

Soil seed bank of the invasive *Robinia pseudoacacia* in planted *Pinus nigra* stands

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Abstract – *Pinus nigra* and *Robinia pseudoacacia* are exotic trees used for afforestation in Hungary. *Pinus nigra* was non-invasive, however *R. pseudoacacia* escaped from cultivation and invaded several vegetation types including pine plantations. It has recently been planned to cut *P. nigra* plantations and replace them by native tree stands, especially in nature reserves. The scattered presence of *R. pseudoacacia* specimens in pine stands might place constraints on planned tree replacement because of their vegetative resprouting and recolonization from an established seed bank. The aim of this study was to investigate the soil seed bank under the canopy of solitary *R. pseudoacacia* specimens found in *P. nigra* plantations. Altogether 250 soil samples were collected from the 0–6 and 6–12 cm soil layers under solitary *Robinia* trees of varying ages (with basal areas between 62.4 and 1089.3 cm²). Seeds were separated by sieving then scarified and germinated. Seed bank density ranged between 640 and 2285 seeds m⁻² with an average distribution of 82.7% and 17.3% in the upper and lower soil layer, respectively. Total density of the seed bank and also the seed bank ratio of the lower soil layer increased with tree age. The accumulated seed bank of *R. pseudoacacia* should be considered in the careful planning of tree replacement operations in *Pinus nigra* stands.

Keywords: Afforestation, dormancy, *Pinus nigra*, plantation, *Robinia pseudoacacia*, seed germination, soil seed bank

Abbreviations: BA – basal area of tree, DSB – density of seed bank, SBR – seed bank ratio of the lower soil layer, USB – seed bank density in the upper soil layer

Introduction

Black locust (*Robinia pseudoacacia* L.) is native to the eastern part of North America. Its introduction to Hungary dates back to 1710 and it has been intensively used in afforestation practices since then, due to its diversified utilization (WALKOVSKI 1998). At present black locust stands cover about 400,000 hectares, or 23% of the total forested lands of Hungary, a higher coverage than in other European countries (RÉDEI et al. 2008). The black lo-

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cust has a remarkable spreading capacity due to its rapid vegetative propagation, high adaptability and nitrogen-fixing character (SWAMY et al. 2002, RICE et al. 2004). The invasion success of *Robinia* and of other woody *Fabaceae* species is further enhanced by their persistent soil seed bank and the physical dormancy of seeds (RICHARDSON and KLUGE 2008). The life expectancy of *Robinia* seeds in soil is considerably prolonged by antimicrobial proteins accumulated in the seed tissues, making them resistant to most pathogens (TALAS-OGRAŠ et al. 2005). This species has accordingly become one of the most important woody plant invaders (CRONK and FULLER 1995) and naturalized in Asia and Australia as well as in the Western and Central part of North America (HOLLE et al. 2006).

In Hungary, black locust spreads spontaneously mainly in semi-arid sandy areas with average annual rainfall around 550–600 mm (RÉDEI et al. 2001). Its ever increasing use in afforestation gives a further impulse to its spread by creating new sources of invasion. As a consequence, the black locust is now recorded as one of the most dangerous invasive neophyte species in Hungary (BALOGH et al. 2004).

The Austrian pine (*Pinus nigra* Arn.) is an indigenous tree in the Balkan-Mediterranean region. In Hungary, it was first introduced in the second half of the 19th century (TAMÁS 2003), and nowadays its stands cover 63,000 hectares (source: Hungarian Forest Management Inventory). Alien tree plantations are usually characterized by a poorly developed herb layer due to the absence of a well organized community of accompanying grasses, forbs and other species. This phenomenon is especially characteristic of Hungarian *P. nigra* plantations, where the strong canopy shading and litter accumulation eliminate the species-rich vegetation existing prior to afforestation with the pine (CSONTOS et al. 1996, 2007). The lack of competitive herb- and shrub-layers makes Austrian pine plantations more susceptible to invasion by aggressively spreading aliens than diverse native forest communities (ALPERT et al. 2000, MANDRYK and WEIN 2006). In Hungary, the widespread invasion of *R. pseudoacacia* in Austrian pine stands is a typical example of this phenomenon, but further species like *Ailanthus altissima* (Mill.) Swingle, *Asclepias syriaca* L. and *Phytolacca americana* L. could also be mentioned in this respect.

The low quality timber of the Austrian pine has a limited applicability and it has lost any economic importance during recent decades. Therefore in Hungary – as in other European countries (AUGUSTO et al. 2001) – the replacement of alien pine plantations by native forests or grass vegetations has been begun and intensively executed, especially in nature reserves and national parks. Obviously, the soil seed banks of alien invasive plants form a real threat to successful conversion of Austrian pine plantations to native vegetation types. Among the invasive species listed above, *R. pseudoacacia* deserves especial attention because of its pronounced ability to form a long-term persistent seed bank in the soil (THOMPSON 1993), due to the physical dormancy of seeds caused by hardseededness (CZIMBER 1980). The soil seed bank of the non-native black locust may cause potential nature conservation and forest management problems on the clear-felled pine stands, even if stubs are pulled out to prevent re-sprouting from roots and trunks. Therefore, the first aim of our studies was to quantify the seed bank of the black locust in the soil Austrian pine plantations that it has invaded.

Black locust seeds are known to remain viable for some decades (up to 40 years) in the soil (TOOLE and BROWN 1946), and thus their accumulation under mature specimens is ex-

pected. The second aim of our studies was to highlight the relationship between the age of individual trees (represented by basal area) and the soil seed bank density beneath their canopy.

Seeds buried in deeper soil layers generally remain viable for longer period of time than those positioned close to the soil surface (FENNER and THOMPSON 2005). If this applies for black locust, then the ratio of viable seeds in the lower soil layer compared to the total amount of soil seed bank should be increased under older tree individuals. Investigation of this presumption formed the third aim of the present studies.

Materials and methods

For soil seed bank studies, five Austrian pine stands invaded by black locust in the North Hungarian region covered by sandy soil were selected: on the boundary of the settlements of Ács, Isaszeg, Csévharaszt, Tárkány and Komárom (Fig. 1). The information about localities and main characteristics of the pine plantations were obtained from local forest inventories (Tab. 1).

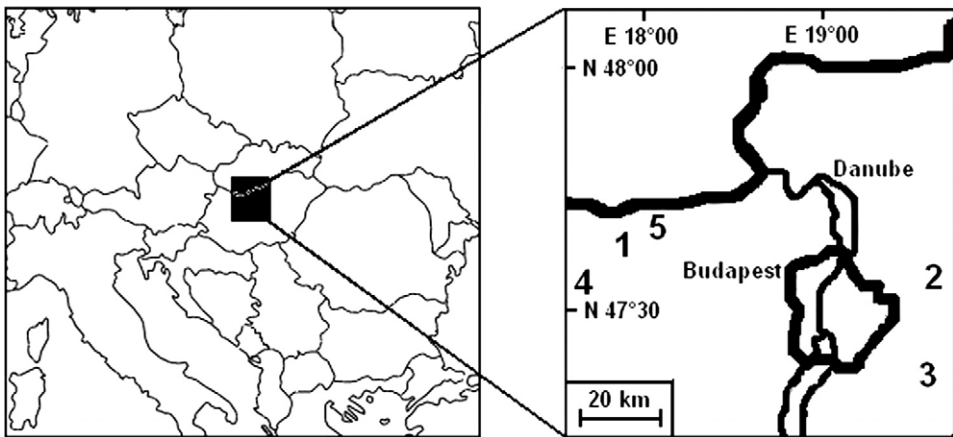


Fig. 1. Geographical position of *Pinus nigra* stands involved in soil seed bank sampling. 1 – Ács; 2 – Isaszeg; 3 – Csévharaszt; 4 – Tárkány; 5 – Komárom

Tab. 1. Characteristics of the studied *Pinus nigra* stands invaded by *Robinia pseudoacacia* in Hungary.

No. of sites	Location	GPS positions	Stand age (years)	Stand area (hectares)
1	Ács	N 47°44'51.6"; E 18°01'00.2"; 120 m asl.	28	5.12
2	Isaszeg	N 47°31'27.1"; E 19°21'25.4"; 205 m asl.	48	4.23
3	Csévharaszt	N 47°16'53.3"; E 19°24'56.2"; 129 m asl.	51	2.47
4	Tárkány	N 47°35'06.6"; E 17°56'05.1"; 143 m asl.	57	7.76
5	Komárom	N 47°44'52.4"; E 18°02'14.8"; 127 m asl.	68	12.35

Soil sampling was carried out between 17 July and 12 August 2009. In each sampling site five black locust trees were chosen in the interior of the plantation, i.e. at least 20 m distance from the edge of the pine plantation (to avoid the edge effect). Also, attention was paid to selecting solitary black locust individuals, thus ensuring that the soil seed bank beneath the targeted tree was not influenced by neighbouring black locust specimens. Afterwards the basal area at breast height (BA) was determined by trunk perimeter measurement. Five sampling points were marked out around each tree at a distance of 1.5–2.0 m from the trunk base. In the sampling points leaf litter and freshly fallen legumes were removed from the soil surface, then soil cores of 80 cm² surface area and 480 cm³ volume were cut from the upper (0–6 cm) and the lower (6–12 cm) soil layers. The five-five subsamples originating from the same vertical layers were bulked, thus forming a 2400 cm³ total soil volume per layer (4800 cm³ per tree). Penetration of *Robinia pseudoacacia* seeds into soil layers deeper than 12 cm is negligible (MARJAI 1995), thus the applied sampling depth was considered to be sufficient.

Samples were transported to the laboratory and washed through a metal sieve with mesh size of 1.5 mm. After room-temperature (22 °C) drying, black locust seeds were hand-sorted from the debris and counted. The number of viable seeds (which actually forms the seed bank) was calculated after performing germination tests. Prior to the germination procedure, seeds were surface sterilized by soaking them for two minutes in a 20% ethanol solution to deal with mould (CHUANREN et al. 2004). Hardseededness was reduced by mechanical scarification carried out with emery paper (BASKIN and BASKIN 1998). Seeds were placed on filter paper moistened with tap water in Petri dishes and kept at 24 °C temperature for 21 days. Evaporated water was supplied as necessary, and the germinated seeds were removed daily. On the fifth day the non-swollen seeds were re-scarified and re-germinated. Numbers of germinated and non-germinated seeds were determined on the 21st day.

The mean seed bank density of each studied Austrian pine stand was calculated (as seeds m⁻²) by averaging the results of the five black locust trees sampled. Consequently, these values can be regarded as the outgrowths of a 12,000 cm³ sample volume, which amount exceeds 2–3-fold the required minimal volume, which is generally 4000–6000 cm³ in climax forest vegetation (CSONTOS 2007). MORIMOTO et al. (2010) collected a 12,500 cm³ sample volume from the upper 5 cm soil layer for investigation of the *R. pseudoacacia* seed bank.

The relationship between basal area of tree (BA; cm²) and density of seed bank (DSB; seeds m⁻²) was evaluated by standard regression methods: linear, power and logarithmic regression. DSB was calculated by adding the numbers of viable seeds found in the upper and lower soil layer. Further regression analyses were also carried out in order to relate the BA (*i*) to density of seed bank found in the upper layer only (USB), and (*ii*) to seed bank ratio of lower soil layer (SBR; %). In the latter case, SBR was calculated as seed density of the lower soil layer divided by the DSB. Statistical significance was assessed at P = 0.05, and among the regression types the one serving the best fit (greatest R²) was accepted.

Results

Altogether 250 soil core subsamples were collected (125 from each soil layer), thus 50 samples (25 from each soil layer) were analysed after bulking 5–5 subsamples of each tree sampled. The black locust seed bank was present in both soil layers of each sampled Aus-

trian pine stands. On average 48.6 (ranging from 20 to 96) and 10.1 (ranging from 0 to 38) seeds per tree were found in the upper (0–6 cm) and the lower (6–12 cm) soil layer, respectively. Among the 25 black locust trees sampled there was only one in the pine stand of Ács, and below it no seeds were found in the lower soil layer. This black locust tree has the smallest BA among the studied specimens.

Altogether 1466 seeds were washed out of soil samples and subjected to germination testing. The seeds expressed a high germination rate: 1371 seeds, or 93.5% of the total amount, proved to be viable. Seeds showed a greater germination rate (95.6%) in the lower than in the upper soil layer (93.1%). On the basis of germination results, 1398 seeds m⁻², the mean soil seed bank density was calculated: 1156 and 242 seeds m⁻² in the upper and the lower soil layer, respectively (Fig. 2). Seed bank densities varied among different Austrian pine stands. The greatest seed bank, 2285 seeds m⁻² was developed in the pine stand of Komárom (in 0–6 cm soil layer: 1660 seeds m⁻²; in 6–12 cm soil layer: 625 seeds m⁻²), while the smallest, 640 seeds m⁻², was found in the pine stand of Ács (only 600 and 40 seeds m⁻² average densities were calculated in the upper and lower soil layer, respectively).

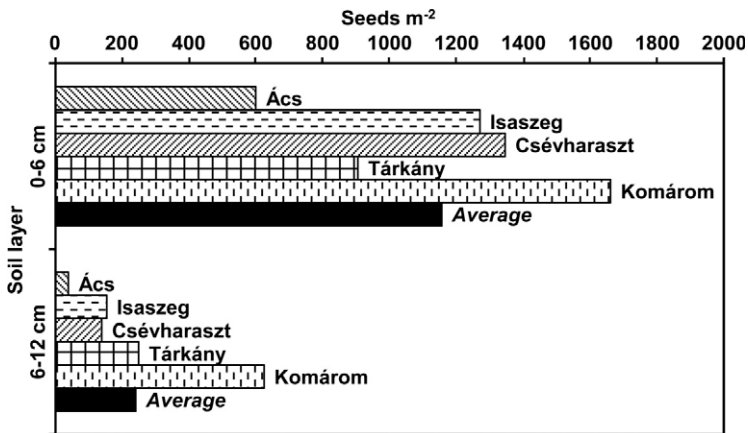


Fig. 2. Average densities of soil seed bank in the upper (0–6 cm) and lower (6–12 cm) soil layers under *Robinia pseudoacacia* trees growing in *Pinus nigra* stands, in Hungary. Sampling sites are ordered according to age of the plantations (solid bars show the average values of the five sites).

Seed distribution between the two soil depths also had a great variability. The mean percentage ratio of seeds found in the upper soil layer was 82.7% of the total amount, consequently 17.3% of seeds appeared from the lower soil layer (Fig. 3). The highest (27.4%) and smallest (6.3%) seed bank ratio in 6–12 cm soil depths were detected in the pinewoods of Komárom and Ács, respectively (the same plantations that were responsible for the highest and the smallest total seed bank densities).

The basal area of the sampled black locust trees ranged from 62 to 1089 cm² (average value was 488 cm²). The simple regression analysis showed a positive correlation ($P < 0.001$) between the the basal area of tree (BA, cm²) and the total density of seed bank (DSB; seeds m⁻², Fig. 4). The obtained regression equation [1] indicates a curvilinear relationship:

$$DSB = 61.87 BA^{0.504} \quad [1]$$

adjusted $R^2 = 0.7003$. Relationship between BA and seed bank density in the upper soil layer (USB) proved to be similar ($USB = 78.33 BA^{0.433}$) but at a lower fit ($R^2 = 0.5596$).

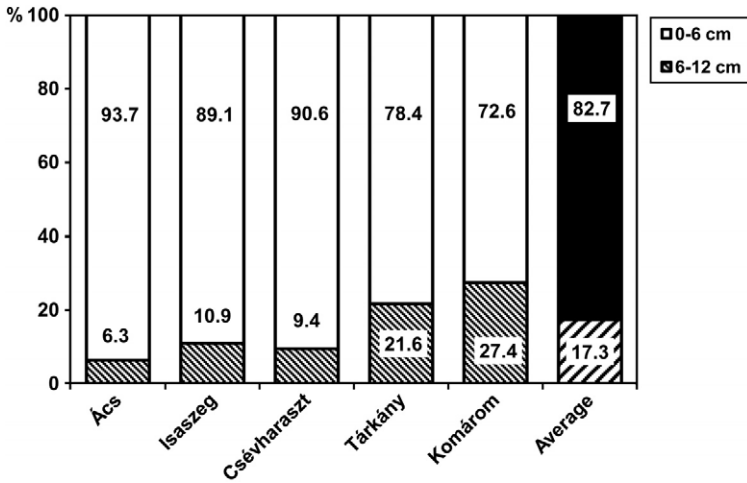


Fig. 3. Percentage share of soil seed bank distribution between the upper (0–6 cm) and lower (6–12 cm) soil layers under *Robinia pseudoacacia* trees growing in *Pinus nigra* stands. Sampling sites are ordered according to age of the plantations (right end bar shows the average of the five sites).

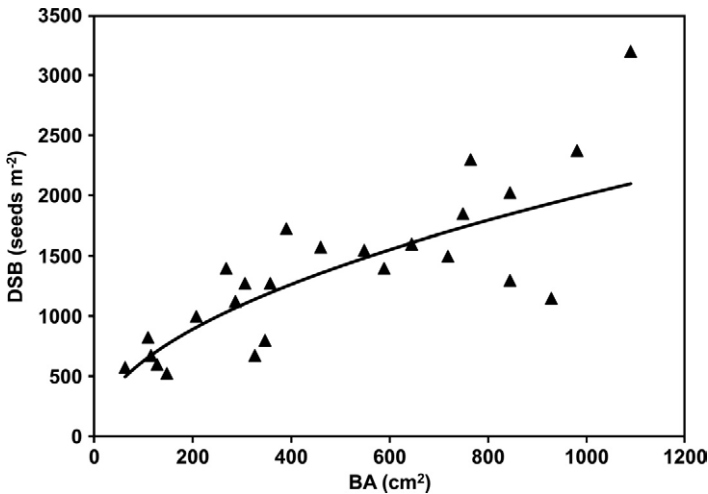


Fig. 4. Correlation between the total density of seed bank (DSB; seeds m^{-2}) and the basal area (BA; cm^2) of *Robinia pseudoacacia* trees growing in *Pinus nigra* stands, in Hungary. The equation of curve: $DSB = 61.87 BA^{0.504}$ ($R^2 = 0.7003$)

Regression analysis resulted in a linear and positive relationship ($p < 0.001$) between the BA and the seed bank ratio of lower soil layer (SBR) (Fig. 5). Increase in BA contributes to the increase of SBR (in %), according to equation [2]:

$$\text{SBR} = 0.0204 \text{ BA} + 3.0576 \quad [2]$$

adjusted $R^2 = 0.5579$.

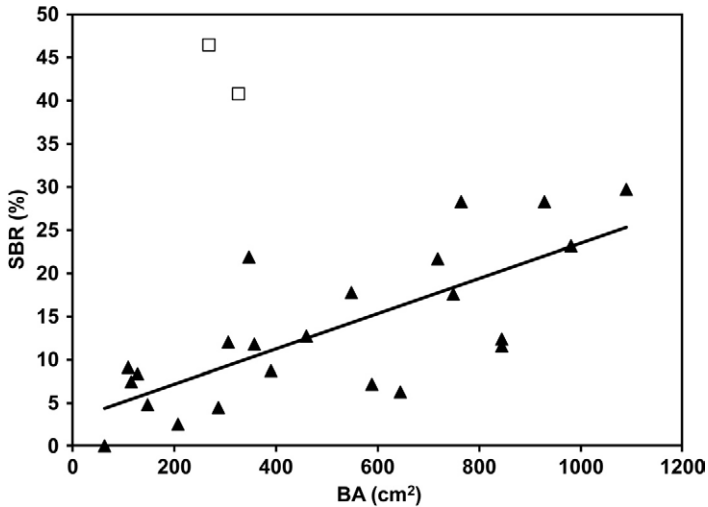


Fig. 5. Correlation between the seed bank ratio of lower (6–12 cm) soil layer (SBR; %) and the basal area (BA; cm²) of *Robinia pseudoacacia* trees growing in *Pinus nigra* stands, in Hungary. (Empty squares show outliers excluded from the regression analysis.) The equation of line: $\text{SBR} = 0.0204 \text{ BA} + 3.0576$ ($R^2 = 0.5579$).

Two SBR data had to be excluded from statistical evaluation, because of the performed outlier analysis: one black locust tree in the sampled pinewood of Komárom and Tárkány as well.

Discussion

Our investigation verified the ability of the black locust to form a persistent soil seed bank in Austrian pine stands, and also showed a correlation between seed bank density and age of tree. The density of *Robinia pseudoacacia* seed banks ranged from 640 to 2285 seeds m⁻² in the pine plantations studied; its mean value was about 1400 seeds m⁻². Under black locust trees in several of the city parks of Budapest, an 871 seeds m⁻² soil seed bank density and – after mechanical scarification – a 94% germination rate were reported by SIMKÓ and CSONTOS (2009). Our mean germination result (93.5%) also agrees with the 92–98% value stated by MASAKA and YAMADA (2009) on the basis of their ecophysiological research executed in Japan. An extensive field study showed a 96% germination rate and a 2000–12 000 seeds m⁻² seed bank density in monodominant black locust stands

in Hungary (MARJAI 1995). In this case, the monodominant character of the studied *Robinia* plantations can explain the relatively high seed bank density. As opposed to MARJAI's (1995) research, our investigation was carried out under the canopies of solitary black locust trees in Austrian pine stands, where the amount of seeds incidentally scattered to greater distances, was not counterbalanced by seed rain of neighbouring trees.

Black locust seed bank was detected in both soil layers. Since the penetration of seeds to deeper soil layers is time-consuming, a high seed density ratio in the 6–12 cm soil depth was mainly observable under old black locust trees, i.e. having large basal area (Fig. 5). The extremely high seed density ratios (shown by the two outlier data) was probably due to local disturbances caused by forest animals or human activities, that by turning up lower soil layers resulted in a local »seed bank profile inversion« (CSONTOS 2007).

Higher germination rate was detected in the lower (95.6%) than in the upper one soil layer (93.1%). Most probably this resulted (i) partly from the favourable environmental conditions for seed survival associated with deep burial (WITKOWSKI and GARNER 2000, FENNER and THOMPSON 2005), and (ii) partly from the higher decomposition rate of non-viable seeds in the deeper soil layer. Both the growing seed production of elderly trees and the long-term accumulation of dormant seeds may interpret the age-dependent increase in soil seed bank density (MARJAI 1995).

Though the fecundity of *R. pseudoacacia* generally declines after about 40–50 years, corresponding to the decrease in tree vigour, the soil seed bank density may continue to increase owing to the considerable accumulation of dormant seeds (MASAKA et al. 2010). *Robinia* seed longevity does not exceed 40 years in general (TOOLE and BROWN 1946), therefore the net seed accumulation rate becomes regressive in the course of time, leading to a curvilinear relationship between seed bank density and basal area of tree (Fig. 4).

Different natural and human disturbances also facilitate the spreading of the black locust. The lifetime of Austrian pine is up to hundreds of years in its natural area, but in the Hungarian stands the decline of trees begins much earlier, mainly as a consequence of the suboptimal environmental conditions. The resistance of 40–50 year old pines declines quickly in shallow-soiled habitats and thus rapid destruction can be initiated by a permanent drought or infested pathogen fungi (KOLTAY 1990, 1997). These factors promote the penetration and spread of invaders in Austrian pine stands. The black locust was naturalized in East Asia in the second half of the 19th century; nowadays the invader regularly appears in indigenous *Pinus thunbergii* stands and broad-leaved forests, chiefly in the vicinity of severely disturbed urban regions (MAEKAWA and NAKAGOSHI 1997, LEE et al. 2004, SONG et al. 2005, TANIGUCHI et al. 2007, MORIMOTO et al. 2010). Frequent fire events prove that pine plantations in Hungary are highly susceptible to forest fires in view of their accumulated resinous needle litter (CSERESNYÉS and CSONTOS 2004; CSERESNYÉS et al. 2006, 2011). The heat effect is capable of breaking the physical dormancy of black locust seeds (MARJAI 1995, MASAKA and YAMADA 2009), thus initiating a quick colonization by the invader in the burnt area (AULD and DENHAM 2006, JUNG et al. 2009).

Within the scope of sustainable forest management, the replacement of Austrian pine stands by natural vegetation types began in nature reserves of Hungary in the past few decades. After coniferous stands are clear cut, the recolonisation of indigenous species from nearby native forests and grasslands is generally slow, even if the appropriate propagulum sources are available in its surroundings (MATLACK 1994, RYDGREN et al. 1998, TAMÁS

2001). Under these circumstances regeneration of the native vegetation primarily happens from the locally available soil seed bank. In soil of *Pinus nigra* stands the seed bank of species of the native flora (that existed prior to afforestation by pine) became impoverished; its density and richness steadily declined because of their short-term persistent seeds (CSONTOS et al. 1996, AUGUSTO et al. 2001). Consequently the diverse natural seed bank could be replaced gradually by the high density persistent seed bank of non-indigenous species including the black locust. The germination of *R. pseudoacacia* seeds can be continued for decades, strongly inhibiting or preventing completely the restoration of the native flora after the removal of an Austrian pine stand. The effective root sprouting also promotes the rapid spread of locust trees, especially after local disturbances, and thus finally it can happen that an area just cleared from the non-native pine is occupied by another non-native tree, the locust, which would be a rather unacceptable result. Since both the density of the seed bank and the ratio of seeds detected in the lower soil layer increases with the diameter at breast height of the tree, the threat of a spontaneous black locust stand establishment is particularly to be expected in areas already sporadically occupied by old black locust trees. The black locust is responsible for irreversible changes in the physicochemical and biological soil properties by N-fixation, leading to the formation of species-poor nitrophilous weed associations in the herb layer (TOBISCH et al. 2003). Therefore, establishment of the black locust should be prevented by careful planning of the replacement of Austrian pine stands by native tree species.

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