\right)$ dataset compared to M2 ( $\mathrm{R}^{2}=0.455, P<0.001$, SI Fig. 6 ).

## Discussion

In this study, we aimed to simultaneously produce a phylogeographic framework on which to base conservation strategies for C. carassius in Europe, and compare the relative suitability of genomewide SNP markers and microsatellite markers for such an undertaking. Through comparison of the inferred population structure from microsatellite and genome-wide SNP data, we show that there are important differences in the results from each data type, attributable predominantly to marker type, rather than within population sampling or spatial distribution of samples. However, despite these differences, all three data types used (mitochondrial, microsatellite and SNP data) agree that, unlike many other European freshwater fish for which phylogeographic data is available, $C$. carassius has not been able to cross the Danubian catchment boundary into northern Europe. This has resulted in two, previously unknown, major lineages of C. carassius in Europe, which we argue should be considered as separate conservation units.

## Phylogeography and postglacial recolonisation of C. carassius in Europe

The most consistent result across all three marker types (mtDNA sequences, microsatellites and RADseq) was the identification of two highly-divergent lineages of $C$. carassius in Europe. The distinct geographic distribution of these lineages; Lineage 1 being widely distributed across north and eastern Europe and Lineage 2 generally only in the River Danube catchment, indicates a long-
standing barrier to gene flow between these geographic regions. Bayesian inference based on mtDNA phylogeny and ABC analysis of RADseq data showed remarkable agreement, estimating that these lineages have been isolated for $2.15 \mathrm{MYA}(95 \% \mathrm{CI}=1.30-3.22)$ and $2.18(95 \% \mathrm{CI}=2-$ 6.12) MYA respectively, which firmly places the event at the beginning of the Pleistocene (2.6 MYA; (Gibbard \& Head 2009). This pattern differs substantially from the general phylogeographic patterns observed in other European freshwater fish. Indeed, previous studies have shown that the Danube catchment has been an important source for the postglacial recolonisation of freshwater fish into northern Europe or during earlier interglacials in the last 0.5 MYA . For example, bullhead Cottus gobio (Hänfling \& Brandl 1998; Hänfling et al. 2002), chub Leuciscus cephalus (Durand et al. 1999), Eurasian perch Perca fluviatilis (Nesbø et al. 1999), riffle minnow Leuciscus souffia (Salzburger et al. 2003), grayling Thymallus thymallus (Gum et al. 2009), European barbel Barbus barbus (Kotlík \& Berrebi 2001), and roach Rutilus rutilus (Larmuseau et al. 2009) all crossed the Danube catchment boundary into northern drainages such as those of the rivers Rhine, Rhône and Elbe during the mid-to-late Pleistocene. The above species occur in lotic habitats, and most are capable of relatively high dispersal. In contrast C. carassius has a very low propensity for dispersal, and a strict preference for the lentic backwaters, isolated ponds and small lakes (Holopainen et al. 1997; Culling et al. 2006; Copp 1991). We therefore hypothesise that these ecological characteristics of C. carassius have reduced its ability to traverse the upper Danubian watershed, which lies in a region characterised by the Carpathian Mountains and the Central European Highlands. This region may have acted as a barrier to the colonisation of C. carassius into northern European drainages during the Pleistocene. It should be noted, however, that the phylogeography of two species, the spined loach Cobitis taenia and European weatherfish Misgurnus fossilus, does not support this hypothesis as a general pattern for floodplain species (Janko et al. 2005; Culling et al. 2006). The former is the only species that we know of other than C. carassius showing long-term isolation between the Danube and northern European catchments, but has lotic habitat preferences and good dispersal abilities (Janko et al. 2005; Culling et al. 2006), whereas the latter inhabits
similar ecosystems as C. carassius, with low dispersal potential, but has colonised northern Europe from the Danube catchment (Bohlen et al. 2006, 2007).

There is one notable exception to the strict separation between Danubian and northern European $C$. carassius populations. The population CZE1, located in the River Lužnice catchment (Czech Republic), which drains into the River Elbe, clusters with Danubian populations in both the microsatellite and mtDNA data. This sample site, from the River Lužnice, is very close to the Danubian catchment boundary and is situated in a relatively low lying area. Therefore, some recent natural movements across the watershed between these river catchments, either through river capture events or ephemeral connections, could have been possible. A similar pattern has been shown in some European bullhead Cottus gobio populations along the catchment Danube/Rhine catchment border (Riffel \& Schreiber 1995). We also observed the presence of two mtDNA haplotypes from Lineage 2 in some individuals from northern German populations (GER1, GER2, GER8), however, one of these haplotypes was shared with Danubian individuals and the results were not confirmed by nuclear markers. Overall this is most likely to be the result of occasional human mediated long-distance dispersal for the purposes of intentional stocking.

Population structure within Lineage 1 is characterised by a pattern of IBD and a loss of allelic richness from eastern to western Europe. This is consistent with the most likely colonisation scenario identified by the DIYABC analysis, indicating a general southeast to northwest expansion from the Ponto-Caspian region towards central and northern Europe (Fig. 4). The Ponto-Caspian region, and in particular the Black Sea basin, was an important refugium for freshwater fishes during the Pleistocene glacial cycles, and a similar colonisation route has been inferred for many other freshwater species in northern Europe (Nesbø et al. 1999; Durand et al. 1999; Culling et al. 2006; Costedoat \& Gilles 2009). The DIYABC analysis also suggests that there was an interval of $>$ 200000 years between the split of the Don population ( $\approx 270000$ years ago) and the next split in
the scenario (approx. 33600 years ago), which marks the main expansion across central and northern Europe. It appears that no further population divergence can be dated back to the time interval between the Riss/Saalian and the Würm/Weichelian glacial periods. This may be because the range of $C$. carassius has not undergone a major change during that time interval, but it is more likely that the signal of expansion during the Riss-Würm interglacial has been eradicated through a subsequent range contraction during the Würm/Weichelian glacial period. The model also suggests that the Würm/Weichelian period was accompanied by a sustained but moderate reduction in population size over almost 9000 years (Bottleneck A, Fig. 4), which may reflect general population size reductions during the Riss glaciations or a series of shorter bottlenecks during subsequent range expansion (Ramachandran et al. 2005, Simon et al 2015, Hewitt 2000).

DIYABC analyses inferred the colonisation of northern Europe by two sub-lineages within the mtDNA Lineage 1, which were isolated from each other approximately 33600 years ago. These sub-lineages may reflect two glacial refugia resulting from the expansion of the Weichselian ice cap to its maximum extent roughly 22000 years ago (see hypothetical refugia II and III in Fig. 4). The western sub-lineage underwent a second long period of population decline (Bottleneck B, Fig. 4), which may again represent successive founder effects during range expansion. There is then evidence of secondary contact between these sub-lineages (node b, approximately $\approx 15940$ years ago), contributing to the genetic variation now found in Poland. This inferred admixture event may represent one of the numerous inundation and drainage capture events, which resulted from the melting of the Weichselian ice cap, that are known to have occurred around this time (Grosswald 1980; Gibbard et al. 1988; Arkhipov et al. 1995). However, as the colonisation of Europe was likely to have occurred via the expansion of colonisation fronts (i.e. dashed contour lines in Fig. 4), rather than along linear paths, it could also be indicative of the known IBD gradient between the inferred western and eastern sub-lineages. Such a gradient (eg. between northwestern and
northeastern Europe) may give false signals of admixture between intermediate populations, such as those in Poland.

The colonisation of the Baltic sea basin also seems to have been complex, with three independent routes inferred by DIYABC scenario 14d; one recent route through Denmark into southern Sweden, one to the east of the Baltic Sea, through Finland, and one across the Baltic Sea, from populations related to those in Poland (Pool 4). The first of these agrees well with the findings of Janson et al. (2014), whereby populations, including SWE8 from our study (SK3P in Janson et al. 2014), in this region were found to be distinct from those in central Sweden. The eastern route shows similarities to the colonisation patterns of $P$. fluvilatilis, which is hypothesised to have had a refugium east of Finland (Nesbø et al. 1999) during the most recent glacial period. This is certainly also plausible in C. carassius and may account for the distinctiveness of Finnish populations seen in microsatellites and RADseq DAPC analysis. The last colonisation route, across the Baltic Sea from mainland Europe, may have coincided with the freshwater Lake Ancylus stage of the Baltic Sea's evolution, which existed from $\approx 10600$ to 7500 years ago (Björck 1995; Kostecki 2014). The Lake Ancylus stage likely provided a window for the colonisation of many of the species now resident in the Baltic, and has been proposed as a possible window for the colonisation of T. thymallus (Koskinen et al. 2000), C. taenia, (Culling et al. 2006), C. gobio (Kontula \& Väinölä 2001) and four Coregonus species (Svärdson 1998). Consistent with this, we found strong similarity between populations from Fasta Åland, southern Finland and central Sweden, suggesting that shallow regions in the central part of Lake Ancylus (what is now the Åland Archipelago), may have provided one route across Lake Ancylus.

It is also likely that the contemporary distribution of C. carassius in the Baltic has been influenced by human translocations. C. carassius were often used as a food source in monasteries in many parts of Sweden (Janson et al. 2014), and the Baltic island of Gotland (Rasmussen 1959; Svanberg et al. 2013) was an important trading port of the Hanseatic League - a commercial confederation
that dominated trade in northern Europe from the $13^{\text {th }}$ to $17^{\text {th }}$ centuries. Previous data suggest that C. carassius was transported from the Scania Province, southern Sweden, where C. carassius aquaculture was common at least during the $17^{\text {th }}$ century, to parts further north (Svanberg et al. 2013; Janson et al. 2014).

## Implications for the conservation of C. carassius in Europe

The two C. carassius lineages exhibit highly-restricted gene flow between them and are the highest known organisational level within the species. They therefore meet the genetic criteria for Evolutionarily Significant Units (ESUs) as described in (Fraser \& Bernatchez 2001). This is especially important in light of the current $C$. carassius decline in the Danubian catchment (Bănărescu 1990; Navodaru et al. 2002; Lusk et al. 2010; Savini et al. 2010). The conservation of C. carassius in central Europe must therefore take these catchment boundaries into consideration, as opposed to political boundaries. A first step would be to include C. carassius in Red Lists, not only for individual countries, but at the regional (e.g. European Red List of Freshwater Fishes; (Freyhof \& Brooks 2011) and global (IUCN 2015) scales, and we hope that the evidence presented here will facilitate this process. Within the northern European lineage, the Baltic Sea basin shows high levels of population diversity, likely owing to its complex colonisation history. As such, the Baltic represents an important part of the C. carassius native range. Although C. carassius is not currently thought to be threatened in the Baltic region, C. gibelio is invading this region and is considered a threat (Urho \& Lehtonen; Deinhardt 2013).

## Microsatellites vs RADseq for phylogeography

Broad conclusions drawn from each of our RADseq-derived SNPs, full or partial microsatellite datasets are consistent, demonstrating deep divergence between northern and southern European populations and an IBD pattern of population structure in northern Europe. This similarity in spatial signal between marker types was also observed by (Bradbury et al. 2015). However, two striking differences exist in the phylogeographic results produced by RADseq compared to those of the microsatellite datasets. Firstly, the IBD pattern inferred from RADseq data was considerably stronger than for any of the microsatellite datasets. This effect was also found by Coates et al. (2009) when comparing SNPs and microsatellites, who postulated that it was driven by the differences in mutational processes of the markers. The second major difference between RADseq and microsatellite results was that clusters inferred by DAPC from the RADseq data were considerably more distinct compared to the full microsatellite dataset, emphasising the fine scale structure in the data (which is particularly apparent in the northern Finnish populations). We ruled out the possibility of these differences being caused by the reduction in number of populations, their spatial uniformity or number of individuals per population used in RADseq by creating two partial microsatellite datasets and comparing these to results from the RADseq-SNPs. Differences between marker types were consistently reproducible whether full or partial microsatellite datasets were used in the analyses.

It is also worth noting that the number of populations or the number of samples per population had no apparent impact on IBD and DAPC results between the microsatellite datasets. This is in contrast to predictions of patchy sampling of IBD made by Schwartz and McKelvey (2009), perhaps because of the strong population structure in C. carassius, and likelihood that a sufficiently informative number of populations was included even in the reduced datasets.

SNP loci provide several advantages over microsatellites additional to those highlighted here. SNPs are more densely and evenly distributed across the genome (Xing et al. 2005) and have been shown to display lower error rates during genotyping (Montgomery et al. 2005). For example, Morin et al. (2009a) showed that HW proportions are very sensitive to microsatellite genotyping errors. SNPs also lend themselves to a plethora of evolutionary applications, including the identification of outlier loci (Hohenlohe et al. 2012) or small regions of introgression in the genome (Hohenlohe et al. 2013). Lastly, SNPs are also much less susceptible to homoplasy than microsatellites (Morin et al. 2004). Van Oppen et al. (2000) found evidence of homoplasy in 10 out of 13 microsatellite loci, which had accumulated in approximately 700,000 years and Cornuet et al. (2010) show that such homoplasy makes microsatellites unreliable and error prone when used in DIYABC for inference over long time scales. For these reasons, SNPs have a clear advantage over microsatellites for the purposes of characterising population divergence over long time scales. This may explain why preliminary microsatellite analyses in DIYABC showed insufficient power to identify a most likely colonisation scenario.

## Conclusions

We have identified the most likely routes of post-glacial colonisation in C. carassius, which deviate from the general patterns observed in other European freshwater fishes. This has resulted in two, previously-unidentified major lineages in Europe, which future broad-scale monitoring and conservation strategies should take into account.

Although our RADseq sampling design included only $17.6 \%$ of samples included in the full microsatellite dataset this was sufficient to produce a robust phylogeography in agreement with the microsatellite dataset, and emphasised the fine scale structure among populations. We therefore
conclude that, if made to choose between the comprehensively sampled microsatellite approach or the RADseq approach with fewer samples but many more loci, the RADseq approach presents the better option for the phylogeography of C. carassius, with the huge number of SNP loci overcoming the limitations imposed by reduced sample number. We also predict that this will hold true for systems with similar genetic characteristics to ours, i.e. strong population structure characterised by IBD.

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## Author contributions

DLJ collected, samples, performed lab work, analysed data and wrote the manuscript. BH was involved with conception of the project, advised on all steps of analyses and commented on the manuscript. GHC was involved in the conception of the project, contributed samples and commented on the manuscript. LJLH advised on the analysis and commented on the manuscript. CDS and KHO provided samples and commented on manuscript.

## Data accessibility

Genbank accession numbers for mtDNA sequences are provided in Table 1 of this manuscript. The microsatellite data files for all DAPC analyses, mantel test matrices, mtDNA raw sequences,
sequence alignments, model testing outputs, tree files, RADseq loci catalog and VCF files used for analyses and the DIYABC project files (containing all inputs, scenarios and parameter priors for each analysis stage) have now been uploaded to Dryad (http://dx.doi.org/10.5061/dryad.t2j45). All scripts used for DAPC, mantel tests and comparisons of RADseq and microsatellite datasets can be found on GitHub (https://github.com/DanJeffries/Jeffries-et-al-2016-crucian-phylogeography). All demultiplexed RADseq reads have been uploaded to the short read archive (Project accession: SRP063043).

Figure 1. Population structure of C. carassius in Europe. a) Sampling locations (sites sampled with nuclear and mtDNA markers $=$ red dots, mtDNA only $=$ blue dots) and population cluster memberships from DAPC analysis. Pie chart size corresponds to microsatellite allelic richness. Pie chart colours for Danubian populations and RUS1 correspond to clusters in the broad scale DAPC analysis b) and for all northern European populations colours correspond to clusters in the northern European DAPC analysis (mtDNA lineage 1 only) c). The Danube river catchment is shaded dark grey.

Figure 2. Maximum credibility tree calculated in BEAST for 100 C. carassius cytb sequences. For the three maximally supported nodes, age is given above and the posterior probability distribution is given below, with $95 \%$ CI's represented by blue bars.

Figure 3. Comparison of DAPC results using a) RADseq dataset, b) M2 dataset and c) M3 dataset. Colours correspond between DAPC scatter plots and maps within but not between panels.

Figure 4. The postglacial recolonisation of C. carassius in Europe. Arrows represent the relationships between population pools used in DIYABC (grey circles) as inferred from Stage 1, scenario 9 (arrows outlined in black) and Stage 3, scenario 14d (arrows with no outline) analyses on RADseq data. Bottlenecks are represented by white-striped sections of arrows. Posterior time estimates in years for each demographic event are given in black, and estimates of Ne are given in blue. Blue diamonds represent ancestral populations inferred by DIYABC and the labels (a-f) correspond to their mention in the text. Hypothetical expansion fronts are represented by dashed contour lines and the Danube river catchment is shaded red. Hypothetical glacial refugia are represented by dashed blue circles (I-III). The blue dashed box (?) represents our inference that $C$. carassius expanded into central and perhaps northern Europe during the Riss-Würm interglacial, however we cannot estimate this range.

SI Figure 1. DIYABC scenarios used in broad-scale analysis (Stage 1). See text for population poolings. See Table 3 for population poolings and prior parameter values.

SI Figure 2. All scenarios tested in stage 2 a) and stage 3 b) of DIYABC analysis. See Table 3 for population poolings and prior parameter values.

SI Figure 3. Filtering out merged ohnologs. a) Distribution of SNP locus coverage prior to removing loci that had observed heterozygosity higher than 0.5 in one or more population. b) Distribution of locus coverage after filtering, showing a loss of many high coverage loci and a reduction in mean SNP coverage. Note the loss of loci with high coverage.

SI Figure 4. Linear regressions for all samples a) $A \mathrm{r}$ against latitude; b) $A \mathrm{r}$ against longitude and for only samples in mtDNA lineage 1 c) Ar against latitude; d) Ar against longitude.

SI Figure 5. DAPC analysis of a) full microsatellite dataset (Excluding NOR2); for results used in Fig. 1) and b) Full RADseq dataset.

SI Figure 6. Isolation by distance a) in M1 dataset for mtDNA lineage 1 only (excluding NOR2), b) full RADseq dataset, c) M2 dataset and d) M3 dataset.

SI Figure 7. DAPC scatter plot for the 1000 SNP RADseq dataset used in the DIYABC analysis, showing the same population structure as inferred from the full RADseq dataset.

SI Figure 8. Broad scale DIYABC analysis (Stage 1) results. a) Direct approach (left) and Logistic regression (right) showing support for scenario 9.b) Model checking for scenario 9, showing that the observed data fall well within the cloud of datasets simulated from the posterior parameter distribution. c) Scenario 9 schematic.

SI Figure 9. Fine scale DIYABC analysis in northern Europe. a) Stage 2 - major topological variants of scenarios. Direct approach (top left) and Logistic regression (top right) showing support for scenario 14 and 13 respectively. Model checking (Middle) for scenario 14 (bottom), showing that the observed data fall well within the cloud of datasets simulated from the posterior parameter distribution. Note the model checking placed the observed data outside of the cloud of posterior datasets for scenario 13. b) Stage 3 - Minor scenario variants of scenario 14 from stage 2. Direct
approach (top left), logistic regression (top right) and model checking (middle) all support scenario 14d (bottom).

SI Figure 10. Comparison of spatial patterns of uniformity in geographic sampling regimes of the full M1 dataset locations ( $\mathrm{a}, \mathrm{c} \mathrm{)} \mathrm{and} \mathrm{the} \mathrm{sampling} \mathrm{location} \mathrm{subset} \mathrm{used} \mathrm{in} \mathrm{M2}, \mathrm{M3}$, $(\mathrm{b}, \mathrm{d})$. Estimates of G and L from true sampling locations are plotted using the black solid lines. Estimates of G and L from simulated locations based on random Poisson distribution is represented by the red dashed line. Grey shaded areas are the $95 \%$ confidence intervals around the random estimates. Both the G and L function estimates show that there is more clustering of sampling locations in the M1 dataset than in the M2, M3 and RAD subsets.

SI Figure 11. Change in a) number of RAD tags and b) average tag coverage for three individuals used in the preliminary Stacks tag mismatch parameter (M) tests.

SI Figure 12. Results of parameter tests for the Stacks module Populations. a) Number of SNP loci in final dataset for incrementing values of parameters $-\mathrm{p},-\mathrm{r}$ and -m ; b ) average coverage per SNP and per sample for the same parameter values; c) the number of loci which drop out in each population for each test value of the -p parameter

Table 1. Location, number, genetic marker sampled, and accession numbers of samples and sequences used in the present study for microsatellite and mitochondrial DNA analyses. mtDNA sequence accession numbers can be found in SI table 2.

| Code | Accession | Location | Country | Drainage | Coordinates |  | Microsatellites | mtDNA | RADseq |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | lat | long |  |  |  |
| GBR1 |  | London | U.K. | U.K | 51.5 | 0.13 | 9 |  |  |
| GBR2 |  | Reading | U.K. | U.K | 51.45 | -0.97 | 4 |  |  |
| GBR3 |  | Norfolk | U.K. | U.K | 52.86 | 1.16 | 7 |  |  |
| GBR4 |  | Norfolk | U.K. | U.K | 52.77 | 0.75 | 27 |  | 9 |
| GBR5 |  | Norfolk | U.K. | U.K | 52.77 | 0.76 | 14 |  |  |
| GBR6 |  | Norfolk | U.K. | U.K | 52.54 | 0.93 | 29 | 3 |  |
| GBR7 |  | Norfolk | U.K. | U.K | 52.9 | 1.15 | 24 | 1 | 10 |
| GBR8 |  | Hertfordshire | U.K. | U.K | 52.89 | 1.1 | 37 | 3 | 9 |
| GBR9 |  | Norfolk | U.K. | U.K | 52.8 | 1.1 | 27 |  |  |
| GBR10 |  | Norfolk | U.K. | U.K | 52.89 | 1.1 | 14 |  |  |
| GBR11 |  | Norfolk | U.K. | U.K | 52.92 | 1.16 | 20 |  |  |
| BEL1 |  | Bokrijk | Belgium | Scheldt River | 50.95 | 5.41 | 13 | 1 |  |
| BEL2 |  | Meer van Weerde | Belgium | Scheldt River | 50.97 | 4.48 | 12 |  |  |
| BEL3 |  | Meer van Weerde | Belgium | Scheldt River | 50.97 | 4.48 | 8 |  |  |
| GER1* |  | Kruegersee | Germany | Elbe River | 52.03 | 11.97 |  | 3 |  |
| GER2 |  | Münster | Germany | Rhine River | 51.89 | 7.56 | 21 | 3 |  |
| GER3 |  | Bergheim | Germany | Danube River | 48.73 | 11.03 | 9 | 3 |  |
| GER4 |  | Bergheim | Germany | Danube River | 48.73 | 11.03 | 8 | 3 |  |
| CZE1 |  | Lužnice | Czech Republic | Danube River | 48.88 | 14.89 | 9 | 3 |  |
| POL1 |  | Sarnowo | Poland | Vistula River | 52.93 | 19.36 | 33 |  |  |
| POL2 |  | Kikót-Wies | Poland | Vistula River | 52.9 | 19.12 | 34 |  |  |
| POL3 |  | Tupadly | Poland | Vistula River | 52.74 | 19.3 | 17 | 3 | 10 |
| POL4 |  | Orzysz | Poland | Vistula River | 53.83 | 22.02 | 13 | 3 | 10 |
| EST1 |  | Tartu | Estonia | Baltic Sea | 58.39 | 26.72 | 5 | 3 |  |
| EST2 |  | Vehendi | Estonia | Baltic Sea | 58.39 | 26.72 | 5 |  |  |
| RUS4* |  | Small lake, Velikaya river | Russia | Baltic Sea | 55.9 | 30.25 | 29 | 3 |  |
| FIN1 |  | Joensuu | Finland | Baltic Sea | 62.68 | 29.68 | 32 | 3 |  |
| FIN2 |  | Helsinki | Finland | Baltic Sea | 60.36 | 25.33 | 32 |  |  |
| FIN3 |  | Jyväskylä | Finland | Baltic Sea | 62.26 | 25.76 | 37 | 3 | 10 |
| FIN4 |  | Oulu | Finland | Baltic Sea | 65.01 | 25.47 | 7 | 3 | 8 |
| FIN5 |  | Salo | Finalnd | Baltic Sea | 60.37 | 23.1 | 10 | 3 |  |
| FIN6 |  | Åland Island | Sweden | Baltic Sea | 60.36 | 19.85 | 8 | 3 |  |
| SWE1 |  | Grānbrydammen | Sweden | Baltic Sea | 59.87 | 17.67 | 25 |  |  |
| SWE2 |  | Stordammen | Sweden | Baltic Sea | 59.8 | 17.71 | 21 | 3 | 10 |
| SWE3 |  | Östhammar | Sweden | Baltic Sea | 60.26 | 18.38 | 27 | 3 |  |
| SWE4 |  | Umeå | Sweden | Baltic Sea | 63.71 | 20.41 | 9 | 3 |  |
| SWE5 |  | Kvicksund | Sweden | Baltic Sea | 59.45 | 16.32 | 9 |  |  |
| SWE7 |  | Grillby | Sweden | Baltic Sea | 59.64 | 17.37 | 10 |  |  |
| SWE8 |  | Skabersjo | Sweden | Baltic Sea | 55.55 | 13.15 | 19 | 3 | 10 |
| SWE9 |  | Märsta | Sweden | Baltic Sea | 59.6 | 17.8 | 31 | 3 |  |
| SWE10 |  | Norrköping | Sweden | Baltic Sea | 58.56 | 16.27 | 29 |  | 9 |
| SWE11 |  | Gotland Island | Sweden | Baltic Sea | 57.85 | 18.79 | 11 | 3 |  |
| NOR1 |  | Oslo | Norway | North Sea | 60.05 | 9.94 |  | 2 |  |
| NOR2 |  | Lake Prestvattnet, Tromsø | Norway | North Sea | 69.65 | 18.95 | 16 |  | 9 |
| BLS |  |  | Belarus | Dnieper | 52.47 | 30.52 | 7 | 1 |  |
| RUS1 |  | Proran Lake | Russia | Don River | 47.46 | 40.47 | 10 | 3 | 9 |
| DEN1 |  | Copenhagan | Denmark | Baltic Sea | 60.21 | 17.79 | 12 |  | 10 |
| DEN2 |  | Pederstrup | Denmark | Baltic Sea | 55.77 | 12.55 | 14 |  | 8 |
| DEN3 |  | Gammel Holte | Denmark | Baltic Sea | 56 | 12.5 | 14 |  |  |
| DEN4 |  | Bornholm Island | Denmark | Baltic Sea | 55.17 | 14.86 |  |  | 5 |
| SWE12 |  | Osterbybruk Mansion | Sweden | Baltic Sea | 55.73 | 12.34 | 14 |  | 9 |
| SWE14 |  | Wenngarn Castle | Sweden | Baltic Sea | 59.66 | 18.95 | 16 |  | 9 |
| RUS2* |  | Karma | Russia | Volga River | 52.9 | 58.4 |  | 2 |  |
| RUS3* |  | Saygach'yedake | Russia | Volga River | 47.5 | 48.5 |  | 4 |  |
| TNO |  |  | Netherlands | North Sea | - | - |  | 1 |  |
| HUN1 |  | Gödöllő | Hungary | Danube River | 47.61 | 19.36 |  | 2 | 6 |
| HUN2 |  | Vörösmocsár | Hungary | Danube River | 46.49 | 19.17 |  |  |  |
|  |  |  |  |  |  |  | 848 | 83 | 160 |
| Genbank mtDNA Sequences |  |  |  |  |  |  | Total number of fish $=867$ |  |  |
| Code | Accession | Reference | Country | Drainage |  |  |  |  |  |
| GER6 | DQ399917 | Kalous et al. (2007) | Germany | Baltic sea |  |  |  |  |  |
| GER6 | DQ399918 | Kalous et al. (2007) | Germany | Baltic sea |  |  |  |  |  |
| GER6 | DQ399919 | Kalous et al. (2007) | Germany | Baltic sea |  |  |  |  |  |
| GER7 | JN412540 | Rylková et al. (2013) | Germany | Hunte River |  |  |  |  |  |
| GER7 | JN412541 | Rylková et al. (2013) | Germany | Hunte River |  |  |  |  |  |
| GER7 | JN412542 | Rylková et al. (2013) | Germany | Hunte River |  |  |  |  |  |
| GER7 | JN412543 | Rylková et al. (2013) | Germany | Hunte River |  |  |  |  |  |
| GER8* | JN412537 | Rylková et al. (2013) | Germany | Lahn River |  |  |  |  |  |
| GER8* | JN412538 | Rylková et al. (2013) | Germany | Lahn River |  |  |  |  |  |
| CZE2 | GU991399 | Rylková et al. (2013) | Czech Republic | Elbe drainage |  |  |  |  |  |
| Milevsko | DQ399938 | Kalous et al. (2012) | Czech Republic | Elbe drainage |  |  |  |  |  |
| AUS1 | JN412534 | Rylková et al. (2013) | Austria | Danube river |  |  |  |  |  |
| AUS1 | JN412533 | Rylková et al. (2013) | Austria | Danube river |  |  |  |  |  |
| AUS2 | JN412535 | Rylková et al. (2013) | Austria | Danube river |  |  |  |  |  |
| AUS3 | JN412536 | Rylková et al. (2013) | Austria | Danube river |  |  |  |  |  |
| GBR12 | JN412539 | Rylková et al. (2013) | U.K. | U.K |  |  |  |  |  |


| GBR12 | GU991400 | Kalous et al. (2012) | U.K. | U.K |
| :--- | :--- | :--- | :--- | :--- |
| SWE15 | JN412545 | Rylková et al. (2013) | Sweden | Baltic sea |
| SWE16 | JN412544 | Rylková et al. (2013) | Sweden | Baltic sea |
| Ccarp1 | AB158807 | Mabuchi et al (2005) | Japan | - |
| Ccarp2 | DQ868875 | Tsipas et al. (2009) | Greece | - |
| Ccarp3 | KF574490 | Unpublished | India | - |

† Also present

* Location on Map (Fig. 1.a) is approximate

| Analysis stage | Population Pools | Scenarios tested | Parameter priors | Most likely Scenario | Median of posterior distributions of most likely scenario |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Pool 1-GBR4, GBR7, GBR8, DEN1, DEN2, DEN3, FIN3, FIN4, POL3, POL4, SWE2, SWE8, SWE9, SWE10, SWE12, SWE14, NOR2 Pool 2 - DEN1, DEN2, DEN3 Pool 3 - FIN3, FIN4 | $1-11$ | $\mathrm{N} 1=10 \mathrm{E}+03-500 \mathrm{E}+03$ | 9 | N1 $=34700$ |
|  |  |  | $\mathrm{Nb} 1=10-100 \mathrm{E}+03$ |  | $\mathrm{Nb} 1=23700$ |
|  |  |  | $\mathrm{N} 2=100-100 \mathrm{E}+03$ |  | $\mathrm{N} 2=74900$ |
|  |  |  | $\mathrm{N} 3=100-200 \mathrm{E}+03$ |  | N3 $=140000$ |
|  |  |  | $\mathrm{tl}=1 \mathrm{E}+03-1 \mathrm{E}+06$ gens |  | $\mathrm{tl}=135000$ |
|  |  |  | $\mathrm{t} 2=1 \mathrm{E}+03-3 \mathrm{E}+06$ gens |  | $\mathrm{db}=4460$ |
|  |  |  | $\mathrm{ra}=0.001-0.999$ |  |  |
|  |  |  | $\mathrm{rb}=0.001-0.999$ |  | $\mathrm{t} 2=1090000$ |
|  |  |  | rc $=0.001-0.999$ |  |  |
|  |  |  | $\mathrm{db}=10-10 \mathrm{E}+03$ gens |  |  |
| $2$ | Pool 1 - GBR4, <br> GBR7, GBR8 <br> Pool 2 - DEN1, <br> DEN2, DEN3 <br> Pool 3 - FIN3, FIN4 | 12-16 | $\mathrm{N} 1=10-4 \mathrm{E}+03$ | 14 | N1 $=3670$ |
|  |  |  | $\mathrm{N} 2=10-10 \mathrm{E}+03$ |  | $\mathrm{N} 2=7520$ |
|  |  |  | $\mathrm{N} 3=10-20 \mathrm{E}+03$ |  | N3 $=17400$ |
|  |  |  | $\mathrm{N} 4=10-50 \mathrm{E}+03$ |  | N4 $=19400$ |
|  |  |  | $\mathrm{N} 5=10-20 \mathrm{E}+03$ |  | N5 $=11800$ |
|  |  |  | N6 = 10-400 |  | N6 $=210$ |
|  |  |  | $\mathrm{tl}=100-10 \mathrm{E}+03$ gens |  | $\mathrm{t} 1=6790$ |
|  |  |  | $\mathrm{tla}=100-10 \mathrm{E}+03$ gens |  | $\mathrm{t} 1 \mathrm{a}=2510$ |
|  |  |  | t2 $=100-10 \mathrm{E}+03$ |  |  |
|  |  |  | $\mathrm{t} 2 \mathrm{a}=100-5 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{t} 2 \mathrm{~b}=500-20 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{t} 2 \mathrm{c}=100-10 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{t} 2 \mathrm{~d}=100-10 \mathrm{E}+03$ gens |  | $\mathrm{t} 2 \mathrm{~d}=6780$ |
|  |  |  | $\mathrm{t} 3=500-20 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{t} 3 \mathrm{c}=100-10 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{t} 3 \mathrm{~d}=100-10 \mathrm{E}+03$ gens |  | t3d $=8910$ |
|  |  |  | $\mathrm{t} 4=500-20 \mathrm{E}+03$ gens |  | t4 $=12000$ |
|  |  |  | $\mathrm{ra}=0.001-0.999$ |  |  |
|  |  |  | $\mathrm{rb}=0.001-0.999$ |  | $\mathrm{rb}=0.668$ |
| $3$ | Pool 4 - POL3, POL4 <br> Pool 5 -SWE2, <br> SWE8, SWE9, SWE10, <br> SWE12, SWE14 <br> Pool 6 - NOR2 | $14 a-14 f$ | $\mathrm{N} 1=10-4 \mathrm{E}+03$ | $14 d$ | N1 $=2390$ |
|  |  |  | $\mathrm{Nb} 1=10-10 \mathrm{E}+03$ |  | $\mathrm{Nb} 1=935$ |
|  |  |  | $\mathrm{N} 2=10-10 \mathrm{E}+03$ |  | N2 $=8140$ |
|  |  |  | $\mathrm{N} 3=10-20 \mathrm{E}+03$ |  | N3 $=9360$ |
|  |  |  | $\mathrm{Nb} 3=10-10 \mathrm{E}+03$ |  |  |
|  |  |  | N4 $=10-50 \mathrm{E}+03$ |  | N4 $=17000$ |
|  |  |  | N5 = 10-20E+03 |  | N5 $=11000$ |
|  |  |  | N6 = 10-400 |  | N6 $=138$ |
|  |  |  | $\mathrm{Nb} 6=10-10 \mathrm{E}+03$ |  |  |
|  |  |  | $\mathrm{tl}=100-10 \mathrm{E}+03$ gens |  | t1 $=3750$ |
|  |  |  | $\mathrm{tla}=100-10 \mathrm{E}+03$ gens |  | $\mathrm{t} 1 \mathrm{a}=2460$ |
|  |  |  | $\mathrm{t} 2 \mathrm{~d}=100-10 \mathrm{E}+03$ gens |  | $\mathrm{t} 2 \mathrm{~d}=5900$ |
|  |  |  | $\mathrm{t} 3 \mathrm{~d}=100-10 \mathrm{E}+03$ gens |  | $\mathrm{t} 3 \mathrm{~d}=7970$ |
|  |  |  | $\mathrm{t} 4=500-20 \mathrm{E}+03$ gens |  | t4 $=16800$ |
|  |  |  | $\mathrm{rb}=0.001-0.999$ |  | $\mathrm{rb}=0.619$ |
|  |  |  | $\mathrm{da}=10-10 \mathrm{E}+03 \mathrm{gens}$ |  |  |
|  |  |  | $\mathrm{db}=10-10 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{dc}=10-10 \mathrm{E}+03 \mathrm{gens}$ |  | $\mathrm{dc}=9070$ |
|  |  |  | $\mathrm{dd}=10-10 \mathrm{E}+03$ gens |  |  |
|  |  |  | $\mathrm{de}=10-10 \mathrm{E}+03$ gens |  |  |

Table 2. Population pools, parameter priors used and median posterior parameter values inferred in the three stages of DIYABC analysis.

Table 3. Summary statistics for M1, M2, M3 and RADseq datasets. RAD contains all RADseq data, M1 contains all microsatellite data, M2 contains only microsatellite for the individuals used in the RADseq, and M3 contains all microsatellite data for all individuals that were available in populations that were used in RADseq.

| Dataset | Description | N samples | Mean N samples/pop | N. loci | Mean N.alleles/pop | Mean N.alleles/locus |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| RAD | RADseq data only | 149 | $8.95 \pm 1.4$ | 13189 | 6723 | 2 |
| M1 | Full Microsatellite dataset | 848 | $17.2 \pm 9.5$ | 13 | $27 \pm 8.8$ | 7.6 |
| M2 | Microsatellites for RADseq samples only | 146 | $9.13 \pm 0.8$ | 13 | $24.4 \pm 7.3$ | $7.84 \pm 5.1$ |
| M3 | Microsatellites for all samples in populations used in RADseq | 313 | $19.6 \pm 9.0$ | 13 | $27.4 \pm 8.1$ | $11.23 \pm 7.6$ |

Table 4. Pearson's product-moment correlation coefficients and paired t-tests comparing heterozygosities and $F_{\mathrm{ST}}$ S between M2, M3 and RADseq datasets. ${ }^{* * *} P=<0.001$, ${ }^{* *} P=<0.005$, ${ }^{*} P=<0.05$.

| Heterozygosities ( $\mathrm{df}=18$ ) | Pearsons correlation coefficient (t) |  |  |
| :---: | :---: | :---: | :---: |
|  | M2 | 11.13*** | 3.85** |
| Paired T-tests | -2.4* | M3 | 3.86** |
|  | $-9.71^{* * *}$ | -9.29*** | RAD |
| $F_{\text {ST }}(\mathrm{df}=105)$ | Pearsons correlation coefficient (t) |  |  |
|  | M2 | 46.26*** | 10.09*** |
| Paired T-tests | -6.21*** | M3 | 9.05*** |
|  | 13.74*** | 15.12*** | RAD |




B) M2-13 Microsatellites, $\mathrm{n}=146$.



# Comparing RADseq and microsatellites to infer complex phylogeographic patterns, a real data informed perspective in the Crucian carp, Carassius carassius, L. 

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## Supporting Information

## Detecting hybrids

## Methods

In total we acquired tissue samples of 1354 fish from 72 populations. All samples were first genotyped using multiplex 1 (SI table 1) which contained the 6 species diagnostic microsatellite loci. These data were then analysed using the NewHybrids v. 1.1 (Anderson \& Thompson 2002) software package in order to determine whether each fish was C. carassius, C. auratus, C. gibelio or a hybrid between any of these species.

NewHybrids uses allele frequencies to give a likelihood probability that an individual belongs to one species or another, or if the individual belonged to one of several hybrid classes (F1, F2 or backcross). Data from 20 C. carassius samples, which were confidently identified as pure from both morphology and genotypes, and were not sympatric with non-native species, were included in each analysis as baseline data. Priors were then added to the analyses specifying that these individuals were indeed pure in order to give the software more power with which to assess allele frequencies associated with C. carassius. To be sure to account for allele frequency differences between different geographic regions, only pure individuals from regions neighbouring the hybrid population
were used. Individuals which had more than a $25 \%$ chance of being an F1 hybrid, F2 hybrid, or a backcross were removed from population structure analyses and were not genotyped at the additional 7 microsatellite loci (Multiplexes 2.1 and 2.2, SI table 1).

## Results

Of the 1354 fish which were genotyped with microsatellites, 942 individuals across 55 populations (86.7\%) were identified as pure C. carassius using the first set of 6 species diagnostic loci in NewHybrids analyses. 19 (1.8\%) from 2 different populations were identified as C. auratus, 15 fish (1.4\%) from 4 populations were identified as. C. gibelio and 10 fish ( $0.93 \%$ ) from two populations were identified as C. carpio. NewHybrids identified 60 (5.5\%) C. carassius x C. auratus hybrids, $25(2.2 \%)$ C. carassius x C. gibelio hybrids, and $16(1.5 \%)$ C. carassius x C. carpio hybrids. Of the 942 fish identified as pure C. carassius, 867 in existed in locations (49 populations) where hybrids or non-native species were not detected by microsatellite genotyping. To safeguard against cryptic introgression which may produce erroneous results only these 867 pure C. carassius were used for the main phylogeographic analyses and tests using either microsatellites, mtDNA or RADseq.

## RADseq data filtering and Stacks analysis parameter testing

RADseq analyses were performed using only the first-end reads from the paired-end sequencing, as coverage across the length of the second-end contigs was not consistent enough to call SNPs in all individuals. For these first-end reads, raw data was first quality checked using FastQC (Andrews 2010), which assesses the per-base sequence quality and content of reads, and provides comprehensive graphical outputs with which to assess the overall quality of raw sequencing data. These analyses did not identify any individuals that had low overall sequence quality, therefore all samples were retained for further analyses.

Preliminary analyses were also carried out using PyRAD (Eaton 2014), which allows for the incorporation of allelic variants resulting from insertions and deletions. However, no significant difference in the number of usable loci was shown. As Stacks provides more downstream populations genetics facilities, this program was used for the final analyses.

Raw RADseq reads were first, demultiplexed using the "process_radtags" module distributed with Stacks and our inline barcodes. Second, reads were filtered for any sequences containing Illumina adapters or primers and trimmed to a length of 92 bp . Third, PCR duplicates introduced during library preparation were removed using the "clone_filter" program (also distributed with Stacks). Finally, preliminary tests of parameter values for each module of the de novo stacks pipeline were performed in order to identify "optimal" parameter values (i.e. where loci number and read depth were stable) for use in the final Stacks analysis. These tests were carried out for 5 sets of 3 randomly chosen individuals from the RADseq dataset and, for each test, all non-test parameters were kept as default. In the ustacks module, which groups identical reads into stacks and then stacks into loci, Parameters M and m were tested (See Catchen et al. 2013 for detailed description of parameters). M values were increased in increments of 2 from 0 to 10 . The efficiency of ustacks in finding real loci was then examined with simple counts of the number of constructed loci at each M parameter value and the read coverage of these loci. The expectation was that, at low parameter values, divergent alleles (percentage divergence $>\mathrm{M}$ ) at a locus will not merge (under-merging), thus increasing the number of loci overall and decreasing the average coverage. In contrast high parameter values could cause over-merging of paralogous loci and have the opposite effects on the number of loci and coverage (Catchen et al. 2013). SI Fig. 11 shows the outputs for a single subset of C. carassius samples, which was typical of all 5 subsets tried. In ustacks, an 'm' parameter value of zero (minimum of 0 reads required to form a stack) resulted in a very large number of tags (49000-54000) as expected. Likely due to many single reads containing sequencing error being called as loci. The number of loci decreased by approximately $3000-4000$ tags in the samples
tested at a required read depth of 2 (approx. 50,000 ), after which further increases in ' m ' resulted in small decreases in the number of tags. This likely reflects merging of paralogous loci, or low coverage loci. Mean coverage across all loci within an individual of course reflected the ' m ' parameter increase, jumping initially from approx. 16 reads per locus with zero read depth required, to 20-35 at a minimum required depth of two reads. On the basis of these results we chose an $m=8$, to ensure high power for SNP calling.

Incrementing over values of ' $M$ ' again met our expectations, with the number of loci dropping significantly as the ' M ' parameter was increased from zero to 2 mismatches allowed, and then dropping more slowly with higher mismatch allowance. These further drops may again be allowing for paralog merging between loci. The mean coverage of loci behaved as expected, with higher mismatch allowance, more divergent reads can be added to existing stacks, inflating coverage for those loci. On the basis of these results $\mathrm{M}=2$ was chosen for final analyses.

Parameter tests were also performed for the cstacks parameter N, which is responsible for setting the maximum mismatch threshold allowed between homologous loci among individuals in the locus catalog. First, ustacks was run using chosen "optimal" parameters to obtain the inputs necessary for cstacks. Cstacks was then run separately on each of the 5 sample subsets with values of N between $0-10$, with increments of 2 .

Finally, we tested three core parameters in the Populations module of Stacks, -m which is analogous to the parameter of the same name in the ustacks module, -r, which specifies the number of individuals within a give population that a locus must be present in, and -p which specifies the number of populations that a locus must be present in (above the -r threshold) for it to be retained in the final dataset (SI Fig. 12). -p was tested for values of between $13-19$ populations, -r was tested for values between $0.5-1.0$ and -m was tested for values between 1-8 however, a the dataset had previously been filtered at previous stages for loci present with a depth of 8 reads or higher, the
tests of -m in the populations stage were redundant.

## Final running parameters used

For all parameter tests, the optimal values were taken to be those where the rate of change in either RAD tag number, or coverage began to decrease. In ustacks, a maximum of two mismatches were allowed between alleles at a given locus $(M=2)$ and at least eight identical reads per stack $(\mathrm{m}=8)$ were required. Default values were used for all other parameters. ustacks also called SNPs within individuals at each locus. The cstacks module was then used to merge loci across individuals into a catalog, where $\mathrm{N}=2$ mismatches were allowed between individuals at a given locus. Individuals were then searched against this catalog using Sstacks to determine their genotype at each catalog locus. For the Populations module, optimal values were chosen so that loci that were shared between at least $70 \%$ of individuals in each population $(-r=0.7)$, allowing loci to drop out in one or two individuals in a population for reasons of low DNA sample quality or low coverage. Loci must have also been present in 17 of the 19 populations $(-\mathrm{p}=17)$, and have read depth of at least $8(-\mathrm{m} 8)$ in each individual.

## DAPC \& Running parameters

## Methods

Population structure was examined using Discriminant Analyses of Principal Components (DAPC, (Jombart et al. 2010)) in adegenet. Similar to the more commonly used program, STRUCTURE (Pritchard et al. 2000), DAPC is an individual-based approach that uses Principal Components Analysis (PCA) to transform population genetic data and Discriminant Analysis (DA) to identify clusters. The number of clusters is assessed using the K-means method, which is also used in STRUCTURE (Pritchard et al. 2000). Unlike STRUCTURE, DAPC does not assume underlying population genetics models such as Hardy-Weinberg Equilibrium (Jombart et al. 2010) and is
therefore more suitable for analysing C. carassius since populations are often bottlenecked (Hänfling et al. 2005). An additional benefit of DAPC is that it maximizes between-group variation, while minimizing variation within groups, allowing for optimal discrimination of betweenpopulation structure (Jombart et al. 2010).

## Results

For the full microsatellite dataset (M1), BIC scores indicated that between 11 and 19 genetic clusters (Error! Reference source not found.) would be an appropriate model of the variation in the data. We therefore chose 11 clusters to use in the discriminant analysis, retaining 8 principal components as recommended by the spline interpolation a-scores (Error! Reference source not found.c) and we kept 2 linear discriminants for plotting (Error! Reference source not found.b).

Three major lineages were found, one located in the Danube, one in the Don, and one spread across northern Europe. However the large amount of divergence between them masked the population structure present in northern Europe. We therefore subsetted the data, separating NEU populations from RUS1, GER3, GER4, CZE1 (and SWE9, which was an outlier within NEU, Error! Reference source not found.b) and reanalysed them with DAPC in order to better infer fine population structure between them.

For the RADseq dataset, BIC scores suggested between 9 and 14 genetic clusters, similar to the range inferred in the microsatellite data, we therefore chose 9 clusters to take forward in the analysis. As recommended by spline interpolation, we retained 7 principal components and we kept 2 of the linear discriminants from the subsequent discriminant analysis

## Methods

In order to assess the geographic uniformity of the sampling regimes in each data subset, we used two measures of spatial patterns. The nearest neighbour distance distribution function (G), measures the distance of each sampling location to its nearest neighbour (Ripley 1991). The L-function is a transformation (for ease of interpretation) of Ripley's K-function (Ripley 1991), which measures the number of sampling locations within a given radius from each point. K has the advantage of assessing the uniformity of the sampling regime over multiple scales, as opposed to only measuring distances between closest neighbours as with G . In both cases, the estimates of G or K from our sampling locations were compared against random Poisson distributions, which would represent uniformly spaced sampling locations. 5\% and $95 \%$ confidence thresholds for these Poisson distributions were also calculated to allow us to determine whether our sampling regimes significantly deviated from random ( $\mathrm{p}<0.05$ ). These calculations were performed using the Gest and Lest functions (for G and L respectively) in the package "spatstats" in R (Baddeley \& Turner 2005).

Results
Both methods used for the assessment of geographic uniformity of sampling locations shows that the M1 dataset locations are more patchily distributed than those of the M2, M3 and RAD datasets (Error! Reference source not found.).

| Locus | Multiplex \# | Primer mix <br> Ratios* | \# Alleles | Allele size range | Ho | GenBank <br> Accession <br> no. | Reference |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GF1 | 1 | 0.1 | 1 | 299 | 0 | U35614 | Zheng et al. 1995 |
| GF17 | 1 | 0.1 | 2 | $182-186$ | 0.024 | U35616 | Zheng et al. 1995 |
| GF29 | 1 | 0.2 | 8 | $191-226$ | 0.348 | U35618 | Zheng et al. 1995 |
| J7 | 1 | 0.07 | 10 | $202-228$ | 0.109 | AY115095 | Yue \& Orban 2002 |
| MFW2 | 1 | 0.1 | 1 | 161 | 0 | - | Croojimans et al. 1997 |
| Ca07 | 1 | 0.2 | 9 | $122-140$ | 0.286 | D85428 | Yue \& Orban 2004 |
| TE Buffer | 1 | 0.23 |  |  |  |  |  |
| J69 | 2.1 | 0.4 | 14 | $213-241$ | 0.404 | AY115106 | Yue \& Orban 2002 |
| HJLY17 | 2.1 | 0.1 | 9 | $152-168$ | 0.223 | DQ378986 | Zhi-Ying et al. 2006 |
| HJLY35 | 2.1 | 0.1 | 18 | $261-307$ | 0.377 | DQ403242 | Zhi-Ying et al. 2006 |
| TE Buffer | 2.1 | 0.4 |  |  |  |  |  |
| J20 | 2.2 | 0.2 | 9 | $171-218$ | 0.149 | AY115099 | Yue \& Orban 2002 |
| J58 | 2.2 | 0.1 | 14 | $119-147$ | 0.398 | - | Yue \& Orban 2002 |
| MFW7 | 2.2 | 0.35 | 25 | $160-206$ | 0.464 | - | Croojimans et al. 1997 |
| MFW17 | 2.2 | 0.35 | 26 | $185-262$ | 0.41 | - | Croojimans et al. 1997 |

## Additional discussion

## Population structure in northwest Europe

An intriguing result lies in the genetic similarity between populations in England with those in Belgium and Germany. C. carassius has been designated as native to England, however this status has been contentious in the past (Maitland 1972). Under the assumption that it is native, and considering the observed diversity and divergence times between populations across mainland Europe, we would expect to see stronger population structure between English and continental Europe, which have been separated for approximately 7800 years (Coles 2000). Given the observed diversity between populations across mainland Europe, which, according to DIYABC analysis, has arisen relatively recently. Clearly further examination of this issue is warranted and molecular data would be a value addition to the current evidence, which is predominantly anecdotal.

SI table 1. Microsatellite loci used, grouped by their combinations in multiplex reactions. Multiplex primer mix ratios for PCR were chosen so as to give even peak strengths when analysing PCR products. Allele size ranges are those present in C . carassius for all 43 putatively pure crucian populations.

SI table 2. Genbank accession numbers for the mtDNA sequences used in this study.

|  | 190 |
| :---: | :---: |
| Sample code | Accession number |
| FIN5_01 | KT630314 |
| FIN5_02 | KT630315 |
| FIN5_03 | KT630316 |
| EST1_02 | KT630317 |
| GER1_01 | KT630318 |
| EST1_01 | KT630319 |
| GER1_03 | KT630320 |
| FIN6_01 | KT630321 |
| FIN6_02 | KT630322 |
| FIN6_03 | KT630323 |
| BEL1_03 | KT630324 |
| EST1_03 | KT630325 |
| GER2_02 | KT630326 |
| GER2_03 | KT630327 |
| GER4_02 | KT630328 |
| NOR1_01 | KT630329 |
| NOR1_02 | KT630330 |
| SWE11_01 | KT630331 |
| SWE11_02 | KT630332 |
| SWE11_03 | KT630333 |
| RUS2_02 | KT630334 |
| RUS4_01 | KT630335 |
| RUS4_03 | KT630336 |
| FIN1_01 | KT630337 |
| FIN1_02 | KT630338 |
| FIN1_03 | KT630339 |
| FIN4_01 | KT630340 |
| FIN4_02 | KT630341 |
| FIN4_03 | KT630342 |
| POL4_01 | KT630343 |
| POL4_02 | KT630344 |
| POL4_03 | KT630345 |
| RUS1_01 | KT630346 |
| RUS1_02 | KT630347 |
| RUS1_03 | KT630348 |
| SWE8_01 | KT630349 |
| SWE8_02 | KT630350 |
| SWE8_03 | KT630351 |
| POL3_01 | KT630352 |
| POL3_02 | KT630353 |
| POL3_03 | KT630354 |
| SWE4_01 | KT630355 |
| SWE4_02 | KT630356 |
| SWE4_03 | KT630357 |
| RUS3_01 | KT630358 |
| RUS3_03 | KT630359 |
| RUS3_04 | KT630360 |
| RUS2_01 | KT630361 |
| RUS4_02 | KT630362 |
| BLS_03 | KT630363 |
| RUS3_02 | KT630364 |
| SWE3_01 | KT630365 |
| SWE3_02 | KT630366 |
| SWE3_03 | KT630367 |
| SWE2_01 | KT630368 |
| SWE2_02 | KT630369 |
| SWE2_03 | KT630370 |
| SWE9_01 | KT630371 |
| SWE9_02 | KT630372 |
| SWE9_03 | KT630373 |
| GBR7_01 | KT630374 |
| GBR6_01 | KT630375 |
| GBR8_01 | KT630376 |
| GBR8_02 | KT630377 |
| GBR8_03 | KT630378 |
| GBR6_02 | KT630379 |
| GBR6_03 | KT630380 |
| CZE1_01 | KT630381 |
| CZE1_02 | KT630382 |
| CZE1_03 | KT630383 |
| GER4_01 | KT630384 |
| GER4_03 | KT630385 |
| GER1_02 | KT630386 |
| GER2_01 | KT630387 |
| FIN3_01 | KT630388 |
| FIN3_02 | KT630389 |
| FIN3_03 | KT630390 |
| HUN1_02 | KT630391 |
| GER3_01 | KT630392 |
| GER3_02 | KT630393 |
| GER3_03 | KT630394 |

SI table 3. Haplotype memberships for 101 Cytochrome B sequences used in Fig. 2.

| Lineage | Haplotype | N | Drainage (n populations) | Sample code |
| :---: | :---: | :---: | :---: | :---: |
| $1$ | 1 | 3 | Baltic | FIN5 1-3 |
|  | 2 | 1 | Baltic | EST1 2 |
|  | 3 | 49 | Elbe(2), Baltic(9), <br> Scheldt(1), Rhine(2), North sea(2), Vistula(6), Volga(4), <br> Don(3), Danube(1), <br> Hunte(4) | GER1 1,3, EST1 1, 3, SWE6 1 -3, BEL1 3 , GER2 2, 3, GER4 2, NOR 1, 2, <br> SWE11 1-3, RUS2 2, RUS4 1, 3, FIN1 1-3, FIN4 1-3, POL4 1-3, RUS1 1-3, <br> SWE8 1-3, POL5 1-3, SWE4 1-3, RUS3 1, 3, 4, CZE2 1, GER6 1 4, SWE14 1, SWE15 1 |
|  | 4 | 1 | Volga | RUS2 1 |
|  | 5 | 1 | Baltic | RUS4 2 |
|  | 6 | 1 | Dnieper | BLS 3 |
|  | 7 | 1 | Volga | RUS3 2 |
|  | 8 | 3 | Baltic | SWE3 1-3 |
|  | 9 | 2 | Baltic | SWE2 1-3 |
|  | 10 | 3 | Baltic | SWE9 1-3 |
|  | 11 | 13 | UK(4), Rhine(1), Baltic (2) | GBR7 1, GBR6 1-3, GBR8 1-3, NET 1, GER5 1-3, GBR12 1, 2 |
|  | 12 | 3 | Baltic | FIN3 1-3 |
| $2$ | 13 | 3 | Danube | GER4 1, 2, AUS3 1 |
|  | 14 | 3 | Elbe(1), Rhine(1), Danube(1) | GER1 2, GER2 1, AUS2 1 |
|  | 15 | 1 | Danube | CZE1 1 |
|  | 16 | 1 | Danube | CZE1 2 |
|  | 17 | 1 | Danube | CZE1 3 |
|  | 18 | 2 | Danube | HUN 1, 2 |
|  | 19 | 3 | Danube | GER3 1-3 |
|  | 23 | 1 | Elbe | CZE2 2 |
|  | 24 | 2 | Danube | AUS1 1, 2 |
|  | 25 | 2 | Lahn | GER7 1, 2 |
| Outgroup | 20 | 1 |  | Ccarp 1 |
|  | 21 | 1 |  | Ccarp 2 |
|  | 22 | 1 |  | Ccarp 3 |

















 $\begin{array}{lllllllllllllllll}0.129 & 0.695 & 0.389 & 0.515 & 0.496 & 0.202 & 0.555 & 0.37 & 0.36 & 0.253 & 0.195 & 0.463 & 0.362 & 0.15 & 0.345 & 0.641 & 0.201 \\ 0\end{array}$ $\begin{array}{lllllllllllllll}0.318 & 0.439 & 0.281 & 0.136 & 0.389 & 0.193 & 0.255 & 0.108 & 0.157 & 0.489 & 0.2 & 0.29 & 0.201 & 0.368 & 0.168 \\ 0.0121 & 0.838 & 0.768 & 0.686 & 0.753 & 0.706 & 0.756 & 0.65 & 0.699 & 0.776 & 0.334 & 0.829 & 0.753 & 0.828 & 0.661 \\ 0.702\end{array}$


 $\begin{array}{lllllllllll}0.48 & 0.288 & 0.265 & 0.118 & 0.155 & 0.405 & 0.314 & 0.271 & 0.295 & 0.561 & 0.178 \\ 0.1929\end{array}$ 0.230 .166
 $\begin{array}{lll}\text { Ns } & \\ \text { Na } & \text { Na } \\ \text { Ns } & \text { Ns }\end{array}$


| 0.364 | 0.245 | 0.269 | 0.265 | 0.406 | 0.138 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 0.0 .181 |  |  |  |  |  |
|  | 0.483 | 0.467 | 0.503 | 0.646 | 0.005 |

198 SI table 5. Pairwise FST values calculated using the RADseq dataset.

|  | GBR8 | BEL1 | GBR4 | FIN3 | DEN1 | GBR7 | SWE12 | FIN4 | DEN2 | POL4 | RUS1 | SWE2 | SWE8 | SWE9 | SWE10 | NOR2 | POL3 | HUN2 | WEN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GBR8 |  | 0.34971 | 0.35695 | 0.49475 | 0.223897 | 0.35613 | 0.406544 | 0.295019 | 0.293628 | 0.211876 | 0.38775 | 0.308973 | 0.273693 | 0.412263 | 0.321365 | 0.650207 | 0.146766 | 0.61801 | 0.397239 |
| BEL1 |  |  | 0.370425 | 0.390916 | 0.098308 | 0.381154 | 0.300836 | 0.225496 | 0.130398 | 0.152617 | 0.343954 | 0.22423 | 0.08032 | 0.326848 | 0.235947 | 0.522507 | 0.103445 | 0.597677 | 0.31111 |
| GBR4 |  |  |  | 0.513779 | 0.231153 | 0.195241 | 0.423664 | 0.302246 | 0.316185 | 0.218776 | 0.392539 | 0.314975 | 0.284155 | 0.422534 | 0.331921 | 0.698989 | 0.149208 | 0.620806 | 0.412409 |
| FIN3 |  |  |  |  | 0.308284 | 0.517114 | 0.341754 | 0.198275 | 0.364426 | 0.222674 | 0.378729 | 0.27048 | 0.328488 | 0.331267 | 0.286862 | 0.562015 | 0.149991 | 0.614832 | 0.341565 |
| DEN1 |  |  |  |  |  | 0.244594 | 0.239562 | 0.194342 | 0.06762 | 0.136982 | 0.356985 | 0.182005 | 0.085513 | 0.266461 | 0.190793 | 0.362014 | 0.102429 | 0.602815 | 0.237037 |
| GBR7 |  |  |  |  |  |  | 0.430574 | 0.31162 | 0.32391 | 0.229621 | 0.396753 | 0.319608 | 0.295939 | 0.433712 | 0.340292 | 0.692819 | 0.157339 | 0.621918 | 0.422803 |
| SWE12 |  |  |  |  |  |  |  | 0.209406 | 0.282835 | 0.173199 | 0.363912 | 0.198857 | 0.259513 | 0.303204 | 0.211775 | 0.459576 | 0.122381 | 0.606576 | 0.250115 |
| FIN4 |  |  |  |  |  |  |  |  | 0.218225 | 0.142389 | 0.328888 | 0.154803 | 0.211809 | 0.203425 | 0.174944 | 0.316929 | 0.099233 | 0.586541 | 0.198636 |
| DEN2 |  |  |  |  |  |  |  |  |  | 0.153556 | 0.362177 | 0.212179 | 0.101801 | 0.307702 | 0.222051 | 0.459347 | 0.108029 | 0.60623 | 0.284015 |
| POL4 |  |  |  |  |  |  |  |  |  |  | 0.321777 | 0.128672 | 0.150743 | 0.192186 | 0.138734 | 0.250273 | 0.073063 | 0.579543 | 0.161299 |
| RUS1 |  |  |  |  |  |  |  |  |  |  |  | 0.341129 | 0.358602 | 0.368371 | 0.349288 | 0.396194 | 0.278006 | 0.516158 | 0.358584 |
| SWE2 |  |  |  |  |  |  |  |  |  |  |  |  | 0.19768 | 0.218326 | 0.145195 | 0.325228 | 0.094924 | 0.591579 | 0.151258 |
| SWE8 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.289356 | 0.208013 | 0.401551 | 0.110433 | 0.604134 | 0.262799 |
| SWE9 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.257245 | 0.429544 | 0.136442 | 0.607715 | 0.29715 |
| SWE10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.350951 | 0.100275 | 0.598136 | 0.184722 |
| NOR2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.165304 | 0.625602 | 0.426179 |
| POL3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.547371 | 0.111018 |
| HUN2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.604399 |


|  |  |  | Scenario 8 |
| :---: | :---: | :---: | :---: |
| Scenario 3 | Scenario 4 | Scenario 9 <br> - N Pool 1 <br> - N Pool 2 <br> - N Pool 3 <br> - Nb Pool 1 <br> Pool 1 <br> Pool 2 <br> Pool 3 |  |
|  |  |  | Pool 1 - Northern Europe <br> Pool 2 - Don river <br> Pool 3 - Danube River |

a) Stage 2. NEU Major variants


Pool 1 - UK
Pool 2 - Denmark / S.Sweden
Pool 3 - Finland
Pool 4 - Poland
Pool 5 - Sweden
Pool 6 - Tromsø
b) Stage 3. Scenario 14 Minor variants



a)

Variance explained by PCA


Score optimisation - spline interpolation


Discriminant analysis eigenvalues

b) $\begin{gathered}\text { Value of BIC } \\ \text { versus number of clu }\end{gathered}$



Discriminant analysis eigenvalues


DAPC Scatter plot - Full RADseq dataset

a) M1, Lineage 1 only (excluding NOR2)



## b) RADseq data IBD







C)


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a) Stage 2. NEU major variants



Molecular Ecology ${ }^{\text {¢ }}$ Stage 3. NEU minor variants






## Spatial patter4psc|trr|titav M1 datasets

M1 Gest


M1 Lest


M2 Gest


M2 Lest

a) Change in number of tags with incrementing $M$ value

b) change in average tag coverage with incrementing $M$ value



b)



c)

SNP dropout across populations


