

Applicability of the “Frame of Reference” approach for environmental monitoring of offshore renewable energy projects

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Abstract

This paper assesses the applicability of the Frame of Reference (FoR) approach for the environmental monitoring of large-scale offshore Marine Renewable Energy (MRE) projects. The focus is on projects harvesting energy from winds, waves and currents. Environmental concerns induced by MRE projects are reported based on a classification scheme identifying stressors, receptors, effects and impacts. Although the potential effects of stressors on most receptors are identified, there are large knowledge gaps regarding the corresponding (positive and negative) impacts. In that context, the development of offshore MRE requires the implementation of fit-for-purpose monitoring activities aimed at environmental protection and knowledge development. Taking European legislation as an example, it is suggested to adopt standardized monitoring protocols for the enhanced usage and utility of environmental indicators. Towards this objective, the use of the FoR approach is advocated since it provides guidance for the definition and use of coherent set of environmental state indicators. After a description of this framework, various examples of applications are provided considering a virtual MRE project located in European waters. Finally, some conclusions and recommendations are provided for the successful implementation of the FoR approach and for future studies.

Keywords: marine renewable energy; large-scale projects; environmental monitoring; environmental indicators; monitoring framework; frame of reference

1. Introduction

Offshore winds, waves and currents have a large potential for long-term electricity generation world wide (Pelc and Fujita, 2002; Thresher and Musial, 2010). The wind industry is leading the way, whilst devices to harvest offshore wave and current energy are still under development (Sutherland et al., 2008; Inger et al., 2009; Bedard et al., 2010). Offshore wind energy is harvested by turbines rotating about a horizontal axis, which are derived from the well-established technology used on land. Nowadays, commercial offshore wind turbines have seafloor foundations, the most common ones being monopiles driven into the bed, gravity-based foundations, tripod foundations and jacket foundations. However, with wind parks moving towards deeper water, various types of floating foundations are being developed (Butterfield et al., 2007; Main(e) International Consulting, 2012). For waves, the technology is relatively immature and no commercial design has emerged yet amongst the very large variety of existing concepts (see Drew et al., 2009; Bald et al., 2010). Regarding currents, the most significant technology offshore consists

of rotating devices on horizontal axes (similar to wind turbines), even though other designs including vertical axes are also considered (see O'Rourke et al., 2010; Polagye et al., 2011).

As wind energy projects are moving further offshore, they are also increasing in size (see EWEA, 2012). The worlds largest (in surface area) Marine Renewable Energy (MRE) project currently operating offshore is the Greater Gabbard (southern North Sea), covering 146 km² with a nominal capacity of 504 MW; it should be soon exceeded by the 1,000 MW London array project (230 km² surface area) which is currently being developed in two phases (Phase 1: 175 turbines over 121 km² generating 630 MW is fully operating since April 2013). The future of both wave and tidal energy converters is also to cover such large areas including hundreds of devices (see Johnson et al., 2012). In addition, the offshore energy industry is considering large-scale (i.e., area > 10 km², at least) multi-platform projects combining various MRE devices (e.g., wind turbines and wave converters) or activities (e.g., energy conversion and aquaculture), in order to increase the utilisation factor per site and the overall revenue. That effort is testified by the relatively large number of recent EU-funded projects related to this domain (e.g., MARINA; MERMAID; ORECA; TROPOS; H2OCEAN).

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47 Multi-platform or not, **MRE** projects are also expected to
48 cumulate at specific locations offshore because of grid and
49 land access considerations, together with site-specificity
50 regarding the resource (especially for waves and currents).
51 In the Irish Sea, for example, three windfarms are cur-
52 rently operating within a radius of less than 20 *km* (Wal-
53 ney, Barrow and Ormonde, covering an area of 73 *km*²,
54 10 *km*² and 8.7 *km*², respectively) and a fourth very large
55 one is proposed (West Duddon, 66 *km*²). The develop-
56 ment of these large-scale projects, and their addition to
57 other anthropogenic activities offshore, is accompanied by
58 environmental concerns (Pelc and Fujita, 2002; Gill, 2005;
59 Michel et al., 2007; Sutherland et al., 2008; Inger et al.,
60 2009; Masden et al., 2010; Simas et al., 2010; Wilhelmsson
61 et al., 2010; Shields et al., 2011).

62 The evaluation of environmental effects in the offshore
63 realm is a difficult task, because the marine environ-
64 ment is a highly complex system where physical, chemi-
65 cal and biological properties interact at several spatial
66 and temporal scales. Although being ambiguously defined
67 (Heink and Kowarik, 2010), “environmental indicators”
68 generally reduce the complexity of a problem, or of a
69 large number of parameters, to a smaller number of key-
70 parameters that enable the description or quantification
71 of the status and trends of (entire or partial) ecosystems.
72 As such, indicators may facilitate management decisions
73 as they provide the necessary information for decision-
74 makers about where, when and how to act (Gubbay, 2004;
75 Davidson et al., 2007). They are also useful for the com-
76 munication of overall progress on stated goals and bench-
77 marks.

78 During the last decade, indicators have been increasingly
79 developed, including for the marine environment (Dav-
80 ies et al., 2001; Gubbay, 2004), and used at global (e.g.
81 World Bank, United Nation, Organization for Economic
82 Co-operation and Development), regional (e.g., European
83 Environment Agency), national and local levels, as well as
84 in the private sector. For example, environmental indicat-
85 ors are commonly used by the offshore oil and gas industry
86 to assess the impact of exploitation on the benthic ecology
87 and water quality (e.g., Olsgard and Gray, 1995; Andrade
88 and Renaud, 2011).

89 Indicators are commonly defined and organized in frame-
90 works that facilitate their understanding and interpret-
91 ation ensuring at the same time the appropriate match
92 between end-users and scientists (Gabrielsen and Bosch,
93 2003; Gubbay, 2004). Frameworks can also help to
94 understand the inter-relations between various indicat-
95 ors (Stegnastam, 1999). Several environmental frame-
96 works have been proposed, depending on the applica-
97 tion and scale of the problem considered. For example,
98 the Drivers-Pressures-Status-Impacts-Response (**DPSIR**)
99 model provides an overall approach for analysing envi-
100 ronmental issues, generally with regards to sustainable de-
101 velopment (Borja et al., 2006). This framework is useful

102 as a descriptive method reporting the environmental im-
103 pacts of a particular sector through the use of indicators;
104 as such, it is largely used to report indicators set at na-
105 tional levels and is able to provide a link between the socio-
106 economic aspects of an activity and the induced environ-
107 mental changes. **DPSIR** may be therefore well-adapted for
108 the strategic development of the offshore **MRE** industry
109 (Elliott, 2002). However, this type of framework might
110 not be relevant -or difficult to implement- if the focus is
111 on environmental monitoring of specific projects, where
112 guidance is required to select specific indicators. In this
113 case, other prescriptive and fully quantitative frameworks
114 that explicitly link objectives and quantitative parameters
115 are more adequate.

116 This paper assesses the applicability of the Frame of Refer-
117 ence (**FoR**) approach for the environmental monitoring of
118 offshore **MRE** projects. Even though the proposed method
119 is applicable to any type of offshore large-scale project, the
120 focus is upon projects harvesting energy from winds, waves
121 and currents (multi-platform or not).

122 2. Environmental effects and impacts of offshore 123 **MRE** projects

124 2.1. Classification of environmental effects

125 Given the complexity of the marine environment and the
126 multiplicity of technologies to harvest **MRE**, it is conveni-
127 ent to classify environmental effects within a framework.
128 The framework used in the present paper is based on the
129 one proposed by McMurray (2008) for wave converters,
130 subsequently modified by Boehlert and Gill (2010) and
131 by Polagye et al. (2011) for application to various **MRE**
132 devices.

133 The framework describes environmental concerns in terms
134 of stressors, receptors, effects and impacts. Stressors are
135 features that may induce environmental changes. Recept-
136 ors are elements of the ecosystem that may (or may not)
137 respond to the stressor. Effects describe how the receptor
138 is affected by the stressor, but do not indicate magnitude
139 or significance. Impacts deal with severity, intensity or
140 duration of the effect, and also with its direction (i.e., pos-
141 itive or negative). Impacts are generally recognized when
142 the effects induce changes in specific variables that are
143 used to define the status of the concerned receptor. These
144 impacts can be either direct or indirect (the latter obvi-
145 ously being more difficult to evaluate). Indicators can be
146 used to determine if the effects are strong enough to induce
147 impacts and if a response is required.

148 2.2. Stressors and receptors

149 In the context of offshore wind, wave and current projects,
150 six distinct **stressors** are identified:
151

	Physical presence of device	Dynamics	Release of chemicals	Generation of sound	Electro-Magnetic Fields
Physical environment	Artificial reef	Scouring Seabed disruption Hydrodynamic changes Aerodynamic changes Sediment dynamic changes			
Marine mammals and turtles	Collision potential Aggregation effect Obstruction of migratory route			Hearing injuries Site avoidance Stress increase Acoustic masking	Behavioural change
Pelagic habitat and communities	Collision potential Artificial reef Aggregation effect No take zone Steppingstone effect	Hydrodynamic changes Aerodynamic changes Pressure effects near rotating devices		Hearing injuries Site avoidance Stress increase Acoustic masking	Behavioural change
Benthic habitat and communities	Artificial reef No take zone Steppingstone effect Flora and fauna impact by moorings	Scouring Seabed disruption Hydrodynamic changes Aerodynamic changes Sediment dynamic changes	Pollution from dredging	Acoustic masking	Behavioural change Sediment temperature increase
Marine birds	Collision potential Aggregation effect Obstruction of migratory route			Site avoidance	
Water quality	Artificial reef Light reduction Sediment re-suspension by moorings	Seabed disruption Hydrodynamic changes Aerodynamic changes Sediment dynamic changes	Leaching Spilling Pollution from dredging Pollution from maintenance		

Table 1: Potential effects of stressors (top row) upon receptors (far left column), associated to offshore MRE devices. For simplicity, the stressor “cumulative impacts” and the receptor “ecosystem interactions” are not included. Environmental effects and main potential impacts are discussed in more detail in subsection 2.3.

1. *physical presence* of (fixed and moving) parts of the devices in the water and in the air (including the introduction of material or substrate at the bed);
 2. *dynamics*, which relates to (near- and far-field) changes in the air and water pressure fields and in sediment dynamics (including changes in sediment distribution due to seabed disruption during construction);
 3. *release of chemicals* in the area from the equipment and vehicles linked to the activity and from seabed removal;
 4. *generation of sound*, both above and under water;
 5. *Electro-Magnetic Fields*, produced by cables (during the operational phase); and
 6. *cumulative impacts* of stressors from several large-scale projects and other human activities.
- are not considered in this paper (for detailed information, see Wilhelmsson et al., 2010).
- Seven groups of *receptors* are considered with respect to MRE activities:
1. *the physical environment*, i.e. the atmospheric and marine (wave and current) climates and the bed sediment (near-field and far-field);
 2. *marine mammals and sea turtles*;
 3. *pelagic habitat and communities*, including planktonic and nektonic organisms (excluding marine mammals and sea turtles);
 4. *benthic habitat and communities*, including macrophytes, invertebrates and vertebrates living in association to bed sediment;
 5. *marine birds*, living or migrating near the project area;
 6. *water quality*, measured based on its physical and chemical properties; and,
 7. *ecosystem interactions*, such as (but not only) food web interactions, and trophic dynamics.

For each stressor, the stage of development of the project (survey, construction, operation and maintenance, and decommissioning) should be considered, together with the time scale (duration and frequency) and spatial extent. Both the time scale and spatial extent are highly project- and site-specific (e.g., the construction phase may take years to complete in the case of very large projects), and

194 Amongst these receptors, marine birds, marine mammals
195 and sea turtles are often protected by specific environ-
196 mental policies, conventions and international agreements.
197 For this reason (and also due to public perception) they are
198 of particular importance for the development of offshore
199 MRE projects. Due to their potential impacts on specific
200 pelagic and benthic receptors (habitats and communities),
201 large scale projects may indirectly change ecological pro-
202 cesses and dynamics of marine food webs (e.g., cascading
203 effects). The response of this receptor group (“ecosystem
204 interactions”) to stresses is the most difficult to evaluate,
205 because of its complexity and also because impacts may
206 occur even if no discernible changes are observed on other
207 receptors.

208 2.3. Environmental effects and main potential im- 209 pacts

210 The potential effects of stressors upon receptors at off-
211 shore MRE projects were identified based on a literature
212 review and reported in Table 1. Information from the fol-
213 lowing sources was used: CMACS (2003); Gill (2005); Gill
214 et al. (2005); Hastings and Popper (2005); Zucco et al.
215 (2006); BSH (2007); Linley et al. (2007); Brostrom (2008);
216 Evans (2008); OSPAR (2008); Vize et al. (2008); Gill
217 et al. (2009); Inger et al. (2009); USDOE (2009); Bald
218 et al. (2010); Boehlert and Gill (2010); Mueller-Blenkle
219 et al. (2010); Wilhelmsson et al. (2010); Normandeau et al.
220 (2011); Shields et al. (2011); Smith et al. (2012). Although
221 the potential effects of stressors on most receptors are iden-
222 tified, there are large knowledge gaps regarding the cor-
223 responding (positive and negative) impacts (Zucco et al.,
224 2006; Inger et al., 2009; Bald et al., 2010; Boehlert and
225 Gill, 2010; Wilhelmsson et al., 2010). To date, results from
226 only few long-term (years) monitoring surveys at wind-
227 farms (all with fixed foundations) are available (e.g., Dan-
228 ish Energy Authority, 2009; Degraer and Brabant, 2009;
229 Stenberg et al., 2011). The lack of knowledge of individual
230 stressor impacts inhibits the realisation of adequate cumu-
231 lative effects assessments (Polagye et al., 2011).

232 Despite large uncertainties, most of the *negative* envir-
233 onmental impacts of a single offshore MRE project are
234 considered of small intensity, short-term and/or of lim-
235 ited spatial extent (see Wilhelmsson et al., 2010). One
236 often cited potential negative impact upon marine birds
237 results from the risk of collision with (fixed or moving)
238 parts of the devices. Available studies indicate, however,
239 that collisions have small impacts at a population scale
240 level, although they can be significant for certain species
241 (Desholm, 2009; Wilhelmsson et al., 2010). Concerns for
242 marine birds are higher in case of fragmentation of co-
243 herent ecological units and habitat loss that can be in-
244 duced by an avoidance behaviour due to the presence of
245 the devices and to the production of noise; likewise, by
246 deflection of migration routes, especially for daily com-
247 muting species which might not have enough energy to

248 cope with the changes (e.g., Larsen and Guillemette, 2007;
249 Masden et al., 2009). In general, the greatest risks faced
250 by marine mammals are hearing injuries and habitat loss
251 due to the production of sounds during the construction
252 phase (Bald et al., 2010; Wilhelmsson et al., 2010), even
253 though strikes by the blades of current devices may also be
254 of concern in some cases (Boehlert and Gill, 2010). Fur-
255 thermore, the production of noise during operation may
256 mask bio-acoustics for communication and navigation of
257 long-distance migrating whales and sea turtles (Samuel
258 et al., 2005; Wilhelmsson et al., 2010). The newly con-
259 structed structures may serve as steppingstones for in-
260 vasive species (dispersal effect), which might pose as a
261 threat for local benthic and pelagic communities (Bulleri
262 and Airoldi, 2005; Glasby et al., 2007; Wilhelmsson et al.,
263 2010). The production of magnetic fields by cables may
264 also modify the behaviour of resident or migratory species
265 that use geomagnetic field for localisation and orientation
266 (CMACS, 2003; Gill, 2005; Gill et al., 2009; Normandeau
267 et al., 2011; Wilhelmsson et al., 2010). Overall, oil slick
268 resulting from aircraft or ship accident is considered to
269 have the largest potential negative impact upon all recep-
270 tors in terms of duration, spatial extent and intensity (Bald
271 et al., 2010).

272 The main potential *positive* impacts are associated with
273 the physical presence of MRE devices. The exclusion of
274 fishing activities, including trawling, within the project
275 area should act as a “no take zone”, with positive impacts
276 for pelagic species (e.g., increase of fish stocks) and benthic
277 communities with a more favourable environment for long-
278 lived rather than opportunistic species (Defew et al., 2012;
279 Fayram and De Risi, 2007; Wilhelmsson et al., 2010). Fur-
280 thermore, the introduction of hard structures (e.g., piles,
281 foundations, scouring protection, buoys) will provide ad-
282 ditional (or new) settlement surface/habitat for benthic
283 organisms and fishes (Langhamer et al., 2009), thus act-
284 ing as an “artificial reef” (Langhamer et al., 2009); ob-
285 servations at artificial reefs and wind farms suggest that
286 this effect generally results in positive impacts in terms
287 of ecosystems and biodiversity (Petersen and Malm, 2006;
288 Seaman, 2007; but see also Inger et al., 2009). In par-
289 ticular, the new settled communities may attract pelagic
290 and nektonic organisms, forming a so-called “fish aggreg-
291 ation device” (Wilhelmsson et al., 2006). The resulting
292 modification of pelagic and benthic habitats, communities
293 and prey distributions may in turn enhance feeding op-
294 portunities for certain species of seabirds, cetaceans and
295 pinnipeds (Wilhelmsson et al., 2010).

296 3. Indicators implementation

297 3.1. Importance of environmental monitoring pro- 298 grams

299 The key environmental regulations of offshore MRE activ-
300 ities are similar in principles worldwide, as they derive

301 from various international agreements and conventions. In
302 particular, an Environmental Impact Assessment (EIA) is
303 generally required prior to project consent, in order to en-
304 sure that the responsible authority makes a decision with
305 the full knowledge of any significant effects (cumulative,
306 positive and negative) on the environment. In this paper,
307 Europe is taken as an example since it is where the off-
308 shore wind power sector is the most developed (EWEA,
309 2010; Tuebke and Hernandez Guevara, 2011; Madariaga
310 et al., 2012).

311 The European Directive 2011/92/EU requires that Mem-
312 ber States carry out an EIA for the consent of projects
313 which are considered to have significant effects on the en-
314 vironment, including offshore wind project (e.g., CEFAS,
315 2004). For other MRE projects, the necessity to conduct
316 an EIA is at the discretion of the Member States, even
317 though in practice it is assumed that an EIA will be also
318 required (Huertas-Olivares et al., 2007; Woolf, 2011; Mar-
319 gheritini et al., 2012). The findings of an EIA are reported
320 in an Environmental Statement (ES), where the environ-
321 mental factors that may be affected by the proposed pro-
322 ject are described (e.g., Talisman, 2004), ideally, from con-
323 struction to decommissioning. The ES also indicates the
324 measures to implement for the mitigation of the potential
325 negative impacts (see CEFAS, 2004). Moreover, an Envir-
326 onmental Management Plan (EMP) should be provided for
327 the follow-up of effects that may threaten the environment
328 (Huertas-Olivares et al., 2007).

329 The exact requested content of the EIA is highly variable
330 in between Member States (and also sometimes in between
331 projects within the same country) (Huertas-Olivares et al.,
332 2007). A cumulative impact assessment should also be
333 undertaken as part of the EIA process, but such assess-
334 ments are rarely considered satisfactorily (Masden et al.,
335 2010). In any case, with regards to the large knowledge
336 gaps about the impacts of offshore MRE projects, it can
337 be considered that the general policy is to “deploy and
338 monitor”, as opposed for example to the “precaution prin-
339 ciple” which is applied to a large range of other activities
340 and supported by EU regulation (Johnson et al., 2012).
341 This is because the need to perform long-term research
342 on environmental impacts is dominated by the imperative
343 to develop marine energy which is driven by urgent polit-
344 ical, economic and climate change considerations (Athanas
345 and McCormick, 2013; SEL, 2010). Such a “deploy and
346 monitor” policy requires the implementation of effective
347 monitoring activities aimed at environmental protection
348 and knowledge development. Therefore the EMP should
349 be regarded as one of the most important outputs of the
350 EIA, with the following general objectives:

- 351 • to provide feedback and early warning of potential
352 environmental damages;
- 353 • to ensure that impacts do not exceed legal standards;
354 and,

- to check the implementation of mitigation measures
in the manner described in the ES report.

The use of environmental indicators to report the results
of the EMP represents a great asset for the development
of the incipient offshore MRE industry because indicat-
ors:

- conveniently convey information to government and
industries about environmental effects (negative as
well as positive);
- help to determine whether observed effects are ac-
ceptable or not through the upfront specification of
thresholds and trigger levels;
- allow the effectiveness assessment of mitigation meas-
ures;
- allow comparisons with similar (MRE) projects and
with other human activities; and,
- may be used as a tool for communication with other
stakeholders, such as non-governmental organizations,
the general public, etc.

The need to implement indicators for the evaluation of
the environmental impacts of offshore windfarms has been
recognized (Degraer and Brabant, 2009), but rarely put
into practice. Some examples include the clicks’ records
of acoustic porpoise’s detectors (T-Pods) which are used
to define various indicators for density, abundance, activ-
ity, etc. (Rye et al., 2008; Lindeboom et al., 2011). At
the Horns Rev 1 offshore windfarm, sand eels have been
also used as indicators of the ecosystem health (Sten-
berg et al., 2011). However, existing indicators are gen-
erally site-specific and not explicitly linked to objectives
and quantitative parameters or thresholds (e.g., Henriksen
et al., 2003). With tighter environmental legislation that
promotes the use of environmental indicators in marine
areas, standardized monitoring protocols should be adop-
ted to enhance their usage and utility (see Johnson, 2008;
Degraer and Brabant, 2009). In that context, the use of
tools such as the FoR approach may be useful for the defin-
ition and use of coherent sets of environmental indicators
at offshore MRE projects. An additional benefit of these
tools is the possibility to compare between different applic-
ations of the same indicator, in a process of gradual im-
provement. These approaches also help to evaluate if the
cost of measuring the indicator is justified by the expected
gain (increased level of environmental protection).

3.2. The Frame of Reference approach

The Frame of Reference (FoR) approach was developed to
help researchers from different fields to use one method
generically applicable to embed their results in a practical
decision context (Van Koningsveld et al., 2003; Van Kon-
ingsveld and Mulder, 2004; Van Koningsveld et al., 2005a).

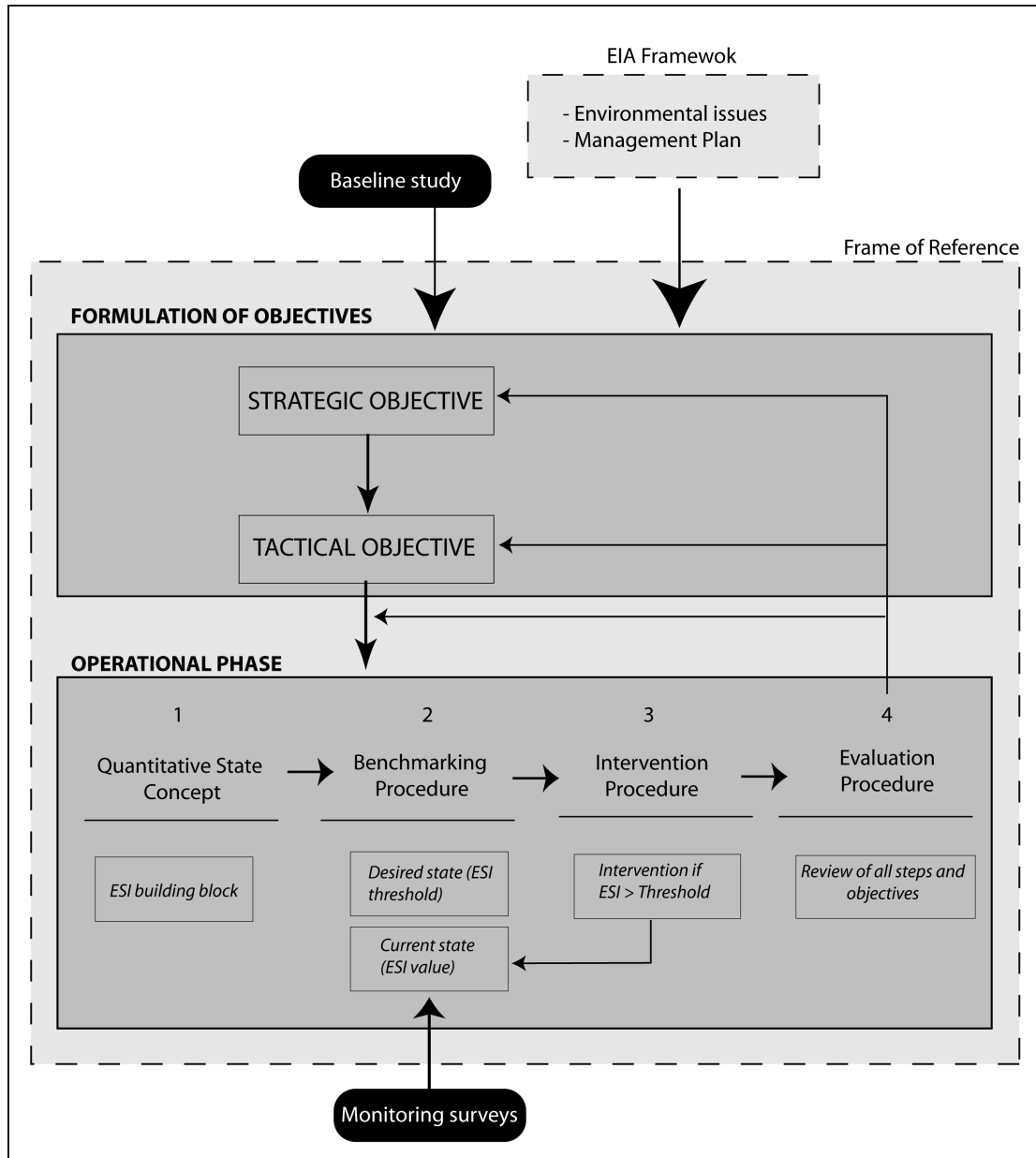


Figure 1: The Frame of Reference framework and application to offshore MRE projects (adapted from Van Koningsveld et al., 2007). The basic actions which are required at each steps of the operational phase are indicated in italics.

405 The approach is characterised by the definition of clear ob- 418
 406 jectives at strategic and operational (or tactical) levels and 419
 407 an operational phase where indicators are defined to verify 420
 408 whether or not these objectives are met (Figure 1).

409 The FoR framework has been used so far for the imple- 421
 410 mentation of coastal state indicators that help decision 422
 411 making with respect to the protection of eroding coasts, 423
 412 through enhanced communication between scientists and 424
 413 coastal managers (Van Koningsveld et al., 2007; Davidson 425
 414 et al., 2007; Ciavola et al., 2011; Marchand et al., 2011; 426
 415 De Vriend and Van Koningsveld, 2012). In fact, it has 427
 416 been used implicitly over the last decade in The Nether- 428
 417 lands for the successful development and implementation 429
 430

of a coastal erosion policy (Van Koningsveld and Mulder, 418
 2004; Van Koningsveld and Lescinski, 2007; Mulder et al., 419
 2011). 420

The first step of the FoR approach is the formulation of 421
 “strategic objectives” based on the long-term vision about 422
 the desired status of the system (Figure 1). In a second 423
 step, the means of satisfying (at least partly) each strategic 424
 objective at the short-term are made explicit through the 425
 definition of one or several “operational objectives”. Fol- 426
 lowing Marchand et al. (2011) and Mulder et al. (2011), it 427
 might be more adequate to qualify these objectives as “tac- 428
 tical” -rather than “operational”- because this step implies 429
 a choice between distinct expedients to realise the corres- 430

431 ponding strategic objective. The words and phrases used
432 for the formulation of the objectives should be considered
433 with extreme caution as they steer all consequent thinking;
434 iteration is required in order to think through the process
435 several times and ensure that the objectives are conceptu-
436 alised adequately regarding the environmental issues which
437 are at stake.

438 At the next level, an operational decision recipe consisting
439 of four stages is applied in order to meet each of the pre-
440 defined objectives:

- 441 1. the Quantitative State Concept (QSC);
- 442 2. the Benchmarking procedure;
- 443 3. the Intervention procedure, and
- 444 4. the Evaluation procedure.

445 The QSC defines for each tactical objective one or more
446 quantifiable parameters that will be used in the decision
447 making. This step is determinant regarding the actions
448 to be implemented at the next stages, as it specifies the
449 quantitative building block that is used to construct indic-
450 ators (second stage), to establish the intervention proced-
451 ure (third stage), and to help to assess whether or not the
452 objectives are met (fourth stage).

453 Threshold values are attributed to the parameters defined
454 at the QSC during the benchmarking procedure stage.
455 These thresholds determine the desired state of the sys-
456 tem, whilst the current state is established based on mon-
457 itoring. The benchmarking procedure is therefore the in-
458 dicator since it is at this stage that impacts are indicated
459 through comparison of the current and desired state. In
460 case of impact, the intervention procedure (third stage)
461 prescribes the actions to implement for restoring the sys-
462 tem toward the desired state. The current state is updated
463 with (new) data from monitoring surveys, prior to another
464 benchmarking procedure. It is important to design inter-
465 ventions as such that after implementation they influence
466 the indicators status as desired. This may seem trivial,
467 but it is not. Finally, the success of the actions undertaken
468 at the three previous stages is assessed through an eval-
469 uation procedure, which determines whether the strategic
470 and operational objectives are being met. At this stage,
471 all the various steps of the framework should be subject to
472 re-assessment. In particular, new parameters or threshold
473 levels can be defined through new QSC and benchmarking
474 procedures; both the strategic and operational objectives
475 may also be modified if required. A FoR may also be dis-
476 carded if it is found irrelevant or uneconomic for a given
477 project or objective.

4. Applicability of the Frame of Reference to off- 478 shore MRE projects 479

4.1. General aspects 480

481 The complexity of ecosystem processes and interactions
482 in the offshore environment may result in a mismatch
483 between scientists and decision makers' needs regarding
484 the development of offshore MRE projects. Although origi-
485 nally developed for the implementation of coastal policy,
486 the FoR approach could help through the construction
487 of Environmental State Indicators (ESI) used to verify
488 whether or not formulated objectives are met.

489 Obviously, the strategic objectives should derive from the
490 key environmental issues identified through the EIA pro-
491 cess. They might follow from (national or international)
492 legislation, conventions or treaties. Strategic objectives
493 might as well derive from binding conditions set by envi-
494 ronmental agencies and local authorities for project con-
495 sent, or by informal commitment of project managers. To
496 ensure that all separate elements of the generally long last-
497 ing EIA process still fit together at the end of the pro-
498 cess, a common framework for impacts classification, such
499 as the stressor-receptor framework described in Section 2
500 (Boehlert and Gill, 2010), should be adopted during both
501 the EIA and FoR procedures. This facilitates the com-
502 munication between the various parties involved at any
503 stage of the EMP. For example, stressor-receptor matrixes
504 can be drawn to represent impacts severity, the temporal
505 and spatial scales and uncertainties (see Polagye et al.,
506 2011).

507 The operational phase should rely on the specific actions
508 proposed in the EMP regarding the identification, follow-
509 up and mitigation of impacts. At offshore wind farms
510 (and presumably at any type of future MRE projects), the
511 EMP commonly follows the Before-After Control-Impact
512 (BACI) approach (Green, 1979), where the current state
513 of the site is compared to previous and/or pristine environ-
514 mental conditions known from the baseline study and from
515 concurrent measurements at "reference areas". The defin-
516 ition of thresholds representing the "desired state" might
517 be one of the most difficult tasks of impacts evaluation,
518 since natural temporal and spatial variability of paramet-
519 ers must be considered. In some cases, indicator thresholds
520 are fixed by legal requirements. In most cases, however,
521 they must be established based on robust expertise to-
522 gether with a good knowledge of the natural environmental
523 conditions at various spatial and temporal scales. Simil-
524 arly, monitoring surveys alone might not be enough to em-
525 brace the natural variability (both spatial and temporal)
526 of the measured parameters. The establishment of the cur-
527 rent state may also be based on statistical and numerical
528 models. Not only do these tools allow the interpretation of
529 a limited number of measurements over broader areas and
530 longer time-scales, but they can also be useful in defining
531 environmental policies.

Receptor	Stressor	Effect	Environmental Issue	Description
Mammals	Sound	Avoidance	Porpoise protection	Porpoises may escape the proposed area due to the production of sound
Benthos	Presence of devices	Artificial reef	Habitat conservation	The shell deposits of the newly settled blue mussel may induce a change from soft- to hard-bottom substrates
Birds	Presence of devices	Collision	Common Eider ducks protection	Migrating Eider ducks population may suffer large loss from collision with rotating blades
Water quality	Chemical	Spilling	Water pollution	The activity induces a risk of oil spilling from device components and vessels

Table 2: Example of Environmental Issues identified during the EIA of a MRE project. The environmental concerns are explained in more detail in [Subsection 4.2](#).

4.2. Examples of application

In order to illustrate how the FoR approach works, a virtual offshore MRE project in Europe is considered, for which the EIA has reported the following environmental concerns (see also [Table 2](#)):

- *Mammals*: Harbour porpoises (*Phocoena phocoena*) are abundant in the area and may suffer hearing injuries and death due to the emission of underwater sound from devices and vessel activities, during construction, operation and maintenance, and decommissioning. Harbour porpoises are strictly protected in Europe under the Habitat Directive, and several conventions.
- *Benthos*: The proposed project sites on sandy shoals, which are markedly distinct from the surrounding in terms of benthic habitat. The benthic communities of these shoals are considered to have a high ecological value, as they provide key ecological services at multiple trophic levels (e.g., [Dubois et al., 2009](#)). Blue mussels (*Mytilus edulis*) dominate the regional hard bottom fauna communities and are expected to settle on the immersed structures and scour protections around foundations. The deposit of mussel shells at the bed may induce a shift from sandy towards hard substrate benthic habitat with potential negative impacts on the structure of the ecosystem. The key environmental issue is the conservation of sandy habitats.
- *Birds*: The migrating route of Common Eider ducks (*Somateria mollissima*) passes in the vicinity of the proposed area. Available studies show that Eider ducks generally avoid flying close or into (single) wind farms ([Masden et al., 2009](#)). However, this behaviour relies on vision and there are large uncertainties

about their flight patterns during periods of darkness and conditions of poor visibility such as fog or snow ([Larsen and Guillemette, 2007](#)). Under these bad visibility conditions, the probability of collision with wind turbines may be significantly enhanced, especially if birds are attracted by illuminated turbines ([Fox et al., 2006](#)). Eider ducks are protected under the EU Bird Directive.

- *Water quality*: The water quality may be affected as a result of oil spilling from components (e.g., gear boxes, hydraulic pumps) of MRE devices, and also from vessels and helicopters supporting the activity.

Regarding harbour porpoise protection against underwater sound, the long-term strategic objective of the FoR could be ‘to preserve the regional population given the planned activity’ ([Table 3](#)). Studies have shown that wind farm related sound, for example, has the potential to affect the behaviour and physiology of harbour porpoises at considerable distances. Physiological effects include Temporary and Permanent Hearing Threshold Shifts and more severe injuries up to death, depending of the distance of the individual to the source. Hence, one tactical objective could be that ‘no porpoise should suffer from sound related to the activity’ ([Table 3](#)). More specifically this objective could be achieved by either reducing the sound at the source or by physically keeping the porpoise away from areas where sound levels are potentially harmful. The underwater sound hazard is greatest during the construction phase, when porpoises are present in the area. Past experience has shown that porpoises avoid areas where piling activities take place; lethal hearing injuries may occur if they are located too close to the source when hammering starts. In this example we will focus on keeping the porpoises at a safe distance from the sound source. Recent studies have indicated that during piling, severe injuries are estimated to occur in a radius of 1.8 km from the source ([Thomsen et al., 2006](#)). The ESI in this case might be derived from the observation of the ‘number of

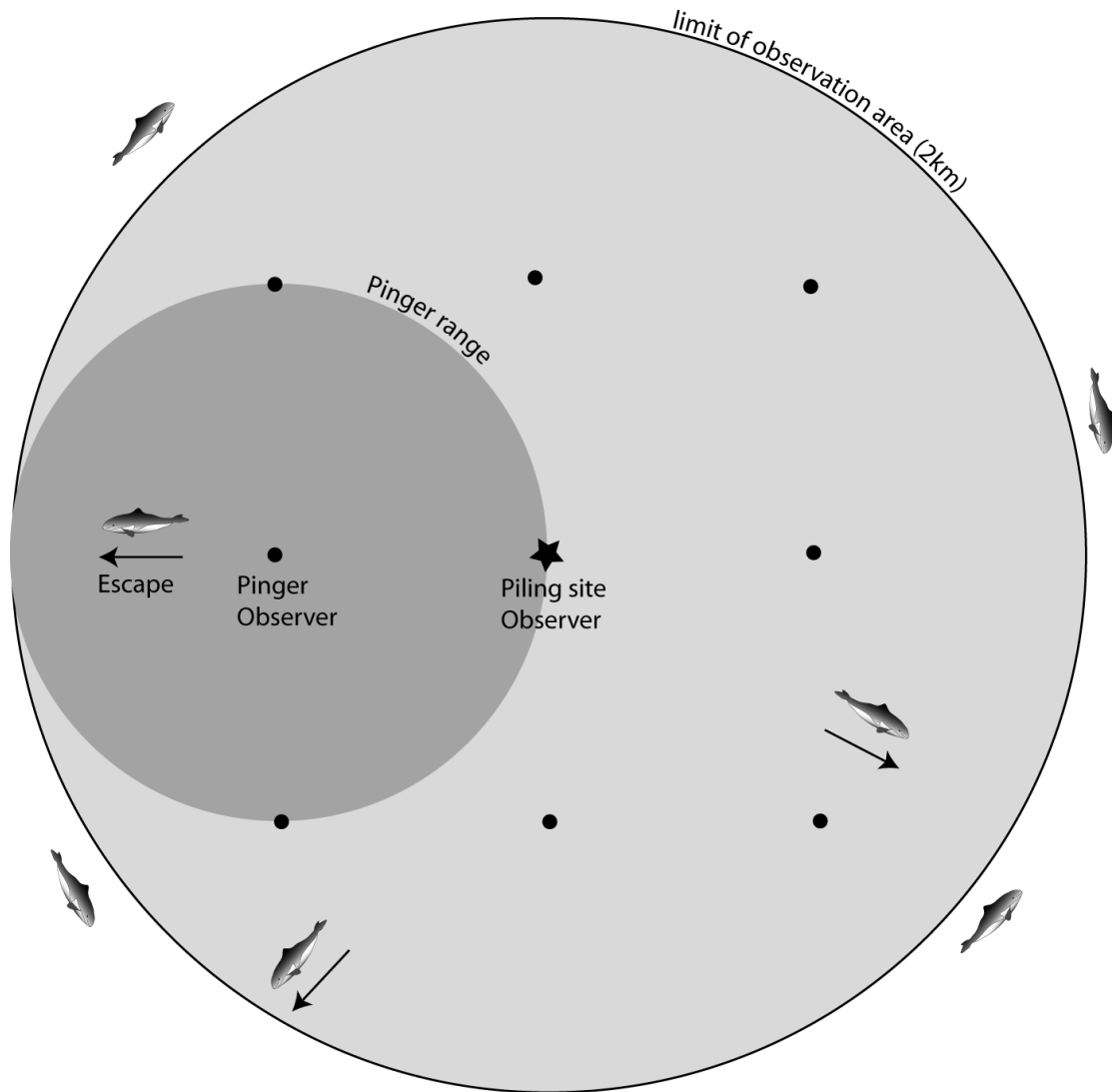


Figure 2: Example of field set up for an operational phase designed to scare porpoises during piling (see tactical objective). Eight pingers (dots) with 1 km range each (see the dark grey area, represented for one pinger only) are used to scare porpoises before piling starts to allow them to escape the area. Marine mammal observers, one at each pinger location plus one at the piling site, establish the current state (benchmarking), by checking for the presence of porpoise in the 2 km radius area of potential injuries by sound (light grey area). Piling is allowed to start if no porpoise is observed in this area during the preceding 2 hr (ESI).

605 individuals within 2 km from the source, after the deployment of acoustic harassment (or scaring) devices, during a certain time (e.g., 2 hr) prior to conduct the piling operations' (QSC, Table 3; Figure 2). Several devices might be necessary, depending on their range of effectiveness (e.g., Cox et al., 2001; Culik et al., 2001). The desired state is the absence of individual within the 2 km radius during this time interval (Table 3). The current state could be established based on marine mammal observers deployed in order to visually cover the total area of restriction (Figure 2; Table 3). If no individual is observed, piling can start without concerns for porpoises. Otherwise, operations should be postponed until reaching the desired state; the use of additional or other types of repelling devices might be necessary (Intervention procedure, Table 3). Evaluating the proposed procedure it seems likely that this FoR

will contribute to its strategic objective. The tactical objective is vulnerable to marine mammals observers missing porpoises that still are present despite the period of harassment. Put in practice procedures should be optimized to minimize this risk.

The proposed project will undoubtedly induce local physical changes of habitat (if only for the introduction of hard structures in the water). The development of organisms such as mussels on the structures may create locally "hot spots" of biological activity (e.g., Norling and Kautsky, 2008) that could be beneficial to the ecosystem (including the shoal benthic community). Thus, one strategic objective could be to enhance biodiversity and productivity, providing there is no negative impact -for example due to

Environmental issue	Strategic Objective	Tactical Objective	Quantitative State Concept	Benchmarking Desired State	Benchmarking Current State	Intervention Procedure
Harbour porpoise protection	To preserve the regional population given the planned activity	No porpoise should suffer from sound related to the activity	Number of individuals observed in a radius of 2 km from the source, during 2 hr of acoustic harassment, prior to the start of piling	No individual in the 2 km radius area, during 2 hr prior to piling starts	Marine Mammals Observers distributed in order to cover by eye the 2 km radius area	Do not start piling, increase the number of repelling devices
Conservation of sandy habitat	To maintain the existing sandy habitat within the project area	Shell beds should not cover > 25% of the project area during operations	Relative percentage of surface area covered by shells in a random subset area being 10% of the total project area	Less than 25% of the sandy surface within the subset area covered	Ground-truthed side-scan sonar data over the entire subset area	Restore habitat (environmental dredging) along corridors in between the devices within the entire project area
Common Eider duck protection	To preserve the population of Common Eider Duck passing over the region	The project activities should not increase the population mortality rate	Percentage of duck population at risk of collision with the structures	No more than 5% of the population predicted to collide	Predictions from stochastic model at 95% confidence interval	Stop turbines
Water pollution by oil	To preserve favourable water quality for local flora and fauna	Timely proactive maintenance of operating devices containing oil	Timely replacement of components containing oil, at time defined by the preventive maintenance strategy	Timely substitution of components and visual inspections during maintenance	Actual substitution of components and visual inspections during maintenance	Substitution of defective components and re-evaluation of the maintenance strategy for the relevant component

Table 3: Examples of application of the FoR approach to MRE projects. The gray column (QSC) corresponds to the quantitative building block that is used for the construction of the indicator at the benchmarking procedure stage. For explanations, see text.

636 greater predation- on the shoal benthic community (another FoR may be defined to tackle this latter issue); this
637 objective could be achieved through the selection of under-
638 water structures designed to favour the colonisation of
639 selected species (see Martins et al., 2010; De Vriend and
640 Van Koningsveld, 2012; De Vriend et al., 2014). Another
641 (more defensive) approach, taken as example here, could
642 be to ‘maintain the existing sandy habitat within the project
643 area’ (Strategic objective, Table 3). This approach
644 supposes that it has been previously evidenced that reduction
645 of this habitat induces negative impacts (at present, this
646 effect is generally not a major concern, but it could
647 become substantial in the case of farms with several
648 hundreds of devices operating for decades). It is technically
649 difficult and costly to inhibit the colonisation of organisms
650 on newly introduced material. Thus, the tactical objective
651 could be that ‘shell beds should not cover more than
652 25% of the project area during operation’ (Table 3). The
653 QSC stage may define the parameter to quantify the ex-

655 tension of shell beds as the ‘relative percentage of surface
656 area which is covered by shell deposits within a subset
657 region (selected randomly) corresponding to 10% in sur-
658 face area of the proposed project’ (Table 3). This requires
659 that the distinction between sandy and mussel bed habi-
660 tats is clearly defined at the benchmarking procedure, as
661 it depends on the method used to establish the current
662 state. For example, if mechanical sediment sampling is
663 involved, the classification of sand mixed with mussels as
664 “sandy” or “shell” habitats requires the definition of limit
665 values, e.g., in terms of relative weight of shells or grain size
666 parameters. Our example contemplates (ground-truthed)
667 side-scan sonar images since they generally allow a clear
668 distinction of hard and soft beds based on their tonal con-
669 trast (Table 3). The desired state corresponds to ‘less than
670 25% of shell beds within the subset area’. More complex
671 proactive approaches could rely on the outputs from mod-
672 els of mussels growth and deposition (e.g., Maar et al.,
673 2009). The intervention procedure could encompass envir-

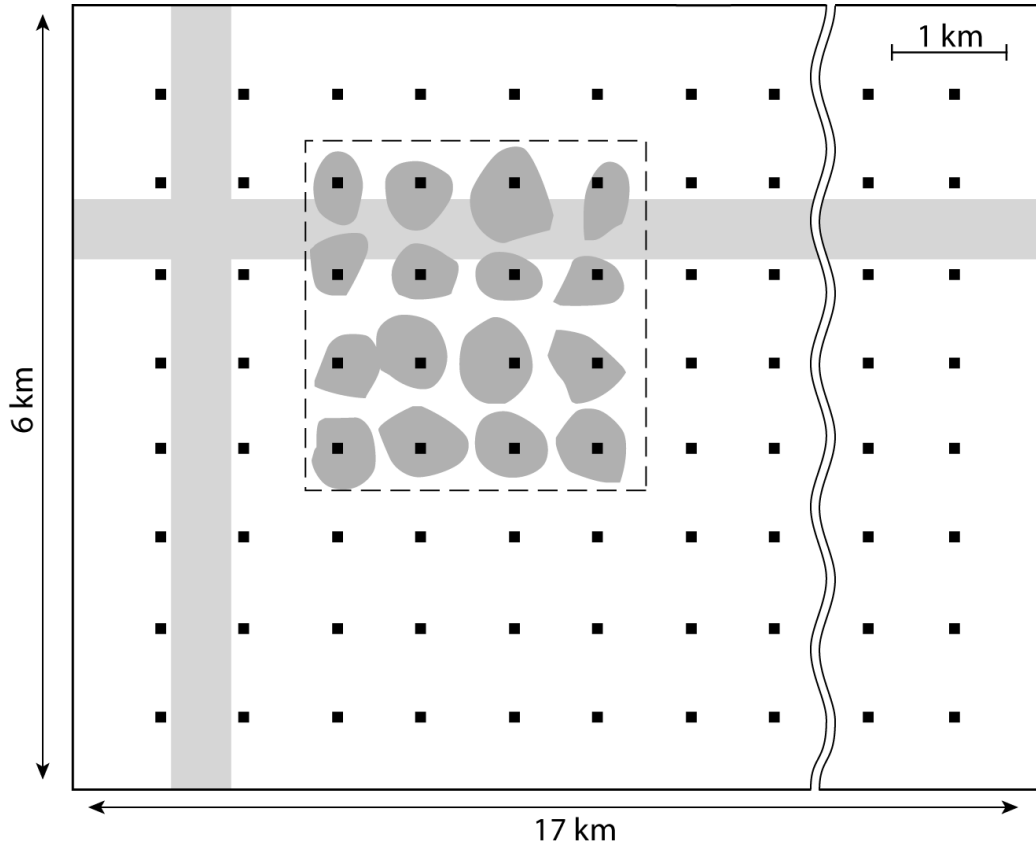


Figure 3: Implementation of the operational phase designed to control the development of mussel beds within the project area (Tactical objective). The total area (plan view) is 100 km^2 and includes 176 MRE devices (dots). Ground-truthed side-scan sonar surveys are conducted in a 10 km^2 area (dashed line) to distinguish sandy bed (white) from shell beds (dark grey). If shell beds represents $> 25\%$ of the survey area, environmental dredging is performed along parallel corridors over the entire project area to remove the mussel layer (examples of these corridors are shown in light grey).

674 onmental dredging guided by video to remove the excess of
 675 deposited layer of material and restore the original habitat
 676 along corridors in between the devices (Table 3; Figure 3).
 677 Such a mitigation option requires strong awareness regard-
 678 ing the financial implication of its implementation. For
 679 example, economic feasibility will be dependent of the re-
 680 quired dredging frequency (every year? every ten years?),
 681 estimated based on modelling tools (e.g., coupled hydro-
 682 dynamic and mussel deposit models). Careful economic
 683 assessment whether the cost associated with a measure can
 684 be justified by the environmental gain is crucially import-
 685 ant. In the end it might be concluded that environmental
 686 dredging is too expensive to be implemented, in which case
 687 more reachable objectives must be set at the start of the
 688 FoR process. Another issue could be that the proposed
 689 intervention (environmental dredging) itself is hampering
 690 the strategic objective (preservation of the existing sandy
 691 habitat). Such potential points of contention highlights
 692 the importance of adopting an iterative approach when for-
 693 mulating the strategic and operational objectives.

694 Any impact of the project on migrating Common Eider
 695 ducks must be analysed at a population level. As a strate-
 696 gic objective, the project activities should ‘preserve the

697 population of Common Eider ducks passing over the re-
 698 gion’ (Table 3), where the extension of the “region” is
 699 clearly defined. One way of meeting this objective could be
 700 ‘to prevent any increase of their mortality rate related to
 701 the project activities’ (Tactical objective, Table 3). At the
 702 QSC stage, this objective may lead to the development of
 703 a parameter representing the percentage of the duck pop-
 704 ulation risking collision with the structures. For selected
 705 periods, the current state can be predicted with a level of
 706 certainty (for example, 95% confidence interval) based on
 707 stochastic models built from compilations of observations
 708 (Figure 4; Table 3). In particular, reliable model predic-
 709 tions require estimates of the number of individual Eider
 710 duck collisions within the project area and of their fluxes
 711 throughout the project area (e.g., Band, 2000; Petersen
 712 et al., 2006; Troost, 2008). Collision estimates can be ob-
 713 tained using non-contact sensors (e.g., acoustic sensors,
 714 microphones) deployed on a number of turbines, especially
 715 during periods of heavy migration (spring and autumn).
 716 Likewise, surveillance radars are useful to measure the
 717 volume of bird movement and to track their altitude and
 718 trajectories through the area (visual observations are also
 719 necessary to calibrate the radar signal for species distinc-

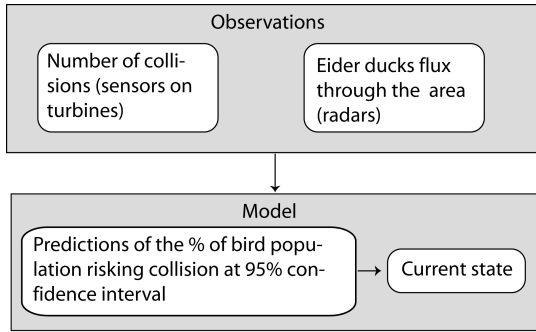


Figure 4: Illustration of the definition of the current state for selected periods of Eider duck migration. Observations (radar and collision monitoring data) are compiled to build a stochastic model. The current state is derived from model prediction of the percentage of bird risking collision with the turbines at a 95% confidence interval.

tion). The desired state (Table 3), for example ‘no more than 5% of the population predicted to collide’, should be fixed considering the effects of the increased mortality on the population over longer time periods (Fox et al., 2006). As a proactive intervention measure, it could be possible to shut down turbines during the periods when the indicator threshold is exceeded (Table 3).

With respect to water pollution, one obvious strategic objective is ‘to preserve favourable water quality for local flora and fauna’ (Table 3). One of the various tactics that can contribute to meet this objective is to ensure a ‘timely proactive maintenance of the operating devices containing oil’, e.g., gear boxes, hoses, in order to prevent oil leaks from happening (Table 3). In this case, the QSC can make use of the maintenance task which is generally established for each component (based on reliability figures like failure rate) for preventive maintenance throughout the duration of the project (Table 3). The comparison between the desired and current states is then performed by comparing the planned replacement of components with the actual recorded replacement. In addition to the preventive replacement of device components, the mitigation procedure may request to revise the maintenance strategy (e.g., frequency) in order to avoid future oil leaking from the relevant component (see the example in Figure 5). Compared to the harbour porpoise example this FoR is likely to achieve its tactical objective. The strategic objective, however, remains vulnerable as other potential causes for leakage are not addressed. This issue should be addressed with the definition of additional tactical objectives.

5. Conclusions and recommendations

It is in the interest of the incipient offshore MRE industry to carefully address the environmental impacts induced by large-scale projects. This task is presently difficult to achieve satisfactorily due to large knowledge gaps about

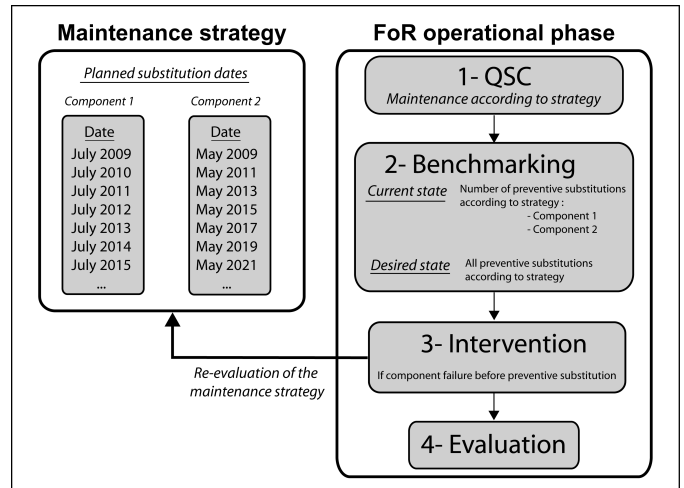


Figure 5: Example of the operational phase of a FoR (right panel) designed for a proactive maintenance of components containing oil (tactical objective). The maintenance strategy (left panel) defines the dates of component substitution (for simplicity, two components are considered). The QSC (Stage 1) is based on the maintenance strategy to indicate impacts during the benchmarking procedure (Stage 2). If failure occurs before the planned substitution of the components, the intervention procedure (Stage 3) requires a revision of the maintenance strategy.

impacts. Thorough long-term (years) EMPs should be implemented in order to enhance scientific knowledge regarding impacts. The implementation of environmental indicators within these programs is recommended as they generally describe in a convenient (simplified) manner the status of (complex) systems and thus may facilitate management decisions.

For the implementation of indicators within EMPs, the FoR approach is advocated over other frameworks due to its prescriptive nature. The FoR provides clear guidance for the selection of indicators that are linked directly to specific management issues. This framework also makes sure that predefined intervention procedures will be implemented if mitigation or remediation actions are required.

The examples presented in this contribution describe the use of FoR as a remediation tool. However, the most effective options to mitigate environmental impacts are generally available during the design phase of the project, i.e., during the selection of the site, of the technology to be used, and of the project layout. It is recommended to implement the FoR approach at various phases of the lifecycle of a project.

The development of a FoR framework is recommended for each of the potential environmental issues. The possible interaction between management issues from different FoRs must be addressed. In particular, future research should seek to optimise the collaborative effort not only between scientists of distinct disciplines, but also between all the parties at stake (non-governmental organ-

isations, nature conservation organisations, stakeholders, managers, policy makers). Furthermore, the development of several FoRs may require the integration of various time and space scales. It is therefore also recommended to investigate the articulation between the distinct management scales (see Mulder et al., 2006).

The occurrence of many environmental issues may also lead to the development of FoRs with conflicting objectives. In such a case, optimising a particular ESI may be detrimental to other objectives. Hence, it would be helpful to have some methods that help to decide what the best option is. Some tools should be developed to rank or prioritise the FoR with conflicting objectives.

Furthermore, it is important that an open policy regarding data access is implemented at a national, European and international level (e.g., the “OpenEarth” approach; Van Koningsveld et al., 2005b; Baart et al., 2012; De Boer et al., 2012; Van Koningsveld et al., 2013). Such a policy would be highly beneficial for research about environmental impacts, and for the industry to establish cost-efficient EMPs while demonstrating a strong commitment toward environmental protection.

Management decisions may have strong ecological influences and substantial financial implications. This fact often leads to reluctance to embrace new, unproven methodologies. It is therefore essential to rigorously test the FoR approach against real cases in order to firmly demonstrate how it can improve the management of specific environmental issues. At last, there should be a strong awareness of the potential financial implications of the managing decisions proposed in the various indicator schemes. As a final recommendation, any set of indicators should be always, as much as possible, scrutinized and tested for practical applicability in relation to the overall protection objective.

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