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## Colonization and dispersal patterns of the invasive American brine shrimp Artemia franciscana (Branchiopoda: Anostraca) in the Mediterranean region --Manuscript Draft--

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Abstract:	Cysts of the brine shrimp Artemia franciscana are harvested from the Great Salt Lake (GSL) and San Francisco Bay saltworks (SFB) in the U.S.A, and marketed worldwide to provide live food for aquaculture. This species has become invasive across several countries. We investigated (1) if the introduced populations in the Mediterranean region could have originated from these U.S.A. populations, (2) how the genetic diversity of Mediterranean compares with that at GSL and SFB, and (3) if genetic patterns in the Mediterranean can shed light on colonization routes. We sequenced a fragment of the Cytochrome c Oxidase Subunit I and screened microsatellites loci from Mediterranean populations were identical or closely related to those from SFB and GSL, and not related to other available American populations. Microsatellite analyses showed a reduced population diversity for most Mediterranean populations suggesting bottleneck effects, but few populations showing similar or higher genetic diversity than native ones, which are likely to be admixed from both GSL and SFB due to multiple introductions. Our analyses show that all invaded populations could have originated from those commercialised U.S.A. populations.
Response to Reviewers:	Dear Editor (Mr. Deepan Selvaraj),
	Please find below our reply to the reviewer's comments on our manuscript (minor changes) "Colonization and dispersal patterns of the invasive American brine shrimp Artemia franciscana (Branchiopoda: Anostraca) in the Mediterranean region" by Muñoz et al. submitted to Hydrobiologia.

We believe that we have addressed the minor changes suggested in the current revised version.

With kind regards, Joaquin Munoz.

Comments for the Author:

Two reviewers found the manuscript interesting and well written. They suggest only minor changes that should be considered in a revised version of the manuscript.

#### Reviewer #1:

The Manuscript Number: HYDR-D-13-00660 "Colonization and dispersal patterns of the invasive American brine shrimp Artemia franciscana (Branchiopoda: Anostraca) in the Mediterranean region" is in scope of the journal Hydrobiologia. It regards sufficient range of relevant material; the aim of the paper is clear and the results are confirmed by good statistical base. The used literature is relevant, including recent research papers. I suggest that the manuscript could be improved adding data from other north-american populations not included in the present manuscript (see page 13, lines 12-15 from Discussion.

I conclude that this work can be published after the suggested corrections.

REPLY: We have re-analysed the phylogenetic relationship including a representative number of COI haplotypes from our study published in PeerJ 1: e200. http://dx.doi.org/10.7717/peerj.200, coming from Mono Lake (U.S.A.), Mexico, Cuba,

Jamaica, Puerto Rico and others. The results stand the same. Additionally, we have rewritten the corresponding parts in the main text.

### Reviewer #2:

The authors describe the genetic structure of invasive Artemia franciscana populations in the Mediterranean in relation to the genetic structure of potential source populations. I think this is an interesting study with an appropriate design. It is well written and the data are analysed according to the current state of the art. It is a mature manuscript that will only require minor revisions before it can be accepted for publication. I have added my comments directly in a pdf file attached in this review. (and I can assure the editor that the fact that these comments are few in number is not the result of personal laziness). I am looking forward to see this manuscript out in print.

### REVIEWER #2 PDF FILE'S COMMENTS

We followed most of the suggestions of this reviewer.

1.- We tried it and found heaps of loci for Branchinella longirostris but because of complex variation in ploidy levels we did not work further on it for now. We had mixed populations of di tri and tetraploids. Can you say something about the ploidy level of A. franciscana? I heard from some sources that Artemia might also have weird ploidy levels. But perhaps not franciscana. Did you ever get more than two microsat peaks?

REPLY: Artemia franciscana is a sexual diploid organism. Although within the Artemia genus there are asexual lineages with different ploidy levels, they are easy to differentiate morphologically and are not included in our current study.

1 TITLE PAGE 2 Colonization and dispersal patterns of the invasive American brine shrimp 3 Artemia franciscana (Branchiopoda: Anostraca) in the Mediterranean region 4 Joaquín Muñoz<sup>1,¶,\*</sup>, Africa Gómez<sup>2,¶</sup>, Jordi Figuerola<sup>1</sup>, Francisco Amat<sup>3</sup>, Ciro Rico<sup>1,4</sup>, 5 Andy J. Green<sup>1</sup> 6 7 8 1 Department of Wetland Ecology, Estación Biológica de Doñana (CSIC), Isla de La 9 Cartuja, Av. Americo Vespucio, s/n, 41092 – Seville, Spain 10 2 School of Biological, Biomedical and Environmental Sciences, University of Hull. 11 Cottingham Rd, HU6 7RX, Hull, United Kingdom 12 3 Instituto de Acuicultura de Torre de la Sal (CSIC), 12595 - Ribera de Cabanes, 13 Castellón, Spain 14 4 School of Marine Studies, The University of the South Pacific, Lower Laucala 15 Campus, Suva, Fiji Islands 16 17 ¶ These authors contributed equally to this work. \* Corresponding author: Joaquín Muñoz, Department of Wetland Ecology, Estación 18 19 Biológica de Doñana (CSIC), Isla de La Cartuja, Av. Americo Vespucio, s/n, 41092 -20 Seville, Spain, E-mail: quini@ebd.csic.es; Telephone number: +34 954 23 23 40; Fax 21 number: +34 954 62 11 25 22 23 Keywords: Aquatic ecosystems, biological invasion, human- and bird-mediated 24 dispersal, microsatellites, mtDNA, population structure 25 26 Running title: Artemia franciscana invasion in the Mediterranean

# 1 Abstract

2	Cysts of the brine shrimp Artemia franciscana are harvested from the Great Salt Lake
3	(GSL) and San Francisco Bay saltworks (SFB) in the U.S.A, and marketed worldwide
4	to provide live food for aquaculture. This species has become invasive across several
5	countries. We investigated (1) if the introduced populations in the Mediterranean region
6	could have originated from these U.S.A. populations, (2) how the genetic diversity of
7	Mediterranean compares with that at GSL and SFB, and (3) if genetic patterns in the
8	Mediterranean can shed light on colonization routes. We sequenced a fragment of the
9	Cytochrome c Oxidase Subunit I and screened microsatellites loci from Mediterranean
10	populations and the two putative U.S.A. sources. Haplotypes from Mediterranean
11	populations were identical or closely related to those from SFB and GSL, and not
12	related to other available American populations. Microsatellite analyses showed a
13	reduced population diversity for most Mediterranean populations suggesting bottleneck
14	effects, but few populations showing similar or higher genetic diversity than native
15	ones, which are likely to be admixed from both GSL and SFB due to multiple
16	introductions. Results suggest natural dispersal via flamingos between two Spanish
17	populations. Our analyses show that all invaded populations could have originated from
18	those commercialised U.S.A. populations.
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## 1 Introduction

2 Aquatic environments are especially vulnerable to biological invasions (Sakai et al., 3 2001; Grosholz, 2002). The introduction and rapid spread of invasive species in both 4 marine and freshwater ecosystems are of worldwide concern (Ruiz et al., 2000; Roman 5 & Darling, 2007; Gherardi, 2007). For instance, the human-mediated dispersal rate for 6 crustacean zooplankton at an intercontinental scale has been estimated to be up to 7 50,000 times greater than the natural dispersal rate (Hebert & Cristescu, 2002). Aquatic 8 invertebrate invasions often remain undetected for decades due to the difficulty of 9 identifying cryptic species (Knowlton, 1993; Knowlton & Weigt, 1997; Lee, 2000), and 10 are often only discovered using molecular approaches (Mergeav et al., 2005; Mergeav 11 et al., 2007). Non-marine aquatic ecosystems contain many passively dispersed taxa 12 such as copepods, rotifers, ostracods, bryozoans and branchiopods, which produce 13 resting eggs (i.e., encysted embryos in arrested state of development) that allow survival 14 during unfavourable environmental conditions and facilitate dispersal (Hairston, 1996), 15 and are often the stages involved in accidental anthropogenic introductions (Bailey et 16 al., 2003; Gray et al., 2005).

17 The anostracan Artemia franciscana Kellogg, 1906, is a sexual brine shrimp 18 native to the Americas (Amat et al., 2004) that inhabits hypersaline ecosystems such as 19 lakes, lagoons and salt ponds. Since the 50s, A. franciscana commercially harvested 20 cysts (i.e., diapausing or resting eggs) have been exported worldwide from two U.S.A. 21 populations, San Francisco Bay (SFB) and the Great Salt Lake (GSL) in Utah for use as 22 live food in aquaculture and the aquarium pet trade (Lavens & Sorgeloos, 2000, and 23 references therein), leading to accidental or deliberate introductions into ecosystems 24 outside the native range. Artemia franciscana was also intentionally inoculated into 25 salterns worldwide to provide local sources of cysts (e.g., Camara, 2001). Thus, by the 26 80s many other commercial sources of cysts became available (Vanhaecke &

1 Sorgeloos, 1983), and notably from coastal China (Van Stappen et al., 2007). However, 2 Bengtson et al. (1991) estimated that over 70% of marketed Artemia cysts originated 3 from the GSL population. In addition to anthropogenic introductions, Artemia cysts can 4 be effectively dispersed by migratory birds (Sánchez et al., 2007; Sánchez et al., 2012) 5 and over short and mid distances by wind and motor vehicles (Vanschoenwinkel et al., 6 2008a, b; Waterkeyn et al., 2010). In particular, there is potential for spread of this 7 invasive anostracan to hypersaline sites unaffected by aquaculture facilities via 8 flamingos and other waterbirds.

9 The recent expansion of aquaculture in the Mediterranean region has led to the 10 release of A. franciscana into sites previously occupied by native Artemia species. 11 Artemia franciscana was first reported in Portugal around the 1980's (Hontoria et al., 12 1987), and later in France (Thiery & Robert, 1992), Spain, Italy, and Morocco (Amat et 13 al., 2007, and references therein; Mura et al., 2006). Its establishment was followed by 14 rapid local extinctions of the native A. salina and A. parthenogenetica. A combination 15 of habitat loss and the establishment of A. franciscana have resulted in the loss of 55 -16 74% of native Artemia populations across Spain, Portugal and France (Amat et al., 17 2007). Although the initial colonization of Mediterranean habitats is assumed to have 18 originated from GSL and SFB in the U.S.A., this hypothesis has not been tested yet 19 using genetic markers. Furthermore, although genetic studies can shed light on the role 20 of aquatic birds as effective dispersal vectors of invertebrates (Figuerola et al., 2005), 21 no such studies have yet been carried out on A. franciscana. 22 Artemia franciscana's high genetic diversity, phenotypic plasticity, high 23 fecundity, and a large native geographic range could explain its high invasiveness 24 (Amat et al., 2007; Ruebhart et al., 2008, and references therein). This invasive 25 anostracan outcompetes the native parthenogenetic Artemia strain in laboratory

- 26 experiments (Browne, 1980). In addition, propagule pressure can be extremely high.

1 Shrimp and finfish hatchery effluent leading to accidental releases and intentional 2 inoculations involving large number of nauplii (i.e., first larval stage) may regularly 3 have taken place in the form of multiple introductions, not necessarily from the same 4 geographic source. Enemy release may also be involved in the ability of A. franciscana 5 to outcompete native species, as compared to native Mediterranean Artemia, A. 6 franciscana experiences reduced levels of parasitism by avian cestodes, which reduce 7 the fecundity of brine shrimps and increase bird predation (Georgiev et al., 2007; 8 Sánchez et al., 2009). Despite its detrimental effects on native Artemia biodiversity 9 (Muñoz et al., 2008) and calls for import control and management (Ruebhart et al., 10 2008; Amat et al., 2005a), no specific actions have been carried out to contain the 11 spread of A. franciscana. However, a better knowledge of the introduction sources, 12 mode and patterns of invasion and colonisation in A. franciscana could assist the 13 development of future strategies for the management of aquaculture and the 14 conservation of hypersaline ecosystems.

15 The relationship between the levels of genetic diversity found in source 16 populations and those found in the established populations after invasion is not 17 straightforward (Roman & Darling, 2007; Darling et al., 2008; Dlugosch & Hays, 18 2008). Genetic diversity of invasive populations has generally been assumed to be 19 reduced compared to those from the native range. Indeed, punctual introductions can 20 result in strong population bottlenecks (Golani et al., 2007). Many invasive species 21 suffer an associated reduction in genetic diversity due to founder effects (e.g., Muñoz-22 Fuentes et al., 2006), while large increases are rare (Dlugosch & Parker, 2008). 23 However, similar or higher genetic diversity to the native range has been reported in 24 some aquatic invasive species (Roman & Darling, 2007), due to introduction events 25 from multiple sources and/or large propagule pressure lead to admixture in the non-26 native range (Roman & Darling, 2007; Wilson et al., 2009). Unlike natural extra-range

1 dispersal events, human-mediated biological invasions are often the result of multiple 2 introductions from different sources to non-indigenous locations (Wilson et al., 2009). 3 Therefore, new populations established directly by human-mediated introductions might 4 have different genetic signatures to those established by natural dispersal from pre-5 existing populations within the introduced range, providing a valuable tool for 6 understanding the colonisation routes of invasive species. For instance, human-7 mediated dispersal and multiple introductions should result in a lack of correlation 8 between genetic and geographic distances (e.g., Elderkin et al., 2004), whereas the 9 opposite (i.e., an isolation-by-distance pattern) or a random genetic distribution (e.g., 10 Dupont et al., 2003) would be more likely under equilibrium conditions (but see 11 Herborg et al., 2007). 12 The invasive history and biological traits of A. franciscana represent a rare

13 opportunity to investigate the interplay between genetic patterns and the relative role 14 that human- and waterbird-mediated dispersal have on population expansion. Here we 15 investigate the origin, mechanisms, and patterns of A. franciscana invasion in the 16 Mediterranean by screening populations with mitochondrial and nuclear markers, using 17 combined phylogenetic, phylogeographic, and population genetic analyses. Specifically, 18 we investigated whether (1) all Mediterranean populations originated from the marketed 19 GSL and SFB populations, and (2) whether current patterns of genetic diversity and 20 differentiation in the non-native range are consistent with scenarios of human- and bird 21 mediated dispersal.

22

### 23 Materials and methods

24 Samples and data collection

25 We collected cyst samples from 16 invaded Mediterranean solar salterns and salt lakes,

26 from a total of 26 populations identified to date, covering the four European countries

1	(i.e., Portugal, Spain, France, and Italy - Hontoria et al., 1987; Amat et al., 2005a)
2	where A. franciscana has been detected. Additionally, we obtained cyst samples from
3	the two native commercially exploited U.S.A. populations (SFB and GSL), and a
4	population of unknown origin in Sal Island in the Cape Verde archipelago (Amat et al.,
5	2010) (see Table 1). All samples were obtained from the 'cyst-bank' of the Instituto de
6	Acuicultura de Torre de la Sal (CSIC, Castellón, Spain). Cysts were preserved in 100%
7	ethanol until needed. In addition, for phylogenetic analyses we downloaded all available
8	A. franciscana COI sequences from GenBank for which we could obtain geographic
9	information, either directly from the GenBank record or from the original publication,
10	including samples from U.S.A., and the rest of its native range in the Americas, as well
11	as India, Vietnam and China.
12	
13	Laboratory procedures
14	DNA was isolated from individual cysts (5-37 individuals per population for
15	mitochondrial analyses, and 29-47 individuals for nuclear analyses), previously rinsed
16	in distilled water, using an alkaline lysis protocol optimized for zooplanktonic
17	diapausing eggs (Montero-Pau et al., 2008). We used specific Artemia primers in the
18	same position as primers LCO1490/HCO2198 from (Folmer et al., 1994) to amplify a
19	fragment of the Cytochrome c Oxidase Subunit I (COI) mitochondrial gene.
20	PCR/sequencing protocols were performed following (Muñoz et al., 2008). Sequences
21	were edited and aligned using Sequencher <sup>TM</sup> version 4.5 (Gene Codes Corp., © 1991-
22	2005). All different sequences (i.e., haplotypes) found in the present study were
23	deposited in DNA Data Bank of Japan (DDBJ) database (Accession No AB859230-
24	AB859239 – see Table 2 for details and link to GenBank Acc Nos). Samples were
25	genotyped for four microsatellite loci (Af_A108, Af_B10, Af_B9, and Af_B11)
26	following Muñoz et al. (2009).

# 1 Mitochondrial DNA analyses

2	Identical sequences were collapsed into haplotypes prior to phylogenetic and
3	phylogeographic analyses using FaBox (http://users-birc.au.dk/biopv/php/fabox/). We
4	reconstructed the evolutionary history of all A. franciscana haplotypes (i.e., from
5	Mediterranean and out of this area) using Neighbor-Joining (NJ) and Maximum
6	Likelihood (ML) approaches in MEGA v.5.2.2. (Tamura et al., 2007) using the
7	evolutionary model best fitting the data. The robustness of the branches was assessed
8	with 1000 pseudo-replicates. DnaSP v.4.90 (Rozas et al., 2003) was used to compute
9	the number of polymorphic sites and of non-synonymous substitutions.
10	TCS v.1.21 (Clement et al., 2000), which follows the statistical parsimony
11	algorithm to generate a haplotype network, was used to display the genealogical
12	relationships among our Mediterranean samples and some public available A.
13	franciscana COI haplotypes. Standard intra-population diversity parameters, haplotype
14	diversity ( <i>H</i> ) and nucleotide diversity ( $\pi$ ), and inter-population pairwise $\phi_{ST}$ values
15	(corrected by a K2-P evolutionary model) were obtained using Arlequin v.3.11
16	(Excoffier et al., 2005). Because differences in sampling can bias genetic diversity
17	comparisons among different populations, a rarefaction analysis adapted for population
18	genetic data conducted by the program RAREFAC v. 1.02 (available from R. Petit at
19	http://www.pierroton.inra.fr/genetics/labo/Software/Rarefac/index.html) was used to
20	calculate standardized allelic richness (A) for each sampled population. RAREFAC
21	requires a rarefaction size (see Petit et al., 1998), which was set to ten in our case ( $n =$
22	10). Thus, three populations with n<10 (i.e., CBU, FVO, and RFR) were not used in
23	such analyses. All those estimates (i.e., $H$ , $\pi$ , $A$ and $\phi_{ST}$ ) were used to assess the
24	population genetic diversity after introduction and to identify the likely origin of each
25	invaded population.

1 Microsatellite analyses from U.S.A. and Mediterranean invaded sites

2	Arlequin was used to compute observed $(H_0)$ and expected $(H_E)$ heterozygosity, number
3	of alleles (Na), linkage disequilibrium (LD) between loci, Hardy-Weinberg equilibrium
4	(HWE) of each locus. The Fst-statistic may not be appropriate for assessment of genetic
5	structure and differentiation among populations (Jost, 2008; Dupont et al., 2009),
6	therefore, we calculated both Fst and Dest pairwise values using GenAlEx ver.6.5
7	(Peakall & Smouse, 2012). In addition, we used a Bayesian multi-locus method (with a
8	non-equilibrium method, individual-based admixture analysis) implemented in BAPS
9	v.5.2 (Corander et al., 2003; Corander et al., 2008) to infer population structure and to
10	group the data by a stochastic optimization model to infer the posterior probability of
11	the number of distinct clusters, $K$ . In particular, we used the spatial model for genetic
12	discontinuities, running five replicates with upper bound values of $K = 5$ , 10, 20 and 25.
13	Furthermore, to assess the most likely grouping of individuals in clusters, we used
14	Principal Component Analysis (PCA) as a different clustering approach. PCA-GEN
15	software (http://www2.unil.ch/popgen/softwares/pcagen.htm) is a program that does not
16	require assumptions of equilibrium within populations, correlates genotypes and allele
17	frequencies among all individuals using no information regarding population
18	identification, and plots genetic structure among populations.
19	

## 20 **Results**

21 Global mitochondrial phylogeography of invasive Artemia franciscana

22 The sequence alignment used in both phylogenetic and phylogeographic analyses were

trimmed to 477 bp. The COI sequences aligned (including 274 generated in the present

study – collected from 19 sites from GSL, SFB, Cape Verde and Mediterranean region;

see Table 1) collapsed into 71 haplotypes. Overall, 94 variable sites and 62 parsimony

informative sites were revealed, with no indels or stop codons. Five non-synonymous
substitutions were found in positions.

Both phylogenetic reconstruction methods (ML and NJ) recovered a virtually identical tree topology and support values, so only the ML reconstruction is shown (Fig. 1), with geographically concordant branches and over ten lineages, similarly to results from Muñoz et al (2013). All Mediterranean and other invasive populations had identical haplotypes to SFB and UTAH populations or highly related haplotypes to these.

9 The median-joining haplotype network showed three disjoined networks. One 10 included the Cape Verde haplotypes, another for the Mexican and Chilean/Argentinean 11 phylogenetic subclades (data not shown, but see Fig. 1), and a major network formed by 12 the rest of haplotypes encompassing U.S.A., invasive and some Chilean haplotypes. In 13 this latter network (see Fig. 2), when excluding relatively divergent Chilean haplotypes, 14 a total of 12 closely related haplotypes, no more than five substitutions apart, were 15 detected. A total of ten haplotypes were present in invaded populations in the 16 Mediterranean, six of them found also in GSL and SFB (Fig. 2). The remaining four 17 haplotypes were only found in invaded populations, although they were closely related 18 to the most common haplotypes in GSL and SFB (1 or 2 substitutions apart). Both GSL and SFB shared the three most common haplotypes (i.e., HAf01, HAf02, and HAf04), 19 20 but they were found at different frequencies. Haplotype HAf02 was the most common 21 in GSL (79.3% of individuals), whereas haplotype HAf04 was the most common in 22 SFB (70.3% of individuals). Furthermore, HAf02 and HAf04 were the most common 23 haplotypes in the invaded populations. Amongst the 16 Mediterranean populations 24 analysed, HAf04 was present at 14 sites, while HAf02 was present at six.

25

26 Mitochondrial genetic diversity in U.S.A. and Mediterranean populations

1	Intra-population haplotype diversity, A (Table 2; note that for three populations, the
2	value A was not estimated due to low sample sizes), and inter-population pairwise
3	genetic diversity $\phi_{ST}$ (Table 4B) indicated: 1) A high and significant level of population
4	differentiation between both native populations SFB and GSL ( $\phi_{ST}$ value of 0.546); 2)
5	GSL had lower haplotype diversity than SFB; 3) Ten Mediterranean populations
6	appeared to be related to SFB with non-significant $\phi_{ST}$ values and lower diversity values
7	than SFB, except for FPI, which had similar diversity values; 4) Only one population,
8	TRI, showed a non-significant $\phi_{ST}$ value when compared to GSL; 5) Four populations
9	(ESM, GER, LTA, and SPA) were significantly different to both U.S.A. populations as
10	indicated by $\phi_{ST}$ values, but of these four only LTA contained a haplotype not found in
11	SFB or GSL.
12	
13	Nuclear genetic diversity, regional structure and demographic patterns in the
14	Mediterranean
15	All loci used to screen Mediterranean and North American samples (714 cysts) were
16	unlinked (results not shown) and only the Af_108 locus was in Hardy-Weinberg
17	disequilibrium for most populations, with significant homozygote excesses probably
18	due to null alleles (Muñoz et al., 2009). Nevertheless, no population was found to be
19	under disequilibrium for all loci. Af_108 was monomorphic for two populations, BFI
20	and RFR, and RFR population could not be genotyped for the Af_B9 locus (see Tables
21	1 and 3 for details). The number of alleles per locus ranged from 13 (Af_A108) to 40 $$
22	(Af_B11). The mean number of alleles (Na) and gene diversity ( $H_E$ ) was similar in both
23	commercialised native populations (13.5 and 14.0, and 0.753 and 0.847 in SFB and
24	GSL, respectively, with SFB showing two private alleles in Af_B10). However, Na and
25	$H_E$ showed wide differences in the Mediterranean, ranging from 3.0 to 17.0 and 0.256 to

1 0.840, respectively, with private alleles in nine populations. Several introduced

2 populations showed equal or higher Na than native ones.

3	Most pairwise $F_{ST}$ and <i>Dest</i> values were highly significant (see Table 4A), even
4	between both native populations SFB and GSL (0.076 and 0.593, respectively).
5	Contrary to the results for mitochondrial $\phi_{ST}$ values, all Mediterranean populations
6	showed significantly high pairwise $F_{ST}$ and <i>Dest</i> values with SFB and GSL except ESM
7	population, which showed non-significant values compared to GSL. Four
8	Mediterranean populations (ESM, SLU, GER and CBU) showed no genetic
9	differentiation between them based on their $F_{ST}$ and <i>Dest</i> pairwise values.
10	Bayesian clustering analysis (BAPS) and Principal Component Analysis (PCA)
11	gave similar population structure, but produced different numbers of clusters. BAPS
12	analysis resulted in ten clusters with a probability higher than 0.97 (see Fig. 3 for
13	details). Four clusters contained more than one population, while six populations were
14	identified as single clusters. As expected, two multi-population clusters included the
15	two native populations. The cluster containing GSL had five populations, and the
16	cluster containing SFB had one. However, two clusters made up of two populations
17	each were inferred with independence from the native populations (BMA-FVO, and
18	LTA-FPI). The first two axes of the PCA explained 70.24% of the total variation.
19	Unlike BAPS, PCA analysis did not consider SPA as belonging to the GSL genetic
20	group, and did not group ALC with SFB. Both population structure analyses clearly
21	show that most populations group around SFB and GSL, or in the space between them,
22	indicating introductions from single sources or a range of admixture. However, three of
23	the populations (BFI, RFR or AIG) were outside this admixture gradient.
24	

# **Discussion**

2	Our results strongly suggest that A. franciscana invasive populations across the
3	Mediterranean region and other parts of the world originate from the commercialised
4	populations at GSL and/or SFB in the U.S.A. Other genetic lineages in the native range
5	are geographically restricted and genetically divergent, and have clearly played no part
6	in the Mediterranean and the non-native range included in our study. Although we
7	included all the available populations in U.S.A., we want to highlight that GSL and SFB
8	are the only ones of importance for exporting cysts on the world market. In addition,
9	Muñoz et al. (2013) have recently confirmed our results by surveying a continental
10	phyogeography for A. franciscana including additional American haplotypes present on
11	this study. The low frequency of private alleles (between 1.1% and 7.9% for
12	microsatellites, data not shown) in the invasive populations, also suggests that SFB and
13	GSL could be the original source populations.
14	However, we cannot rule out that some of the Mediterranean populations were
15	established by secondary introductions from Asia, given the dominance of SFB and
16	GSL haplotypes in China, India and Vietnam (Fig. 2) and the commercial availability of
17	A. franciscana cysts from Bohai Bay in China on the world market (Van Stappen et al.
18	2007, http://www.bhb-artemia.com/). Surprisingly, Cape Verde haplotypes form a
19	highly supported independent mitochondrial clade, even though this region has
20	previously been assumed to be part of the invasive range of this species due to its
21	isolation from the Americas (see Muñoz & Pacios, 2010). Our results suggest that A.
22	franciscana may be native in the Cape Verde islands.
23	The three most common haplotypes from SFB and GSL (HAf01, HAf02, and
24	HAf04) are extremely similar, indicating that either: 1) both populations were formed
25	very recently; 2) one of them was used to 'seed' the other one (e.g., A. franciscana

26 colonizing the salt ponds created at SFB may have originated from GSL, these sites

1	being connected via migratory waterbirds); or 3) after some relatively recent population
2	differentiation there has been a lot of admixture. Despite the fact that our analyses
3	included only a few nuclear loci, our results indicate a significant genetic divergence
4	between these two populations (see also Muñoz et al., 2009). Although there are 10
5	microsatellites developed for this species, only the four used in this study amplified
6	consistently and provided repeatable banding patterns. We recognize that this small
7	number gives little power in our PCA and BAPS analyses. However, developing
8	microsatellite markers for Anostraca is notoriously difficult, and we are not aware of
9	any other studies that have used them for these crustaceans (but see Deiner et al., 2013),
10	despite a range of studies using mitochondrial markers.
11	Mitochondrial markers have been very useful in inferring the origin and invasion
12	pathways of introduced vertebrate and invertebrate species (Kelly et al., 2006; Ashton et
13	al., 2008; Ficetola et al., 2008; Mabuchi et al., 2008; Gaubert et al., 2009). However,
14	since the same three commonest haplotypes in A. franciscana are shared by GSL and
15	SFB, but with different relative frequencies, the resolution offered by mtDNA is
16	insufficient to make clear conclusions on which of these U.S.A. populations is involved
17	as the ultimate source of invasions or estimate the level of admixture. The most
18	common mtDNA haplotype from SFB (HAf04) is present in most invaded populations
19	(see Table 2), and genetic drift is likely to be involved in changing haplotype
20	frequencies of the invaded populations. In addition, many Mediterranean populations
21	were not significantly different from the SFB population as measured with $\phi_{ST}$ (see
22	Table 4B).
23	Different colonization and dispersal patterns can be expected to leave specific
24	genetic signatures across the invaded range (Dupont et al., 2009; Willson et al., 2009).
25	For instance, under a scenario of mass introduction, high genetic diversity is expected in

26 the invaded populations. In addition, homogenization of the gene pool of invaded

1	populations or low population differentiation between invaded populations and the
2	source population can be expected due to continuous or frequent introduction events,
3	which is likely for easily accessible geographic areas. Examples of this mass
4	introduction pattern occurs in our microsatellite analyses where we found one
5	Mediterranean A. franciscana population (ESM) is not significantly different than the
6	GSL population as measured with the G-statistics <i>Dest</i> and $F_{ST}$ (see Table 4A).
7	Furthermore, this Mediterranean population does not show significant differentiation
8	with SLU or GER, and SLU is not differentiated from CBU. These four populations
9	from different parts of the Iberian Peninsula also cluster together with GSL in the BAPS
10	population structure analyses (Fig. 3), and do not show signs of loss of genetic diversity
11	when compared with native populations (see Table 3).
12	On the other hand, as expected given the relatively reduced number of
13	mitochondrial haplotypes in the native populations (and the smaller effective population
14	size of mtDNA), most invasive populations showed reduced mtDNA diversity likely
15	due to population bottlenecks, founder effects, and genetic drift during the colonisation
16	process, which might reflect habitat monopolisation by a few highly successful
17	individuals (see De Meester et al., 2002).
18	Our microsatellite results could fit with punctual human introductions resulting
19	in population bottlenecks for at least three Mediterranean populations (BFI, AIG and
20	MSA), which show the lowest genetic diversity (i.e., expected heterozygosity and
21	number of alleles), but non-significant differentiation with the native population SFB at
22	mitochondrial level (see Table 4B). All of these populations were strongly differentiated
23	genetically with the rest of the populations according to the microsatellite analyses
24	(Table 4A). BFI and AIG also appear as single population clusters in the PCA analysis,
25	away from the admixture gradient between SFB and GSL where the rest of the
26	populations are distributed (Fig. 3).

1	In addition, our results suggest natural dispersal between two populations in
2	South Spain, La Tapa salt ponds (LTA) and Fuente de Piedra lagoon (FPI). Both
3	mitochondrial and nuclear data show a close relationship between these populations
4	(e.g., no significant $F_{ST}$ value, clustering analyses, both share a unique mitochondrial
5	haplotype – HAf05 in Table 2). LTA is a coastal saltpan population in Cadiz Bay with
6	an intensive aquaculture industry and where A. franciscana was fully established around
7	2002 (Amat et al., 2005a; Amat et al., 2007). In contrast, FPI is a natural inland closed-
8	basin lake situated 140 km away from LTA, and where the native A. salina occurred
9	until A. franciscana was detected in 2005 (Amat et al., 2007). FPI holds the most
10	important breeding colony of the greater flamingo Phoenicopterus ruber in Spain, is a
11	highly protected Nature Reserve and has no influence from the aquaculture industry, but
12	flamingos breeding there regularly fly to LTA to feed (Amat et al., 2005b). Flamingos
13	are the most abundant waterbirds in saltpans along the Iberian coast by biomass
14	(Rodríguez-Pérez & Green, 2006; Sánchez et al., 2013), and are effective dispersers of
15	Artemia cysts (MacDonald, 1980; Sánchez et al., 2012. In addition, a mechanistic
16	model of dispersal of Artemia cysts by waterbirds estimated that ducks may disperse
17	them over distances of 230-1209 Km (Viana et al., 2013). Although anostracan cysts
18	can also be dispersed a short distance by wind, this appears to be limited to a maximum
19	of a few hundred metres Vanschoenwinkel et al., 2008a, b). Therefore, the most likely
20	explanation for the colonization of FPI by A. franciscana is through natural dispersal
21	via birds, rather than by direct human intervention. Unfortunately, our dataset does not
22	have the necessary resolution to shed light into the invasion patterns of the other
23	Mediterranean populations such as SPA, TRI from Spain; SGU from France; and ALC,
24	BMA, FVO, RFR from Portugal, which likely involve a combination of several patterns
25	described and also including admixture.

1	In conclusion, our results confirm previous indications that the worldwide
2	invasion of A. franciscana is based on the spread of cysts originally from two
3	commercially exploited U.S.A. populations (i.e., SFB and GSL). As in other aquatic
4	invaders (Rius et al., 2008), high genetic diversity found in several Mediterranean
5	populations point to an establishment as a result of multiple introductions from different
6	populations of origin and/or high propagule pressure (see Wilson et al., 2009).
7	Furthermore, high genetic diversity is usually linked to both adaptive potential and
8	physiological plasticity, helping an introduced species to success as an invader
9	(Dlugosch & Parker, 2008). Our results, and previous studies (Browne &
10	Wanigasekera, 2000), indicate that A. franciscana possesses high genetic diversity, and
11	high adaptive potential and plasticity, facilitating the successful colonisation of suitable
12	habitats through the world. Future research using genomics approaches is desirable to
13	provide better information on the relationships between populations in the native and
14	non-native ranges and the role of local adaptation in the invasive process (e.g. to
15	variation in water chemistry or temperature).
16	
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# **References**

2	Amat, F., R. G. Cohen, F. Hontoria & J. C. Navarro, 2004. Further evidence and
3	characterization of Artemia franciscana (Kellogg, 1906) populations in
4	Argentina. Journal of Biogeography 31: 1735-1749.
5	Amat, F., F. Hontoria, O. Ruiz, A. J. Green, M. I. Sánchez, J. Figuerola & F. Hortas,
6	2005a. The American brine shrimp as an exotic invasive species in the western
7	Mediterranean. Biological Invasions 7: 37-47.
8	Amat, J. A., M. A. Rendon, M. Rendon-Martos, A. Garrido & J. M. Ramirez, 2005b.
9	Ranging behaviour of greater flamingos during the breeding and post-breeding
10	periods: Linking connectivity to biological processes. Biological Conservation
11	125: 183-192.
12	Amat, F., F. Hontoria, J. C. Navarro, N. Vieira & G. Mura, 2007. Biodiversity loss in
13	the genus Artemia in the Western Mediterranean Region. Limnetica 26: 177–
14	194.
15	Amat, F., F. Hontoria, E. Redon, M. Maccari, I. Varo, J. C. Navarro & L. Ballell, 2010.
16	Biodiversidad de Artemia en Macaronesia. XV Congreso de la Asociación
17	Ibérica de Limnología. Ponta Delgada, San Miguel, Azores. 4-11 Julio.
18	Ashton, G. V., M. I. Stevens, M. C. Hart, D. H. Green, M. T. Burrows, E. J. Cook & K.
19	J. Willis, 2008. Mitochondrial DNA reveals multiple Northern Hemisphere
20	introductions of Caprella mutica (Crustacea, Amphipoda). Molecular Ecology
21	17: 1293-1303.
22	Bailey, S. A., I. C. Duggan, C. D. A. van Overdijk, P. T. Jenkins & H. J. MacIsaac,
23	2003). Viability of invertebrate diapausing eggs collected from residual ballast
24	sediment. Limnology and Oceanography 48: 1701-1710.

1	Bengtson, D. A, P. Léger & P. Sorgeloos, 1991. Use of Artemia as a food source for
2	aquaculture. In: Browne RA, Sorgeloos P, Trotman CAN (eds). Artemia
3	biology. CRC Press, Boca Raton, FL. Pp. 255-285.
4	Browne, R. A., 1980. Competition experiments between parghenogenetic and sexual
5	strains of the brine shrimp, Artemia salina. Ecology 31: 471-474.
6	Browne, R. A. & G. Wanigasekera, 2000. Combined effects of salinity and temperature
7	on survival and reproduction of five species of Artemia. Journal of Experimental
8	Marine Biology and Ecology 244: 29-44.
9	Camara, M. R., 2001. Dispersal of Artemia franciscana Kellogg (Crustacea; Anostraca)
10	populations in the coastal saltworks of Rio Grande do Norte, northeastern Brazil.
11	Hydrobiologia 466: 145-148.
12	Clement, M., D. Posada & K. A. Crandall, 2000. TCS: a computer program to estimate
13	gene genealogies. Molecular Ecology 9: 1657–1659.
14	Corander, J., P. Waldmann & M. J. Sillanpaa, 2003. Bayesian analysis of genetic
15	differentiation between populations. Genetics 163: 367-374.
16	Corander, J., P. Marttinen, J. Sirén & J. Tang, 2008. Enhanced Bayesian modelling in
17	BAPS software for learning genetic structures of populations. BMC
18	Bioinformatics 9: 539.
19	Darling, J. A., M. J. Bagley, J. Roman, C. K. Tepolt & J. B. Geller, 2008. Genetic
20	patterns across multiple introductions of the globally invasive crab genus
21	Carcinus. Molecular Ecology 17: 4992-5007.
22	De Meester, L., A. Gómez, B. Okamura & K. Schwenk, 2002. The Monopolization
23	Hypothesis and the dispersal-gene flow paradox in aquatic organisms. Acta
24	Oecologica 23: 121-135.
25	Deiner, K., J. Hull & B. May, 2013. Eight novel microsatellite loci developed from
26	vernal pool fairy shrimp. Journal of Fish and Wildlife Management 4: 134-138.

1	Dlugosch, K. M. & C. G. Hays, 2008. Genotypes on the move: some things old and
2	some things new shape the genetics of colonization during species invasions.
3	Molecular Ecology 17: 4583-4585.
4	Dlugosch, K. M. & I. M. Parker, 2008. Founding events in species invasions: genetic
5	variation, adaptive evolution, and the role of multiple introductions. Molecular
6	Ecology 17: 431-449.
7	Dupont, L., D. Jolliver & F. Viard, 2003. High genetic diversity and ephemeral drift
8	effects in a successful introduced mollusc (Crepidula fornicata: Gastropoda).
9	Marine Ecology Progress Series 253: 183-195.
10	Dupont, L., F. Viard, M. J. Dowell, C. Wood & J. D. D. Bishop, 2009. Fine- and
11	regional-scale genetic structure of the exotic ascidian Styela clava (Tunicata) in
12	southwest England, 50 years after its introduction. Molecular Ecology 18: 442-
13	453.
14	Elderkin, C. L., E. J. Perkins, P. L. Leberg, P.L. Klerks & R. F. Lance, 2004. Amplified
15	fragment length polymorphism (AFLP) analysis of the genetic structure of the
16	zebra mussel Dreissena polymorpha, in the Mississippi River. Freshwater
17	Biology 49: 1487-1494.
18	Excoffier, L., G. Laval & S. Schneider, 2005. Arlequin (version 3.0): An integrated
19	software package for population genetics data analysis. Evolutionary
20	Bioinformatics 1: 47–50.
21	Ficetola, G. F., A. Bonin & C. Miaud, 2008. Population genetics reveals origin and
22	number of founders in a biological invasion. Molecular Ecology 17: 773-782.
23	Figuerola, J., A. J. Green & T. C. Michot, 2005. Invertebrate eggs can fly: evidence of
24	waterfowl-mediated gene flow in aquatic invertebrates. The American Naturalist
25	165: 274-280.
26	Folmer, O., M. Black, W. Hoeh, R. Lutz & R. Vrijenhoek, 1994. DNA primers for

1	amplification of mitochondrial cytochrome C oxidase subunit I from diverse
2	metazoan invertebrates. Molecular Marine Biology and Biotechnology 3: 294-
3	299.
4	Gaubert, P., J. A. Godoy, I. del Cerro & F. Palomares, 2009. Early phases of a
5	successful invasion: mitochondrial phylogeography of the common genet
6	(Genetta genetta) within the Mediterranean Basin. Biological Invasions 11: 523-
7	546.
8	Georgiev, B. B., M. I. Sánchez, G. P. Vasileva, P. N. Nikolov & A. J. Green, 2007.
9	Cestode parasitism in invasive and native brine shrimps (Artemia spp.) as a
10	possible factor promoting the rapid invasion of A. franciscana in the
11	Mediterranean region. Parasitology Research 101: 1647-1655.
12	Gherardi, F., 2007. Biological invaders in inland waters: Profiles, distribution, and
13	threats. Edited by Francesca Gherardi. Published by Springer, The Netherlands.
14	ISBN: 978-1-4020-6028-1. Pp. 733.
15	Golani, D. G., E. Azzurro, M. Corsini-Foka, M. Falautana, F. Andaloro & G. Bernardi,
16	2007. Genetic bottlenecks and successful biological invasions: the case of a
17	recent Lessepsian migrant. Biology Letters 3: 541-545.
18	Gray, D. K., S. A. Bailey, I. C. Duggan & H. J. MacIsaac, 2005. Viability of
19	invertebrate diapausing eggs exposed to saltwater: implications for Great Lakes'
20	ship ballast management. Biological Invasions 7: 531–539.
21	Grosholz, E., 2002. Ecological and evolutionary consequences of coastal invasions.
22	Trends Ecology and Evolution 17: 22-27.
23	Hairston, N. G., 1996. Zooplankton egg banks as biotic reservoirs in changing
24	environments. Limnology and Oceanography 41: 1087-1092.
25	Hebert, P. D. N. & M. Cristescu, 2002. Genetic perspective on invasions: the case of the
26	Cladocera. Canadian Journal of Fisheries and Aquatic Science 59: 1229-1234.

1	Herborg, L. M., D. Weetman, C. van Oosterhout & B. Hänfling, 2007. Genetic
2	population structure and contemporary dispersal patterns of a recent European
3	invader, the Chinese mitten crab, Eriocheir sinensis. Molecular Ecology 16:
4	231-242.
5	Hontoria, F., J. C. Navarro, I. Varo, A. Gonzalbo, F. Amat & N. Vieira, 1987. Ensayo
6	de caracterización de cepas autóctonas de Artemia de Portugal. Seminario
7	Aquac. Inst. Ciencias Biom. "Abel Salazar" Porto (Portugal). Publ Inst C
8	Biomed. Pp. 10.
9	Jost, L., 2008. Gst and its relative do not measure differentiation. Molecular Ecology
10	17: 4015-4026.
11	Kelly, D. W., J. R. Muirhead, D. D. Heath & H. J. Macisaac, 2006. Contrasting patterns
12	in genetic diversity following multiple invasions of fresh and brackish waters.
13	Molecular Ecology 15: 3461-3653.
14	Knowlton, N., 1993. Sibling species in the sea. Annual Review of Ecology and
15	Systematics 24: 189–216.
16	Knowlton, N. & L. A. Weigt, 1997. Species of marine invertebrates: a comparison of
17	the biological and phylogenetic species concepts. In M. F. Claridge, H. A.
18	Dawah, and M. R. Wilson (eds.). Species: the units of biodiversity. Chapman
19	and Hall, New York. Pp. 199–219.
20	Lavens, P. & P. Sorgeloos, 2000. The history, present status and prospects of the
21	availability of Artemia cysts for aquaculture. Aquaculture 181: 397-403.
22	Lee, C. E., 2000. Global phylogegraphy of a cryptic copepod species complex and
23	reproductive isolation between genetically proximate "populations". Evolution
24	54: 2014-2027.
25	Mabuchi, K., H. Senou & M. Nishida, 2008. Mitochondrial DNA analysis reveals
26	cryptic large-scale invasion of non-native genotypes of common carp (Cyprinus

1	carpio) in Japan. Molecular Ecology 17: 796-809.
2	MacDonald, G. H., 1980. The use of Artemia cysts as food by the flamingo
3	(Phoenicopterus ruber roseus) and the shelduck (Tadorna tadorna). In G.
4	Persoone, P. Sorgeloos, O. Roels, and E. Jaspers (eds.). The Brine Shrimp
5	Artemia. Ecology, Culturing, Use in Aquaculture. Universa Press, Wetteren. Pp.
6	97-104.
7	Mergeay, J., D. Verschuren & L. De Meester, 2005. Cryptic invasion and dispersal of
8	an American Daphnia in East Africa. Limnology and Oceanography 50: 1278–
9	1283.
10	Mergeay, J., J. Vanoverbeke, D. Verschuren & L. De Meester, 2007. Extinction,
11	recolonization, and dispersal through time in a planktonic crustacean. Ecology 88:
12	3032–3043.
13	Montero-Pau, J., A. Gómez & J. Muñoz, 2008. Application of an inexpensive and high-
14	throughput genomic DNA extraction method for the molecular ecology of
15	zooplanktonic diapausing eggs. Limnology and Oceanography Methods 6: 218-
16	222.
17	Mura, G., I. Kappas, A. D. Baxevanis, S. Moscatello, Q. D'Amico, G. M. Lopez, F.
18	Hontoria, F. Amat & T. J. Abatzopoulos, 2006. Morphological and molecular
19	data reveal the presence of the invasive Artemia franciscana in Margherita di
20	Savoia salterns (Italy). International Review of Hydrobiology 91: 539-554.
21	Muñoz, J., A. Gómez, A. J. Green, J. Figuerola, F. Amat & C. Rico, 2008.
22	Phylogeography and local endemism of the native Mediterranean brine shrimp
23	Artemia salina (Branchiopoda: Anostraca). Molecular Ecology 17: 3160-3177.
24	Muñoz, J., A. J. Green, J. Figuerola, F. Amat & C. Rico, 2009. Characterization of
25	polymorphic microsatellite markers in the brine shrimp Artemia (Branchiopoda,
26	Anostraca). Molecular Ecology Resources 9: 547-550.

1	Muñoz, J. & F. Pacios, 2010. Global biodiversity and geographical distribution of
2	diapausing aquatic invertebrates: the case of the cosmopolitan brine shrimp,
3	Artemia (Branchiopoda, Anostraca). Crustaceana 83: 465-480.
4	Muñoz, J., F. Amat, A. J. Green, J. Figuerola & A. Gómez, 2013. Bird migratory
5	flyways influence the phylogeography of the invasive brine shrimp Artemia
6	franciscana in its native American range. PeerJ 1: e200.
7	http://dx.doi.org/10.7717/peerj.200.
8	Muñoz-Fuentes, V., A. J. Green, M. D. Sorenson, J. J. Negro & C. Vilà, 2006. The
9	ruddy duck Oxyura jamaicensis in Europe: natural colonisation or human
10	introduction? Molecular Ecology 15: 1441-1453.
11	Peakall, R. & P. Smouse, 2012. GenAlEx 6.5: genetic analysis in Excel. Population
12	genetic software for teaching and research - an update. Bioinformatics 28: 2537-
13	2539.
14	Petit, R. J., A. El Mousadik & O. Pons, 1998. Identifying populations for conservation
15	on the basis of genetic markers. Conservation Biology 12: 844-855.
16	Rius, M., M. Pascual & X. Turon, 2008. Phylogeography of the widespread marine
17	invader Microcosmus squamiger (Ascidiacea) reveals high genetic diversity of
18	introduced populations and non-independent colonizations. Diversity and
19	Distribution 14: 818-828.
20	Rodríguez-Pérez, H. & A. J. Green, 2006. Waterbird impacts on widgeongrass Ruppia
21	maritima in a Mediterranean wetland: comparing bird groups and seasonal
22	effects. Oikos 112: 525-534.
23	Roman, J. & J. A. Darling, 2007. Paradox lost: genetic diversity and the success of
24	aquatic invasions. Trends in Ecology and Evolution 22: 454-464.

1	Rozas, J., J. C. Sánchez-DelBarrio, X. Messeguer & R. Rozas, 2003. DnaSP, DNA
2	polymorphism analyses by the coalescent and other methods. Bioinformatics 19:
3	24969-2497.
4	Ruebhart, D. R., I. E. Cock & G. R. Shaw, 2008. Invasive character of the brine shrimp
5	Artemia franciscana Kellogg 1906 (Branchiopoda: Anostraca) and its potential
6	impact on Australia inland hypersaline waters. Marine & Freshwater Research
7	59: 587-595.
8	Ruiz, G. M., P. W. Fofonoff, J. T. Carlton, M. J. Wonham & A. H. Hines, 2000.
9	Invasion of coastal marine communities in North America: apparent patterns,
10	processes, and biases. Annual Review of Ecology and Systematics 31: 481-531.
11	Sakai, A. K., F. W. Allendorf, J. S. Holt, D. M. Lodge, J. Molofsky, K. A. With, S.
12	Baughman, R. J. Cabin, J. E. Cohen, N.C. Ellstrand, D. E. McCauley, P. O'Neil,
13	I. M. Parker, J. N. Thompson & S. G. Weller, 2001. The population biology of
14	invasive species. Annual Review of Ecology and Systematics 32: 305-332.
15	Sánchez, M. I., A. J. Green, F. Amat & E. M. Castellanos, 2007. Transport of brine
16	shrimps via the digestive system of migratory waders: dispersal probabilities
17	depend on diet and season. Marine Biology 151: 1407-1415.
18	Sánchez, M. I., F. Hortas, J. Figuerola & A. J. Green, 2009. Sandpipers select red brine
19	shrimps rich in both carotenoids and parasites. Ethology 115: 196-200.
20	Sánchez, M. I., F. Hortas, J. Figuerola & A. J. Green, 2012. Comparing the dispersal
21	potential of a native and an invasive brine shrimp via waterbirds. Freshwater
22	Biology 57: 1896–1903.
23	Sánchez, M. I., P. N. Nikolov, D. D. Georgieva, B. B. Georgiev, G. P. Vasileva, P.
24	Pankov, M. Paracuellos, K. Lafferty & A. J. Green, 2013. High prevalence of
25	cestodes in Artemia spp. throughout the annual cycle: relationship with
26	abundance of avian final hosts. Parasitology Research. 112: 1913-1923.

1	Stamatakis, A., P. Hoover & J. Rougemont, 2008. A rapid bootstrap algorithm for the
2	RAxML web-servers. Systematic Biology 75: 758-771.
3	Tamura, K., J. Dudley, M. Nei & S. Kumar, 2007. MEGA4: Molecular Evolutionary
4	Genetics Analysis (MEGA) software version 4.0. Molecular Biology and
5	Evolution 24: 1596–1599.
6	Thiery, A. & F. Robert, 1992. Bisexual populations of the brine shrimp Artemia in Sète-
7	Villeroy and Villeneuve Saltworks (Languedoc, France). International Journal of
8	Salt Lake Research 1: 47-63.
9	Van Stappen, G., H. Y. Yu, X. M. Wang, S. Hoffman, K. Cooreman, P. Bossier & P.
10	Sorgeloos, 2007. Occurrence of allochthonous Artemia species in the Bohai Bay
11	area, PR China, as confirmed by RFLP analysis and laboratory culture tests
12	Fundamental and Applied Limnology 170: 21-28
13	Vanhaecke, P. & P. Sorgeloos, 1983. International study on Artemia XIX. Hatching
14	data for ten commercial sources of brine shrimp cysts and re-evaluation of the
15	"hatching efficiency" concept. Aquaculture 30: 43-52.
16	Vanschoenwinkel, B., S. Gielen, M. Seaman & L. Brendonck, 2008a. Any way the
17	wind blows - frequent wind dispersal drives species sorting in ephemeral aquatic
18	communities. Oikos 117: 125-134.
19	Vanschoenwinkel, B., S. Gielen, H. Vandewaerde, M. Seaman & L. Brendonck, 2008b.
20	Relative importance of different dispersal vectors for small aquatic invertebrates
21	in a rock pool metacommunity. Ecography 31: 567-577.
22	Viana, D. S., L. Santamaria, T. C. Michot & J. Figuerola, 2013. Migratory strategies of
23	waterbirds shape the continental-scale dispersal of aquatic organisms.
24	Ecography 36: 430-438.
25	Waterkeyn, A., B. Vanschoenwinkel, S. Elsen, M. Anton-Pardo, P. Grillas & L.
26	Brendonck, 2010. Unintentional dispersal of aquatic invertebrates via footwear

1	and motor vehicles in a Mediterranean wetland area. Aquatic Conservation:
2	Marine and Freshwater Ecosystems 20: 580-587.
3	Wilson, J. R. U., E. E. Dormontt, P. J. Prentis, A. J. Lowe & D. M. Richardson, 2009.
4	Something in the way you move: dispersal pathways affect invasion success.
5	Trends in Ecology and Evolution 24: 136-144.
6	
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## 1 Figure Legends

2 Figure 1: Tree showing the evolutionary history of Artemia franciscana inferred from the Cytochrome Oxidase Subunit I dataset obtained from the total geographic range 3 4 analysed. Topology shown was obtained using Maximum Likelihood (see text for 5 details). The haplotypes found in this study are labelled HAf01 to HAf10 (see Table 2). 6 Only tips of the U.S.A.-invasive and Cape Verde lineages are shown for simplicity. For 7 further information about the remaining lineages see Muñoz et al (2013). Tips from 8 U.S.A.-invasive clade include geographical information in parenthesis. Haplotypes 9 found in invaded non-American sites are indicated in red, and those exclusive from the 10 U.S.A. commercialised populations analysed (i.e., GSL and SFB) are indicated in bold. 11 Bootstrap supports over 50, after 1000 pseudo-replicates, are shown for the main 12 branches. 13

14 Figure 2: Haplotype network displayed by TCS software. Circles (i.e., haplotypes) are 15 scaled to the number of individuals observed with that haplotype. Grey circles indicate 16 haplotypes exclusive to Mediterranean invaded populations. Higher divergent Chilean, 17 Cape Verde, Argentinean, and Mexican haplotypes were removed from the network 18 analysis, as they were not included in the 95% Confidential Interval of the parsimony 19 algorithm in TCS software. Only the closest Chilean haplotypes are shown. Each 20 connection represents a single nucleotide difference. Black circles correspond to 21 unsampled or missing haplotypes. Haplotypes obtained in this study are labelled HAf01 22 to HAf10. GenBank haplotypes are labelled with their ARC code or corresponding 23 Accession number. The geographic origin is indicated next to each haplotype. For 24 population codes see Table 1. Haplotypes sharing one or more Mediterranean 25 populations are labelled as Med.

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Figure 3: Principal Component Analysis computed by PCA-GEN software plotted with the two main axes. Populations enclosed within lines were identified to have NO significant  $F_{ST}/Dest$  values in genetic differentiation analyses, but they group into the same cluster in a mutation-migration-drift equilibrium model (i.e., BAPS). Population codes (in the left column next to the figure) and points (inside the figure) sharing the same colour come from the cluster analysis computed by BAPS. A map of sampling sites is also included (see Table 1 for population codes and geographic coordinates).

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#### Tables

3 4 5 Table 1: Artemia franciscana populations sampled for this study, population codes, geographic

coordinates, and sampling date. Population codes are listed by alphabetical order within each

country. Sequences for additional populations from the Americas range used in the phylogenetic

analyses were obtained from GenBank (see methods).

Logelitz	Cala Istitada		Longitud	<u>G 1' 1. (</u>		
Locanty	Code	Latitude	Longitude	Sampling date		
Native						
San Francisco Bay, California	SFB	37°39' N	122°25' W			
Great Salt Lake, Utah	GSL	40°45' N	111°54' W			
Non-native						
Alcochete, Portugal	ALC	38°44' N	08°58' W	2004		
Esmolas, Aveiro, Portugal	ESM	40°39' N	08°41' W	1991		
Bonfim, Portugal	BFI	38°24' N	08°34'01" W	1996		
Bella Mandil, Portugal	BMA	37°01' N	07°52' W	2005		
Cerro Bufo, Portugal	CBU	37°13' N	07°26' W	2002		
F.M. Vontade, Portugal	FVO	37°00' N	07°54' W	1987		
Rio Frio, Portugal	RFR	38°24' N	08°34' W	1993		
Santa Luzia, Tavira, Portugal	SLU	37°06' N	07°38' W	2004		
Fuente de Piedra, Málaga, Spain	FPI	37°06' N	04°45' W	2007		
Gerri de Sal, Lleida, Spain	GER	42°20' N	01°04' E	2004		
La Tapa, Cádiz, Spain	LTA	36°36' N	06°13' W	2004		
San Pascual, Cádiz, Spain	SPA	36°30' N	06°09' W	2003		
Trinitat, Ebro Delta, Tarragona, Spain	TRI	40°35' N	00°40' E	2004		
Aigües Mortes, France	AIG	43°34' N	04°11' E	2002		
Saillé-Guérande, France	SGU	47°20' N	02°26' W	2007		
Margherita di Savoia, Italy	MSA	41°22' N	16°05' E	2004		
Pedra de Lume, Sal Island, Cape Verde	PLU	16°46' N	22°53' W	2005		

Table 2: Mitochondrial diversity for Artemia franciscana for a 477 bp COI fragment from the two native, Cape Verde (PLU), and the 16 non-native Mediterranean populations utilized

in this study. Note that PLU<sup>†</sup> population has an uncertain origin. (n) = number of haplotypes per population;  $\hat{H}$  = gene diversity;  $\pi$  = nucleotide diversity; A = standardized allelic

richness; N = number of individuals analyzed per population. Bold numbers indicate the two main native haplotypes. Bold and italic numbers indicate haplotypes found exclusively in

non-native Mediterranean populations. Asterisks indicate those non-native populations with higher standardised mitochondrial diversity than native ones. DDBJ Acc. No = DNA Data

Bank of Japan Accession Number.

	Haplotype #							Diversity						
Locality (n)	HAf01	HAf02	HAf03	HAf04	HAf05	HAf06	HAf07	HAf08	HAf09	HAf10	H	π	A	N
Native														
SFB (4)	6	4	0	26	0	0	0	0	1	0	0.48	0.0019	1.877	37
GSL (5)	2	23	2	1	0	0	0	0	0	1	0.37	0.0022	1.847	29
Non-native														
ALC (2)	0	1	0	11	0	0	0	0	0	0	0.17	0.0007	0.833	12
ESM (3)	6	3	1	0	0	0	0	0	0	0	0.60	0.0033	2.000*	10
BFI (1)	0	0	0	10	0	0	0	0	0	0	0.00	0.0000	0.000	10
BMA (1)	0	0	0	12	0	0	0	0	0	0	0.00	0.0000	0.000	12
CBU (1)	0	0	0	6	0	0	0	0	0	0	0.00	0.0000	N.C.	6
FVO (2)	0	0	0	7	0	0	0	1	0	0	0.25	0.0005	N.C.	8
RFR (1)	0	0	0	5	0	0	0	0	0	0	0.00	0.0000	N.C.	5
SLU (1)	0	0	0	12	0	0	0	0	0	0	0.00	0.0000	0.000	12
FPI (3)	0	1	0	7	5	0	0	0	0	0	0.60	0.0025	1.759	13
GER (2)	16	0	0	1	0	0	0	0	0	0	0.12	0.0005	0.588	17
LTA (4)	2	4	0	3	3	0	0	0	0	0	0.80	0.0033	2.985*	12
SPA (3)	9	4	1	0	0	0	0	0	0	0	0.54	0.0029	1.713	14
TRI (2)	0	13	0	5	0	0	0	0	0	0	0.42	0.0018	0.993	18
AIG (1)	0	0	0	11	0	0	0	0	0	0	0.00	0.0000	0.000	11
SGU (1)	0	0	0	16	0	0	0	0	0	0	0.00	0.0000	0.000	16
MSA (1)	0	0	0	16	0	0	0	0	0	0	0.00	0.0000	0.000	16
PLU <sup>†</sup> (2)	0	0	0	0	0	15	1	0	0	0	0.12	0.0003	0.625	16

**DDBJ Acc. No** \*AB859230 \*AB859231 \*AB859232 \*AB859233 AB859234 \*AB859235 \*AB859236 AB859237 \*AB859238 \*AB859239

Due to shutdown occurred in U.S.A. and the stop of PubMed service, we followed the *Hydrobiologia* Editor's suggestion to send our sequences to DDBJ database. Asterisks (\*) correspond to haplotypes with identical nucleotide sequence, but different length, to GenBank Acc No as follow: AB859230 = KF662968; AB859231 = KF662970; AB859232 = KF662971; AB859233 = KF662960; AB859235 = KF663036; AB859236 = KF663043; AB859238 = KF662975; and AB859239 = KF662977.

1 **<u>Table 3:</u>** Genetic characteristics of each native and Mediterranean sampled site for the four *Artemia franciscana* microsatellites used. N = number of

2 individuals;  $H_0$  = observed heterozygosity;  $H_E$  = expected heterozygosity; P = p-value of exact test using Markov Chain Monte Carlo with a confidence

3 interval of 95%; Na = number of alleles (pa = number of private alleles). Bold numbers indicate significant departure from HWE after sequential

4 Bonferroni correction (*p*-value = 0.0083). Non-native populations marked with the symbol ‡ indicate those with higher mean Na values than native

5 populations.

_											Loci											
Locality			Af_A108	;				Af_B10					Af_B9					Af_B11	l		Me	an
	Ν	$H_O$	$H_E$	Р	Na (pa)	Ν	$H_O$	$H_E$	Р	Na (pa)	Ν	$H_O$	$H_E$	Р	Na (pa)	Ν	$H_O$	$H_E$	Р	Na (pa)	$H_E$	Na
Native																						
SFB	27	0.296	0.797	0.000	7	42	0.357	0.355	0.204	8 (2)	38	0.579	0.923	0.000	20	34	0.882	0.935	0.861	19	0.753	13.5
GSL	44	0.705	0.724	0.225	9	44	0.818	0.817	0.598	9	40	0.825	0.903	0.285	17	37	0.865	0.944	0.015	21	0.847	14.0
Non-native																						
ALC	35	0.486	0.813	0.000	7	43	0.558	0.694	0.232	7	37	0.676	0.929	0.001	18 (1)	41	0.805	0.922	0.152	14	0.840	11.5
ESM	45	0.667	0.655	0.924	9	44	0.750	0.737	0.659	10	45	0.889	0.930	0.366	21 (1)	44	0.841	0.948	0.029	24 (1)	0.818	16.0‡
BFI	30	Ν	Ionomorphi	ic	1	44	0.477	0.463	0.872	3	32	0.531	0.833	0.009	9	42	0.738	0.734	0.242	9	0.507	5.2
BMA	46	0.630	0.568	0.258	6	46	0.761	0.709	0.384	7	46	0.870	0.875	0.937	21 (1)	46	0.804	0.855	0.235	15	0.752	12.2
CBU	41	0.610	0.655	0.576	8	40	0.725	0.681	0.807	8	32	0.875	0.890	0.854	13	38	0.868	0.938	0.007	22	0.791	12.7
FVO	45	0.578	0.537	0.611	6	45	0.444	0.709	0.000	5	42	0.929	0.928	0.138	21 (1)	43	0.605	0.828	0.000	10	0.751	10.5
RFR	36	Ν	Ionomorphi	ic	1	44	0.386	0.446	0.195	3	0	N.A.	N.A.	N.A.	0	44	0.659	0.615	0.799	9	0.265	3.0
SLU	42	0.476	0.601	0.032	7	42	0.786	0.749	0.519	9	38	0.947	0.934	0.165	22 (1)	41	0.829	0.924	0.281	20	0.802	14.5‡
FPI	39	0.590	0.803	0.000	8	40	0.675	0.652	0.819	6	40	0.800	0.871	0.671	12	38	0.842	0.925	0.118	18	0.813	11.0
GER	43	0.814	0.751	0.954	12(1)	43	0.651	0.718	0.628	9	37	0.811	0.924	0.039	20	34	0.823	0.942	0.008	22	0.834	15.7‡
LTA	38	0.500	0.780	0.000	8	39	0.769	0.686	0.504	7	37	0.784	0.850	0.827	14	30	0.900	0.905	0.398	15	0.805	11.0
SPA	42	0.643	0.745	0.669	8	43	0.674	0.703	0.385	10(1)	44	0.727	0.936	0.027	23 (1)	41	0.805	0.954	0.056	27 (1)	0.835	17.0‡
TRI	42	0.357	0.692	0.000	5	47	0.468	0.553	0.499	5(1)	42	0.476	0.812	0.000	9	45	0.844	0.910	0.058	17	0.742	9.0
AIG	29	0.000	0.133	0.000	3	36	0.472	0.469	1.000	3	33	0.788	0.868	0.070	13	32	0.812	0.8649	0.369	12(1)	0.583	7.7
SGU	28	0.321	0.675	0.000	4	29	0.034	0.034	1.000	2	23	0.304	0.907	0.000	11	29	0.828	0.829	0.202	9	0.611	6.5
MSA	32	0.125	0 569	0.000	3	41	0 195	0.182	1.000	3	37	0 784	0.860	0 308	12	40	0.775	0 781	0.521	9	0 598	67
TOTAL	684	0.120	01007	0.000	13	752	0.170	0.102	1.000	15	643	5.70	0.000	0.000	37	699	01770	0.701	0.021	40	0.0270	0.7

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**Table 4:** A) Pairwise population matrix of G-Statistics Analysis for *Dest* and *Fst* values from nuclear loci above and below the diagonal, respectively,2calculated by GenAlEx for 17 *Artemia franciscana* populations (two native from U.S.A. and 15 non-native from the Mediterranean. Values for the3Portuguese population, RFR, could not be calculated, as only three loci were available. Values with NO statistical significance (*p*-value >0.05) are4shown in bold and italics. B)  $\phi_{ST}$  values from mitochondrial COI locus for 18 *A. franciscana* populations (two native from U.S.A. and 16 non-native5from the Mediterranean). Population codes are those indicated in Table 1. Values in bold are statistically significant (*p*-value < 0.05).</td>

<u>A)</u>																		
	SFB	GSL	ALC	ESM	BFI	BMA	CBU	FVO	RFR	SLU	FPI	GER	LTA	SPA	TRI	AIG	SGU	MSA
SFB	-	0.593	0.092	0.703	0.644	0.700	0.761	0.573	NA	0.708	0.427	0.666	0.501	0.228	0.143	0.333	0.115	0.119
GSL	0.076	-	0.390	0.021	0.863	0.140	0.085	0.140	NA	0.034	0.315	0.036	0.236	0.173	0.564	0.786	0.720	0.782
ALC	0.020	0.042	-	0.052	0.636	0.066	0.508	0.431	NA	0.057	0.049	0.047	0.056	0.020	0.030	0.082	0.055	0.062
ESM	0.094	0.008	0.462	-	0.911	0.120	0.046	0.197	NA	0.014	0.051	0.006	0.042	0.212	0.083	0.162	0.787	0.158
BFI	0.166	0.177	0.140	0.194	-	0.190	0.886	0.870	NA	0.187	0.181	0.185	0.191	0.177	0.195	0.123	0.200	0.213
BMA	0.106	0.023	0.497	0.022	0.767	-	0.086	0.086	NA	0.014	0.053	0.021	0.046	0.336	0.107	0.159	0.158	0.170
CBU	0.108	0.016	0.061	0.012	0.200	0.019	-	0.173	NA	0.009	0.045	0.011	0.038	0.042	0.094	0.175	0.158	0.174
FVO	0.094	0.024	0.059	0.032	0.210	0.020	0.032	-	NA	0.022	0.049	0.034	0.046	0.038	0.094	0.165	0.141	0.151
RFR	NA	NA	NA	NA	NA	NA	NA	NA	-	NA	NA	NA	NA	NA	NA	NA	NA	NA
SLU	0.099	0.010	0.486	0.008	0.840	0.054	0.021	0.106	NA	-	0.046	0.012	0.036	0.275	0.092	0.164	0.811	0.162
FPI	0.064	0.038	0.422	0.418	0.820	0.356	0.323	0.315	NA	0.340	-	0.049	0.012	0.315	0.064	0.132	0.571	0.113
GER	0.087	0.010	0.423	-0.000	0.885	0.116	0.035	0.213	NA	0.050	0.4115	-	0.312	0.236	0.079	0.158	0.752	0.154
LTA	0.074	0.031	0.476	0.321	0.861	0.292	0.253	0.285	NA	0.250	0.041	0.040	-	0.292	0.075	0.146	0.645	0.132
SPA	0.036	0.023	0.132	0.028	0.838	0.048	0.316	0.250	NA	0.036	0.039	0.030	0.038	-	0.037	0.118	0.070	0.083
TRI	0.032	0.074	0.172	0.591	0.773	0.665	0.630	0.570	NA	0.632	0.425	0.579	0.501	0.229	-	0.130	0.243	0.065
AIG	0.085	0.142	0.393	0.862	0.313	0.723	0.887	0.752	NA	0.841	0.662	0.864	0.730	0.612	0.550	-	0.352	0.369
SGU	0.037	0.126	0.251	0.143	0.600	0.761	0.828	0.658	NA	0.152	0.112	0.134	0.126	0.343	0.064	0.116	-	0.075
MSA	0.036	0.137	0.293	0.872	0.641	0.816	0.914	0.699	NA	0.863	0.572	0.871	0.670	0.418	0.251	0.121	0.213	-

**B**)

	SFB	GSL	ALC	ESM	BFI	BMA	CBU	FVO	RFR	SLU	FPI	GER	LTA	SPA	TRI	AIG	SGU	MSA
SFB	-	0.546	0.017	0.433	0.088	0.100	0.046	0.078	0.027	0.100	0.149	0.556	0.241	0.446	0.415	0.094	0.312	0.100
GSL		-	0.653	0.265	0.717	0.728	0.689	0.697	0.680	0.728	0.486	0.624	0.217	0.300	0.049	0.723	0.233	0.728
ALC			-	0.611	-0.016	-0.000	-0.069	-0.017	-0.093	-0.000	0.224	0.828	0.365	0.620	0.544	-0.008	0.380	-0.000
ESM				-	0.705	0.728	0.643	0.658	0.622	0.728	0.337	0.209	0.084	-0.091	0.289	0.717	0.294	0.728
BFI					-	0.000	0.000	0.029	0.000	0.000	0.308	0.921	0.460	0.702	0.643	0.000	0.469	0.000
BMA						-	0.000	0.053	0.000	0.000	0.335	0.927	0.489	0.721	0.661	0.000	0.493	0.000
CBU							-	-0.040	0.000	0.000	0.237	0.907	0.387	0.652	0.596	0.000	0.408	0.000
FVO								-	-0.069	0.053	0.269	0.892	0.417	0.665	0.610	0.042	0.433	0.053
RFR									-	0.000	0.212	0.903	0.362	0.636	0.581	0.000	0.387	0.000
SLU										-	0.335	0.927	0.489	0.721	0.661	0.000	0.493	0.000
FPI											-	0.555	0.053	0.368	0.361	0.322	0.267	0.335
GER												-	0.394	0.172	0.649	0.924	0.579	0.927
LTA													-	0.119	0.122	0.475	0.137	0.489
SPA														-	0.324	0.711	0.331	0.721
TRI															-	0.652	0.117	0.661
AIG																-	0.481	0.000
SGU																	-	0.493
MSA																		-

### Figure 1 Click here to download Figure: FigNEW. 1.ppt





Figure 3 Click here to download high resolution image



PCA 2 (22.08%)