

The antagonistic and synergistic effects of temperature during solar disinfection of synthetic secondary effluent



Stefanos Giannakis^{a,b,c}, Efthymios Darakas^a, Antoni Escalas-Cañellas^{b,d}, César Pulgarin^{c,*}

^a Laboratory of Environmental Engineering and Planning, Department of Civil Engineering, Aristotle University of Thessaloniki, Thessaloniki 54624, Greece

^b Laboratory of Control of Environmental Contamination, Institute of Textile Research and Industrial Cooperation of Terrassa (INTEXTER), Universitat Politècnica de Catalunya, Colom 15, Terrassa 08222, Catalonia, Spain

^c Swiss Federal Institute of Technology, Lausanne, Institute of Chemical Sciences and Engineering, Lausanne 1015, Switzerland

^d Department of Chemical Engineering & Terrassa School of Engineering, Universitat Politècnica de Catalunya, Colom 1, Terrassa 08222, Catalonia, Spain

ARTICLE INFO

Article history:

Received 6 January 2014

Received in revised form 7 February 2014

Accepted 8 February 2014

Available online 19 February 2014

Keywords:

Solar disinfection

Wastewater

Full factorial design

E. coli

Modeling

ABSTRACT

A 4-factor, multilevel, full factorial design of 240 experiments was performed in order to investigate the effect of temperature on the inactivation efficiency of spiked *Escherichia coli* in simulated solar disinfection of a synthetic secondary effluent. The initial population of the microorganisms was 10^3 , 10^4 , 10^5 and 10^6 CFU/mL, the exposure time 1, 2, 3 and 4 h, the treatment temperature 20, 30, 40, 50 and 60 °C and the sunlight intensity 0, 800 and 1200 W/m². Radical changes in bacterial behavior, process efficiency and remaining populations were observed, while treating effluents in discreet temperatures. Elevating treatment temperature from 20 to 40 °C drastically impaired disinfection. Thermal inactivation with no regrowth predominated at 50 °C and total inactivation of microorganisms was observed at 60 °C in non-irradiated samples. Irradiation at 800 and 1200 W/m² much increased inactivation efficiency, especially at 50 and 60 °C, proving sensitive light-temperature synergy at those temperatures. Total inactivation was achieved within 4 h under a range of treatment conditions, including all samples at 1200 W/m², or 60 °C samples at 800 W/m². Also, 99.9–100% efficiencies and final population below 1000 CFU/100 mL were obtained at 800 W/m² and temperatures of 50 °C and above. Treatment time, temperature and intensity are the critical parameters for the disinfection process, while initial population is insignificant for removal efficiency. An explanation of the mechanism of the process as well as a general linear model predicting the outcome of the experiments is also suggested.

© 2014 Elsevier B.V. All rights reserved.

1. Introduction

The scientific basis of solar disinfection was established in the 80s by Acra et al. [1], marking an era of important advances in solar water purification. Gradually, the laboratory work was implemented in the field, with studies performed by Wegelin et al. [2] or McGuigan et al. [3], which set the milestones for solar disinfection (SODIS) of water. More specific studies have followed throughout the years, which highlighted the important parameters of the process, as the UV-A dose, boosting efficacy and rendering SODIS a safe practice [4–6], by explaining the acute inactivation of microorganisms after a few hours of exposure to sunlight.

In parallel, many studies have initiated a cycle of investigations over the efficacy of solar disinfection for wastewater. This field was relatively unexplored and several aspects needed to be studied; this knowledge area welcomed works conducted by Kositzi et al. [7] and Polo-Lopez et al. [8] and Rizzo et al. [9], that have investigated several aspects of solar photolytic and photocatalytic treatment in different microorganisms (*E. coli*, *Fusarium*). Interest was also given in the enhancement of the process by technical means, such as compound parabolic collector (CPC) solar photo-reactors [10,11], with special focus given to the application and reuse of wastewater.

Although an interesting practice, there has not been enough focus on the possibility of treating wastewater exclusively with sunlight. Works that have demonstrated potential application margins, such as Davies-Colley et al. [12] and Craggs et al. [13] in waste stabilization ponds, have indicated the efficiency of sunlight in disinfecting wastewater as well. However, the high retention times make them less attractive than catalytic processes as far as

* Corresponding author. Tel.: +41 216934720; fax: +41 216936161.
E-mail address: cesar.pulgarin@epfl.ch (C. Pulgarin).

Nomenclature

CFU	Colony Forming Units
W	Watts
DOE	Design of Experiments
SS	Sum of Squares
Seq SS	Sequential Sum of Squares
Adj SS	Adjusted Sum of Squares
DF	Degrees of Freedom
C_i	Concentration (at time = i)
$\log_{10} U$	logarithmic Units
F	F -test
P	P -value
t	treatment time
T	temperature
C	initial bacterial population
I	light intensity
S	Standard Error of the Regression
PRESS	Prediction Sum of Squares
R-Sq	Sum of Squared Residuals
R-Sq(adj)	Adjusted Sum of Squared Residuals
R-Sq(pred)	Predicted Sum of Squared Residuals

the application point of view is concerned. However, developing countries benefited a lot from SODIS and can possibly benefit from solar disinfection of wastewater. Sanitation conditions in many African countries are marginally non-existent and untreated or poorly treated sewage end up polluting the drinking water supplies [14]. It also occurs that the pre-mentioned regions are areas with a vast number of sunny days per year, so an application of the disinfecting action of light without other technological means could be attractive.

Solar wastewater disinfection follows the same principles as water disinfection; the effect of light against pathogens is the same, but practically, one of the major differences lies in the support microorganisms find in this water matrix. The presence of ions and nutrients, organic matter, etc. provides solid ground for their survival and growth [15]. The process depends on several parameters, which complicate the study more than the drinking water one. Another important aspect is the temperature conditions that are present during the treatment. SODIS applications have reported elevated temperatures and synergistic actions of light and UV [2,3], in otherwise simpler water matrices. Reed [16] highlighted, among others, the presence of organic substances in SODIS water; the case of wastewater is an even enhanced one.

Hence, since the number of examined parameters is high, it is useful to employ experimental design techniques, which permit the extraction of information otherwise not visible [17]. This tool has been proven efficient in works that study wastewater disinfection [18,19], by creating a pre-designed set of experiments, which explains the process and the interactions between the studied parameters.

Under this prism, the current work focuses on the disinfection of wastewater by solar light alone and a statistical approach has been done, to investigate the behavior of microorganisms in synthetic secondary wastewater, when exposed to sunlight. In summary, a full factorial design has been employed to further investigate the effects of (i) exposure time, (ii) treatment temperature, (iii) initial bacterial population and (iv) sunlight intensity on *E. coli*, in batch tests, simulating solar disinfection of secondary treated wastewater. The efficiency of the process was measured, as well as a construction of a general linear model, working as an indicator of the process efficiency.

2. Materials and methods

2.1. Preparation of the synthetic secondary effluent

The pre-experimental processes involved with the preparation of the synthetic wastewater included two significant parts: the preparation of the *E. coli* suspension and the actual wastewater, performed as follows.

2.1.1. Preparation of the bacterial cultures

The selected microorganism was an *E. coli* K12 strain (MG 1655) and was provided from “Deutsche Sammlung von Mikroorganismen und Zellkulturen”. Pre-cultures supplied a colony intended for loop-inoculation in sterile Luria-Bertani broth (10 g Bacto™ Tryptone, 5 g Yeast extract, and 10 g NaCl, per liter of distilled water). After incubation overnight and collection in the stationary phase, bacteria were washed three times, by centrifugation at 5000 rpm, with a neutral pH pre-sterilized saline solution, containing 8 g/L NaCl and 0.8 g/L KCl. The result was a bacterial suspension of 10^9 CFU/mL, approximately.

2.1.2. Composition of the synthetic wastewater

The wastewater employed was described analytically elsewhere [20]. The preparation of the synthetic wastewater took place as follows: 160 mg/L peptone, 110 mg/L meat extract, 30 mg/L urea, 28 mg/L K_2HPO_4 , 7 mg/L NaCl, 4 mg/L $CaCl_2 \cdot 2H_2O$ and 2 mg/L $Mg_2SO_4 \cdot 7H_2O$. The initial COD was 250 mg/L COD. In order to better approximate the values of secondary effluent, a 10% dilution was used [21]. 1 mL of concentrated (10^9) bacterial solution per liter was dispersed in the solution, to reach an initial population of 10^6 CFU/mL. Consecutive dilutions were done to achieve the lower initial populations.

2.2. Simulated solar light specifications

The light source was a bench-scale Suntest solar simulator from Hanau, employing a 1500 W air-cooled Xenon lamp, with effective illumination surface of 560 cm². A portion of 0.5% of the emitted photons fall within a range shorter than 300 nm (UVB) and 7% in the UVA area (320–400 nm). After 400 nm, the emission spectrum follows the solar spectrum. The solar simulator also contains an uncoated quartz glass light tube and cut-off filters for UVC and IR wavelengths. The three intensity levels employed in this study (0, 800 and 1200 W) were monitored by a Global and UV radiometer (Kipp & Zonen Mod. CM3 and CUV3). Concerning the applied intensities, 800 W/m² is a feasible value of solar irradiance, in the areas candidate for solar disinfection, in general. On the other hand, 1200 W/m² is a relatively high value chosen in purpose, defining (i) a neighboring value to the highest intensity able to reach earth's crust and (ii) a value with profound results, in order to stress the modifications in bacterial kinetics; our investigations (data not shown) indicated that values around 1000 W/m² had the desired effect, but not as obvious as the presented ones.

2.3. Reactor configuration

The batch tests that withheld the bacterial samples were cylindrical double-wall Pyrex glass bottle reactors (outer diameter 7.5 cm, inner diameter 6.5 cm, height 9 cm, irradiation surface 20.41 cm²), which allow control of the temperature and UVB transmission, as well as mild stirring with magnetic stirrer. Water was taken from the body of the irradiated sample, still under stirring.

Table 1
Experimental design parameters, levels and respective units.

Parameters		Levels	Units
Time	4	1, 2, 3, 4	h
Initial population	4	10^3 , 10^4 , 10^5 , 10^6	CFU/mL
Temperature	5	20, 30, 40, 50, 60	°C
Light intensity	3	0, 800, 1200	W/m ²

2.4. Bacterial enumeration

Bacterial colonies were enumerated by the pour-plating method on 9-cm petri dishes containing PCA agar. Samples were properly diluted to maintain measurable counts on the Petri dishes (15–150 colonies per plate). The detection limit for diluted samples is 10 CFU/ml, and 1 CFU/mL for the undiluted [22]. In all cases, even under 15 colonies per dish, the actual reading of CFU/mL is reported here. In each measurement, plating was done in duplicates, and 5% difference (maximum 10% in low numbers) was obtained. Therefore, for clarity, error bars of the average counts will not be plotted.

2.5. Design of Experiments (DOE) set-up

Full factorial DOE was chosen to investigate the influence of the important parameters of treatment time, temperature and initial bacterial population, on the disinfection process and their possible synergies and/or interactions. When a full factorial DOE is chosen, the responses are measured at all the combinations in the different experimental levels. Combining the different factor levels reflects the conditions in which the various responses are measured by the actual experiments. It was chosen over fractional factorial design to prevent confounding and data credibility loss. Table 1 presents the parameters under study and their respective levels.

MINITAB for Windows was used for the data analysis. The DOE was configured as a Multilevel Full Factorial Design, because of the different levels within the parameters. The timespan of the experiment is 4 h, initial bacterial population was chosen to vary from 10^3 to 10^6 CFU/mL and temperature was analyzed for five levels, from 20 to 60 °C. Data analyses are presented grouped by light intensity levels: (i) 0 W, (ii) 800 W and (iii) 1200 W. Table 2 summarizes the DOE for each intensity level. In our work, temperature and irradiation were varied in order to obtain a range of conditions. Some of the conditions tested are hardly feasible in natural conditions, but achievable, for instance, with mechanical assistance by CPCs or solar collectors. The experimental set-up allowed controlling

Table 2
Design of experiments set-up.

Run	A	B	C	Run	A	B	C	Run	A	B	C	Run	A	B	C
1	1	1	1	21	2	1	1	41	3	1	1	61	4	1	1
2	1	1	2	22	2	1	2	42	3	1	2	62	4	1	2
3	1	1	3	23	2	1	3	43	3	1	3	63	4	1	3
4	1	1	4	24	2	1	4	44	3	1	4	64	4	1	4
5	1	1	5	25	2	1	5	45	3	1	5	65	4	1	5
6	1	2	1	26	2	2	1	46	3	2	1	66	4	2	1
7	1	2	2	27	2	2	2	47	3	2	2	67	4	2	2
8	1	2	3	28	2	2	3	48	3	2	3	68	4	2	3
9	1	2	4	29	2	2	4	49	3	2	4	69	4	2	4
10	1	2	5	30	2	2	5	50	3	2	5	70	4	2	5
11	1	3	1	31	2	3	1	51	3	3	1	71	4	3	1
12	1	3	2	32	2	3	2	52	3	3	2	72	4	3	2
13	1	3	3	33	2	3	3	53	3	3	3	73	4	3	3
14	1	3	4	34	2	3	4	54	3	3	4	74	4	3	4
15	1	3	5	35	2	3	5	55	3	3	5	75	4	3	5
16	1	4	1	36	2	4	1	56	3	4	1	76	4	4	1
17	1	4	2	37	2	4	2	57	3	4	2	77	4	4	2
18	1	4	3	38	2	4	3	58	3	4	3	78	4	4	3
19	1	4	4	39	2	4	4	59	3	4	4	79	4	4	4
20	1	4	5	40	2	4	5	60	3	4	5	80	4	4	5

Factors	3
Replicates	2
Base runs	80
Total runs	160
Base blocks	1
Total blocks	1
No. of levels	A;B;C; = 4;4;5
A	Time (h)
B	Initial population (CFU/mL)
C	Temperature (°C)
where:	
A	1, 2, 3, 4 (h)
B	10^3 , 10^4 , 10^5 , 10^6 (CFU/mL)
C	20, 30, 40, 50, 60 (°C)

the temperature at desired levels. The point of this study was to investigate the potential synergies and antagonistic effects that temperature would create and influence, during solar disinfection. In any case, with this method of artificial temperature control, we expect to observe the possible combined effects, positive or negative, according to the potential acquired temperatures in a solar disinfection application.

Also, MINITAB was used to display both the sequential sums of squares (Seq SS) and adjusted sums of squares (Adj SS), after the presentation of the Degrees of Freedom (DF). Since the model is orthogonal and does not contain covariates, these two SS values will be the same. The SS quantifies the variability between the groups of interest, here being the values of the first column, the control variable (process efficiency %). In other words, the difference between the source means and the grand mean is represented. Variation between individual scores and the mean of every group is presented by the values; the greater this value is, the bigger the effect of changing that factor on our control variable is.

3. Results

3.1. Dark experiments (0 W/m²) – effects of reaction time, temperature and initial population in absence of light

Fig. 1a presents the evolution of bacterial population over time, within the varying initial population and the corresponding temperature conditions. The figure can be split into two major groups of curves showing clearly different behavior and their respective sub-groups: (i) for temperatures 20–40 °C and (ii) 50–60 °C. In the absence of light, the driving force of the reaction is temperature alone. The initial bacterial population sets the bar from which we observe the initiation of the thermal impact. The contour plot of the removal (Fig. 1b, % process efficiency) over time and provides an overview with a clear ineffective area (20–40 °C) and a thermal effect one. However, the main effects plot (Fig. 1c) does not clearly present the effect of the different temperature ranges and provide a false, rather masked image by the overall means; time for instance seems to be biased by the different efficiencies noticed in Fig. 1a and b and presents quite mild influence in the process, which is not true. Therefore, the graphs are presented divided according to the temperature range they belong, in Fig. 2.

For the first group of graphs, in Fig. 2a (20–40 °C), it can be noticed that there is a slight increase in the bacterial count.

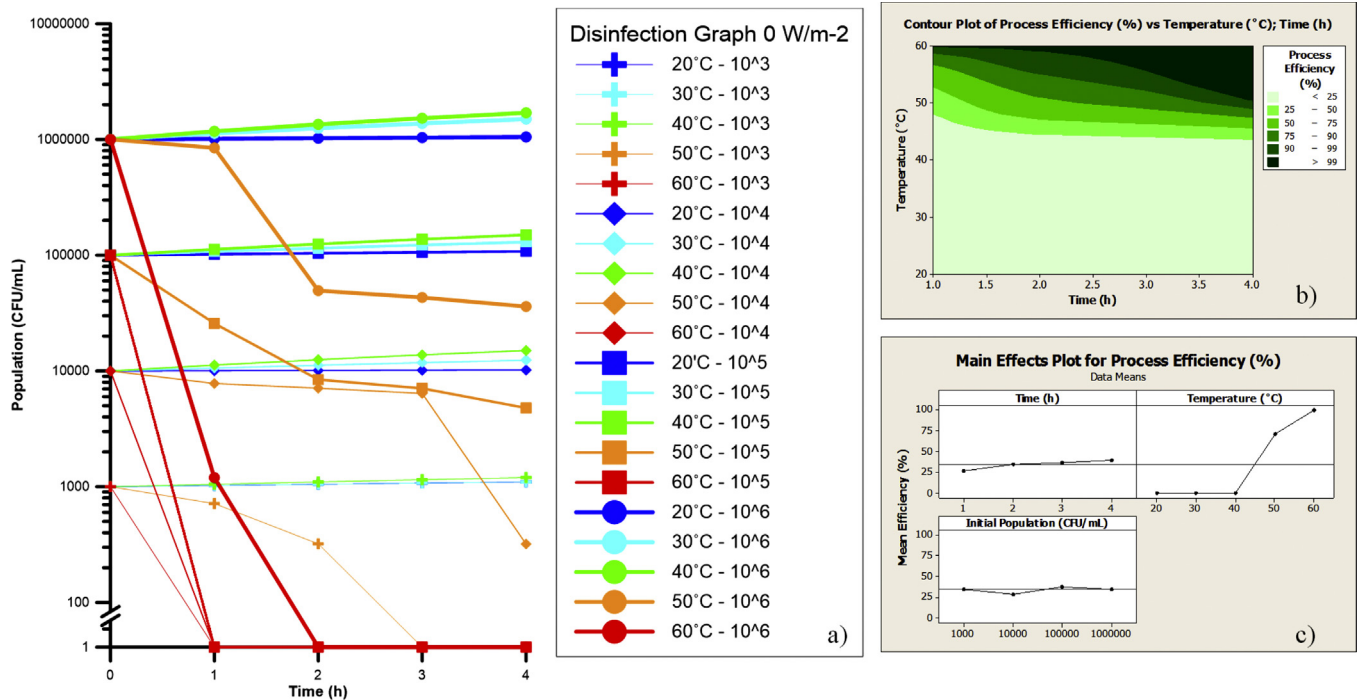


Fig. 1. Main results of non-irradiation experiments for synthetic secondary effluent at different temperatures and initial *E. coli* populations. (a) Disinfection kinetic curves; (b) contour plot of process efficiency vs. temperature and time; (c) main effects plot (control variable: Process Efficiency).

The water matrix supporting the bacterial population is synthetic wastewater and, due to the existing nutrient sources, growth is expected. This observation is valid for all initial bacterial levels and within this increase, there are two tendencies present: Firstly, there is a correlation between the temperature of the reaction and the final bacterial numbers. The 40°C traces are always higher than the 30°C traces and them, always higher than the 20°C traces. This behavior agrees with the fundamental findings of Johnson and Lewin [23] that attribute higher cell division rates in higher temperatures. Secondly, there is a statistical observation presented in Table 3 that generally, higher initial numbers lead to larger percentile increases of population, when the critical temperature is reached.

This behavior changes radically above 40°C. In the same table (Table 3), we observe that the curves indicate bacterial inactivation instead of bacterial growth. *E. coli* are mesophilic microorganisms, that typically thrive between 20 and 45°C [24]. Above this temperature, there is a thermal stress applied to the cells, altering the cell wall plus damaging the proteins and nucleic acids, leading to easier bacterial death [25]. This effect becomes more visible (Fig. 2b) as we increase the temperature from 50°C to 60°C. Treating *E. coli* within high temperatures can result to total inactivation as it can be seen for low initial populations, but slightly more difficult when the initial population is high. Also, we observe that temperatures as high as 60°C lead to fast inactivation. This is attributable to the increased degradation of the vital components of the cell, by

Table 3
Percentile change of bacterial concentration after 4 h of treatment in absence of solar light.

	10 ³	10 ⁴	10 ⁵	10 ⁶
20°C	10	2	8	5
30°C	10	24	30	50
40°C	20	50	50	70
50°C	-100	-96.8	-95.2	-95
60°C	-100	-100	-100	-100

the decomposition mechanisms characterized many decades ago [23,26] (Table 4).

As far as the efficiency of the process is concerned, in terms of removal percentage, we notice the variation in Fig. 2c, which demonstrates the modification of the process, when temperature is increased from 40 to 60°C. We observe that maximum efficiency is achieved at 60°C after 1 h and as the time passes, the thermal threshold is lower, reaching 51°C, for a 4-h period of treatment. An increase of 10°C achieves dramatic enhancement in removal rates (up to 75%) and the last 10°C increase ensures total inactivation (Fig. 2d). The significance of temperature is verified by the P values of the ANOVA table (Table 5) produced by all data from MINITAB, which validates the previous results; in order of significance, temperature is the most important factor that influences the outcome, then treatment time, while the initial population is the least significant among the three.

3.2. 800-W/m² experiments – effects of reaction time, temperature and initial population for intensity of 800 W/m²

The second group of experiments utilizes solar energy to inactivate *E. coli*, with the irradiance of the solar simulator set at 800 W/m². The same batch test configurations were used as the control experiments, to ensure comparability among the conditions. Many authors have demonstrated that there is a synergistic action between light and temperature in different media

Table 4
Percentile removal of bacterial concentration after 4 h of treatment under 800 W/m² light.

	10 ³	10 ⁴	10 ⁵	10 ⁶
20°C	90.0	88.0	87.5	93.3
30°C	87.0	86.7	68.8	93.3
40°C	47.4	30.0	15.8	25.0
50°C	100.0	100.0	99.9	99.9
60°C	100.0	100.0	100.0	100.0

and microorganisms [2–4,22,27]. This test investigates the light-temperature interaction in synthetic secondary effluent.

Fig. 3a demonstrates in overall the evolution of bacterial population over time, grouped by initial numbers and temperature of the process. Within 4 h of treatment, samples that were processed at 20 °C demonstrated a continuous decrease of the population. However, as temperature rises to 30 °C the remaining populations are somewhat equal or higher than the respective ones at 20 °C. The phenomenon is even clearer at temperatures around 40 °C, where insignificant removal rates are demonstrated and presented in Table 5. Fig. 3b presents an overview of the efficiency of the process, in which we notice a gap, around 40 °C. There is a descending trend until 40 °C and then an increase in the efficiency, which is verified in Fig. 3c; temperature is dominating the process and modifies the outcome of the experiment. Therefore, once again we observe the two clear groups of graphs, according to the large temperature groups (i) 20–40 °C and (ii) 40–60 °C.

Within this system there are two opposing forces present that determine the outcome so far. Compared to the 0-W/m² experiments, first of all, (Figs. 3a and 4a–b) light changes the growth phenomenon observed before. What appears in Fig. 3b(i) as an “island” of low efficiency among the average ones, is attributed to the 40 °C area, which provides with increased metabolic rates and thereby higher remaining populations. On the one hand, we have the disinfecting action of light, which tends to inactivate bacteria as seen in 20 °C curves (Fig. 4a), with the number of inactivated bacteria vs. initial population increasing when initial population is increased. On the other hand, submitting the population to temperatures around 37 °C, in a nutrient-enhanced matrix as the wastewater, mesophiles, such as *E. coli*, tend to present their highest reproduction rates [28]. *E. coli* belongs to this category and is encountered in the human gut [29], with the normal human body temperatures being the most favorable for their growth. Normally, *E. coli* are inactivated by exposure to 55 °C for 1 h or 60 °C for 20 min [30]. Hence, as we raise the temperature in the disinfection process, the two concurrent actions tend to balance in favor of the reproduction rates, around 40 °C.

However, a temperature increase over 45 °C would affect *E. coli* metabolic cycles, and lead to cell death. Indeed, as it is observed, the 50 °C curves (Fig. 4b) after an initial shoulder, a common observation at solar disinfection processes [31–33], then present total (10³ and 10⁴ curves) and almost total inactivation (10⁵ and 10⁶ curves). In addition, we verify that increasing the treatment temperature up to 60 °C leads to total inactivation of the microorganisms before 60 min, regardless of the initial bacterial population.

Furthermore, one can notice the synergy between light effects and temperature increase at the graphs, by comparing Fig. 2 with Fig. 4: First of all, at 50 °C without light, only samples with 10³ initial population were inactivated, whereas in presence of light 10³ and 10⁴ were totally inactivated and 10⁵ and 10⁶ presented a 3 or 4 log₁₀ U reduction instead of 1 or 2 log₁₀ U. Secondly, 60 °C treated samples were totally inactivated in less than an hour, slightly faster than in absence of light. Consequently, in the latter case thermal treatment is the main disinfecting force and light is only complementary.

Speaking in terms of efficiency, Fig. 4a-i and b-i, provide information about the effect of each parameter over the total inactivation capability of the process. In the 20–40 °C interval, lower temperatures seem to favor inactivation with the peak appearing between 20 and 25 °C, while treatment time increases the potentials; the 4th hour contributes in the greatest proportion, adding on the inactivation side of the balance. Comparing with the equivalent graphs for 40–60 °C, temperature increase leads to percentile inactivation enhancement, while statistically in both cases, initial bacterial population does not seem to significantly affect the percentage of inactivated bacteria in the process. However, the same

actions manage to inactivate lower bacterial numbers more efficiently (in percentage) but in absolute numbers, removal increases with higher populations, due to larger numbers' correspondence of the removal percentage.

Finally, the ANOVA table reveals the important contribution of time and temperature and the milder one from initial bacterial population. The *P* values presented in Table 5 are also verified by Fig. 4a-ii and b-ii. We draw the information that time almost proportionally increases the total efficiency, while initial population fluctuates around the average inactivated bacteria. What is more important, is the temperature effect on efficiency, which presents what was in detail described before; temperature increase enhances bacterial inactivation, as literature suggested for other water matrices, but only above 40 °C. Otherwise, the disinfection process is delayed by the excessive growth of the microorganisms.

3.3. High irradiance experiments (1200 W/m²) – effects of reaction time, temperature and initial population for high intensity irradiation conditions

The final experimental part consists of the runs that utilized high intensity illumination. Higher supply of photons in the system could result to higher possibility of effective hits in the *n* number of crucial areas of the cell, as described by Harm [31].

Fig. 5a presents an overview of the disinfection reactions of the varied initial population. The main and most profound difference between this set-up and the previous ones, is that all samples regardless of initial population and treatment temperature have been inactivated within the time frame of 4 h. The action of light was more intense and influenced the outcome of the experiments in cases that was not sufficient before. Bacteria have now to cope with higher concurrent light and thermal action, which is expressed by less acute kinetics in the final hour. When the samples were treated at lower temperatures, the disinfection curves again present a lag-phase or shoulder, but considerably lower, varying from 30 to 120 min, compared to the minimum 3-h shoulder presented under 800 W/m² irradiance. Fig. 5b and c, presents once more the erratic behavior around 40 °C, demonstrated as a lower efficiency area (Fig. 5b) or a mean decrease (Fig. 5c), however mitigated, compared to the equivalent of 800 W/m² or even the increase in numbers observed in null intensity experiments.

What is more, the main effects plot of this high irradiance also adds direct information over the main overall efficiency. All parameters concerned, the addition of light initially increased the efficiency from 35% to 65% (from 0 to 800 W/m²), to reach 80% when high intensity is applied. This is a good indicator of the robustness of the system, predicting, at some extent, the success of the group of trials. Finally, it is also shown that the biggest contribution in bacterial inactivation derives from the 1st hour of illumination and the least, but most important from the application point of view, during the 4th hour. Plus, the drop in efficiency around 40 °C is also visible, like each previous case but less intense; high irradiance illumination compensates for the inactivation difference.

Observing Fig. 6a, it is clear that, at 1200 W/m², the equilibrium between the disinfecting action of light and the growth-stimulating effect of increasing temperatures changes within the 20–40 °C range. After a 2–3-h shoulder, bacterial numbers fall sharply to total disinfection at the fourth hour. This means the disinfecting action becomes higher than the growth force and, as far as the cell is concerned, indeed, the growth action is present but is no longer in favor of their survival. Also, the contour plot of efficiency over time (Fig. 6a-i) has a clear area of total inactivation, after 3.5 h, while temperature increase has a mitigated effect of delay in inactivation, compared to all other cases till now.

For higher temperatures, it is shown that at 50 °C, compared to 0 and 800 W/m², the same process at 1200 W/m² is completed

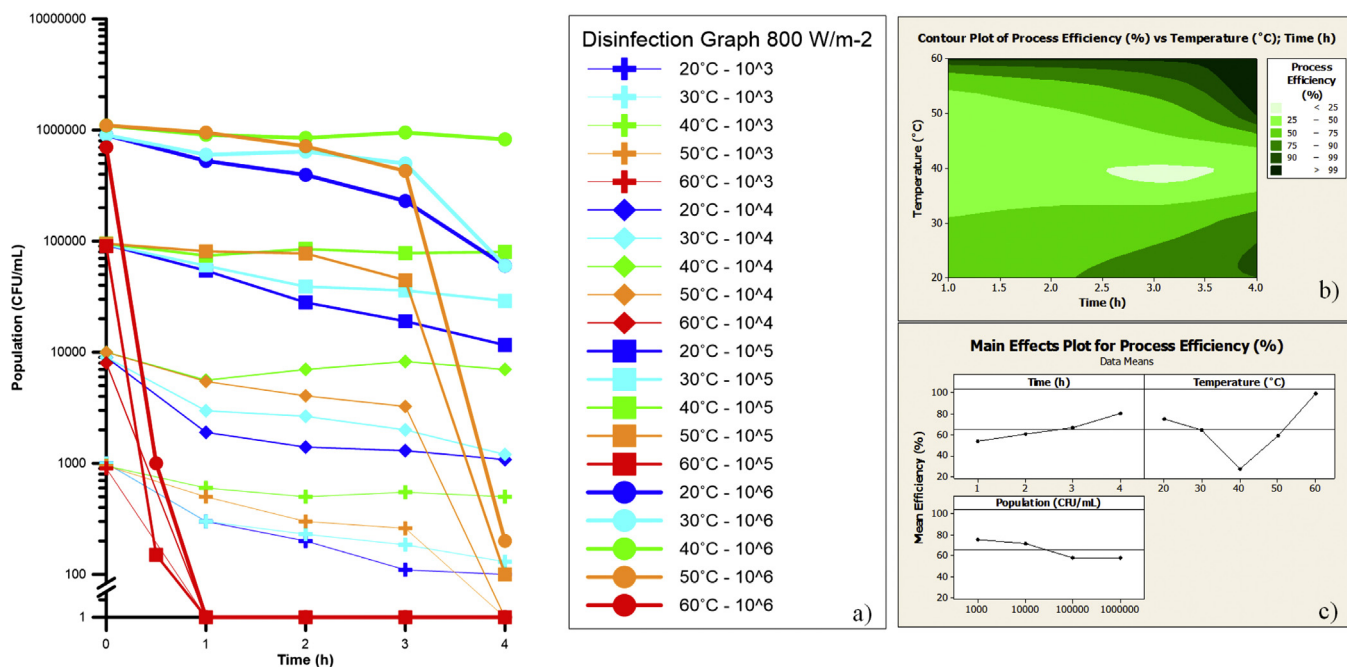


Fig. 3. Main results of 800 W/m² experiments for synthetic secondary effluent at different temperatures and initial *E. coli* populations. (a) Disinfection kinetic curves. (b) Contour plot of process efficiency vs. temperature and time. (c) Main effects plot (control variable: Process Efficiency).

faster, compared with the cases it was completed before, and in total, all cases resolved to total inactivation. As shown in Fig. 6b, the disinfection kinetics at these particular conditions (1200 W/m², 50°C) is very sensitive to the initial bacterial concentration, probably attributable to shielding [34] playing a critical role in these runs. At 60°C and 1200 W/m², complete disinfection is achieved faster than at 0 or 800 W/m². Where in absence of light inactivation time was around an hour, at 800 W/m² slightly less, and now is even less than 30 min. Finally, this outcome is common for all initial populations; all result to total inactivation faster than their respective 800 W/m² curves.

The contour plots of the process efficiency (Fig. 6b-i) indicate clearly the bigger “effective” area of >99%, and the relatively higher rates; no area lies under 50% bacterial inactivation even after only 1 h. As treatment time increases the efficiency increases as well, however, the same cannot be observed for temperature. For instance, at 50°C, only 2.5–3 h are sufficient to achieve total inactivation, demonstrated in Fig. 6b-ii. Also, from the ANOVA table we draw the information that the efficiency is highly correlated only with treatment time and temperature.

3.4. Modeling solar disinfection of secondary treated wastewater

As a result of the statistical interpretation of the experimental data, a simplified model can be proposed. Through the statistical software of MINITAB, a model is suggested, which relates the response factors with the parameters of the process, in order to further analyze the experimental concept, and help facilitate all these experimental runs [19].

In our experiments, the parameters involved in the process were treatment time, temperature, initial population and light intensity. Furthermore, in order to achieve a decent fitting model, the interactions of the parameters were used; the first-order model (20–60°C) without interactions yields R-sq=51.17% (model not shown). The ANOVA tables have indicated initial population as relatively insignificant; however, we choose to model all the

experiments in one equation and include it in the model, expressed as follows:

$$\begin{aligned} \text{process efficiency}(\%) = & -41.60 - 8.43t + 1.76T - 2.20e - 005C \\ & + 0.02I + 0.27t \times T + 5.03e - 006t \times C + 0.036t \times I + 3.77e \\ & - 007T \times C \\ & - 7.26e - 005T \times I + 6.21e - 009C \times I - 6.85e - 008t \times T \times C \\ & - 0.001t \times T \times I - 3.42e - 010T \times C \times I + 4.67e - 011t \times T \times C \times I \end{aligned}$$

$$\begin{aligned} S = 24.4245, R\text{-Sq} = 65.10\%, R\text{-Sq}(\text{adj}) = 62.93\%, \text{PRESS} \\ = 150, 725, R\text{-Sq}(\text{pred}) = 60.81\% \end{aligned}$$

Fig. 7a demonstrates the level of approach. The R-Sq, as a general indicator of the success of the fit, gives a 65% of match. In addition, we present the coefficients and ANOVA table for the model (Table 6), confirming the small contribution of the initial population to the model. This figure represents the 240 experiments conducted in these conditions and X axis presents the order of experimental runs, from 1 to 240. Each X value corresponds to an Efficiency value, shown in Y axis. The difference between the experimental and the calculated value (linear model values) is shown by the distance among the two corresponding marks. We can see that the trends are similar; the values follow the same tendency and are relatively close.

However, following the same principle noticed in the disinfection experiments, we can propose splitting the data in two sets, of lower and equal to 40°C and to higher than 40°C. Even though the use of interactions suggests the introduction of the synergies (especially light and temperature) in the model, we face a possible danger of over-fitting and un-necessary complexity in a simple concept, like the general linear model. For the above reasons, we introduce a temperature-dependent linear model, without the use of interactions between the parameters.

$$\text{process efficiency}(\%) = \begin{cases} 4.21 + 10.47t - 0.89T - 4.29 \times 10^{-6}C + 0.06I, & \text{for } 20 \leq T \leq 40^\circ\text{C} \\ -139.68 + 8.57t + 3.47T - 3.34 \times 10^{-6}C + 0.02I, & \text{for } 40 < T \leq 60^\circ\text{C} \end{cases}$$

First of all, the coefficients are included in Table 6. We observe that the new model has more advantages than the formerly

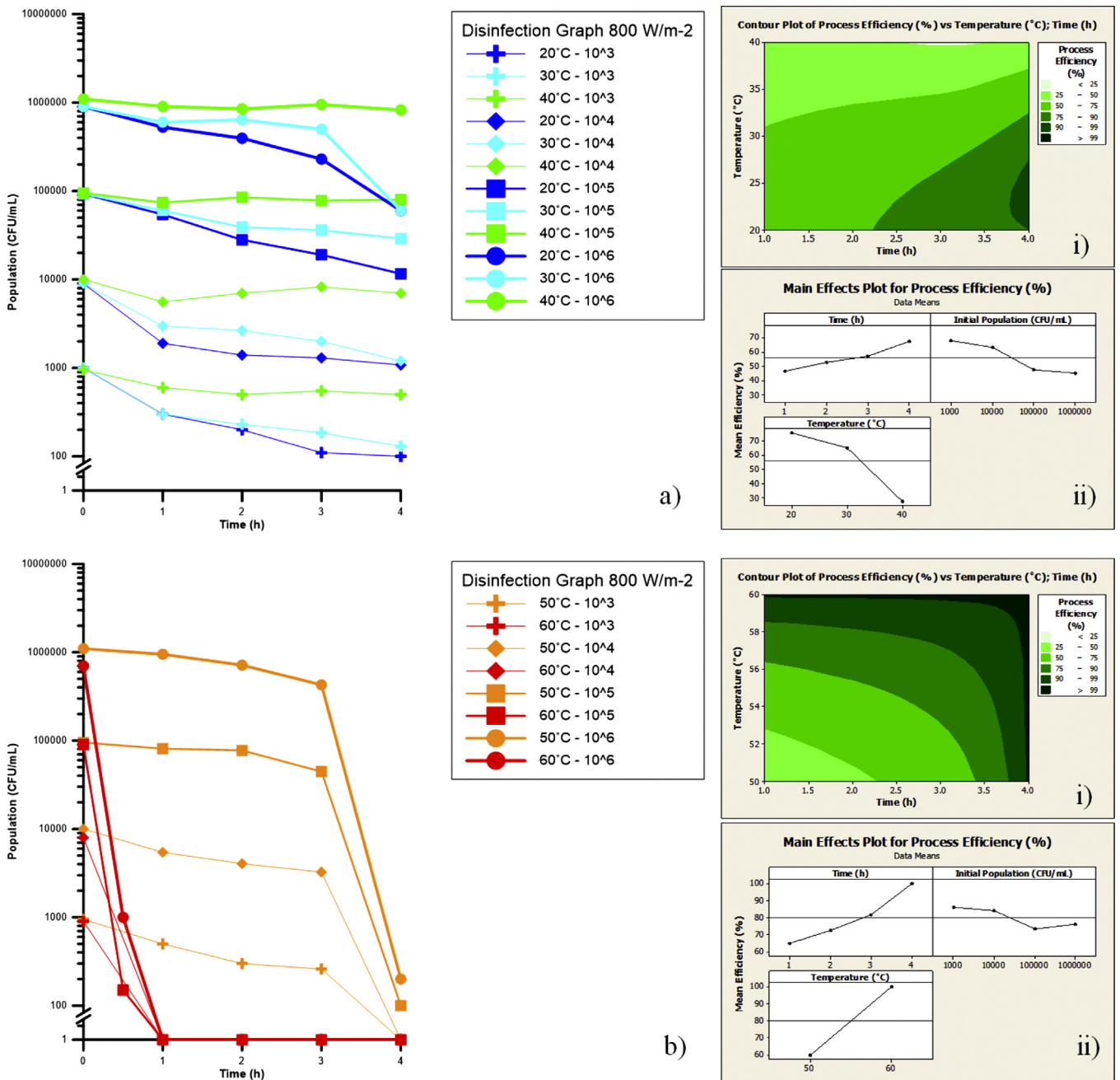


Fig. 4. Main results of 800 W/m² experiments for synthetic secondary effluent at different temperatures and initial *E. coli* populations. (a) 20–40 °C disinfection kinetic curves, (a-i) 20–40 °C contour plot of process efficiency vs. temperature and time, (a-ii) 20–40 °C Main effects plot (control variable: Process Efficiency). (b) 50–60 °C Disinfection kinetic curves, (b-i) 50–60 °C contour plot of process efficiency vs. temperature and time, (b-ii) 50–60 °C Main effects plot (control variable: Process Efficiency).

suggested; the regression standard error (S) is lower, it does not use 2nd level interactions and in addition, it yields higher R^2 values. Therefore, it is a simpler and more accurate model, describing in better extent the evolution of the process efficiency. Fig. 7b and c presents in separate plots the experimental values acquired versus the predicted ones from the model (40 °C plotted in both figures for better demonstration of the temperature evolution). All things considered, we suggest that this temperature-dependent model is a good indicator of the tendencies present in solar disinfection of wastewater or an estimating tool concerning the remaining population within some range, rather than an actual predictor of the efficiency.

4. Discussion

4.1. Inactivation mechanism: light source and bacterial damage

Solar disinfection experiments were conducted under a solar simulator that emits all spectrum from 290 nm and above, excluding infrared wavelengths, due to the existence of cut-off filters. Therefore, the actions expected should be attributed UVB, UVA and visible light.

4.1.1. UVB irradiation

Matallana-Surget et al. [35] have stated the double action of UVB irradiation; in general, UVB damage is considered to mainly

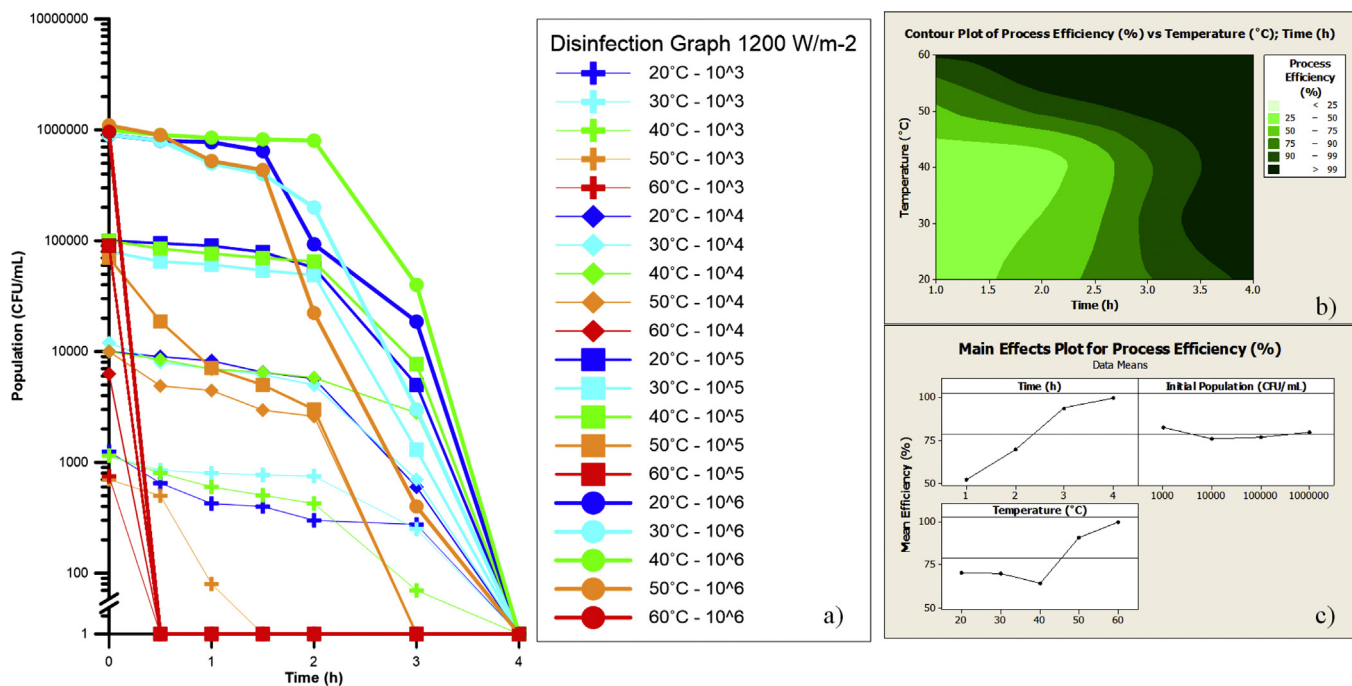


Fig. 5. Main results of 1200 W/m² experiments for synthetic secondary effluent at different temperatures and initial *E. coli* populations. (a) Disinfection kinetic curves. (b) Contour plot of process efficiency vs. temperature and time. (c) Main effects plot (control variable: Process Efficiency).

Table 6

Summary of the statistical parameters of the full interaction model and the temperature-dependent models.

Coefficients	Analysis of variance										Summary of model
	Coef	SE	T	P	DF	Seq SS	Adj SS	Adj MS	F	P	
Model 1											
Constant	-41.6006	21.2256	-1.9599	0.0510	14.0	250386.0	250386.0	17884.7	29.9798	0.0000	S = 24.4245
t	-8.4277	7.5979	-1.1092	0.2690	1.0	28413.0	734.0	734.0	1.2304	0.2685	PRESS = 150725
T	1.7565	0.5057	3.4736	0.0010	1.0	83168.0	7198.0	7197.9	12.0658	0.0006	R-Sq = 65.10%
C	0.0000	0.0000	-0.6980	0.4860	1.0	402.0	291.0	290.6	0.4872	0.4859	R-Sq(adj) = 62.93%
I	0.0178	0.0241	0.7418	0.4590	1.0	84823.0	328.0	328.3	0.5503	0.4590	R-Sq(pred) = 60.81%
t × C	0.0000	0.0000	0.4976	0.6190	1.0	656.0	148.0	147.7	0.2476	0.6192	
t × T	0.2726	0.1815	1.5023	0.1340	1.0	1801.0	1346.0	1346.3	2.2568	0.1344	
t × I	0.0355	0.0085	4.1911	0.0000	1.0	7360.0	10479.0	10478.5	17.5650	0.0000	
T × C	0.0000	0.0000	0.4779	0.6330	1.0	26.0	136.0	136.2	0.2283	0.6332	
T × I	-0.0001	0.0006	-0.1257	0.9000	1.0	37360.0	9.0	9.4	0.0158	0.9001	
C × I	0.0000	0.0000	0.2742	0.7840	1.0	83.0	45.0	44.8	0.0752	0.7842	
t × T × C	0.0000	0.0000	-0.2630	0.7930	1.0	15.0	41.0	41.2	0.0691	0.7928	
t × T × I	-0.0007	0.0002	-3.1913	0.0020	1.0	6122.0	6075.0	6075.4	10.1841	0.0016	
T × C × I	0.0000	0.0000	-0.5146	0.6070	1.0	107.0	158.0	158.0	0.2648	0.6073	
t × T × C × I	0.0000	0.0000	0.2936	0.7690	1.0	51.0	51.0	51.4	0.0862	0.7693	
Error					225.0	134225.0	134225.0	596.6			
Total					239.0	384611.0					
Model 2 (≤40 °C)											
Constant	4.2149	7.7882	0.5420	0.5890	4.0	151682.0	151682.0	151682.0	94.3	0.0	S = 20.0507
t	10.4743	1.4945	7.0086	0.0000	1.0	19748.0	19748.0	19748.0	49.1	0.0	PRESS = 60045.2
T	-0.8980	0.2046	-4.3882	0.0000	1.0	7742.0	7742.0	7742.0	19.3	0.0	R-Sq = 73.08%
C	0.0000	0.0000	-1.0770	0.2830	1.0	466.0	466.0	466.0	1.2	0.3	R-Sq(adj) = 72.30%
I	0.0588	0.0034	17.5428	0.0000	1.0	123726.0	123726.0	123726.0	307.8	0.0	R-Sq(pred) = 71.07%
Error					139.0	55882.0	55882.0	402.0			
Total					143.0	207564.0					
Model 2 (>40 °C)											
Constant	-139.6750	12.3560	-11.3042	0.0000	4.0	144740.0	144740.0	36185.0	75.3	0.0	S = 21.9270
T	8.5700	1.6343	5.2435	0.0000	1.0	13219.0	13219.0	13219.0	27.5	0.0	PRESS = 71869.3
T	3.4670	0.2238	15.4915	0.0000	1.0	115385.0	115385.0	115385.0	240.0	0.0	R-Sq = 68.41%
C	0.0000	0.0000	-0.7650	0.4460	1.0	281.0	281.0	281.0	0.6	0.4	R-Sq(adj) = 67.50%
I	0.0210	0.0000	5.7425	0.0000	1.0	15855.0	15855.0	15855.0	33.0	0.0	R-Sq(pred) = 66.03%
Error					139.0	66831.0	66831.0	481.0			
Total					143.0	211571.0					

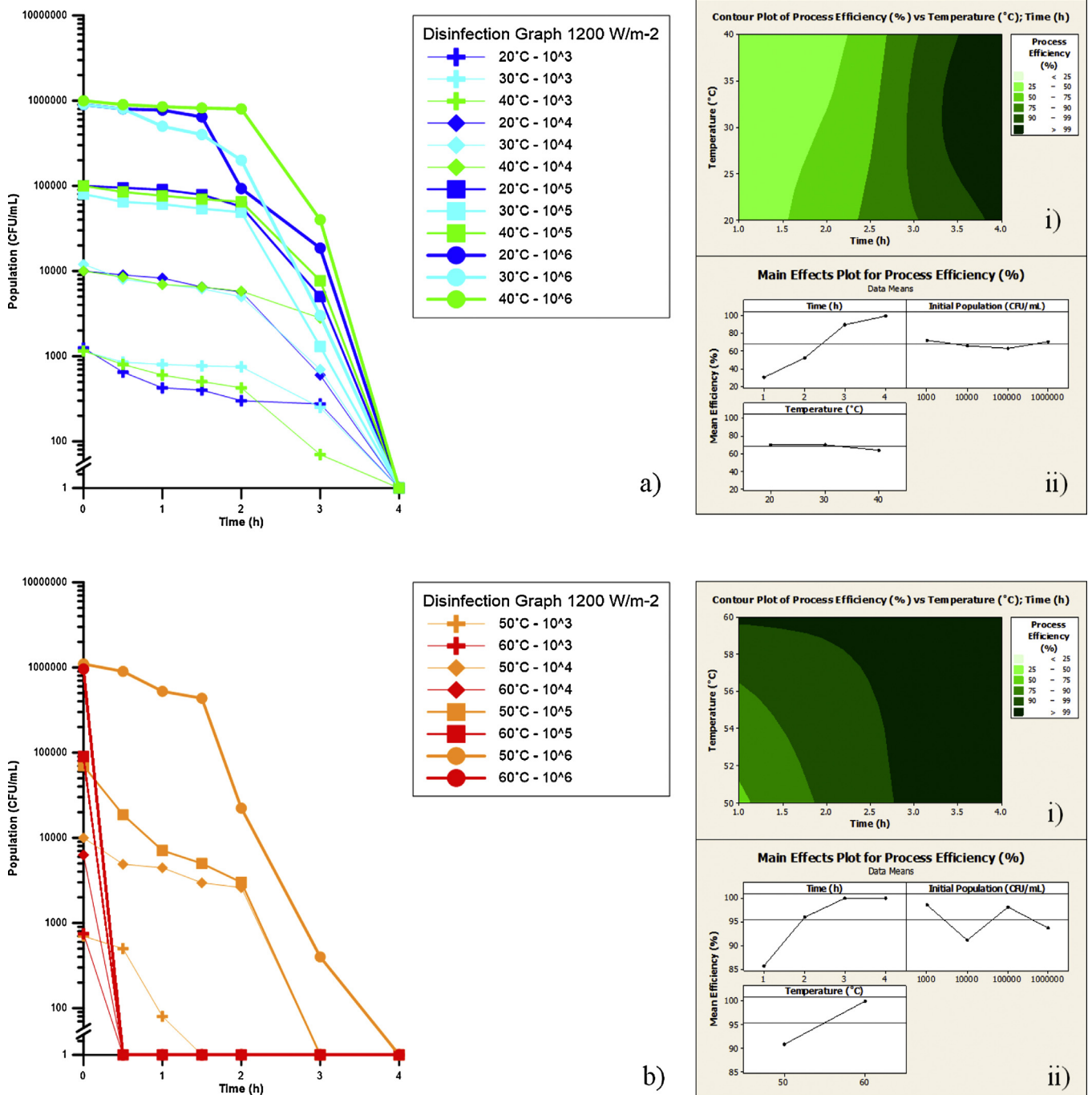


Fig. 6. Main results of 1200 W/m² experiments for synthetic secondary effluent at different temperatures and initial *E. coli* populations. (a) 20–40 °C Disinfection kinetic curves, (a-i) 20–40 °C contour plot of process efficiency vs. temperature and time, (a-ii) 20–40 °C Main effects plot (control variable: Process Efficiency). (b) 50–60 °C Disinfection kinetic curves, (b-i) 50–60 °C contour plot of process efficiency vs. temperature and time, (b-ii) 50–60 °C Main effects plot (control variable: Process Efficiency).

cause direct DNA damage, through the creation of photoproducts (cyclobutane pyrimidine dimer and the pyrimidine (6–4) photoproducts) [36]. They also mention the creation of internal and external reactive oxygen species (ROS) such as hydrogen peroxide (H₂O₂), and more profoundly, the creation of singlet oxygen [37]. These ROS attack nucleic acid, proteins and cell lipids [38]. However, UVB is very often overlooked, although it has a relatively high contribution in bacterial inactivation. The important impact of UVB inactivation of bacteria has been stated since 1974 [39], there are very few works that add up to this wavelength band to attribute part of the bacterial inactivation. Of course, this

happens due to the sensitivity of UVB to meteorological phenomena, but this is far from our case, and we cannot ignore a force two or three orders of magnitude higher than UVA [40,41]. Also, the peak of UVB germicidal activity, roughly among 300 and 310 nm, is clearly within our range [41] and according to previous reports, UVB radiation of 313 nm demonstrates an interaction with the 365 nm, to enhance DNA transformation (Peak et al., 1975; Tyrrell and peak, 1978) [53,54]. Hence, we have a double UVB action, of DNA strand break, and the creation of ROS which have been identified to be implicated in bacterial inactivation through oxidative stress.

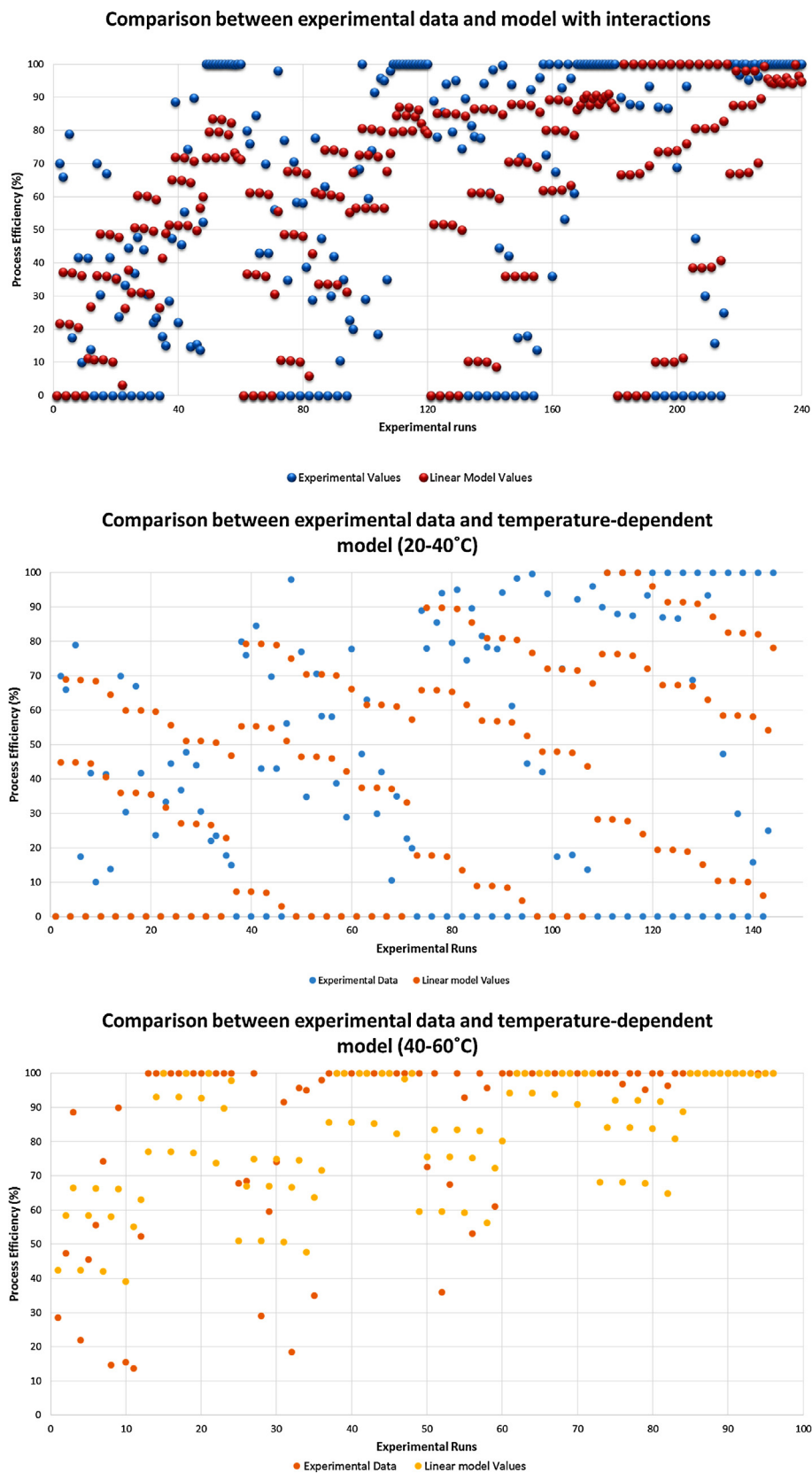


Fig. 7. Fitting of the linear models to the experimental data. (a) Linear model with interactions. (b) Temperature dependent model (20–40 °C). (c) Temperature dependent model (40–60 °C).

4.1.2. UVA irradiation/near-UV visible light

UVA-induced loss of bacterial cultivability is attributed to the catalysis of the formation of ROS. It is the least effective irradiation range to damage bacterial DNA directly, but its proven efficiency [42] comes from the biological effects of internal and external ROS attacks, such as protein destruction or adducts of nucleic acid with membrane proteins with the bacterial envelope escaping key damage, toward cell inactivation [43]. One of the first attacks is the respiratory chain and the cell's potential to produce ATP (Bosshard et al., 2010) [55]. Other attacks include internal photo-Fenton reaction [44] due to loose cell iron sources, disruption of normal internal ROS suppression mechanism (SOD, catalase etc.) (Chiang and Schellhorn, 2012) and others, all related by the ROS production inside and outside the cell. ROS are normal by-products of bacterial respiratory chain, and bacteria possess a big number of suppressive mechanisms (Mishra and Imlay, 2012) [56,57]. Hence, UVA damage is an internal/external oxidative damage, plus the internal/external photo-Fenton contribution, with measurable effects; an increase in dose can inflict greater damage [10].

4.2. Inactivation mechanism: influence of the water matrix

There is no disagreement that the majority of the solar disinfection experiments were conducted in distilled or drinking water, making the inactivation mechanism clear and well established. The main difference of this synthetic secondary effluent is the added salts and organic components. Marugán et al. [15] have explained that during bacterial osmotic stress among the first released ions are calcium and magnesium ones, while Caballero et al. [45] stated the importance of organic substances as nutrient sources for bacteria; therefore, bacterial survival/growth is favored in this matrix. Given the absence of light in the first group of experiments, growth is normal and expected, and as temperature rises, with a peak around 35–39 °C (according to our discreet choice, 40 °C growth will be increased. However, this behavior is expected to change when the irradiation is present and light is applied to the sample. The presence of organic substances can induce an indirect stress. They can either be endogenous, like porphyrins, co-enzymes or cytochromes, or exogenous, synthetic ones, which lead to either internal or external photosensitized matter. After receiving UV irradiation, this effect can cause indirect photolysis, while the photo-sensitizers are in an excited, high energetic state [16,46,47]. Other works however, have demonstrated reduction of cell inactivation, when inorganic and organic compounds were present [48–50].

4.3. Temperature influence and evolution of experiments

As demonstrated in the experimental part, temperature altered the outcome of the inactivation assays in great extent, from level to level. For this reason, the experiments were divided in two parts, below and over 40 °C degrees. Wegelin et al. [2] reported no differences between 12 and 40 °C in water, and Reed [16] explained this behavior by the double effect of temperature range. When temperature is increased, growth is favored, and inactivation as well: Thermally-driven growth is canceled by oxygen depletion, due to its lower solubility at higher temperatures. In our case, growth was favored to a point that the synergic effect was canceled, depletion of oxygen did not occur (samples under mild stirring) and eventually, until the intensity was increased over 1000 W/m², light alone could not overcome the rapid growth. Rincon and Pulgarin [51] suggested the increase in intensity to efficiently remove *E. coli*, and also, the effects of physiological bacterial state; we adopted the same techniques to ensure reproducible results.

When low temperatures were applied, metabolic activity was at its minimum, so the same actions of light battled against less

targets. This, however was not the case in temperatures around 40 °C, where excessive growth was observed, thus providing more targets for incoming photons or ROS. In addition, this excess growth can lead to extensive shielding from one cell to another [34], inducing higher inactivation rates for increased populations. In the first steps of each experiment, a shoulder is observed, and this latency effect is due to initial self-defense mechanisms [51]. As time passes and new generations of bacteria appear due to high reproduction rates, the new generations are more resistant to the disinfecting action of light, having endured the exposure of the original cells toward the actions of light. It has been stated that a greater effectiveness of applying a high intensity for a short time is demonstrated and preferred, rather than applying a lower intensity for a longer period of time [52].

5. Conclusions

Non-irradiated samples of synthetic secondary effluent treated at 20–40 °C showed slight growth during treatment. Significantly, thermal inactivation predominated at 50 °C and was total at 60 °C.

Irradiation at 800 W/m² was sufficient to suppress growth at 20–40 °C, but not for providing proper disinfection in 4 h of treatment, with efficiency decreasing with rising temperatures and showing a minimum around 40 °C. Synergy between light and temperature above 40 °C was evident, with all 60 °C samples undergoing total disinfection in just 1 h, or, at 50 °C, high disinfection efficiencies after 4 h of treatment.

Irradiation at 1200 W/m² resulted in total disinfection (no bacterial counts) in 4 h (20–40 °C), in 1.5–4 h (50 °C) or in just 0.5 h (60 °C), showing again the light-temperature synergy.

The profound actions of UVB and UVA irradiation demonstrated different results, according to the experimental temperature range, with the cases of very low and very high demonstrating the best results, due to either lower metabolic rhythm or synergy between temperature and light, plus thermal modifications of cells' proteins.

A 4-factor, multilevel, complete factorial design of experiments has proved a powerful, useful tool to evaluate the main variables governing disinfection. A linear model with interactions (R-Sq = 65.1%, S = 24.42) has been initially proposed and improved, when it was modified to a temperature dependent one. The new model is simpler (no interactions needed), as well as more accurate (S = 20.0507, R-Sq = 73.08%, for 20 < T ≤ 40 °C and S = 21.9270, R-Sq = 68.41% for 40 < T ≤ 60 °C). While unrelated to any fundamental modeling of the process, it has allowed to statistically determine the significant factors and interactions in the process.

As far as a potential application is concerned, the recommended practice would be to acquire the highest irradiation times possible for the given regional climatological constraints. Given the fact that real applications will be temperature-limited, the design practices should be oriented to acquiring prolonged exposure to sunlight, since we observed that extension of the treatment always favored bacterial disinfection.

Acknowledgements

The authors wish to thank, in order of acquisition, the Mediterranean Office for Youth Program (MOY, call 2011–2014), by means of which Mr. Stefanos Giannakis has received a PhD mobility grant (MOY grant no. 2010/044/01) in the joint Environmental Engineering Doctoral Program. Also would wish to thank the Swiss Government for the Swiss Government Excellence Scholarship, by means of which Mr. Stefanos Giannakis has received a Research Visit fellowship (No. 2012.0499).

References

- [1] A. Acra, Z. Raffoul, Y. Karahagopian, Solar Disinfection of Drinking Water and Oral Rehydration Solutions: Guidelines of Household Application in Developing Countries, 1984.
- [2] M. Wegelin, S. Canonica, K. Mechsner, T. Fleischmann, F. Pesaro, A. Metzler, Solar water disinfection: scope of the process and analysis of radiation experiments, *Aqua* 43 (1994) 154–169.
- [3] K. McGuigan, T. Joyce, R. Conroy, J. Gillespie, M. Elmore-Meegan, Solar disinfection of drinking water contained in transparent plastic bottles: characterizing the bacterial inactivation process, *Journal of Applied Microbiology* 84 (1998) 1138–1148.
- [4] A.-G. Rincon, C. Pulgarin, Field solar *E. coli* inactivation in the absence and presence of TiO₂: is UV solar dose an appropriate parameter for standardization of water solar disinfection? *Solar Energy* 77 (2004) 635–648.
- [5] M. Boyle, C. Sichel, P. Fernández-Ibáñez, G. Arias-Quiroz, M. Iriarte-Puña, A. Mercado, E. Ubomba-Jaswa, K. McGuigan, Bactericidal effect of solar water disinfection under real sunlight conditions, *Applied and Environmental Microbiology* 74 (2008) 2997–3001.
- [6] E. Ubomba-Jaswa, C. Navntoft, M.I. Polo-Lopez, P. Fernandez-Ibáñez, K.G. McGuigan, Solar disinfection of drinking water (SODIS): an investigation of the effect of UV-A dose on inactivation efficiency, *Photochemical & Photobiological Sciences* 8 (2009) 587–595.
- [7] M. Kositzki, I. Poullos, S. Malato, J. Cáceres, A. Campos, Solar photocatalytic treatment of synthetic municipal wastewater, *Water Research* 38 (2004) 1147–1154.
- [8] M.I. Polo-Lopez, I. García-Fernández, T. Velegraki, A. Katsoni, I. Oller, D. Mantzavinos, P. Fernández-Ibáñez, Mild solar photo-Fenton: an effective tool for the removal of *Fusarium* from simulated municipal effluents, *Applied Catalysis B: Environmental* 111 (2012) 545–554.
- [9] L. Rizzo, D. Sannino, V. Vaiano, O. Sacco, A. Scarpa, D. Pietrogiamici, Effect of solar simulated N-doped TiO₂ photocatalysis on the inactivation and antibiotic resistance of an *E. coli* strain in biologically treated urban wastewater, *Applied Catalysis B: Environmental* 144 (2014) 369–378.
- [10] M. Polo-Lopez, P. Fernández-Ibáñez, E. Ubomba-Jaswa, C. Navntoft, I. García-Fernández, P. Dunlop, M. Schmid, J. Byrne, K.G. McGuigan, Elimination of water pathogens with solar radiation using an automated sequential batch CPC reactor, *Journal of Hazardous Materials* 196 (2011) 16–21.
- [11] F. Bichai, M.I. Polo-López, P. Fernández Ibañez, Solar disinfection of wastewater to reduce contamination of lettuce crops by *Escherichia coli* in reclaimed water irrigation, *Water Research* 46 (2012) 6040–6050.
- [12] R. Davies-Colley, A. Donnison, D. Speed, C. Ross, J.a. Nagels, Inactivation of faecal indicator micro-organisms in waste stabilisation ponds: interactions of environmental factors with sunlight, *Water Research* 33 (1999) 1220–1230.
- [13] R.J. Craggs, A. Zwart, J.W. Nagels, R.J. Davies-Colley, Modelling sunlight disinfection in a high rate pond, *Ecological Engineering* 22 (2004) 113–122.
- [14] J. Mwabi, F. Adeyemo, T. Mahlangu, B. Mamba, B. Brouckaert, C. Swartz, G. Offringa, L. Mpenyana-Monyatsi, M. Momba, Household water treatment systems: a solution to the production of safe drinking water by the low-income communities of Southern Africa, *Physics and Chemistry of the Earth, Parts A/B/C* 36 (2011) 1120–1128.
- [15] J. Marugán, R. van Grieken, C. Pablos, C. Sordo, Analogies and differences between photocatalytic oxidation of chemicals and photocatalytic inactivation of microorganisms, *Water Research* 44 (2010) 789–796.
- [16] R.H. Reed, The inactivation of microbes by sunlight: solar disinfection as a water treatment process, *Advances in Applied Microbiology* 54 (2004) 333–365.
- [17] D.C. Montgomery, *Introduction to Statistical Quality Control*, Wiley, New York, 2001.
- [18] R. Mosteo, P. Ormad, E. Mozas, J. Sarasa, J.L. Ovelleiro, Factorial experimental design of winery wastewaters treatment by heterogeneous photo-Fenton process, *Water Research* 40 (2006) 1561–1568.
- [19] J. Rodríguez-Chueca, R. Mosteo, M. Ormad, J. Ovelleiro, Factorial experimental design applied to *Escherichia coli* disinfection by Fenton and photo-Fenton processes, *Solar Energy* 86 (2012) 3260–3267.
- [20] S. Giannakis, A.I. Merino Gamó, E. Darakas, A. Escalas-Cañellas, C. Pulgarin, Impact of different light intermittence regimes on bacteria during simulated solar treatment of secondary effluent: implications of the inserted dark periods, *Solar Energy* 98 (Part C) (2013) 572–581.
- [21] J.J. Vélez-Colmenares, A. Acevedo, E. Nebot, Effect of recirculation and initial concentration of microorganisms on the disinfection kinetics of *Escherichia coli*, *Desalination* 280 (2011) 20–26.
- [22] A.-G. Rincon, C. Pulgarin, Bactericidal action of illuminated TiO₂ on pure *Escherichia coli* and natural bacterial consortia: post-irradiation events in the dark and assessment of the effective disinfection time, *Applied Catalysis B: Environmental* 49 (2004) 99–112.
- [23] F.H. Johnson, I. Lewin, The growth rate of *E. coli* in relation to temperature, quinine and coenzyme, *Journal of Cellular and Comparative Physiology* 28 (1946) 47–75.
- [24] U. Fotadar, P. Zaveloff, L. Terracio, Growth of *Escherichia coli* at elevated temperatures, *Journal of Basic Microbiology* 45 (2005) 403–404.
- [25] R.A. Blaustein, Y. Pachepsky, R.L. Hill, D.R. Shelton, G. Whelan, *Escherichia coli* survival in waters: temperature dependence, *Water Research* 47 (2013) 569–578.
- [26] A.G. Marr, J.L. Ingraham, Effect of temperature on the composition of fatty acids in *Escherichia coli*, *Journal of Bacteriology* 84 (1962) 1260–1267.
- [27] V.G. Petin, G.P. Zhurakovskaya, L.N. Komarova, Fluence rate as a determinant of synergistic interaction under simultaneous action of UV light and mild heat in *Saccharomyces cerevisiae*, *Journal of Photochemistry and Photobiology B: Biology* 38 (1997) 123–128.
- [28] I. Dawes, I. Sutherland, *Microbial Physiology*, Basic Microbiology, Blackwell Scientific Publications, Oxford, 1976, pp. 185.
- [29] N.D. Levine, R.E. Buchanan, N.E. Gibbons (Eds.), *Bergey's Manual of Determinative Bacteriology*, 8th ed., Williams & Wilkins Co, Baltimore, Md, 1974.
- [30] P. Chakraborty, *A Textbook of Microbiology*, New Central Book Agency, 2005.
- [31] W. Harm, *Biological Effects of Ultraviolet Radiation*, Cambridge University Press, Cambridge, 1980.
- [32] L.W. Sinton, R.K. Finlay, P.A. Lynch, Sunlight inactivation of fecal bacteriophages and bacteria in sewage-polluted seawater, *Applied and Environmental Microbiology* 65 (1999) 3605–3613.
- [33] M. Berney, H.U. Weilenmann, A. Simonetti, T. Egli, Efficacy of solar disinfection of *Escherichia coli*, *Shigella flexneri*, *Salmonella Typhimurium* and *Vibrio cholerae*, *Journal of Applied Microbiology* 101 (2006) 828–836.
- [34] S.A. Craik, D. Weldon, G.R. Finch, J.R. Bolton, M. Belosevic, Inactivation of *Cryptosporidium parvum* oocysts using medium-and low-pressure ultraviolet radiation, *Water Research* 35 (2001) 1387–1398.
- [35] S. Matallana-Surget, C. Villette, L. Intertaglia, F. Joux, M. Bourrain, P. Lebaron, Response to UVB radiation and oxidative stress of marine bacteria isolated from South Pacific Ocean and Mediterranean Sea, *Journal of Photochemistry and Photobiology B: Biology* (2012).
- [36] C. Hallmich, R. Gehr, Effect of pre- and post-UV disinfection conditions on photo-reactivation of fecal coliforms in wastewater effluents, *Water Research* 44 (2010) 2885–2893.
- [37] J. Regensburger, A. Knak, A. Felgenträger, W. Bäuml, Generation of singlet oxygen by UVB-irradiation of endogenous molecules, *Photodiagnosis and Photodynamic Therapy* 8 (2011) 152.
- [38] G. Storz, J.A. Imlay, Oxidative stress, *Current Opinion in Microbiology* 2 (1999) 188–194.
- [39] R.B. Setlow, The wavelengths in sunlight effective in producing skin cancer: a theoretical analysis, *Proceedings of the National Academy of Sciences* 71 (1974) 3363–3366.
- [40] O.J. Oppezzo, Contribution of UVB radiation to bacterial inactivation by natural sunlight, *Journal of Photochemistry and Photobiology B: Biology* (2012).
- [41] E.G. Mbonimpa, B. Vadheim, E.R. Blatchley III, Continuous-flow solar UVB disinfection reactor for drinking water, *Water research* 46 (2012) 2344–2354.
- [42] J. Robertson, P.K.J. Robertson, L.A. Lawton, A comparison of the effectiveness of TiO₂ photocatalysis and UVA photolysis for the destruction of three pathogenic micro-organisms, *Journal of Photochemistry and Photobiology A: Chemistry* 175 (2005) 51–56.
- [43] S. Pigeot-Rémy, F. Simonet, D. Atlan, J. Lazzaroni, C. Guillard, Bactericidal efficiency and mode of action: a comparative study of photochemistry and photocatalysis, *Water Research* 46 (2012) 3208–3218.
- [44] D. Spuhler, J.A. Rengifo-Herrera, C. Pulgarin, The effect of Fe²⁺, Fe³⁺, H₂O₂ and the photo-Fenton reagent at near neutral pH on the solar disinfection (SODIS) at low temperatures of water containing *Escherichia coli* K12, *Applied Catalysis B, Environmental* 96 (2010) 126–141.
- [45] L. Caballero, K. Whitehead, N. Allen, J. Verran, Inactivation of *Escherichia coli* on immobilized TiO₂ using fluorescent light, *Journal of Photochemistry and Photobiology A: Chemistry* 202 (2009) 92–98.
- [46] R. Matthews, Environment: photochemical and photocatalytic processes. Degradation of organic compounds, in: *Photochemical Conversion and Storage of Solar Energy*, Springer, 1991, pp. 427–449.
- [47] L. Dünkel, Organic Photochemistry: “A Visual Approach”, *Zeitschrift für Physikalische Chemie* 177 (1992) 122–123.
- [48] C. Sichel, J. Blanco, S. Malato, P. Fernández-Ibáñez, Effects of experimental conditions on *E. coli* survival during solar photocatalytic water disinfection, *Journal of Photochemistry and Photobiology A: Chemistry* 189 (2007) 239–246.
- [49] D. Alrousan, P.S. Dunlop, T.A. McMurray, J.A. Byrne, Photocatalytic inactivation of *E. coli* in surface water using immobilised nanoparticle TiO₂ films, *Water Research* 43 (2009) 47–54.
- [50] P. Dunlop, M. Ciavola, L. Rizzo, J. Byrne, Inactivation and injury assessment of *Escherichia coli* during solar and photocatalytic disinfection in LDPE bags, *Chemosphere* 85 (2011) 1160–1166.
- [51] A. Rincon, C. Pulgarin, Photocatalytic inactivation of *E. coli*: effect of (continuous–intermittent) light intensity and of (suspended–fixed) TiO₂ concentration, *Applied Catalysis B: Environmental* 44 (2003) 263–284.
- [52] B. Sommer, A. Marino, Y. Solarte, M. Salas, C. Dierolf, C. Valiente, D. Mora, R. Rechsteiner, P. Setter, W. Wirojanagud, SODIS – an emerging water treatment process, *AQUA (OXFORD)* 46 (1997) 127–137.
- [53] R.M. Tyrrell, M.J. Peak, Interactions between UV radiation of different energies in the inactivation of bacteria, *Journal of Bacteriology* 136 (1978) 437–440.
- [54] M. Peak, J. Peak, R. Webb, Synergism between different near-ultraviolet wavelengths in the inactivation of transforming DNA, *Photochemistry and photobiology* 21 (1975) 129–131.
- [55] F. Bosshard, M. Bucheli, Y. Meur, T. Egli, The respiratory chain is the cell's Achilles' heel during UVA inactivation in *Escherichia coli*, *Microbiology* 156 (2010) 2006–2015.
- [56] S.M. Chiang, H.E. Schellhorn, Regulators of oxidative stress response genes in *Escherichia coli* and their functional conservation in bacteria, *Archives of biochemistry and biophysics* 525 (2012) 161–169.
- [57] S. Mishra, J. Imlay, Why do bacteria use so many enzymes to scavenge hydrogen peroxide?, *Archives of biochemistry and biophysics* (2012) 145–160.