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1 An attempt to partition stomatal and non-stomatal ozone 2 deposition parts on a short grassland

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6
7 **Abstract** To evaluate the damaging effect of tropospheric ozone on vegetation, it is important
8 to evaluate the stomatal uptake of ozone. Although stomatal flux is a dominant pathway of
9 ozone deposition onto vegetated surfaces, non-stomatal uptake mechanisms as soil
10 deposition, especially when $LAI < 4$, and cuticular deposition are also vital parts. In this
11 study, we partitioned canopy conductance into stomatal and non-stomatal parts. To calculate
12 the stomatal conductance of water vapour for sparse vegetation, firstly, we partitioned the
13 latent heat flux into transpiration and evaporation parts using the Shuttleworth-Wallace (SW)
14 model. Then we derived the stomatal conductance of ozone by the Penman-Monteith (PM)
15 theory based on the similarity to water vapour conductance. The non-stomatal conductance
16 was calculated by subtracting the stomatal conductance from canopy conductance derived
17 from direct flux measurement data. Our results show that for short vegetation ($LAI = 0.25$)
18 dry deposition of ozone was dominated by non-stomatal flux, exceeding stomatal flux even in
19 daytime, while at night stomatal uptake of ozone was negligibly small. In the case of
20 vegetation with $LAI \approx 1$, the daytime stomatal and non-stomatal fluxes were of the same order
21 of magnitude. These results underline that non-stomatal processes have to be considered even
22 in the case of well-developed vegetation where cuticular uptake is comparable in magnitude
23 with stomatal uptake, and especially in the case of vegetated surfaces with $LAI < 4$ where soil
24 uptake takes part in ozone deposition as well.

25
26 **Keywords** Deposition · Eddy covariance · Non-stomatal conductance · Ozone flux · Stomatal
27 conductance

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28 **List of symbols**

29

Symbol	Name	Unit
<i>AGB</i>	above ground green biomass	g m^{-2}
$b_1; b_3$	empirical constants for estimation of r_{ss}	s m^{-1}
b_2	empirical constant for estimation of r_{ss}	--
c	concentration of ozone at the reference height (= 4 m)	nmol m^{-3}
$C_c; C_s$	canopy and surface resistance coefficients in the SW model	--
c_p	specific heat capacity of moist air at constant pressure (= 1,013)	$\text{J kg}^{-1} \text{K}^{-1}$
c_{pv}	specific heat capacity of water vapour at constant pressure (= 1,850)	$\text{J kg}^{-1} \text{K}^{-1}$
c_w	specific heat capacity of water (= 4,220)	$\text{J kg}^{-1} \text{K}^{-1}$
d	displacement height	m
D	vapour pressure deficit in the air	Pa
D_{O_3}/D_w	molecular diffusivity ratio of ozone to water (= 0.608)	--
E	ecosystem evapotranspiration (water vapour flux)	$\text{kg m}^{-2} \text{s}^{-1}$
E_e	evaporation of soil water and other wet surfaces	$\text{kg m}^{-2} \text{s}^{-1}$
E_t	stomatal transpiration	$\text{kg m}^{-2} \text{s}^{-1}$
e	water vapour pressure	Pa
e_s	saturated water vapour pressure	Pa
F	ozone flux	$\text{nmol m}^{-2} \text{s}^{-1}$
G	heat flux into the soil	W m^{-2}
h	vegetation height	m
H	sensible heat flux	W m^{-2}
k	Kármán constant (= 0.4)	--
L	Obukhov length	--
LAI	leaf area index (green fraction)	$\text{m}^2 \text{m}^{-2}$
LW	leaf surface wetness	%
n	eddy diffusivity decay constant (= 2.5)	--
p	air pressure	Pa
PM_c	canopy transpiration in the SW model	W m^{-2}
PM_s	soil evaporation in the SW model	W m^{-2}
r	total resistance to ozone dry deposition	s m^{-1}
r_a	aerodynamic resistance	s m^{-1}
r_b	boundary layer resistance	s m^{-1}
r_c	canopy resistance	s m^{-1}
r_{st}	bulk stomatal resistance including mesophyll resistance r_{mes}	s m^{-1}
r_{nst}	non-stomatal resistance	s m^{-1}
r_{aa}	resistance of canopy height to reference height in the SW model	s m^{-1}
r_{ac}	bulk boundary layer resistance of the vegetative elements in the canopy in the SW model	s m^{-1}
r_{as}	resistance of soil surface to canopy height in the SW model	s m^{-1}
r_{bv}	mean boundary layer resistance per unit area of vegetation in the SW model (= 25)	s m^{-1}
r_{mst}	mean stomatal resistance in the SW model (= 400)	s m^{-1}
r_{sc}	canopy stomatal resistance in the SW model	s m^{-1}
r_{ss}	soil surface resistance in the SW model	s m^{-1}
RH	relative humidity	%
R	available energy input above the canopy	W m^{-2}

R_g	global radiation	$W m^{-2}$
R_n	net radiation	$W m^{-2}$
R_{ns}	net radiation fluxes to soil	$W m^{-2}$
R_s	available energy input above the soil surface	$W m^{-2}$
R_w	specific gas constant for water vapour (= 461.5)	$J kg^{-1} K^{-1}$
Sc/Pr	ratio of the Schmidt to the Prandtl number (= 1.486)	--
$t_a; t_s$	air; soil temperature	$^{\circ}C$
t	time	s
T	air temperature	K
u	wind velocity at the reference height (x)	$m s^{-1}$
u^*	friction velocity	$m s^{-1}$
U	electric voltage	mV
v_d	dry deposition velocity of ozone	$m s^{-1}$
z	reference height of measurements above canopy (= 4)	m
z_0	roughness length	m
β	Bowen-ratio (= $H/\lambda E$)	--
γ	psychrometric constant [= $c_p p/(\lambda \epsilon)$]	$Pa K^{-1}$
δ	water vapour density saturation deficit	$kg m^{-3}$
Δ	slope of the saturation vapour pressure [= $e_s \lambda/(R_w T^2)$]	$Pa K^{-1}$
ϵ	ratio of mean molar mass of water to dry air (= 0.6215)	--
θ	soil water content	volume %
θ_s	saturated soil water content (= 28 at measuring site)	volume %
κ_c	canopy conductance	$m s^{-1}$
κ_{nst}	non-stomatal conductance	$m s^{-1}$
κ_{st}	stomatal conductance including mesophyll conductance	$m s^{-1}$
λ	latent heat of vaporization [$\lambda_0 = 2,500,800$ at $0^{\circ}C$, $\lambda = \lambda_0 + (c_{pv} - c_w)t_a$]	$J kg^{-1}$
λE	latent heat flux	$W m^{-2}$
λE_e	latent heat flux from the soil surface	$W m^{-2}$
λE_t	latent heat flux from the canopy	$W m^{-2}$
ρ	density of moist air (calculated from RH, T, p)	$kg m^{-3}$
ρ_v	density of water vapour	$kg m^{-3}$
ρ_{vs}	density of saturated water vapour	$kg m^{-3}$
$\rho_a; \rho_s; \rho_c$	parameters in calculation of C_c and C_s	$Pa s K^{-1} m^{-1}$
τ	momentum flux	$kg m^{-1} s^{-2}$
Φ	relative ozone flux	$mV m s^{-1}$
ζ	dimensionless height (= z/L)	--

30

31

32 1 Introduction

33 The harmful effects of ozone on plants are well known (Amann et al. 2011, Colette et al.
34 2012). Ozone molecules enter via the stomata; therefore, the risk of ozone damage can be
35 quantified by stomatal uptake, rather than by simple exposure-based indices like SUM06,
36 W126 and AOT40 (Emberson et al. 2000, Massman 2004, Musselman et al. 2006, Mills et al.
37 2011).

38 Ozone flux measurements generally allow the aerodynamic, boundary layer and canopy
39 resistances (r_a , r_b , and r_c , respectively) to be separated on the basis of resistance analogy
40 models. Canopy resistance includes both stomatal and mesophyll components (in this paper
41 the sum of these two parts is referred to as stomatal) and so-called non-stomatal resistance,
42 consisting of the deposition to leaf cuticle, the ground, litter and other parts of the plant, as
43 well as near-surface chemistry.

44 Several examples of methods can be found in the literature to calculate the stomatal
45 conductance of ozone. For example, Rummel et al. (2007) applied a modified Jarvis-type
46 model (Jarvis 1976) derived for water vapour flux, using maximum stomatal conductivity
47 ($\kappa_{st,max}$), *LAI* and functions for specific humidity deficit, t_a , and short wave radiation.
48 According to the compilation of Kelliher et al. (1995), the $\kappa_{st,max}$ is site- and vegetation-
49 specific and ranges between 6-12 mm s⁻¹ at optimum meteorological conditions, which
50 makes it difficult to generalise the method. Another example was published by Granz et al.
51 (1995). They also used the similarity between the stomatal conductance of ozone and water
52 vapour, deriving a simple empirical equation for κ_{st} expressed as a function of
53 photosynthetically active radiation. Massman (2004) described a simple empirical method for
54 a vineyard site, using solar radiation and *LAI* as inputs. The disadvantage of this calculation is
55 that the model is site-specific.

56 The canopy model by Wang and Leuning (1998) used a simple model to partition the
57 available energy and calculate the stomatal conductance for CO₂. The parameterisation of
58 stomatal conductance involves, among others, the net photosynthetic and carboxylation rates,
59 which are not widely available parameters. In this approach, a single-layer canopy model
60 calculates the fluxes of sensible heat, latent heat, and CO₂, separately for sunlit and shaded
61 leaves. Compared to a multi-layer model (assuming ozone deposition takes place separately
62 on different parts of the canopy), the CO₂, latent and sensible heat fluxes predicted usually
63 agreed with a less than 5% difference over a typical range of leaf area index values for a
64 wheat crop grown in a temperate climate.

65 Lamaud et al. (2002) estimated the stomatal conductance for ozone using the mechanism
66 mentioned above, based on the similarity to the water vapour flux, for a pine forest canopy in
67 dry and wet conditions. Ozone fluxes were measured using the eddy covariance (EC)
68 technique above and within the canopy. They demonstrated that the ozone uptake by the
69 understory is a significant proportion of the entire ozone deposition onto the whole pine
70 stand. According to their results, the understory contributes more to the overall ozone flux

71 than to the other measured scalar fluxes (sensible heat and water vapour). Also, during the
72 day, in dry conditions, the canopy stomatal conductance is the major parameter controlling
73 ozone deposition. Furthermore, in winter, the influence of dynamic processes persists during
74 daytime. It was also found that surface wetness associated with dew significantly enhanced
75 ozone deposition during the night as well as in the morning.

76 Lamaud et al. (2009) partitioned ozone deposition over a developed maize crop into
77 stomatal and non-stomatal uptakes using eddy covariance flux measurements and modelling.
78 Data were analysed using a big-leaf model, which was developed based on the current
79 knowledge of ozone deposition. In-canopy aerodynamic resistance, intrinsic ground
80 resistance and cuticular resistance were determined from the relationship between
81 experimental non-stomatal conductance and friction velocity in dry conditions. Non-stomatal
82 conductance was determined as the difference between canopy conductance and stomatal
83 conductance, where the latter was estimated by a method that combines the PM (Penman-
84 Monteith) approach with the use of the similarity to carbon dioxide flux. They showed that
85 the relative contributions of stomatal and non-stomatal uptakes varied strongly with the
86 physiological activity of the maize and the meteorological conditions.

87 Gerosa et al. (2007) compared different algorithms for stomatal ozone flux determination
88 from micrometeorological measurements using the similarity between ozone stomatal fluxes
89 and water vapour stomatal fluxes. A series of observations, made during the growing season
90 over an onion field, were used to show the equivalence of two algorithms from the literature
91 to derive the stomatal fluxes of ozone. One of these algorithms uses the PM approach, where
92 the water vapour pressure deficit is calculated using air temperature. The second calculates,
93 using another formulation, the water vapour deficit based on leaf temperature. As they
94 argued, the two approaches led to the same results if applied properly, both theoretically and
95 numerically.

96 Gerosa et al. (2012) modelled stomatal conductance to estimate the evapotranspiration of
97 natural and agricultural ecosystems on an hourly basis. In these cases, the big-leaf approach,
98 together with the resistance analogy that simulates the gas-exchange between vegetation and
99 atmosphere, is a simple but valid example of a process-based model which includes stomatal
100 conductance behaviour, as well as a basic representation of the canopy features.

101 Coyle et al. (2009) calculated the non-stomatal resistance of ozone as the residual of the
102 difference of canopy resistance and stomatal resistance over a potato field. The stomatal part

103 was estimated using the similarity between the fluxes of water vapour and ozone. In this
104 study, it was assumed that transpiration is the only source of water vapour from the surface.

105 In most of these approximations, it is assumed that water vapour flux consists only of the
106 water loss from stomata through transpiration (E_t), which is true for well-developed
107 vegetation, especially for forests, where the leaf area index is $LAI > 4$ and the surface is dry.
108 In the case of low vegetation (e.g. grass surfaces), however, water vapour flux can also be
109 derived from evaporation (E_e) from other wet surfaces, especially from the ground. Over bare
110 soil there is no transpiration; and in parallel with increasing LAI , the share of transpiration in
111 the total evapotranspiration increases as well. At $LAI = 4$ (a practically closed canopy) the
112 share of transpiration is still 91-94% (Shuttleworth and Wallace 1985), hence evaporation is
113 nearly negligible. However, below $LAI = 4$ water vapour flux cannot be used to estimate the
114 stomatal conductance of ozone, therefore transpiration and evaporation rates have to be
115 separated.

116 In the ÉCLAIRE EU 7th Framework Program project (Sutton et al. 2013) we monitored
117 the ozone flux by the eddy covariance method above short vegetation (grassland) between
118 August 2012 and January 2014. As a result of the mean leaf area index ($LAI_{\text{mean}} = 0.5$) in the
119 observation period, when calculating the different deposition parts, in addition to the
120 transpiration, we also had to take into account the potential effect of evaporation.

121 The aim of the current study is to derive stomatal conductance (κ_{st}) based on the
122 partitioning of water vapour flux. This also lets us calculate the stomatal flux of ozone, which
123 is an important factor in the estimation of the damage caused by the direct uptake of ozone. In
124 addition, once κ_{st} is obtained, non-stomatal conductance (κ_{nst}) can also be derived as the
125 residual term: $\kappa_{\text{nst}} = \kappa_c - \kappa_{\text{st}}$. The κ_{nst} values estimated in this way can serve as a basis for
126 future work, for finding empirical equations that express κ_{nst} . Hence the bulk canopy
127 conductance and dry deposition velocity can be calculated as the function of meteorological
128 variables (including calculated $r_a + r_b$). In this way, we were able to obtain the total ozone
129 fluxes using only data from a slow ozone monitor instead of eddy covariance flux
130 measurement. Such an approach would be useful during gap-filling when eddy covariance
131 ozone flux measurements are not available or when assumptions for eddy covariance (EC)
132 are not satisfied.

133 Therefore, firstly, we calculated the dry eddy flux of ozone and the canopy resistance.
134 Secondly, we partitioned the latent heat fluxes into fluxes from the canopy and from the
135 surface by the SW (Shuttleworth-Wallace) model (Shuttleworth and Wallace 1985, Hu et al.

136 2009) resulting in evaporation and transpiration, respectively. Thirdly, we used the
137 transpiration part to calculate stomatal conductance using the inverted PM equation as
138 suggested by Lamaud et al. (2002). Finally, we partitioned stomatal and bulk non-stomatal
139 conductances and we investigated them under different meteorological conditions.

140

141 **2 Methodology**

142 2.1 Site of Investigations

143 One of the selected grassland stations of the ÉCLAIRE project is Bugacpuszta on the
144 Hungarian Great Plain (46.69° N, 19.60° E, 113 m a.s.l.). A detailed description of the site
145 was given by Machon et al. (2015). The climate of this semi-natural, semi-arid, sandy
146 grassland is temperate continental, the mean annual temperature is 10.7 °C and the average
147 yearly precipitation is 550 mm. The region has Chernozem-type sandy soil with a high sand
148 (79%) and low clay (13%) content in the upper 10-cm soil layer. The area within 200 m of
149 the measurement plot has never been ploughed. Apart from grazing by a herd of the ancient
150 Grey Cattle breed at an average grazing pressure of 0.5-0.8 stock ha⁻¹ in the grazing season
151 (220 days each year) – which has been going on for centuries in dynamic equilibrium with
152 the grass ecosystem (Machon et al. 2010) – the soil has been undisturbed. The plant
153 association is semi-arid sandy grassland (*Cynodonti Festucetum pseudovinae*) dominated by
154 *Festuca pseudovina*, *Carex stenophylla*, and *Cynodon dactylon* (Koncz et al. 2014).

155

156 2.2 Measurements

157 Measurements were conducted between August 2012 and January 2014. The fast response
158 ozone monitor was not operating between the middle of May and the beginning of August
159 2013 due to a fault. In this study we used the whole (\approx 15 month) dataset for a general picture
160 as well as short (5-12 days) periods to examine the applicability of the coupled SW and PM
161 models to estimate the stomatal conductance of ozone. The list of measured parameters, the
162 methods, and the sampling/logging time are compiled in Table 1.

163 The ultrasonic anemometer and the inlet of the fast response ozone monitor were arranged
164 at a height of 4 m. The air inlet and the sensor were connected by a 3-m PTFE tube. The air
165 flow during sampling and calibration was 2 L min⁻¹. Sensor disks were provided by the
166 manufacturer as described by Schurath et al. (1991).

167

168 **INSERT HERE TABLE 1**

169

170 The HORIBA APOA 350 ozone monitor was calibrated before and after installation (in
171 July 2002 and in January 2004) in the reference laboratory of the Hungarian Meteorological
172 Service by a UV photometric system. During the campaign, we checked the sensitivity and
173 drift of the instrument by gas phase titration on five occasions using a Type 146 multigas
174 calibration system manufactured by Thermo Environmental Instruments Inc. USA. The error
175 caused by zero line drift and change of sensitivity in the measurement period was within 2%.
176 The relative output voltage of the fast sensor was frequently calibrated by a slow response
177 ozone monitor to eliminate the change in sensitivity caused by changing air humidity.

178 *Above ground green biomass (AGB)* was sampled by cutting the plants above the litter
179 layer > 1 cm in five sampling quadrants along a 5-metre-long transect. The total biomass was
180 separated into dead, dry (yellow, brown) and living (green) parts to understand the dynamics
181 of living (fresh) and senescent (dry) biomass. The biomass was oven-dried at 85 °C for 48 h.

182 *Vegetation height (h)* was measured at the four corners of the quadrants. Permanent
183 quadrants (40 × 40 cm) located along 5-m long permanent transects were sampled in one- to
184 two-week intervals during summer, autumn, and spring as well as monthly during the winter.

185 *Leaf area index (LAI)* was estimated from light interception measurements described by
186 Campbell (1986) and Campbell and Norman (1989). Throughout the study we applied the same
187 sampling protocol, measuring device and calculation methodology to estimate *LAI* (for details
188 of *LAI* measurements at the site see Koncz et al. 2015). Therefore, we eliminated the
189 uncertainties which could have been created when using different protocols, devices or
190 analyses (He et al. 2007, Confalonieri et al. 2013). Uncertainties in *LAI* estimation also arise
191 due to the varying leaf area distribution over time in relation to the sun. However, we used the
192 methodology as described by Campbell (1986) and Campbell and Norman (1989). Measured
193 *LAI* was corrected by the ratio of dead/green biomass (*AGB*) to obtain the green fraction.

194 The measurement methods of all other parameters are listed in Table 1.

195

196 2.3 Calculation of Ozone Flux and Dry Deposition Velocity

197 The 30-min mean ozone fluxes were determined based on the eddy covariance technique
198 using a dry chemiluminescence fast response analyser with a typical precision of 0.3-1.0%
199 between 10 and 100 ppbv at a frequency of $f = 10$ Hz (Zahn et al. 2012). The absolute ozone
200 concentration was measured by an ozone monitor (types and manufacturers can be seen in
201 Table 1).

202 We used two methods to calculate turbulent fluxes. The momentum and heat fluxes were
203 calculated according to Nagy et al. (2007), applying the “traditional” planar fit method. These
204 long-term measurements started in 2002.

205 For the calculation of the ozone flux during the ÉCLAIRE campaign we used the 2D
206 coordinate rotation method for the sonic anemometer measurements. Above flat surfaces both
207 methods can be used with the same precision. The high frequency (10 Hz) data series
208 (3D wind, sonic temperature and ozone voltage signal) were despiked (4σ), linear detrended,
209 and wind vectors were rotated to the main wind direction (2D rotation, McMillen 1988). The
210 raw relative ozone time series data (U) were shifted considering the lag time at the inlet,
211 based on the maximum correlation of vertical velocity and relative ozone signal. The default
212 time lag (t_{default}) for the maximum covariance was $t_{\text{default}} = 2$ s based on the statistical analysis
213 of the long term flux dataset and a laboratory experiment performed before the measuring
214 campaign (knowing the tube length, diameter and the mean flow rate). The uncertainty of the
215 time lag was a few tenths of seconds. In each time period, we recalculated the time lag by
216 maximizing the eddy covariance. When the calculated maximum time lag, t_{max} was within
217 $t_{\text{default}} \pm 0.5$ s, Φ_{max} was regarded as a valid relative flux (proportional to the flux expressed in
218 the relative unit: mV m s^{-1}), in other cases Φ_{default} with time lag (t_{default}) was chosen as the
219 valid flux (Φ_{max}) (see also Ocheltree and Loescher 2006, Aubinet et al. 2012).

220 The absolute raw ozone fluxes (F_{raw}) were calculated by the ratio method (Muller et al.
221 2010) using absolute ozone concentrations (nmol m^{-3}), which does not require the
222 determination of a calibration factor obtained from the relative ozone concentration
223 fluctuation measurements (voltage signals). In this calculation, average ozone concentration
224 and the offset of the fast response ozone sensor (U_{off}) during the flux averaging period are
225 needed to obtain absolute fluxes:

$$226 \quad F_{\text{raw}} = \frac{\Phi_{\text{max}} c_{\text{avg}}}{U_{\text{avg}} - U_{\text{off}}}, \quad (1)$$

227 where c_{avg} and U_{avg} are the half-hourly average ozone concentrations from the slow response
228 ozone monitor and the average voltage from the fast response instrument, respectively. The
229 offset (U_{off}) was checked regularly with an active ozone disc by stopping the air flow and it
230 was found to be approximately constant (10 ± 2 mV).

231 The effect of the density fluctuations generated by the closed-path analyser itself was
232 taken into account by the traditional Webb-Pearman-Leuning (WPL) correction (Webb et al.
233 1980, Leuning 2007), using the moisture fluctuation term and neglecting the temperature

234 fluctuation term, which is important only for the open path sensors (Rannik et al. 1997, Lee
235 and Massman 2011).

236 Spectral correction was performed according to two different methodologies.

237 a) Based on the eddy covariance software package TK3 (Mauder and Foken 2011)
238 corrections were applied for i) inadequate frequency response, ii) sensor line averaging, iii)
239 air sampling through tubes, and iv) flux loss at low frequency due to the limited averaging
240 period.

241 b) The other empirical method (Ammann et al. 2006) estimates high frequency loss by
242 determining the maximum difference of the relative ogive function of kinematic heat flux
243 covariance ($\overline{w'T'}$) and the ozone flux ($\overline{w'c'}$) as the first step. Secondly, it calculates the
244 spectral correction of the kinematic heat flux according to the TK3 method.

245 Spectral correction was carried out by using the mean of the two (a and b) methods. In the
246 case of the noisy ogives, when the maximum difference between the ogive functions was
247 higher than 30%, only the TK3 spectral correction (correction a) was used. The final value of
248 ozone flux was denoted as F . The flux calculation program was written in FORTRAN.

249 Spectral correction depends on stability. Higher relative values were observed during
250 stable stratification. The mean values and standard deviations of the spectral corrections
251 using the methodology of TK3 software and the semi-empirical corrections based on
252 Ammann et al. (2006) are presented in Table 2 for a test period of May 2013. A total of 589
253 half-hourly measurements were analysed from unstable to stable stratifications in the interval
254 of $-1 < \zeta < 1$.

255

256 **INSERT HERE TABLE 2**

257

258 Co-spectral correction (maximum differences between two relative ogive functions for
259 $\overline{w'T'}$ and $\overline{w'c'}$) slightly depends on stability. The mean values and standard deviations are in
260 the same order of magnitude in each stability category (5-7%). Dependence of both types of
261 spectral corrections on stability is similar. The TK3 methodology (TK3 corr. $\overline{w'c'}$) gives
262 higher mean values for each stability category compared to the semi-empirical methodology.
263 The values of spectral corrections are not negligible.

264 We used a standard flux calculation methodology comparable with other ÉCLAIRE flux
265 sites. The numerical optimisation of the ozone flux ogive function (Sievers et al. 2015) and a

266 more detailed uncertainty analysis of the ozone flux calculation (Zhu et al. 2015) are focuses
 267 of near future investigations.

268 In the present work, the random flux error was estimated as the root mean squared
 269 deviation of the covariance function from the zero line within the two tail ranges, which can
 270 be calculated as (Nemitz 2014):

$$271 \quad \delta F = \pm \sqrt{\frac{1}{N} \sum_{i=1}^N \delta_i \rho_{wi}^2}, \quad (2)$$

272 where ρ_{wi} is the value of the cross-covariance function. The delta function is $\delta_i=1$ for those
 273 indices (i) which are far from the optimum time lags as: a) $(t_{\text{default}} - 90 \text{ s}) < t_i < (t_{\text{default}} - 30 \text{ s})$,
 274 b) $(t_{\text{default}} + 30 \text{ s}) < t_i < (t_{\text{default}} + 90 \text{ s})$, otherwise $\delta_i = 0$. N is the number of samples for which
 275 $\delta_i = 1$. In our case $N = 1200$.

276 The above formula can also be written as:

$$277 \quad \delta F = \pm \sqrt{0.5(\text{std}_{\text{left}}^2 + \text{avg}_{\text{left}}^2 + \text{std}_{\text{right}}^2 + \text{avg}_{\text{right}}^2)}, \quad (3)$$

278 where std_{left} , avg_{left} , $\text{std}_{\text{right}}$ and $\text{avg}_{\text{right}}$ are the mean and standard deviations of the cross-
 279 covariance function of ozone and vertical wind speed using different time delays (t_i) on the
 280 left and right hand side of the auto-covariance function, respectively.

281 Dry deposition velocity and random error of deposition were also calculated based on the
 282 flux dataset (F) as follows:

$$283 \quad v_d = \frac{F}{c_{\text{avg}}}, \quad (4)$$

$$284 \quad \delta v_d = \frac{\delta F}{c_{\text{avg}}}. \quad (5)$$

285 The uncertainty in the measurements of the average ozone mixing ratio was not taken into
 286 account for the calculation. The signals from the fast and slow ozone sensors were recorded
 287 separately. We assumed that uncertainties mostly originated from flux measurement errors
 288 (Nemitz 2014, Zhu et al. 2015).

289 Averaged ozone fluxes were calculated for each half-hour period when real signals were
 290 received (no error message) both from the ultrasonic anemometer and from the fast response
 291 ozone monitor. On the basis of the calculated ozone flux (F) and random flux error (δF),
 292 semi-empirical data filtering was applied removing the average half-hour fluxes when: i) $|\delta F|$
 293 $\gg |F|$, ii) $F < -10 \text{ nmol m}^{-2} \text{ s}^{-1}$, iii) any unrealistic jumping in the values F , δF , and v_d , iv) $|F|$
 294 $> 0.5 \text{ nmol m}^{-2} \text{ s}^{-1}$ and $|\delta F| \geq |F|$. The number of error cases was lower than 5% and occurred
 295 mostly in night-time and transient periods.

296 Spike detection and removal of the raw (10 Hz) data was carried out as suggested by
297 Vickers and Mahrt (1997) and linear detrending was performed afterwards. Possible
298 inaccurate levelling of the sonic anemometer was corrected by the “traditional” planar fit
299 method (Wilczak et al. 2001).

300 From the corrected raw data, the momentum flux was calculated by the following
301 equation:

$$302 \quad \tau = \rho u_*^2 = \rho \sqrt{(\overline{u'w'})^2 + (\overline{v'w'})^2}, \quad (6)$$

303 where $\overline{u'w'}$ and $\overline{v'w'}$ denotes the covariances of the two horizontal (u , v) components and the
304 vertical (w) component of wind speed.

305

306 2.4 Estimation of the Effect of Storage Changes and the Flux Divergence Caused by 307 Chemistry on Calculated Fluxes

308 Storage changes are an important source of bias in flux estimation, but in the case of low
309 vegetation, uptake is close to the ground surface, hence the storage changes are generally
310 considered to be negligible (Wohlfahrt et al. 2012).

311 In-canopy chemistry is another sink for ozone (Fuentes et al. 2007). Chemical reactions of
312 ozone, involving biogenic VOCs (Volatile Organic Compounds) should definitely be taken
313 into account in flux calculations, however, they have a dominant role in the case of forested
314 areas (Goldstein et al. 2004). Over grasslands, although the emitted VOCs react with ozone
315 rapidly enough to influence the flux, these emissions are minimal and not measured.
316 Therefore, the strongest potential source of divergence can be the reaction with NO emitted
317 from the soil. However, the influence of NO on the ozone flux profiles is usually weak
318 because the ozone fluxes are typically considerably larger than nitrogen oxide fluxes (Kramm
319 et al. 1995). For a short canopy – even for bare soil – it is generally estimated as negligible,
320 below 1% (Stella et al. 2012). This assumption is supported by the mean soil NO flux (0.025
321 $\text{nmol m}^{-2} \text{s}^{-1}$) measured at our site being two orders of magnitude lower than the ozone flux
322 (measured between 2006 to 2010, Machon et al. 2015). Therefore, the majority of non-
323 stomatal conductance is attributed to the dry deposition and decomposition processes on
324 plant, litter, and soil surfaces.

325

326 2.5 Partitioning of Resistance Terms

327 The reciprocal value of dry deposition velocity equals the sum of aerodynamic, boundary
 328 layer, and canopy resistances:

$$329 \quad \frac{1}{v_d} = r = r_a + r_b + r_c. \quad (7)$$

330 To calculate the canopy resistance (r_c) using Eq. (7) we computed the term of ($r_a + r_b$)
 331 according to Baldocchi and Meyers (1991) and Lamaud et al. (2002) as:

$$332 \quad r_a + r_b = \frac{u}{u_*^2} + \frac{2}{\kappa u_*} \left(\frac{Sc}{Pr} \right)^{\frac{2}{3}}, \quad (8)$$

333 where u_* was derived from momentum flux (τ) calculated using ultrasonic anemometer data
 334 according to Eq. (6) as described in Section 2.3.

335 Canopy resistance r_c can be further divided into stomatal (r_{st}) and non-stomatal (r_{nst})
 336 terms:

$$337 \quad \frac{1}{r_c} = \frac{1}{r_{st}} + \frac{1}{r_{nst}}, \text{ or } \kappa_c = \kappa_{st} + \kappa_{nst}. \quad (9)$$

338 Non-stomatal conductance (κ_{nst}) – as the residual of κ_c after subtracting κ_{st} – represents the
 339 bulk conductance of different processes, namely the effect of air chemistry (virtual loss of O₃
 340 by thermal reaction with NO), leaf surface chemistry, as well as deposition to ground level
 341 (dead parts of plants, litter) and soil (Byun and Dennis 1995, Fares et al. 2012). Partitioning
 342 of κ_c into κ_{st} and κ_{nst} cannot be calculated directly. Parameterisation and modelling of stomatal
 343 resistance generally use the similarity of ozone flux to other gases like CO₂ or water vapour.
 344 Gerosa et al. (2007) proposed an algorithm to calculate the stomatal flux of ozone by the PM
 345 and evaporation-resistance approaches using measured water vapour flux. Those formulae
 346 assume equivalence between the stomatal water vapour flux (E_i) and the total water vapour
 347 flux used for closed canopy with negligible soil evaporation. However, for our open canopy
 348 ($LAI_{mean} = 0.5$) water vapour flux consists not only of stomatal transpiration but also of
 349 evaporation. Shuttleworth and Wallace (1985) described a one-dimensional model (see also
 350 in Hu et al. 2009) to partition the evaporation (E_e) and transpiration (E_t) terms (all the
 351 equations shown below are based on SW model, unless stated otherwise):

$$352 \quad \lambda E = \lambda E_e + \lambda E_t = C_c PM_c + C_s PM_s, \quad (10)$$

353 where

$$354 \quad PM_c = \frac{\Delta R + (\rho c_p D - \Delta r_{ac} R_s) / (r_{aa} + r_{ac})}{\Delta + \gamma [1 + r_{sc} / (r_{aa} + r_{ac})]}, \quad (11)$$

355 and

$$356 \quad PM_s = \frac{\Delta R + [\rho c_p D - \Delta r_{as} (R - R_s)] / (r_{aa} + r_{as})}{\Delta + \gamma [(1 + r_{ss} / (r_{aa} + r_{as}))]}. \quad (12)$$

357 Radiation terms are expressed as (Hu et al. 2009):

358
$$R = R_n - G, \quad (13)$$

359
$$R_s = R_{ns} - G, \quad (14)$$

360 and

361
$$R_{ns} = R_n e^{-0.6 LAI}. \quad (15)$$

362 The soil heat flux was estimated according to Hillel (1998) combining the time lag and
 363 damping deep methods, using the measured soil wetness (θ) and temperature (t_s) at two upper
 364 depths (-0.03 ; -0.30 m). When soil physical measurements were not available (less than 5%
 365 of all the cases) G was estimated from the mean ratio of calculated soil heat flux by Hillel
 366 (1998) and the measured net radiation:

367
$$G = 0.1 \times R_n. \quad (16)$$

368 Canopy and surface resistance coefficients (17)-(21) were calculated (Shuttleworth and
 369 Wallace 1985) as:

370
$$C_c = \frac{1}{1 + \left[\frac{\rho_c \rho_a}{\rho_s (\rho_c + \rho_a)} \right]} \quad (17)$$

371 and

372
$$C_s = \frac{1}{1 + \left[\frac{\rho_s \rho_a}{\rho_c (\rho_s + \rho_a)} \right]}, \quad (18)$$

373 where

374
$$\rho_a = (\Delta + \gamma) r_{aa}, \quad (19)$$

375
$$\rho_c = (\Delta + \gamma) r_{ac} + \gamma r_{sc}, \quad (20)$$

376
$$\rho_s = (\Delta + \gamma) r_{as} + \gamma r_{ss}. \quad (21)$$

377 The resistances in the SW model were estimated as follows: r_{aa} and r_{as} were calculated
 378 according to Shuttleworth and Wallace (1985) from the parameters z , d , z_0 , h , k , u , and n
 379 (assuming that $d = 0.63 \times h$ and $z_0 = 0.13 \times h$, Shuttleworth and Wallace 1985). For a fully
 380 developed crop ($LAI > 4$):

381
$$r_{aa}(\alpha) = \frac{\ln\left(\frac{z-d}{z_0}\right)}{k^2 u} \left\{ \ln \frac{z-d}{h-d} + \frac{h}{n(h-d)} e^{n\left[1 - \frac{(d+z_0)}{h}\right]} - 1 \right\}, \quad (22)$$

382 and

383
$$r_{as}(\alpha) = \frac{\ln\left(\frac{z-d}{z_0}\right)}{k^2 u} \frac{h}{n(h-d)} \left[e^n - e^{n\left(1 - \frac{d+z_0}{h}\right)} \right]. \quad (23)$$

384 For bare soil:

385
$$r_{aa}(0) = \frac{\ln^2\left(\frac{z}{z_0}\right)}{k^2 u} - r_{as}(0), \quad (24)$$

386 and

387
$$r_{as}(0) = \frac{\ln \frac{z}{z_0} \ln \frac{(d+z_0)}{z_0}}{k^2 u}, \quad (25)$$

388 where $z'_0 = 0.01m$.

389 For a canopy with $0 < LAI < 4$ (as in our case) the two resistance terms in the model are
390 (Shuttleworth and Wallace 1985):

391
$$r_{aa} = \frac{LAI r_{aa}(\alpha)}{4} + \frac{(4-LAI)r_{aa}(0)}{4}, \quad (26)$$

392 and

393
$$r_{as} = \frac{LAI r_{as}(\alpha)}{4} + \frac{(4-LAI)r_{as}(0)}{4}, \quad (27)$$

394 furthermore (according to Shuttleworth and Wallace 1985):

395
$$r_{ac} = \frac{r_{bv}}{2 LAI}, \quad (28)$$

396 and

397
$$r_{sc} = \frac{r_{mst}}{2 LAI}. \quad (29)$$

398 The r_{ss} resistance term was derived according to Hu et al. (2009) as:

399
$$r_{ss} = b_1 \left(\frac{\theta_s}{\theta} \right)^{b_2} + b_3, \quad (30)$$

400 where $b_1 = 2.63 \text{ s m}^{-1}$, $b_2 = 1.32$, and $b_3 = 4.87 \text{ s m}^{-1}$. Empirical constants (b_1 , b_2 , and b_3)
401 were applied for a temperate steppe similar to our site, optimised with a Monte Carlo
402 simulation (Hu et al., 2009) We also tested wide intervals of b_1 , b_2 , and b_3 constants for all
403 types of surface ($b_1 = 1-5$; $b_2 = 1-5$; $b_3 = 1-500$). The variation of b_1 (1-5), b_2 (1-2.6), and b_3
404 (1-5) caused 1.8% and 0.97% variances in the calculated transpiration and evaporation terms,
405 respectively. The increase of b_2 and b_3 to the upper limit (5 and 500, respectively) resulted in
406 a 40-70% increase in the latent heat flux from the canopy and a proportional decrease in the
407 latent heat flux from the soil. In parallel, the correlation also decreased between the
408 calculated and measured latent heat flux towards the upper limit. We obtained maximum
409 correlation by using $b_1 = 2.63 \text{ s m}^{-1}$, $b_2 = 1.32$, $b_3 = 4.87 \text{ s m}^{-1}$, hence we accepted these as
410 optimised values.

411 After calculation of the PM and C terms stomatal transpiration ($C_c PM_c$) and soil
412 evaporation ($C_s PM_s$) can be separated. Using the calculated transpiration rate the stomatal
413 conductance can be computed by inverting the PM equation as suggested by Lamaud et al.
414 (2002):

415
$$\kappa_{st} = \frac{D_{O_3}}{D_w} \frac{\frac{E_t}{\delta}}{1 + \frac{E_t}{\delta}(r_a + r_b) \left(\frac{\beta \Delta}{\gamma} - 1 \right)}, \quad (31)$$

416 where

417
$$\delta = \rho_{vs} - \rho_v; \rho_{vs} = \frac{e_s}{R_w T}, \quad (32)$$

418
$$\rho_v = \frac{e}{R_w T}; e = e_s RH, \quad (33)$$

419
$$e_s = 611 \times 10^{\frac{at}{b+t}} \quad (34)$$

420 where $a = 7.5/9.5$; $b = 237.3/265.5$ °C for water/ice, respectively (Magnus-Tetens formula),

421 and

422
$$s = \frac{e_s \lambda}{R_w T^2}, \quad (35)$$

423
$$\gamma = \frac{c_p p}{0.6215 \lambda}, \quad (36)$$

424 where 0.6215 is the molecular weight ratio of water to dry air.

425 Stomatal flux was derived according to Mészáros et al. (2009) by using the different
426 resistances as:

427

428
$$-F_{st} = c \kappa_{st} \left(\frac{r_c}{r} \right). \quad (37)$$

429 **3 Results**

430 3.1 Validation of the Model

431 Direct validation of the coupled SW-PM model is not possible due to the lack of measured
432 stomatal conductance. The last but one step in the modelling is the calculation of water
433 vapour flux before partitioning it into evaporation and transpiration terms. Hence, we can
434 compare the measured and modelled water vapour fluxes. Fig. 1 shows the regression and
435 correlation between measured and modelled water vapour fluxes for the whole period. The
436 regression parameters suggest a close relationship between the measured and modelled
437 values.

438

439 **INSERT HERE FIGURE 1**

440

441 3.2 Response of model output to the change of main input parameters

442 We examined how predicted stomatal conductance responds to the change of the most
443 effective physical parameters, such as leaf area index (*LAI*), available energy input (*R*), and
444 relative humidity (*RH*) of air (Fig. 2). We tested these variables in the model by changing the
445 value of the investigated variable whilst keeping the others constant. The increase of the
446 available energy input increased the stomatal conductance along a logarithmic scale (left
447 panel).

448 Relative humidity slightly increased the stomatal conductivity at low RH values, while at
449 higher RH , an exponential increase of κ_{st} was observed (middle panel). The strongest
450 dependence was observed in the case of LAI (right panel). At lower LAI values the model
451 output was quite sensitive to an increase of LAI , following a saturation curve towards the high
452 leaf area indices. As it is generally accepted, above $LAI = 4$ the vegetation is regarded as fully
453 developed. In this case soil evaporation does not make a significant contribution to latent heat
454 flux, hence the share of the evaporation term in evapotranspiration decreases, leading to
455 $\lambda E \cong \lambda E_t$. Therefore, the increase of κ_{st} above $LAI = 4$ is weak

456

457 **INSERT HERE FIGURE 2**

458

459 3.3 Daily Fluxes of Ozone

460 Half-hourly average ozone fluxes were calculated according to Eq. 1. Due to the uncertainty
461 of the observations, caused mainly by the lack of turbulence during night hours, the data set
462 was filtered as described in Section 2.3. The seasonal variation of the averaged daily fluxes is
463 illustrated in Fig. 3. Characteristic differences can be seen between the fluxes measured in the
464 growing and dormant periods. It is evident that in the summer half-year (April-September)
465 the role of stomatal uptake is more relevant compared to the dormant season. In the
466 vegetative period, the magnitude of the fluxes greatly depends – among others – on the green
467 biomass, and in particular, on LAI . This can be seen in Fig. 4 where two 12-day periods (see
468 3.4 for details) were compared with different leaf area indices. It can be noted – as described
469 in detail in Section 3.4 – that August 2012 was a dry period in contrast to May 2013 when
470 there was no water limitation affecting the stomatal ozone fluxes (Mészáros et al. 2009).
471 Differences between LAI and moisture characteristics resulted in significantly higher total
472 and stomatal ozone fluxes in May 2013.

473

474 **INSERT HERE FIGURE 3**

475

476 **INSERT HERE FIGURE 4**

477

478 3.4 Partitioning Stomatal and Non-stomatal Conductance

479 The half hourly averages of dry deposition velocities were calculated according to Eq. 4. The
480 canopy conductance κ_c was derived from Eq. (7) and (8). After partitioning the transpiration

481 and evaporation terms according to Eq. (10)-(30), stomatal conductances were calculated by
482 Eq. (31)-(36). At night, the radiation terms have zero or negative values in Eq. (11) ($R_g = 0$
483 and $R, R_s < 0 \text{ W m}^{-2}$) and stomata are practically closed (as it is supposed below in this
484 section); hence, r_{st} is close to infinity and the calculated r_c refers to the non-stomatal
485 resistances, i.e. $r_c = r_{nst}$ or $\kappa_c = \kappa_{nst}$ according to Eq. (9).

486 To evaluate the general pattern of the daily variation of the stomatal and the non-stomatal
487 conductances, we calculated the bulk daily course of these parameters for the total
488 measurement period of August 2012 to January 2014 (Fig. 5), separately for the summer
489 (April-September) and winter (October-March) half-years.

490

491 **INSERT HERE FIGURE 5**

492

493 Night-time transpiration and stomatal conductance were regarded as zero. The summer half-
494 year includes the majority of the growing season; however, growth of above-ground green
495 biomass was also observed at the beginning and at the end of the winter half-year. As it can
496 be seen from the graphs, stomatal conductance is roughly two times higher in the summer
497 half-year. Non-stomatal deposition dominates throughout the day in both seasons, showing a
498 less even pattern than stomatal conductance owing to the great number of physical
499 parameters governing non-stomatal deposition through many different processes. Not only
500 soil deposition, which is dominant for sparse vegetation (characterized by low LAI as
501 observed e.g. by Stella et al. 2013), but also wet leaf surface chemistry, i.e. cuticular
502 deposition, is a sink of ozone. It has to be mentioned here that throughout the modelling
503 period, the observed mean leaf area index was $LAI = 0.5$. The share of stomatal, non-stomatal,
504 and canopy bulk conductances are of the same order of magnitude compared to other
505 investigations (Kelliher et al. 1995, Pio et al. 2000, Tuovinen et al. 2004).

506 A t-test was applied for medium LAI cases when the expected values of κ_{st} and κ_{nst} are
507 similar. In the range of $LAI = 1.0-2.2$ the means of the two conductances were 0.23 and 0.25
508 cm s^{-1} , respectively. Only daytime cases were taken into account since at night $\kappa_{st} = 0$. The
509 parameters of the t-test were $t = -2.06$, $t_{0.05} = 1.96$, $p = 0.04$, and $n = 1011$. Since $|t| > t_{0.02}$ ($p <$
510 0.05) the two datasets are significantly different.

511 The combined effect of low moisture availability and sparse vegetation on the stomatal
512 uptake, calculated by Eq. (37), is well represented by the substantial difference in stomatal

513 flux in the dry season with $LAI = 0.25$ (Fig. 4, left panel) and in the wet period with leaf area
514 indices being 4-times higher (right panel), as it can be followed in Table 3.

515 For a more detailed examination of stomatal and non-stomatal conductances in the
516 growing season, we analysed two 12-day observation periods in August 2012 and May 2013
517 (Table 3). The criteria for selection were: i) a continuous dataset, ii) as large a difference
518 between mean LAI as possible (1.05 vs. 0.25), 3) the period is part of the growing season.

519

520 **INSERT HERE TABLE 3**

521

522 The first investigated period (12-23 August 2012) was a typical, dry summer season with
523 no rain. The daily maximum of global and net radiation was 770-890 $W m^{-2}$ and 550-575 W
524 m^{-2} , respectively (except 12 August, which was a cloudy day), and the daily maximum values
525 of the latent heat fluxes did not exceed 80-130 $W m^{-2}$. The typical daytime Bowen-ratio was
526 $\beta = 1.4$. The mean leaf area index (LAI) was 0.25 with a mean above-ground green biomass
527 (AGB) of 3.2 $g m^{-2}$; other mean physical parameters in August were $RH = 57\%$; leaf wetness
528 $LW = 11\%$, and air temperature $t_a = 20$ °C.

529 The second period (2-13 May 2013) was a typical late spring period with 32 mm of
530 precipitation on 4 rainy days. There was a large variation in the daily maximum values of the
531 global and net radiations. They varied within 200-865 $W m^{-2}$ and 100-600 $W m^{-2}$,
532 respectively. The typical daytime Bowen-ratio values were $\beta = 0.25-0.40$. In this period, there
533 was no water limitation. These 12 days can be characterized as: mean $LAI = 1.05$; AGB
534 (green) = 96 $g m^{-2}$; $RH = 75\%$; $LW = 25\%$, and $t_a = 17$ °C.

535 Conductances and fluxes were selected according to global radiation into daytime (R_g , $R >$
536 $0 W m^{-2}$) and night ($R_g = 0$ and $R < 0 W m^{-2}$) groups. Night-time transpiration was regarded
537 as negligible with zero stomatal conductance.

538 Although incomplete closing of stomata has been observed during the night (Caird et al.
539 2007), very little is understood about this phenomenon. At night the main governing factors
540 for transpiration, e.g. the water vapour pressure difference between leaves and air as well as
541 atmospheric mixing, are much lower than during the daytime; hence, transpiration is lower by
542 one order of magnitude, it represents only 5-15% of the daytime rate. The magnitude of
543 stomatal exchange can also be estimated by comparing the ratio of the mean calculated
544 transpiration terms (E_t) during the daytime (R_g and $R > 0 W m^{-2}$) and at night ($R_g = 0$ and $R <$
545 $0 W m^{-2}$). They were 1.53×10^{-5} (day), $0.0269 \times 10^{-5} kg m^{-2} s^{-1}$ (night) and 3.28×10^{-5} (day),

546 $0.336 \times 10^{-5} \text{ kg m}^{-2} \text{ s}^{-1}$ (night) in August 2012 and May 2013, respectively. The ratios of day
547 to night transpiration rates were 56.7 in August (mean $LAI = 0.25$) and 9.76 in May (mean
548 $LAI = 1.05$). Similarly, the day to night stomatal conductivity ratio for water vapour
549 calculated as $\kappa_w = E_t/\rho_v$ was 51.4 in August 2012 and 16.7 in May 2013. These values verify
550 the at least one order of magnitude lower transpiration rate at night especially for the
551 examined ecosystem. Therefore, in this study we considered the night-time transpiration rate
552 and stomatal conductance as negligible.

553 The first period is represented by a low leaf area index of 0.25. In the second period the
554 vegetation is more developed with an average $LAI = 1.05$ (Table 3). There are further
555 differences between the two periods; namely, in May 2013 the relative humidity and the leaf
556 wetness were higher and a large increase was observed in the mass of above-ground green
557 biomass. Evidently, there are parallel increases in the number of stomata with increasing LAI
558 and AGB (green) which is reflected in the 8.5-times higher stomatal conductivity in the
559 daytime in May compared to August when lower LAI values and drought were observed.
560 There is a factor of 2 in the non-stomatal conductance between lower and higher LAI
561 situations, showing the importance of cuticular deposition, and the relatively wet climate
562 regime in May 2013 that favours not only cuticular uptake but also deposition processes to
563 wet surfaces. In the season represented by $LAI = 0.25$ the ratio of κ_{nst}/κ_{st} is around 4-5 and
564 when the LAI reaches unity ($= 1$), the daytime ratio of these two parameters becomes the
565 same in magnitude.

566 A similar pattern can be seen in the total ozone flux and in the stomatal flux in Table 3.
567 While total ozone flux has doubled due to growth of LAI and other factors, stomatal flux
568 increased by a factor of 5. These variations can also be observed in Fig. 6 and 7, where the
569 variation of stomatal and canopy conductances as well as total and stomatal ozone fluxes are
570 illustrated together.

571

572 **INSERT HERE FIGURE 6**

573

574 **INSERT HERE FIGURE 7**

575

576 When vegetation is completely covered by snow there appears to be no stomatal activity.
577 Table 3 shows this situation on five selected days (15; 16; 26; 27; 28 March 2013) with the
578 highest snow depth episodes (12-16 cm) completely covering the 5-7 cm tall vegetation. In

579 this case $\kappa_{\text{nst}} = \kappa_{\text{c}}$ refers to the ozone surface conductance to snow. Conductances were small,
580 approximately 0.03 cm s^{-1} on average (in agreement with earlier observations, e.g. by Wesely
581 et al. 1980), practically independent of the period of the day.

582 During control days in the same month (12; 13; 19; 20; 21 March, 2013) maximum daily
583 temperatures ranged between 10 and 15 °C, net radiations were below 200 W m^{-2} , the
584 vegetation was still free of snow, but regarding the dormant season stomatal conductance was
585 negligibly small.

586 Interestingly non-stomatal conductivity is as high in magnitude as in the following May.
587 This phenomenon can be explained by the wet surfaces as illustrated in Table 3, by the high
588 relative humidity and soil water content, indicating the importance of surface loss processes
589 in the non-stomatal deposition of ozone.

590

591 **4 Summary and Conclusion**

592 We partitioned canopy conductance into different parts (non-stomatal and stomatal) by
593 calculating the stomatal conductance separately. For well-developed vegetation ($LAI > 4$)
594 evaporation in the evapotranspiration process is practically negligible, hence transpiration can
595 be used to calculate the stomatal conductance of water vapour and ozone, using the similarity
596 between them described by the PM theory. In the case of low, sparse vegetation ($LAI < 4$),
597 evaporation is no longer negligible; therefore, E has to be partitioned into E_t and E_e to
598 estimate stomatal conductance for water and for ozone using the transpiration term in the PM
599 equation. We found that the coupled SW and PM model can simulate and partition stomatal
600 and non-stomatal conductances over short, low, and sparse vegetation, where evaporation is
601 of the same magnitude or even more significant than transpiration. Our result suggests that
602 the non-stomatal part is highly significant in controlling total ozone deposition to sparse
603 vegetation.

604

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611

612

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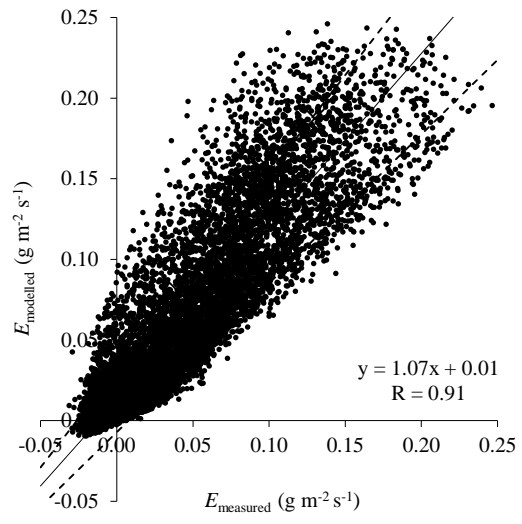
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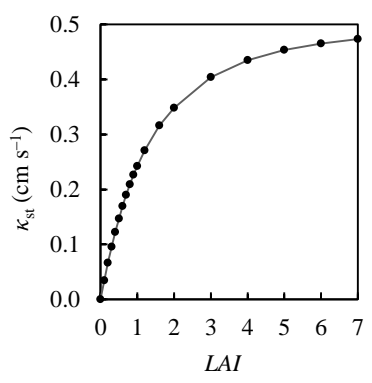
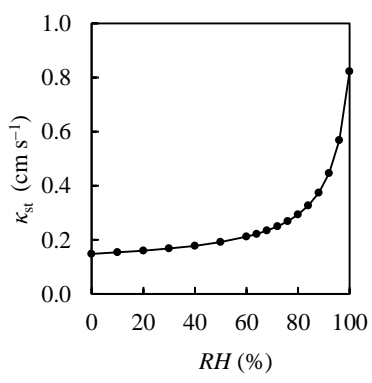
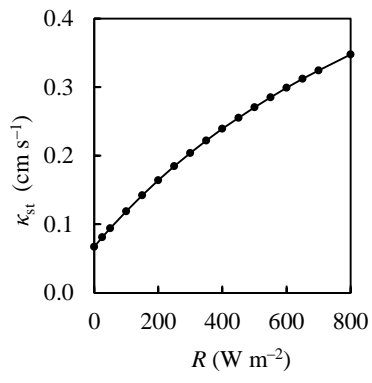
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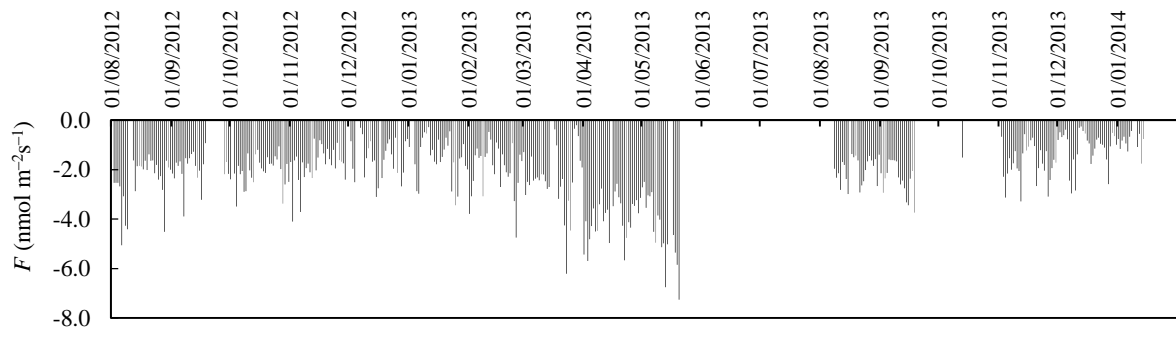
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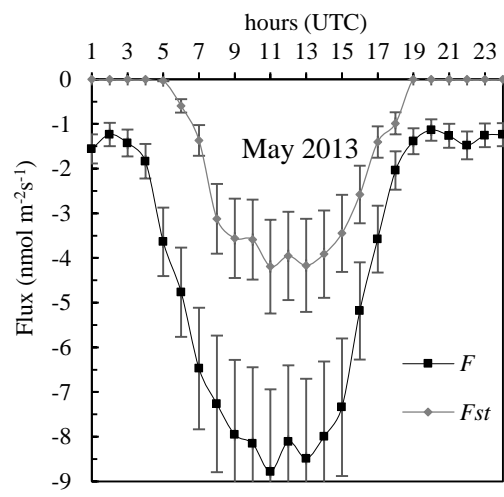
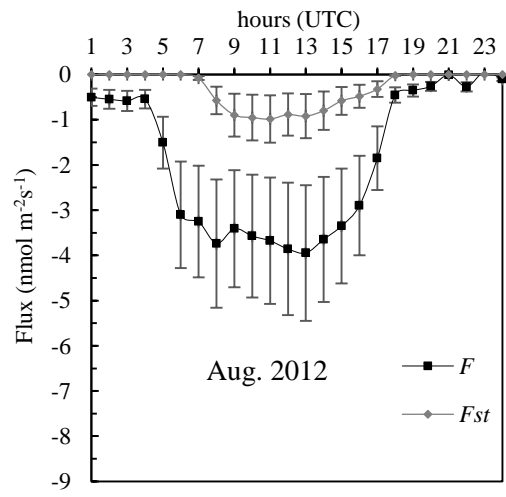
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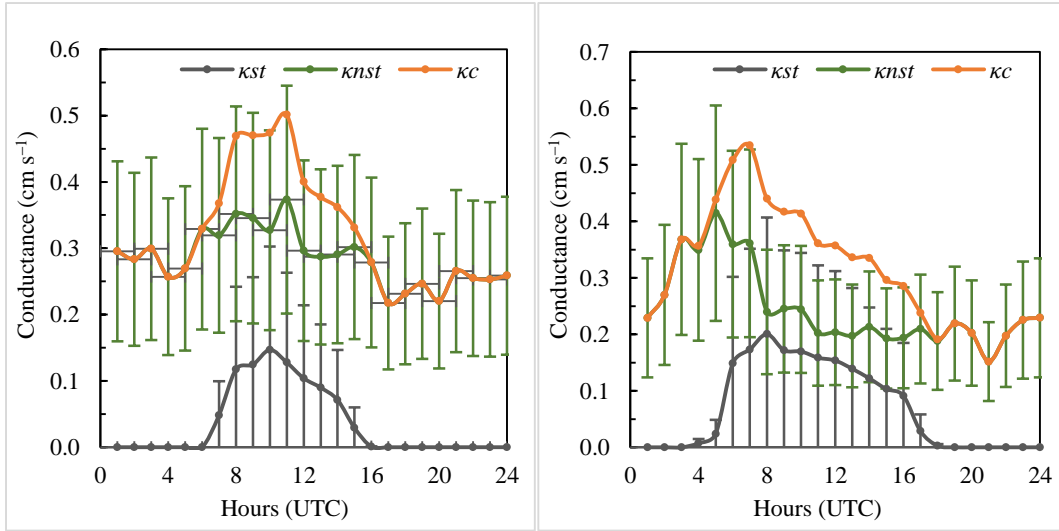
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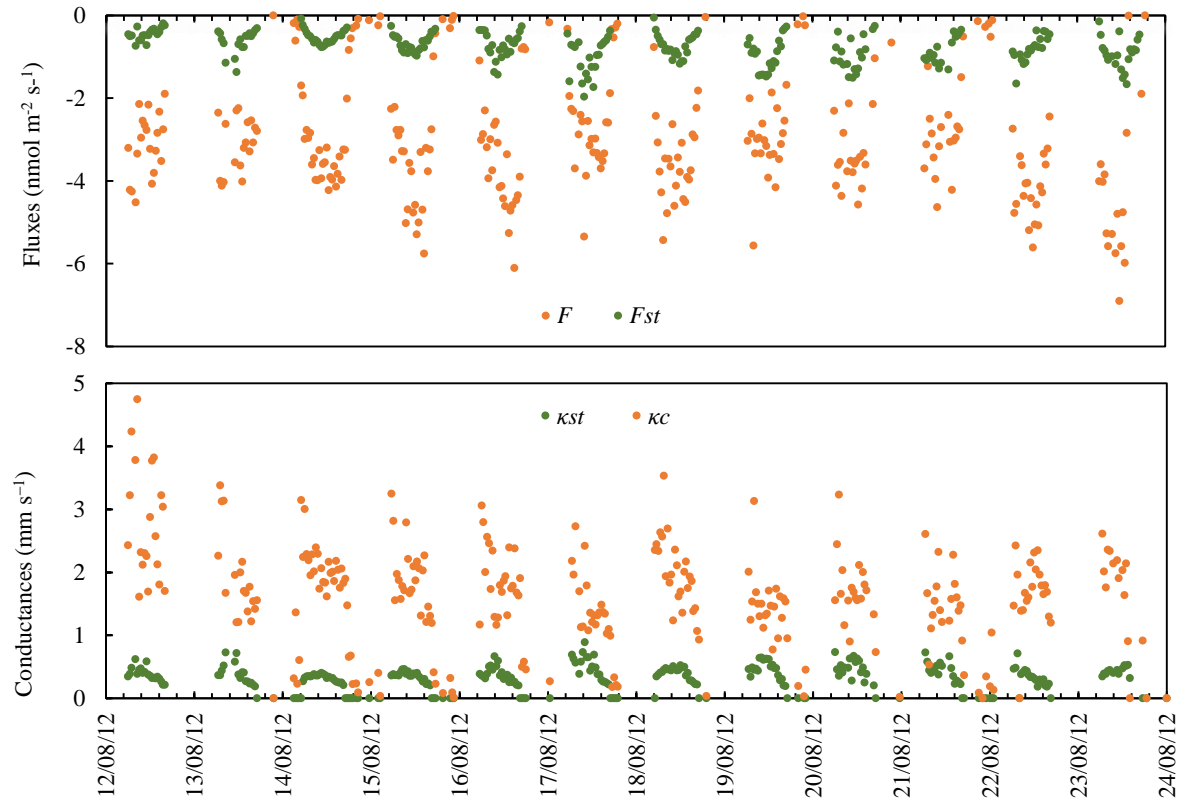
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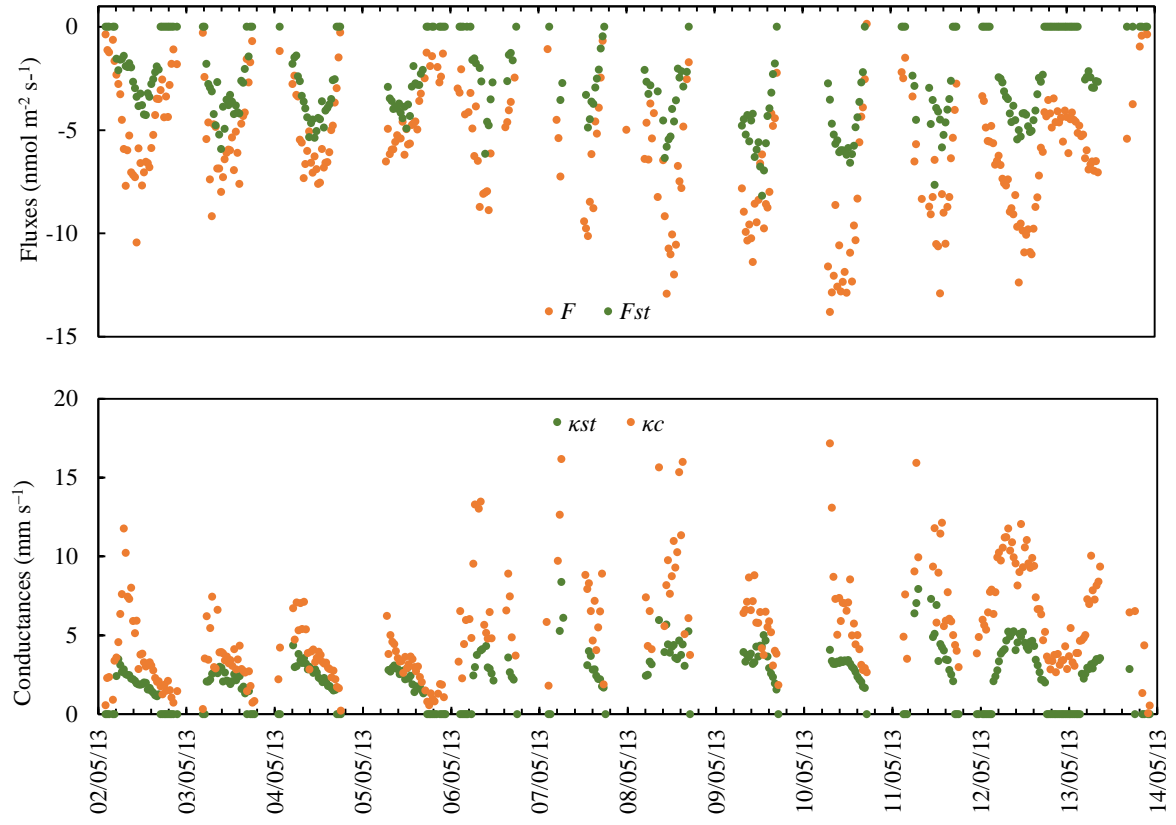
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Parameter	Symbol	Instrument / method	Logging time
sensible heat flux	H	CSAT3 ultrasonic anemometer / eddy covariance (EC)	0.1 s
latent heat flux	λE	CSAT3 ultrasonic anemometer + Li-Cor 7500 (EC)	0.1 s
momentum flux	τ	CSAT3 ultrasonic anemometer (EC)	0.1 s
net radiation	R_n	NR Lite net radiometer	30 min
ozone concentration (slow)	c	HORIBA APOA 350 ozone monitor	10 min
ozone concentration (fast)	c	ENVISCOPE fast response sensor	0.1 s
wind velocity	u	from CSAT3 ultrasonic anemometer	30 min
vegetation height	h	ruler at $5 \times 40 \text{ cm}^2$ quadrats	1-2 weeks
above ground biomass	AGB	balance (samples from quadrats)	1-2 weeks
soil water content	θ	CS616 WC reflectometer at -0.03 ; -0.30 m	30 min
soil temperature	t_s	Campbell 105 T thermocouple at -0.05 ; -0.30 m	30 min
density of air	ρ	calculated from T , RH and p	30 min
relative humidity	RH	Vaisälä HMP35AC	30 min
air temperature	t	Vaisälä HMP35AC	30 min
air pressure	p	Li-Cor 7500	30 min
leaf area index	LAI	CEP-40 ceptometer (Decagon Devices, USA)	1-2 weeks
leaf surface wetness	LW	5 Bayreuth-type clips	30 min

Table 1 List of measurement methods.

Stability	No. of cases	TK3 corr. $\overline{w'c'}$	TK3 corr. $\overline{w'T'}$	Co-spectr. corr.	Semi-empir. corr. $\overline{w'c'}$
Unstable $-1 < \zeta \leq -0.1$	145	10.8±4.0%	3.6±1.3%	6.2±5.2%	10.0±5.4%
Near neutral $-0.1 < \zeta \leq 0.1$	317	16.9±4.3%	5.8±1.6%	7.1±5.5%	13.4±6.1%
Stable $0.1 < \zeta \leq 1$	127	18.8±6.2%	7.3±2.4%	5.6±5.0%	13.3±6.1%

note: TK3 corr. $\overline{w'c'}$ and TK3corr. $\overline{w'T'}$ are the TK3 spectral corrections for covariances (Mauder and Foken 2011); co-spectr. corr. are the maximum differences of relative co-spectrums for covariances (Ammann et al. 2006); semi-empirical corr. $\overline{w'c'}$: semi-empirical ozone flux correction (in %) calculated as:
 $[(1+\text{corr. } \overline{w'T'}/100) \times (1+\text{Co-spectr. corr.}/100) - 1] \times 100$

Table 2 Mean value and relative error of different types of spectral corrections for ozone flux measurements in May 2013.

Season	Period	κ_{st} cm s ⁻¹	κ_{nst} cm s ⁻¹	κ_c cm s ⁻¹	F nmol m ⁻² s ⁻¹	F_{st} nmol m ⁻² s ⁻¹
Vegetation <i>LAI</i> = 0.25	daytime	0.035±0.015	0.151±0.070	0.186±0.066	-3.33±1.17	-0.72±0.37
	night	0	= κ_c	0.032±0.031	-0.13±0.37	0
Vegetation <i>LAI</i> = 1.04	daytime	0.299±0.135	0.335±0.281	0.634±0.335	-6.87±2.69	-3.61±1.48
	night	0	= κ_c	0.311±0.193	-2.90±1.57	0
Winter snow	daytime	0	= κ_c	0.032±0.034	-0.42±0.42	0
	night	0	= κ_c	0.027±0.035	-0.31±0.33	0
Winter no snow	daytime	0.008±0.005	0.467±0.228	0.475±0.227	-4.87±2.20	-0.104±0.052
	night	0	= κ_c	0.233±0.216	-1.80±2.09	0

Table 3 Arithmetic mean and standard deviation ($\pm 1 \sigma$) of stomatal (κ_{st}), non-stomatal (κ_{nst}) and canopy (κ_c) conductances, total flux (F) and stomatal flux (F_{st}) during daytime (R_g , $R > 0 \text{ W m}^{-2}$) and night ($R_g = 0$ and $R < 0 \text{ W m}^{-2}$).

Legends of figures

Fig 1 Comparison of water vapor flux calculated by Eq. (10) to eddy covariance measurements based on 14,688 half hourly measurements (2012 August – 2014 January). Dotted lines show the $\pm 1 \sigma$ intervals.

Fig 2 Variation of the modelled stomatal conductance as a function of the main governing physical parameters (others are kept at constant values: $R = 376 \text{ W m}^{-2}$; $RH = 65\%$; $LAI = 0.93$).

Fig 3 Averaged daily ozone fluxes.

Fig 4 Diurnal variation of total (F) and stomatal (F_{st}) ozone flux in August 2012 with mean $LAI = 0.25$, (left); and in May 2013 with mean $LAI = 1.05$, (right).

Fig 5 Daily course of stomatal (κ_{st}), non-stomatal (κ_{nst}), and canopy bulk (κ_c) conductances for the periods of the winter half-year (October-March), left; and the summer half-year (April-September), right. Each hourly average includes the previous two half-hour measurements. Note: during night $\kappa_c = \kappa_{nst}$. The error bars for κ_{st} and κ_{nst} are illustrated.

Fig 6 Variation of fluxes (top) and conductances of ozone (bottom) between 12-23 Aug. 2012 ($LAI=0.19-0.31$).

Fig 7 Variation of fluxes (top) and conductances of ozone (bottom) between 2-13 May 2013 ($LAI=0.90-1.19$).

Highlights

- evapotranspiration was partitioned into E_t and E_c parts by the SW model
- E_t was used to calculate stomatal conductance (κ_{st}) of ozone by the PM equation
- canopy conductance (κ_c) was calculated from eddy covariance measurement of O_3 flux
- stomatal and non-stomatal conductances were partitioned as $\kappa_c = \kappa_{st} + \kappa_{nst}$
- non-stomatal deposition of O_3 dominates especially for low LAI vegetation