

1 **Grassland management in agricultural vs. forested landscapes drives butterfly and bird**
2 **diversity**

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13 **Abstract**

14 Calcareous grasslands and orchard meadows are among the most species-rich semi-natural
15 habitats in Europe, but they are severely threatened by intensified land use and abandonment.
16 Here, we focus on the effects of management vs. abandonment of these grasslands in
17 agricultural vs. forest-dominated landscapes of Germany. We recorded butterflies and birds
18 and classified them in farmland and woodland species according to their habitat preferences.
19 Species richness and abundance of farmland butterflies were higher on calcareous grasslands
20 than orchard meadows and benefited from forested landscapes in case of orchard meadows.
21 Species richness of woodland butterflies was higher on abandoned than managed grasslands,
22 independent of habitat type and landscape context. Richness and abundance of farmland birds
23 benefited from managed orchard meadows, and were more abundant in agricultural
24 landscapes. On calcareous grasslands, however, the abandonment led to higher richness and
25 abundance of farmland birds. Woodland birds exhibited higher species richness in abandoned
26 than managed grasslands, especially in orchard meadows. Woodland birds and butterflies
27 appeared to be less affected by habitat type, management or landscape context than farmland
28 species. Calcareous grasslands were much more important for butterfly diversity than orchard
29 meadows, but suitability of orchards for butterflies was improved when embedded in forested
30 landscapes. In contrast to butterflies, bird diversity benefited more from orchard meadows
31 than calcareous grasslands, which had higher diversity when management was abandoned. In
32 conclusion, landscape context can shape communities in these two grassland habitat types, so
33 conservation management should consider reserves in both agricultural and forested
34 landscapes and thereby, diversify regional biota.

35

36 **Keywords:** biodiversity; extensive management; farmland species; habitat abandonment;
37 species richness; woodland species

38 **1. Introduction**

39 Land use change by agricultural intensification has led to a widespread biodiversity loss in
40 human-dominated landscapes causing reduced habitat area and increased habitat degradation
41 of natural and semi-natural habitats (Fischer and Lindenmayer, 2007). At small spatial scales,
42 vegetation heterogeneity originates from different management practices and the habitat type,
43 whereas at larger scales, landscape heterogeneity contributes to the local community structure
44 (Stein et al., 2014).

45 Due to their extensive management and high structural diversity, semi-natural habitats
46 such as orchard meadows and calcareous grasslands are among the most species rich habitats
47 in Western Europe and important for butterfly and bird conservation (van Swaay, 2002;
48 Herzog et al., 2005). Orchard meadows are characterised by sparse, old, tall fruit trees,
49 species-rich herbaceous vegetation and greatly differ from intensively managed fruit
50 plantations as trees have a heterogeneous spatial pattern (Mycko et al., 2013). However,
51 extensively managed semi-natural grassland often faces two contrasting trajectories. On the
52 one hand, traditional low-intensity management such as extensive grazing, hay making and
53 fruit harvesting is often abandoned, resulting in massive regeneration of shrubs and ultimately
54 the loss of these habitats (Poschlod and WallisDeVries, 2002; Plieninger et al., 2015). On the
55 other hand, management intensification and land conversion also threatens them (Stoate et al.,
56 2009; Plieninger et al., 2015). Species respond differently to changing environmental
57 variables. For example van Swaay et al. (2006) identified agricultural intensification, such as
58 conversion of grassland into arable land, the fertilisation of grassland as well as abandonment
59 as main threats to butterflies. However, in early succession stages after abandonment insect
60 species richness might increase, but overall habitat quality is changing over time, leading to
61 the decline of specialist species (Balmer and Erhardt, 2000; Kormann et al., 2015).

62 In order to understand the effects of management practices on biodiversity, it is
63 important to consider the landscape-scale (Tschardt et al., 2012). According to the theory of

64 island biogeography, species diversity of fragmented semi-natural habitats such as calcareous
65 grasslands and orchard meadows is influenced by movement between habitat fragments
66 through the surrounding landscape matrix (McArthur and Wilson, 1967). Landscape structure
67 and land use surrounding the habitat fragments may therefore improve or hinder dispersal
68 through the landscape matrix or even provide additional resources to some species, depending
69 on the quality of the matrix and the species' environmental needs (Eycott et al., 2012;
70 Öckinger et al., 2012). Thus, landscape context measures might reflect the quality of the
71 landscape matrix for population movement between the remaining habitat fragments and
72 availability of additional resources.

73 Both butterflies and birds are representatives of the most studied and ecologically best
74 known invertebrates and vertebrates (Schlegel and Rupf, 2010). In particular, butterflies
75 which have been categorized as grassland specialists have been found to decline in
76 distribution across Europe (van Swaay et al., 2006). Birds, on the other hand, often rely on the
77 presence of scattered trees such as fruit trees in orchard meadows, which act as local and
78 landscape keystone structures in intensively managed landscapes that are otherwise poor in
79 landscape elements (Manning et al., 2006). Overall, population declines of birds over the last
80 decades have been widely reported, especially of farmland birds, but of woodland birds as
81 well (Gregory et al., 2005, 2007).

82 This is the first study investigating the potentially complex effects of changing
83 environments and their interactions at three spatial scales: management practices
84 (mowing/grazing or abandonment), habitat type (calcareous grassland or orchard meadow)
85 and landscape context (agricultural or forested landscape) on two flagship taxa of nature
86 conservation. We focused on butterflies and birds, which were classified as farmland or
87 woodland species according to their known principal habitat occurrence. We hypothesise that
88 (i) farmland butterflies and birds prefer calcareous grasslands, whereas woodland species
89 prefer orchard meadows, (ii) regular local management of calcareous grasslands and orchard

90 meadows positively affects farmland species, whereas abandonment benefits woodland
91 species, and (iii) there are more farmland species in agricultural landscapes, whereas there are
92 more woodland species in forest-dominated landscapes (Fig. 1).

93

94 **2. Material and methods**

95 *2.1. Study area*

96 The study area was situated in southern Lower Saxony (Germany) in the districts of Göttingen
97 and Northeim (about 1000 km²; see Appendix A1 in Supplementary online material). The
98 main land use types are arable fields, intensively used meadows and semi-natural deciduous
99 forests. The surveys were conducted in calcareous grasslands (*Mesobrometum erecti* Koch
100 1926) covering only 0.26 % and in orchard meadows (*Arrhenatheretum elatioris* Braun 1915)
101 covering 0.39 % of the study area. Both semi-natural grassland habitats are patchily
102 distributed across the landscape, and managed either by mowing or grazing with sheep, goats,
103 cattle or horses. Many smaller fragments and party also the larger ones are in a process of
104 abandonment with shrub encroachment and/or dye off of old fruit trees.

105

106 *2.2. Study design*

107 We surveyed butterfly and bird species in 20 orchard meadows and 20 calcareous grasslands
108 in a full factorial design (mean \pm SEM distance between sites: 17.9 ± 0.3 km; range of
109 distance between sites: 0.5 – 52.0 km; fruit tree density on abandoned orchard meadows: 38.6
110 ± 4.4 (23.9 – 67.9) and on managed orchard meadows: 36.4 ± 5.5 (16.4 – 70.5)). The habitat
111 fragments were selected according to differences in landscape context (forested vs. agriculture
112 dominated landscapes) and management status (managed vs. abandoned), resulting in five
113 replicates per treatment (Fig. 1). Within a 500 m buffer area around each habitat fragment,
114 forest-dominated landscapes had $44 \pm 2\%$ (mean \pm SEM) forest cover ranging from 28 to
115 63%, whereas agricultural landscapes had $14 \pm 2\%$ forest cover ranging from 0 to 28% (forest

116 cover was measured in ArcGIS 10.4). As many semi-natural habitats are neither fully
117 managed nor completely abandoned, we selected managed habitat fragments to be managed
118 each year by grazing or mowing, whereas abandoned fragments to be irregularly or not
119 managed characterised by high degree of succession to woody shrubs or dead wood. In
120 summary, majority of the managed grasslands were grazed extensively (< 1 LUI/ha) between
121 May and September with different livestock including cattle, sheep, goat, horse or donkey
122 (calcareous grasslands: eight fragments grazed, one mown and one both grazed and mown;
123 orchard meadows: eight fragments grazed and two mown). In order to minimize the effect of
124 habitat size on species richness and abundance, fragments with a similar size were chosen,
125 and species were surveyed on a 0.8 ha patch in each study site. The area of the selected
126 habitat fragments was 2.64 ± 0.27 ha (mean \pm SEM) for calcareous grassland (ranging from
127 0.90 ha to 5.38 ha) and 1.45 ± 0.15 ha for orchard meadows (ranging from 0.85 ha to 3.34 ha).

128

129 *2.3. Sampling methods*

130 Butterflies (Lepidoptera: Hesperidae and Papilionidea) and burnet moths (Lepidoptera:
131 Zygaenidae) were sampled from 24th of May until 19th of August 2015 with three survey
132 rounds (roughly one survey/month) by a 20 minute zig-zag transect-walk (split into 5 four-
133 minute sections) once on each habitat fragment (following Krauss et al., 2003; Brückmann et
134 al., 2010). Butterflies were surveyed visually or using a butterfly net between 10.45 am and
135 5.30 pm, and were identified and released immediately. Surveys were conducted on a 5 m
136 wide corridor under suitable weather conditions for butterfly activity (dry conditions, wind
137 speed less than Beaufort scale 5, and temperature 13 °C or higher if there was at least 60 %
138 sunshine, or more than 17 °C if overcast (Pollard, 1977)). To characterise the availability of
139 nectar resources, the percent cover of flowering plants inside the transect corridor was
140 estimated at the end of each transect walk. We classified butterfly species to farmland or
141 woodland species based on literature (van Swaay et al., 2006; Plattner et al., 2010).

142 We recorded birds between 8th and 22nd of May and between 8th and 28th of June 2015
143 in two survey rounds by a 12 minute point-count on 0.8 ha patches half an hour after sunrise
144 until 4 hours after sunrise under calm and dry weather conditions (Bibby et al., 1992). There
145 were 22 habitat fragments with two 0.8 ha survey patches as they were larger than 1.6 ha. In
146 the 18 remaining, smaller habitat fragments there was one 0.8 ha survey patch in each habitat
147 fragment. Due to the high degree of heterogeneous structures and the different shape of the
148 habitat fragments, each point-count was split into three 4-minute sections placed at points
149 suitable to represent the study design (managed or abandoned). This guaranteed to perceive
150 all acoustic signals of the birds and to detect them visually. To characterise the availability of
151 nesting and foraging sites, the percent bush cover in each 0.8 ha study patch was estimated in
152 the end of each survey. We classified bird species to farmland or woodland species based on
153 literature (Gregory et al., 2005, 2007; Südbeck et al., 2005; Batáry et al., 2012). Species
154 habitat affinity and specialism might change with different European regions suggesting that
155 our classification approach may lead to different results there (see e.g. Koleček et al., 2010).

156

157 *2.4. Statistical analysis*

158 Abundance of butterflies was summed over transects and sampling occasions. For each bird
159 species we pooled the data using the maximum abundance of the two survey rounds per patch.
160 Species richness of birds was calculated as the number of species that were present in the
161 particular sampling patch at least in one survey round.

162 For both taxa we applied linear regression models for analysing the species richness and
163 abundance of farmland and woodland species. Habitat type (calcareous grassland vs. orchard
164 meadow), management status (managed vs. abandoned), landscape context (forested vs.
165 agriculture dominated landscapes) and their two-way interactions were used as explanatory
166 design variables. In case of bird models, the survey patch within habitat fragment was used as
167 random factor. Models were fitted with Poisson distribution or in case of overdispersion with

168 negative binomial distribution using the MASS (for butterflies, Venables and Ripley, 2002)
169 and lme4 packages (for birds, Bates et al., 2015) of R software (R Development Core Team,
170 2017). We calculated all models nested in the global model by the command ‘dredge’ in the
171 package MuMIn (Barton, 2016), and compared them based on Akaike Information Criterion
172 corrected for small sample size (AICc). We performed model averaging (Burnham and
173 Anderson, 2002), if the top model and subsequent models differed less than two units in
174 AICc. Model-averaged parameter estimates were calculated over the subset of models
175 including the parameter (conditional average) to avoid shrinkage towards zero (Grueber et al.,
176 2011). We present the 95% confidence intervals (CI) of parameter estimates and the relative
177 importance of each parameter. Relative importance is 100%, when the parameter is present in
178 all top models.

179 We also performed further linear regression models to test for effects of explanatory
180 design variables on percent flowering plants and percent bush cover (both normal
181 distribution), potentially important for butterflies and birds, respectively. The percent
182 flowering plants was significantly higher in calcareous grasslands than in orchard meadows
183 (Table A1; Fig. A1a). The percent bush cover was mainly determined by the management
184 with about three times higher cover of bushes in abandoned than in managed sites (Fig. A1b).
185 Nevertheless bush cover was also significantly higher in calcareous grasslands than in orchard
186 meadows and in agricultural than in forested landscapes.

187 Furthermore, we applied redundancy analyses (RDA) to assess the variability in species
188 composition of butterfly and bird communities explained by the environmental variables
189 habitat type, management status and landscape context. For the bird analysis we included
190 habitat patch as conditional variable as the study design was nested. The results were
191 presented in ordination biplots to visualise the variability in species composition. Prior to
192 analyses, community data matrices were Hellinger-transformed (Legendre and Gallagher,

193 2001). To assess for statistical significance, a permutation test based on 999 permutations was
194 calculated using the package *vegan* (Oksanen et al., 2017).

195

196 **3. Results**

197 On the 20 calcareous grassland and 20 orchard meadow fragments we recorded 5182
198 individual butterflies belonging to 55 butterfly species (seven of them burnet moths, hereafter
199 also called butterflies) and 1075 individuals of 55 bird species. Classification based on
200 environmental preferences resulted in 35 farmland butterfly species with 3973 individuals and
201 20 woodland butterfly species with 1209 individuals (Table A2) as well as 22 farmland bird
202 species with 272 individuals and 33 woodland bird species with 803 individuals (Table A3).

203 The most abundant farmland butterflies were *Maniola jurtina*, *Polyommatus coridon* and
204 *Melanargia galathea*, whereas the most abundant woodland butterflies were *Aphantopus*
205 *hyperantus*, *Pieris napi* and *Coenonympha arcania* (Table A2). For birds, the most abundant
206 farmland species were *Emberiza citrinella*, *Sylvia communis* and *Columba palumbus*, whereas
207 the most abundant woodland birds were *Parus major*, *Turdus merula* and *Cyanistes caeruleus*
208 (Table A3).

209

210 *3.1. Effects on butterflies*

211 We found habitat type to be the most important factor determining farmland butterfly species
212 richness and abundance with higher values in calcareous grasslands than in orchard meadows
213 (Table 1; Fig. 2a,c). Farmland species richness and abundance depended on an interaction
214 between landscape context and habitat type; high species richness and abundance were found
215 in both agricultural and forest-dominated landscapes of calcareous grasslands, but lower
216 values in orchard meadows with a decrease from forested to agricultural landscapes.

217 Additionally, farmland butterfly abundance was influenced by management in interaction
218 with habitat type. Management increased butterfly abundances in calcareous grasslands, but

219 decreased them in orchard meadows. In contrast, abandonment increased the abundance of
220 woodland butterflies, but not their species richness (Fig. 2c,d).

221 The RDA of butterfly community composition revealed significant associations with
222 habitat type and landscape context (Table 3; Fig. 3a). In the ordination biplot, the first axis
223 separated calcareous grasslands from orchard meadows with e.g. chalkhill blue (*Polyommatus*
224 *coridon*) as characteristic species in calcareous grasslands and ringlet (*Aphantopus*
225 *hyperantus*) as characteristic species in orchard meadows. The second axis separated
226 agricultural from forest-dominated landscapes with small white (*Pieris rapae*) being a
227 characteristic agricultural species and small skipper (*Thymelicus sylvestris*) being a species
228 associated with forest.

229

230 3.2. Effects on birds

231 Performing generalized linear mixed effects models on birds, we found that management type
232 was the variable that most strongly explained both farmland and woodland species richness
233 and abundance (Table 2). This was, however, in an interaction with habitat type in case of
234 farmland species (Fig. 4a,c). Farmland birds preferred managed over abandoned fragments in
235 orchard meadows, and abandoned over managed fragments in calcareous grassland.

236 Additionally, they were more abundant in agricultural than forest-dominated landscapes. In
237 contrast, woodland birds (both richness and abundance) were more common in abandoned
238 than in managed fragments (Fig. 4b,d). Finally, woodland bird abundance was higher in
239 orchard meadows than in calcareous grasslands.

240 In the RDA of bird communities all three variables explained a significant part of the
241 variation in community composition (Table 3). Landscape context explained the smallest part,
242 followed by management and habitat type, explaining the largest part of the variation. The
243 first axis separated in particular abandonment and management, but also orchard meadows
244 and calcareous grasslands (Fig. 3b). For example, chiffchaff (*Phylloscopus collybita*) showed

245 a preference for abandoned orchard meadows, whereas tree pipit (*Anthus trivialis*) was a
246 characteristic species of managed calcareous grassland. The second axis separated agricultural
247 from forest-dominated landscapes with green woodpecker (*Picus viridis*) as characteristic
248 woodland species in grassland fragments of forest-dominated landscapes and yellowhammer
249 (*Emberiza citrinella*) being a characteristic farmland species in grasslands of agricultural
250 landscapes.

251

252 **4. Discussion**

253 We studied the effects of habitat type (calcareous grassland vs. orchard meadow),
254 management (managed vs. abandoned) and landscape context (forested vs. agricultural
255 landscape) in a full factorial design and found that the classification into farmland and
256 woodland traits helps to identify key factors of diversity and abundance patterns for
257 conservation management strategies. Farmland butterflies were more diverse in calcareous
258 grasslands than farmland birds, which exhibited higher species richness in orchard meadows.
259 Woodland butterfly and bird abundance increased with abandonment, whereas regular
260 management affected farmland butterflies in calcareous grassland positively. Surprisingly,
261 landscapes dominated by forest had a positive effect on farmland butterfly richness and
262 abundance, but not on woodland butterflies. Farmland bird abundance was higher in
263 agricultural landscapes, while woodland bird diversity and abundance benefited from
264 abandonment.

265

266 *4.1. Effects on butterflies*

267 Supporting our first hypothesis, species richness and abundance of farmland butterflies was
268 highest in calcareous grasslands. Management such as mowing and grazing leads to high
269 cover of flowering plants as feeding and reproduction resources. This positive relationship has

270 often been reported (e.g. Krämer et al., 2012), and highlights the disproportionate high value
271 of calcareous grassland for butterfly conservation (van Swaay, 2002).

272 As hypothesised, management had a negative effect on woodland butterfly abundance,
273 but not on species richness. Woodland butterflies were more common in abandoned semi-
274 natural grasslands, which can be explained by increasing microhabitat heterogeneity and the
275 availability of plant communities typical for different successional stages (WallisDeVries et
276 al., 2002). On the contrary, farmland butterfly abundance increased with management, but
277 only in calcareous grassland, where abundances were generally higher than in orchard
278 meadows. Abandonment appeared to provide less life-sustaining resources for farmland
279 butterflies such as flowering plants and warm micro-climate (van Swaay, 2002). Surprisingly,
280 in orchard meadows farmland butterfly abundance increased with abandonment. This might
281 have been caused by the fact that management was characterised by high stocking rates,
282 fertilisation and frequent mowing, degrading the diversity of herbs and flowers (Uchida et al.,
283 2016). Abandoned orchard meadows were characterised by additional resources such as
284 flowering forbs or shrubs, for example blackberries, but in the long run, late successional
285 stages may decrease butterfly species richness and abundance (Balmer and Erhardt, 2000;
286 Kesting et al., 2015). There is a lack of target-oriented management in orchard meadows,
287 which should be regularly restored by clearance of shrubs and trees, opening of the canopy for
288 light and warm micro-climate as well as reducing grazing density or intensified hay-making to
289 facilitate larval hosts and nectar-providing plants.

290 In contrast to our hypothesis, farmland butterfly species richness was higher in orchard
291 meadows when embedded in forest-dominated landscapes but not agricultural landscapes.
292 Forest-dominated landscapes are more heterogeneous providing more resources than simple
293 landscapes dominated by agriculture (Öckinger et al., 2012). Compared to calcareous
294 grassland, local habitat conditions in orchard meadows were worse (less food resources) and
295 farmland butterflies appeared to use additional resources in the surroundings (Krämer et al.,

296 2012; Villemey et al., 2015). In this study we found more flowering plants in orchard
297 meadows of forest-dominated than agricultural landscapes, which suggests that non-arable
298 patches may act as buffer against intensive agricultural practices such as chemical weed
299 control (Gonthier et al., 2014; Villemey et al., 2015).

300 As shown in the redundancy analysis, the greatest variability in community composition
301 was explained by habitat type with most butterfly species showing a strong preference for
302 calcareous grasslands especially by farmland species. For example, chalkhill blue is regarded
303 as threatened in the red list of the study region (Lower Saxony, (Lobenstein, 2004)), and was
304 the most characteristic farmland species on calcareous grasslands. The high population
305 density of chalkhill blue is determined by the presence of the larva's food plant *Hippocrepis*
306 *comosa* (Krauss et al., 2005), which is dispersed by the hooves of livestock (Brereton et al.,
307 2008). Hence, this result reflects the need for appropriate habitat management for specialised
308 butterflies in the study region. Contrastingly, the community composition for orchard
309 meadows showed that management can be important habitat for species that are associated
310 with open woodland. For example, ringlet (*Aphantopus hyperatus*) occurred in relatively high
311 abundances in orchard meadows. This species was often shown to be present in grasslands
312 and mixed woodlands, but also in tree lines (van Swaay et al., 2006). Thus, orchard meadows
313 potentially provide habitat for species that are associated with woodland edges and can be
314 assumed to provide habitat to an even wider range of open-woodland butterfly species
315 profiting from improved management practices. Hay-making or low-intensity grazing with
316 reduced fertiliser use and allowance of seed maturation could restore the degraded orchard
317 meadows in the study region.

318

319 4.2. Effects on birds

320 Regarding farmland bird species richness and abundance, our first and second hypotheses
321 were only partly confirmed, because we found an interaction of habitat type and management.

322 Abandonment caused an increase in farmland bird species richness and abundance in
323 calcareous grasslands, but a decrease in orchard meadows. Partly abandoned calcareous
324 grasslands were characterised by less disturbance and provided a wide range of niches,
325 because of their heterogeneous habitat characteristics caused by higher amounts of woody
326 vegetation and heterogeneous sward structures (Hartel et al., 2014). This supported nesting
327 sites and foraging opportunities, e.g. for insects on the ground (Vickery et al., 2001).
328 However, abandonment can benefit farmland birds only on a short term and further
329 succession will exclude farmland birds (Gregory et al., 2007). Contrastingly to calcareous
330 grasslands, farmland bird species richness and abundance were higher in managed compared
331 to abandoned orchard meadows. Scattered trees act as keystone habitat for farmland birds and
332 provide nesting and foraging opportunities as well as song posts (Fischer et al., 2010;
333 Jakobsson and Lindborg, 2017). Since orchard meadows were mostly grazed by livestock,
334 they were suitable for foraging, e.g. of insects on animal dung, or as ground-nesting sites in
335 patches avoided by livestock. Nevertheless, some orchard meadows were frequently used and
336 there might be a higher potential for farmland birds as the positive effect of management on
337 biodiversity may be restricted to low levels of interference. Intensified grassland management
338 decreases the suitability as habitat for feeding and nesting because of higher disturbance
339 levels and a fast growing, homogeneous grassland structure as a consequence (Vickery et al.,
340 2001). Management activities should provide feeding and nesting sites, such as breeding
341 burrows of old trees, shelter of bushes for ground breeding birds and heterogeneous, open
342 sward structures.

343 Corresponding to our hypotheses, woodland bird species richness and abundance were
344 higher in abandoned compared to managed grassland fragments and abundance was also
345 higher in orchard meadows than in calcareous grassland. Abandoned habitat fragments are
346 structurally more similar to forest as they contain a high bush and tree cover. Orchard
347 meadows were characterised by high, old fruit trees representing structurally rich stands

348 important for birds nesting in treetops and hollows (Tworek, 2002), which can be compared
349 with forest structures as well, but being more open. Thus, habitat structural diversity might be
350 reasonable for some parts of the habitat, but probably favours primarily forest species and not
351 characteristic semi-open woodland species. Long-term abandonment should be avoided as
352 orchard meadows would develop into forest.

353 In accordance with the third hypothesis, farmland bird abundance increased in semi-
354 natural grassland located in agricultural landscapes. Similar results were found by Wretenberg
355 et al., (2010) with a positive effect of low-intensity land use on farmland birds in open
356 landscapes with low forest cover. This indicates that farmland birds are using resources from
357 different semi-natural grasslands, but also the surrounding agricultural landscape. Birds
358 experience the landscape at a large scale, which also enables them to react fast to local habitat
359 changes (Tscharntke et al., 2012). Hence, semi-natural grassland can be regarded as a
360 valuable landscape element for landscape-wide conservation management.

361 Analysing the bird community composition, habitat type and management explained the
362 greatest part of its variation. For abandoned orchard meadows, for example, chiffchaff
363 (*Phylloscopus collybita*) was a characteristic woodland species nesting on the ground or in
364 herbaceous woody vegetation structures (Südbeck et al., 2005; Gregory et al., 2007).
365 However, for orchard meadows there are also some semi-open woodland bird species
366 regarded as characteristic due to their ecological requirements (Herzog et al., 2005), but only
367 two of them were found in this study (*Phoenicurus phoenicurus*; *Picus viridis*) and one of
368 them, namely *P. phoenicurus*, only with one individual. This indicates that the ecological
369 requirements of many characteristic species for orchard meadows cannot be fulfilled by the
370 current habitat status, e.g. for ortolan (*Emeriza hortulana*) and hoopoe (*Upupa epops*), which
371 are regarded as threatened in the red list of the study region (Lower Saxony, (Krüger and
372 Nipkow, 2015)). This shows the importance of orchard meadows for a wide range of bird
373 species, but emphasises the urgent need for conservation management to work more target-

374 oriented with land owners. Thus, abandoned orchard meadows should be taken into low-
375 intensity management again, while nest-holes and heterogeneous structures must be preserved
376 at the same time. Another rare open woodland species is tree pipit (*Anthus trivialis*), which is
377 specialised on open semi-natural grassland with single trees and characteristically occurred in
378 managed calcareous grassland of forest-dominated landscapes. High solitary trees are used as
379 perches, and an increasing shrub cover was shown to negatively affect the occurrence
380 (Kumstátová et al., 2004). This suggests that the tree pipit, being in a sharp decline across
381 Europe (Gregory et al., 2007), was favoured by open semi-natural grassland with single
382 perches and would be disadvantaged by abandonment.

383

384 **5. Conclusions**

385 Our results show that the classification of species into farmland and woodland traits can help
386 to disentangle the complex local and landscape effects on butterflies and birds in semi-natural
387 grasslands. Results of this study detail the relative importance of local and landscape
388 management and their complex interaction for understanding and applying best conservation
389 measures. Woodland birds and butterflies appeared to be less affected by habitat type,
390 management or landscape context than farmland species. Calcareous grasslands were much
391 more important for butterfly diversity than orchard meadows, but suitability of orchards for
392 butterflies was improved when embedded in forested landscapes. In contrast to butterflies,
393 bird diversity benefited more from orchard meadows than calcareous grasslands, which had
394 higher diversity when management was abandoned. Hence, short-term abandonment can
395 improve habitats for birds and butterflies, but of course, long-term abandonment would
396 destroy the identity of these openland habitats and their associated community. Landscape
397 context can shape communities in these two grassland habitat types, so conservation
398 management should consider reserves in both agricultural and forest landscapes and thereby,
399 diversify regional biota.

400

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544 **Table 1.** Summary table for generalized linear regression model results on farmland and
 545 woodland butterfly species richness and abundance testing the effects of habitat type (H:
 546 calcareous grassland vs. orchard meadow), management (M: abandoned vs. managed) and
 547 landscape context (L: agricultural vs. forest-dominated) after multimodel averaging of best
 548 candidate models. AB: abandoned, AG: agricultural, CG: calcareous grassland, FO: forest-
 549 dominated, MA: managed, OM: orchard meadow. Significant estimates are in bold characters.

Model ^a	Variable	Relative importance (%) ^b	Multimodel estimate ± 95 % CI ^c			Direction
Species richness						
Farmland	Landscape (L)	100	0.026	±	0.312	
	Habitat (H)	100	-1.217	±	0.415	CG>OM
	Management (M)	100	0.054	±	0.31	
	L × H	100	0.702	±	0.448	
	L × M	100	-0.095	±	0.403	
	H × M	100	-0.042	±	0.436	
Woodland	Landscape (L)	30	0.139	±	0.276	
	Habitat (H)	63	-0.239	±	0.278	
	Management (M)	13	0.099	±	0.276	
Abundance						
Farmland	Landscape (L)	100	0.029	±	0.481	
	Habitat (H)	100	-1.792	±	0.718	CG>OM
	Management (M)	100	0.458	±	0.465	
	L × H	62	0.863	±	0.697	
	H × M	62	0.934	±	0.675	
Woodland	Landscape (L)	20	0.138	±	0.344	
	Habitat (H)	28	-0.198	±	0.341	
	Management (M)	100	-0.554	±	0.345	CG>OM

550 ^a Farmland species richness and abundance and woodland abundance butterfly models were
 551 fitted with negative binomial distribution, whereas woodland species richness with Poisson
 552 distribution

553 ^b Each variable's importance within the best candidate models ($\Delta AIC < 2$)

554 ^c Estimates with 95 % CI values after multimodel averaging of the top-model set ($\Delta AIC < 2$)

555

557 **Table 2.** Summary table for generalized linear mixed-effects model results on farmland and
 558 woodland bird species richness and abundance testing the effects of habitat type (H:
 559 calcareous grassland vs. orchard meadow), management (M: abandoned vs. managed) and
 560 landscape context (L: agricultural vs. forest-dominated) after multimodel averaging of best
 561 candidate models. AB: abandoned, AG: agricultural, CG: calcareous grassland, FO: forest-
 562 dominated, MA: managed, OM: orchard meadow. Significant estimates are in bold characters.

Model ^a	Variable	Relative importance (%) ^b	Multimodel estimate ± 95 % CI ^c			Direction
Species richness						
Farmland	Landscape (L)	100	-0.437	±	0.492	
	Habitat (H)	100	-0.032	±	0.448	
	Management (M)	100	-0.659	±	0.518	AB>MA
	L × H	100	0.09	±	0.612	
	L × M	100	0.022	±	0.61	
	H × M	100	0.853	±	0.602	
Woodland	Landscape (L)	70	-0.043	±	0.329	
	Habitat (H)	52	0.133	±	0.261	
	Management (M)	100	-0.381	±	0.335	CG<OM
	L × H	12	0.286	±	0.418	
	L × M	49	-0.439	±	0.443	
Abundance						
Farmland	Landscape (L)	100	-0.518	±	0.408	AG>FO
	Habitat (H)	100	-0.145	±	0.495	
	Management (M)	100	-0.629	±	0.49	AB>MA
	L × H	28	0.283	±	0.666	
	H × M	100	1.086	±	0.671	
Woodland	Landscape (L)	27	-0.097	±	0.305	
	Habitat (H)	100	0.338	±	0.305	CG<OM
	Management (M)	100	-0.545	±	0.307	AB>MA

563 ^a All models were fitted with Poisson distribution
 564

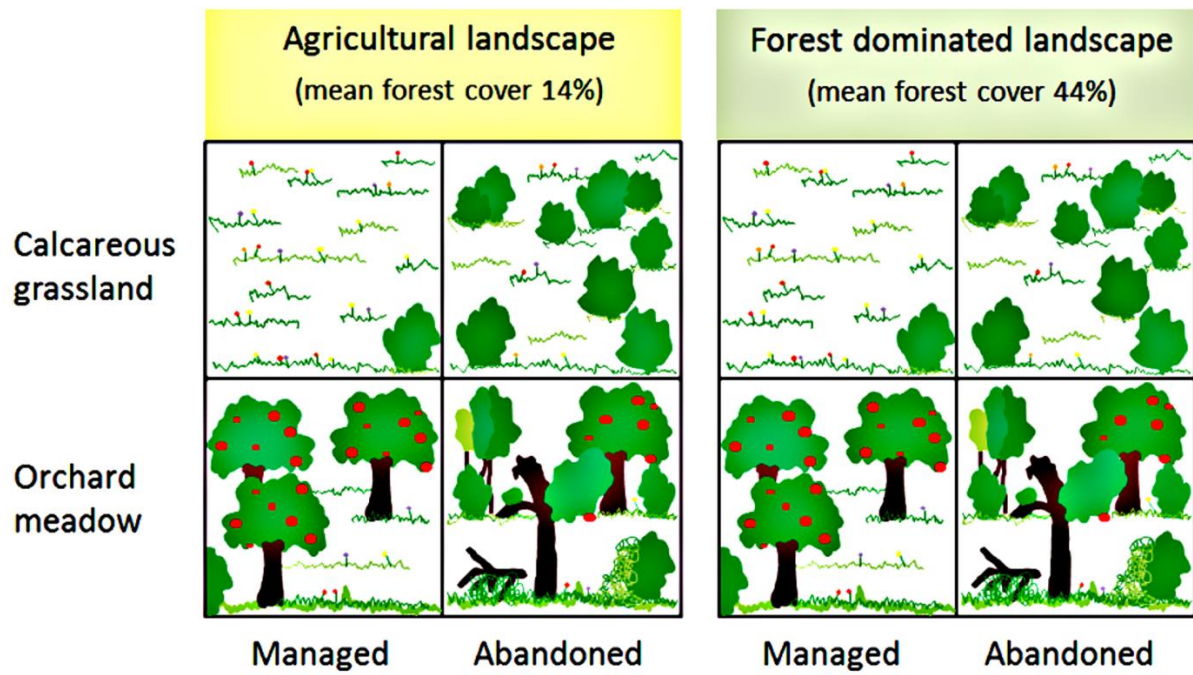
565 ^b Each variable's importance within the best candidate models ($\Delta AIC < 2$)

566 ^c Estimates with 95 % CI values after multimodel averaging of the top-model set ($\Delta AIC < 2$)

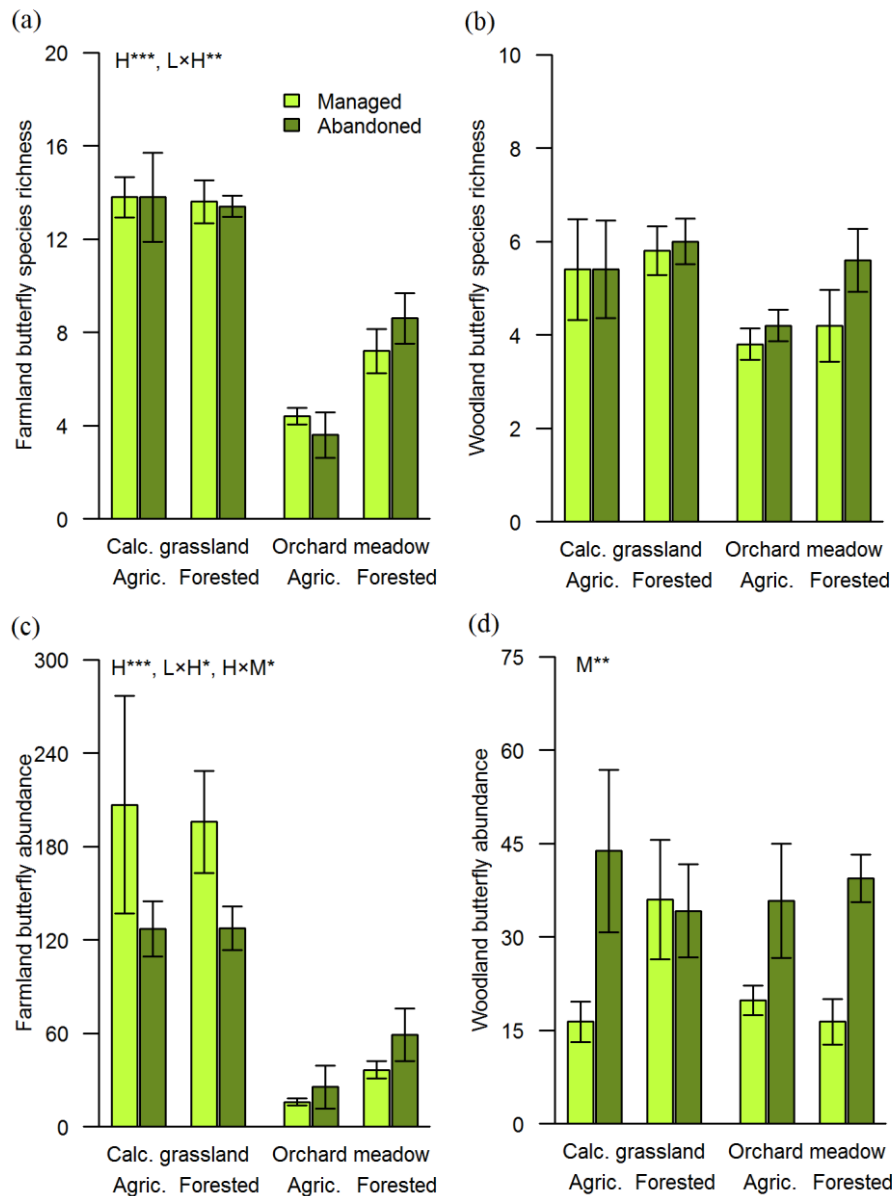
567 **Table 3.** Results of redundancy analyses to test the effect of habitat type (H: calcareous
568 grassland vs. orchard meadow), management (M: abandoned vs. managed) and landscape
569 context (L: agricultural vs. forest-dominated) on the community composition of all butterfly
570 and bird species. % var.: percentage variation explained. *P* values < 0.05 are in bold
571 characters.

	% var.	<i>F</i>	<i>P</i>
Butterfly			
Landscape	5.61	2.86	0.012
Habitat	21.33	10.88	0.001
Management (M)	2.49	1.27	0.229
Total constrained	29.43	5.00	0.001
Bird			
Landscape	2.30	1.52	0.023
Habitat	3.56	2.35	0.001
Management (M)	3.50	2.30	0.001
Total constrained	9.37	2.06	0.001

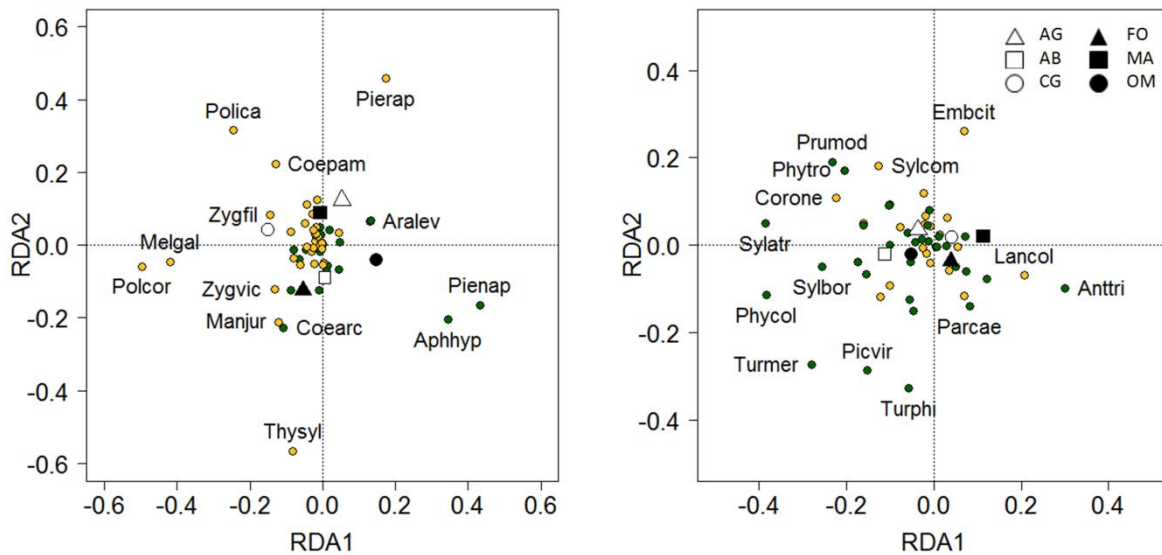
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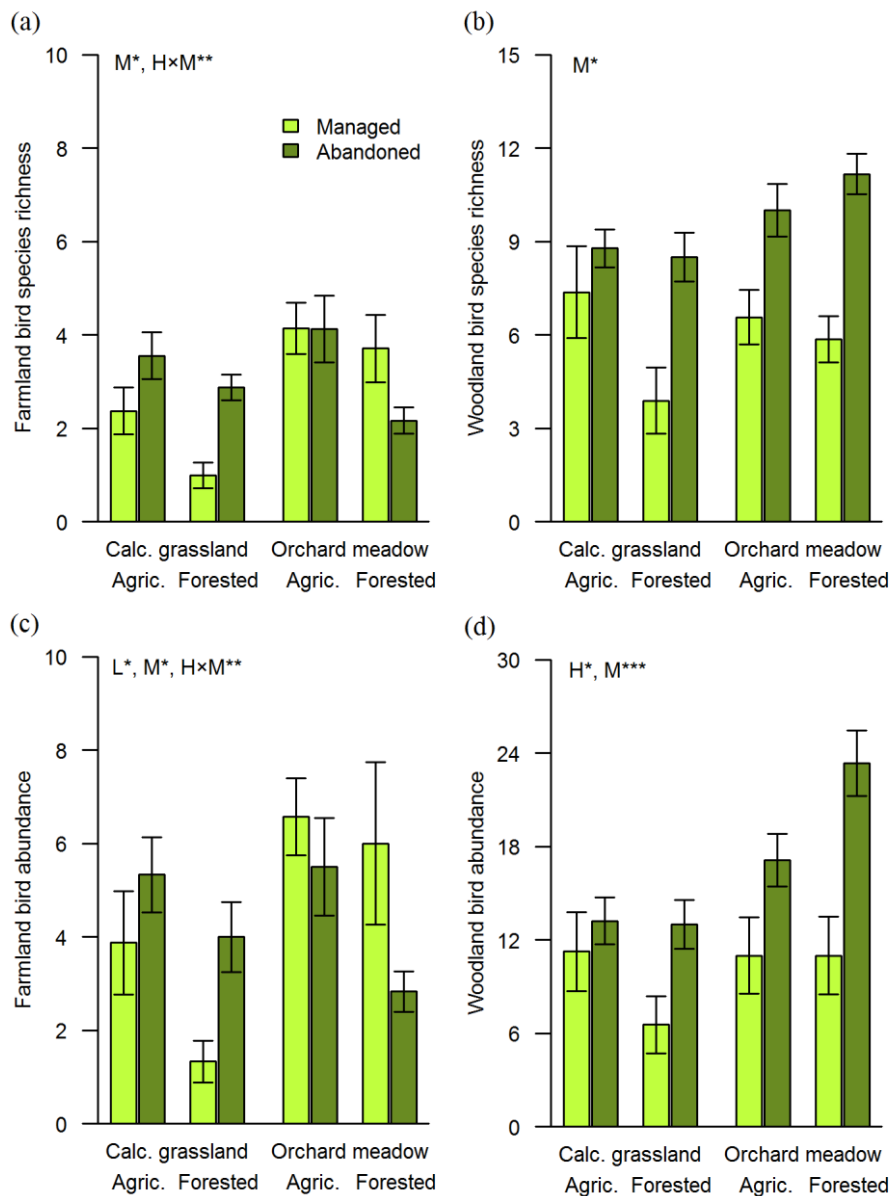
573 **Figure 1.** Schematic figure of the study design representing the study sites. There were five
574 replicates per treatment resulting in 20 calcareous grasslands and 20 orchard meadows located
575 in contrasting landscape context (agricultural or forest-dominated) differing in management
576 (regularly managed or abandoned management).



577 **Figure 2.** Mean (\pm SEM) farmland (a) and woodland (b) butterfly species richness and
 578 farmland (c) and woodland (d) butterfly abundance in managed vs. abandoned calcareous
 579 grasslands and orchard meadows situated in agricultural vs. forest-dominated landscapes.
 580 Results are based on generalized linear regression models (see Table 1) with $*P \leq 0.05$, $**P \leq$
 581 0.01 , $***P \leq 0.001$ (H: Habitat type, L: Landscape context, M: Management type).



582 **Figure 3.** Redundancy analysis biplots for all species of (a) butterfly and (b) bird
 583 communities (yellow circles: farmland species, green circles: woodland species) showing the
 584 effect of habitat type (CG: calcareous grassland, OM: orchard meadow), presence of
 585 management (AB: abandoned, MA: managed) and landscape context (AG: agricultural, FO:
 586 forest-dominated landscape). For visibility, only species with the highest fraction of variance
 587 fitted by the two first RDA axes are indicated. Species code consists of the first three letters of
 588 genus plus the first three letters of species names (Table A2, A3).



589 **Figure 4.** Mean (\pm SEM) farmland (a) and woodland (b) bird species richness and farmland
 590 (c) and woodland (d) bird abundance in managed vs. abandoned calcareous grasslands and
 591 orchard meadows situated in agricultural vs. forest-dominated landscapes. Results are based
 592 on generalized linear mixed-effects models (see Table 2) with * $P \leq 0.05$, ** $P \leq 0.01$, *** $P \leq$
 593 0.001 (H: Habitat type, L: Landscape context, M: Management type).