# Population dynamics and growth of perch in a small, humic lake over a 20 -year period - importance of abiotic and biotic factors 

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Population dynamics and growth of perch (Perca fluviatilis) were studied over a 20-year period in a small, humic, boreal lake (Valkea-Kotinen). The annual estimated total population size of perch ( $>8 \mathrm{~cm}$, total length) varied from 2700-13 400 and the biomass from $16.1-44.2 \mathrm{~kg} \mathrm{ha}^{-1}$. Strong year-classes were born at four-year intervals between 1991 and 2003. A decreasing trend was recorded in the growth of 1 - and 2 -year-old perch, whereas a slight increase appeared in the yearly growth of 4 - and 5 -year-old fish. Our view is that the decrease in the early growth was caused by negative effects of increased organic carbon loads on water colour and hence on light conditions and the general productivity of the plankton community. These changes outweighed the expected positive effect of increasing temperature for young perch but not for the older fish.

## Introduction

The European perch (Perca fluviatilis) is a widely distributed fish species in boreal waters. Apart from the subarctic lakes of Lapland, it is the most common species in Finnish lakes (Tonn et al. 1990, Tammi et al. 2003, Rask et al. 2010). The presence of perch in almost all the lakes of southern and central Finland is related to its wide environmental tolerance: the life cycle of perch can be completed in small acidic ponds (Nyberg et al. 1995, Tammi et al. 2004, Posch et al. 2012) as well as in larger lakes of widely
different trophic state (Olin et al. 2002, Tammi et al. 2003). The wide environmental tolerance is complemented by its ecological flexibility; perch is a generalist carnivore feeder, usually shifting during its life span from zooplankton to zoobenthos and later to fish (Allen 1935, Persson 1994, Horppila et al. 2000). As it is a visual feeder, light conditions are of importance, low light causing decreased feeding efficiency (Bergman 1988, Estlander et al. 2012) and dominance of roach (Rutilus rutilus) over perch in competition for zooplankton food in the turbid water of eutrophic lakes (Persson 1983, Olin et al. 2002)
or in highly humic lakes (Estlander et al. 2010, Olin et al. 2010).

Significant environmental trends were recorded in a small, humic Lake Valkea-Kotinen during 1990-2009. Decreasing acidifying deposition resulted in lower $\mathrm{SO}_{4}$ concentrations and in associated increases in pH (from 5.0 to 5.5 ) and alkalinity (from 5 to $20 \mu \mathrm{eq}{ }^{-1}$ ) of water (Vuorenmaa et al. 2014). The effects of climate change, predicted to increase the temperature and shorten the ice-covered period of the lake (Saloranta et al. 2009), have already been recorded (Lehtovaara et al. 2014, Jylhä et al. 2014). The increase in organic carbon load, apparently due to the afore-mentioned factors (Monteith et al. 2007, Futter et al. 2009, Arvola et al. 2010), caused brownification of the water, from 100 to close to $200 \mathrm{mg} \mathrm{Pt} \mathrm{l}{ }^{-1}$. Such brownification may affect energy and carbon sources and processes within the food webs (Salonen et al. 1992, Kankaala et al. 2006, Jones et al. 2008) and result in a higher proportion of bacterial production via the microbial loop as compared with algal production (Ask et al. 2009, Karlsson et al. 2009). This may in turn affect food availability to higher trophic levels, i.e., zooplankton (Brett et al. 2009) and fish (Estlander et al. 2012).

In lakes recovering from acidification, increased reproductive success of perch is expected to lead to increasing population density, and decreasing growth as well as mean size of fish due to food competition (Nyberg et al. 1995). However, Ohlberger et al. (2011) predicted that increasing temperature, coupled with increasing intraspecific competition, may also result in dominance of younger and smaller perch, which was recorded in some Danish lakes (Jeppesen et al. 2012). On the other hand, perch is a warm-water species, and hence is predicted to benefit from warmer waters being a consequence of climatic change (Lappalainen and Lehtonen 1997, Jeppesen et al. 2012).

In the present study, we related the changes in perch population dynamics and growth to the environmental trends recorded in the lake during the 20-year study period to compare the impact of each stressor on perch. Both environmental and biological factors were included to evaluate their relative importance.

## Material and methods

Small (0.042 km²), mesotrophic, shallow (maximum depth 6.5 m , mean depth 2.5 m ), brownwater Lake Valkea-Kotinen is located in a small headwater catchment $\left(0.22 \mathrm{~km}^{2}\right)$ in a remote protected forest area in southern Finland. It is affected by pollution from airborne sources only (for details see Ukonmaanaho et al. 1998, Ruoho-Airola et al. 2014). During the growing season, steep thermal and oxygen stratification is typical for the lake, resulting in ca. 2-m thick warm and oxygenated epilimnion, and a cold and anoxic hypolimnion (Forsius et al. 2010) restricting the suitable habitat for perch in the lake. The littoral habitat is narrow due to poor light penetration and steep shores, and consists mainly of floating leaved vegetation (Nuphar lutea) and aquatic mosses.

Perch and pike (Esox lucius) are the only fish species in the lake. The size and structure of the perch population have been monitored since 1991 (Rask et al. 1998). Annual population estimates of perch were obtained from two-week continuous marking and recapturing in May without fish removal (Schnabel estimate, Krebs 1989). The fish were caught by wire traps with a $1 \mathrm{~cm}^{2}$ mesh retaining perch $\geq 8 \mathrm{~cm}$ long, which corresponds to $\geq 2$ years of age, and targeting the spawning population. Altogether 34731 perch were captured during the 20 -year study period. The fish were measured to the nearest 1 cm (total length) for length distribution evaluation, after which they were fin-clipped and released. Pelvic fins were clipped, one of the two fins each year, to ensure correct interpretation of cuts from consecutive years. Marking and recapturing was done during spawning, some $2-3$ weeks after ice-out. The mean number of marked fish varied from year to year between 950 and 4100 resulting in recaptures of 111-811 marked perch. The aim was for the $95 \%$ confidence limits to be less than $\pm 20 \%$ of the population estimate, and this was achieved in most years. A population estimate for 1990 (to be used in multivariate analysis) was derived from the 1991 value by using an estimate for annual mortality rate ( $M=0.6$; Thorpe 1977). Annual perch biomass was calculated as follows: first, based on the size distribution of the fish caught in the test fishing, the number of fish



Fig. 1. (A) Mean water temperature ( 1 m depth, June-September) and oxygen concentration ( 2 m depth, week 33) and (B) mean alkalinity and water colour (1 m, June-September) in Lake ValkeaKotinen during 19902009.
belonging to each size class in the whole population was estimated; second, biomass of each size class was estimated using the mean weight of perch from each length group (cm), calculated from the perch sampled for age and growth determination ( $n=1266$ ); these results were then summed to give the total biomass. Corresponding data for pike are not available.

Samples for age determination (50-100 perch per year) were taken from the last recapture catch. In some years additional samples were obtained from late summer gill-net sampling (NORDIC survey nets, CEN 2005). Each fish was measured to the nearest mm (total length) and weighed to the nearest g , and opercular bones were used to determine age and to back-calculate growth according to the Monastyrsky procedure (Raitaniemi et al. 1988). Relative year-class strength was determined using the Svärdson-Kempe method (Kempe 1962).

Water-quality (Fig. 1) and season-length (the number of days with water temperature at 1 m depth $>15{ }^{\circ} \mathrm{C}$ ) data were obtained from the database of the Lammi Biological Station (University of Helsinki). Water-quality data are presented in more detail by Vuorenmaa et al. (2014). Data on primary production of phytoplankton and chlorophyll $a$ (Fig. 2, see Arvola et
al. 2014) and data on main zooplankton groups and Chaoborus larvae (Fig. 2, see Lehtovaara et al. 2014) are from the plankton database of the Lammi Biological Station (University of Helsinki).

A non-parametric Mann-Kendall test (Z-statistics; see Hipel and McLeod 2005) was used for long-term monotonic (i.e. increasing or decreasing) trend analyses of the perch population and growth parameters. No assumption of normality is required by the test, hence the non-transformed original data were used. To analyse the joint effects of abiotic and biotic factors (later called environmental variables) on the perch population and its growth parameters, expressed as annual length increment, a redundancy analysis (RDA, Canoco 4.51 (ter Braak \& Šmilauer 2002) was applied since the gradient lengths of biotic variables were rather short, 3-4 SD units. All variables were standardized as follows: $\left(x_{i}-x\right) \mathrm{SD}^{-1}$. A Monte-Carlo permutation test (1000 permutations) was used to test the significances of the single variables and the RDA axes. As the measurements of a given variable were related to each other (sampled in successive years from the same lake), a time series option was used as a permutation restriction. The perch variables in the RDA analysis included seven variables that

Fig. 2. (A) Primary production of phytoplankton (PP, 0-1 m, June-August) and mean chlorophyll a concentration (Chl a, 0-1 m, June-August) and (B) mean density of cladocerans, copepods and Chaoborus larvae (0-5 m, June-August) in Lake Valkea-Kotinen during 1990-2009.


were assumed to respond to environmental variables on a yearly basis: size of population, relative year-class strength and the length increments of $1-5$-year-old perch in each year. The environmental variables were first selected in order to avoid high ( $>0.8$ ) between-variable correlation causing multicollinearity, and secondly based on maximum extra fit. The eight environmental variables entered in the RDA analysis included five water quality parameters (alkalinity, water colour, oxygen concentration at 2 m depth, total phosphorus ( $\mathrm{P}_{\text {tot }}$ ) and water temperature at 1 m depth), season length, and two biological parameters (primary production of phytoplankton and cladoceran zooplankton density).

## Results

## The perch population

The size of the perch population (individuals of $>8 \mathrm{~cm}$ in total length) varied between 2400 and 13400 (570-3190 perch ha ${ }^{-1}$ ) and the biomass between 16.1 and $44.2 \mathrm{~kg} \mathrm{ha}^{-1}$ with no significant trends during the 20-year monitoring period (Fig. 3 and Table 1). The population size peaked
in 1998 and the biomass in 1997, both due to the recruitment into the spawning population of the 1995 year-class, the strongest recorded during the study period.

During the first half of the study period, a four-year interval was apparent in the occurrence of strong year classes, which were born in 1991, 1995, 1999 and 2003 (Fig. 4). The year class 1991 formed the most abundant length group during 1993-1996 (Fig. 5), until the strongest year class 1995 appeared as numerous perch of size 8 and 9 cm in 1997. In 2001, 9 and 10 cm perch of the strong year class 1999 appeared as most abundant. During the latter part of the monitoring period there was less variability in the strength of year classes than in the 1990s but no significant trend was recorded (Table 1). Correspondingly, after 2003 the variation in size structure of the perch population stabilised with a dominance of small individuals ( $\leq 10 \mathrm{~cm}$ in total length; Fig. 5). This resulted in a significant negative trend in the mean length and weight of perch (Table 1).

## The growth of perch

The growth of perch in Lake Valkea-Kotinen



Fig. 3. Estimates ( $\pm 95 \%$ confidence limits) of perch biomass (circles) and population size (bars) for Lake Valkea-Kotinen during 1991-2009.

Fig. 4. The relative yearclass strength of perch in Lake Valkea-Kotinen during 1990-2009.
was fairly slow as the total length of 15 cm has usually not been exceeded before the fish reached the age of five or six years (Fig. 6). During the study period, a significant decreasing trend in perch growth was recorded in their first
summer and slowing growth also during the 2 nd and 3 rd summers. In contrast, in 4- and 5 -yearold perch the trend was slightly positive (Table 1 and Fig. 7), although the back-calculated length of 5 -year-old perch in 1990 was close to the

Table 1. Mann-Kendall statistics for trends in perch population variables and growth in Lake Valkea-Kotinen during 1990-2009. Significance of $Z$ values are as follows: $p<0.1$ when $|Z|>1.563, p<0.05$ when $|Z|>1.862, p<0.01$ when $|Z|>2.447$.

| Perch variable | Range | $Z$ | $p$ |
| :--- | :---: | :---: | :---: |
| Population size | $2400-13400$ | 0.91 | ns |
| Biomass of perch $\left(\mathrm{kg} \mathrm{ha}^{-1}\right)$ | $16.1-44.2$ | -0.616 | ns |
| Density of perch $>15 \mathrm{~cm}$ (indiv. $\mathrm{ha}^{-1}$ ) | $82-487$ | -1.19 | ns |
| Relative year-class strength | $8-355$ | -0.681 | ns |
| Mean length of perch $(\mathrm{mm})$ | $101-155$ | -2.239 | $<0.05$ |
| Mean weight of perch $(\mathrm{g})$ | $13-50$ | -2.029 | $<0.05$ |
| 1st year growth $(\mathrm{mm})$ | $48.8-63.3$ | -3.439 | $<0.01$ |
| 2nd year growth $(\mathrm{mm})$ | $24.5-45.8$ | -1.752 | $<0.1$ |
| 3rd year growth $(\mathrm{mm})$ | $16.7-29.6$ | -1.849 | $<0.1$ |
| 4th year growth $(\mathrm{mm})$ | $14.5-26.9$ | 1.428 | ns |
| 5th year growth $(\mathrm{mm})$ | $8.5-24.5$ | 1.622 | $<0.1$ |



Fig. 5. The length frequency distribution of perch in Lake Valkea-Kotinen during 1991-2009.



Fig. 6. Back calculated growth of perch from yearclasses 1985-2007 ( $n=$ 1266, 5-245 per yearclass).

Fig. 7. The trends in annual length increment of perch at age $1-5$ years in Lake Valkea-Kotinen during 1990-2009.
records of the last research years indicating that no actual trend existed. Nevertheless, the length at age 5 decreased until the last half of the 1990s, showing slowest growth during the years of highest population size and thus suggesting density-dependent decrease of growth in those years.

## Effects of environmental and biological factors

The RDA analysis of the perch and environmental variables suggested that environmental variables had a rather weak effect on the yearclass strength or population size, but a relatively strong and varying effect on the growth in different age classes (Table 2 and Fig. 8). The relationships between perch variables, and between perch variables and environmental variables were statistically significant ( $p=0.022$ in
both cases; Table 2). However when the environmental variables were considered separately in the RDA analysis, only oxygen concentration had a significant effect ( $p=0.033$ ) on the matrix of perch variables. The first RDA axis explained $26 \%$ of the variation in the perch variables and $51 \%$ of the variation in the relationships between perch variables and environment variables.

On the first RDA axis, water colour had a strong positive score, and primary production and oxygen concentration had high negative scores indicating the gradient of low productivity, limited oxygenated layer and poor light conditions. Growth in the youngest age groups was strongly negatively affected by these variables; this indicates that during the first years perch growth was positively dependent on primary production and oxygen conditions and negatively dependent on water colour. In contrast, in older perch (age 4 and 5) this relation was positive which suggests a lower dependence
of larger fish on pelagic food resources and on light conditions. Furthermore, the relative abundance of small perch increased during the study period providing a better prey-fish resource for large perch. Year class strength and population size were only weakly related to the first RDA axis. The brownification trend of the lake with decreasing oxygen concentration and primary production can also be seen in the RDA analysis, as the later years mostly have higher scores in the first RDA axis than the earlier ones.

The second RDA axis explained $16 \%$ of the variation in the perch variables and $24 \%$ of the variation in the perch and environment relationship (Table 2). On this axis, water colour had a high negative score, and $P_{\text {tot }}$, water temperature and cladoceran density positive scores indicating the gradient of good light, high food resources and warm water. All the perch variables except population size were positively related on this axis, especially growth of $3-5$-year perch. This suggests that growth in older age groups is more dependent on good light conditions and warm water than that in young age groups.

## Discussion

## Perch population

The perch population and biomass estimates from Lake Valkea-Kotinen are in line with those from other studies of similar lakes in the Evo area (Rask 1983, Lappalainen et al. 1988, Horppila et al. 2010). The 20 -year data from Lake Valkea-Kotinen revealed that the size of the perch population and the occurrence of strong year classes were only weakly associated with the environmental and biological factors considered, and were regulated more through intraspecific processes. This was seen in the weak associations of population density and year-class strength with the axes of the RDA analysis. The lake had never acidified to critical levels for perch reproduction in humic waters (see Henriksen et al. 1989), as shown by the occurrence of strong year-classes of perch during the "acidification years" of the 1990s, before the onset of pronounced chemical recovery of the lake around the year 2000 (Vuorenmaa et al. 2014).

Perch is a warm-water species that is predicted to benefit from the warming of waters caused by climate change (Lappalainen and Lehtonen 1997, Jeppesen et al. 2012). However, in the case of Lake Valkea-Kotinen, this was seen neither in the occurrence of strong year-classes nor in the positive responses of early growth of perch (see Jeppesen et al. 2012).

In the first half of the monitoring period, the fluctuation of the perch population size followed the occurrence of strong year-classes that were born at a 4 -year intervals until the early 2000s. This kind of pattern has often been recorded in small perch-dominated lakes where larger cannibalistic perch can prevent the recruitment of YOY fish until the density of large individuals is small enough to enable the development of a new strong year-class (Alm 1952, Persson et al. 2000). In the latter part of the monitoring period, since the early years of the 2000s, no really strong or weak year-classes occurred and the population was dominated by small individuals. The decreased average size of perch is in line

Table 2. Results of the RDA analysis of perch variables, and of environmental variables during 20 years of monitoring in Lake Valkea-Kotinen. Axes 1 and 2 are the first two RDA axes.

| Variable | Axis 1 | Axis 2 | All |
| :---: | :---: | :---: | :---: |
| Perch variables |  |  |  |
| 5 yr . length increment | 0.5975 | 0.3818 |  |
| 4 yr . length increment | 0.5737 | 0.5322 |  |
| Size of population | -0.1583 | -0.2489 |  |
| Year class strength | -0.1941 | 0.3051 |  |
| 3 yr . length increment | -0.2003 | 0.5918 |  |
| 2 yr . length increment | -0.6430 | 0.2875 |  |
| 1 yr . length increment | -0.7747 | 0.2713 |  |
| Environmental variables |  |  |  |
| Water colour | 0.5106 | -0.5987 |  |
| Temperature | 0.2018 | 0.3433 |  |
| Alkalinity | 0.0791 | 0.1237 |  |
| Number of days when water $\mathrm{T}>15^{\circ} \mathrm{C}$ | -0.0186 | -0.2287 |  |
| Total phosphorus | -0.1150 | 0.5547 |  |
| Cladoceran density | -0.1261 | 0.3472 |  |
| Primary production | -0.7632 | 0.2868 |  |
| Oxygen concentration | -0.7863 | 0.1325 |  |
| $F$ | 3.8150 |  | 2.008 |
| $p$ | 0.0220 |  | 0.022 |
| Variance explained (\%) |  |  |  |
| Perch variables | 25.75 | 15.58 | 54.38 |
| Perch-environment relation | 51.29 | 23.73 | 91.64 |



Fig. 8. Biplot of the redundancy analysis for perch variables and environmental variables. Pop. size $=$ size of perch population, YCS = relative year-classstrength of perch, 1 yr . LI = first year length increment etc., Colour $=$ water colour, Temp = water temperature, $\mathrm{P}_{\text {tot }}=$ total phosphorus, $\mathrm{O}_{2}=$ oxygen concentration, Days T>15 ${ }^{\circ} \mathrm{C}$ $=$ number of days with water temperature ( 1 m ) $>15^{\circ} \mathrm{C}, \mathrm{PP}=$ primary production of phytoplankton, Clad = cladoceran density. Scores for the sampling years (1990-2009) are also shown.
with the observations from Denmark (Jeppesen et al. 2012), related to improved recruitment of fish due to higher spring temperatures and increasing survival of fish during winter due to shorter icecover period. Moreover, it has been predicted, by using a physiologically structured consumerresource model that increasing water temperature, connected with enhanced intraspecific competition, may result in dominance of younger and smaller perch (Ohlberger et al. 2011).

One possible explanation for decreased mean size of perch might be reduced predation pressure from pike, although due to the lack of long-term pike data this cannot be ascertained. However, other studies from similar lakes have suggested that the intensity of pike predation can affect the perch population structure (Olin et al. 2010). Predation by pike decreases the perch population density and the intraspecific food competition among perch thus contributing to faster growth and larger mean size of perch (Rask 1983, Persson
et al. 1996, Olin et al. 2010). On the other hand, dark coloured lake water may reduce the feeding efficiency of pike (Horppila et al. 2010).

## Growth of perch

The main environmental factor affecting the early growth of perch in Lake Valkea-Kotinen appeared to be the organic carbon load, resulting in darker water, and decreased light. Overall, the limiting effect of worsening light conditions would outweigh the expected positive effect of increasing temperature on the growth of young perch (Jeppesen et al. 2012). It is possible that the decrease in the early growth of perch was - at least partly - a direct response to the deterioration of the light conditions as perch is a visually-oriented fish species and active especially at dawn and dusk (Helfman 1979, Rask 1986). Recent field and experimental studies
showed the harmful effects of poor light conditions on the ability of perch to compete for food with roach and on the feeding efficiency of perch that may result in slower growth (Estlander et al. 2010, 2012).

However, a more important factor causing the decreased early growth of perch is a general decrease in the biological production of the lake, also related to the increased water colour. In Lake Valkea-Kotinen, there was a significant decreasing trend in the primary production (Arvola et al. 2014) and also in the densities of important food items for small perch, such as cladoceran zooplankton and Chaoborus larvae (Lehtovaara et al. 2014). A significant positive association between the primary production of phytoplankton and the early growth of perch, as indicated in the RDA analysis, suggests that the decrease in general productivity of the lake affected the perch growth via zooplankton as zooplankton is the primary food source of perch fry: first ciliates, rotifers and copepod naupli (Siefert 1972, Guma’a 1978, Zingel et al. 2012), followed by various crustaceans. The RDA analysis also emphasized the importance of oxygen, suggesting that the decreasing volume of the oxygenated habitat may affect the early growth of perch. The reduction of water volume suitable for perch can lead to increased intraspecific competition. At the same time, perch are forced to use the uppermost and warmer water layer which increases metabolism and the need for energy. The decreased early growth of perch may also be linked to the changes in the population structure, as it has been suggested that slow growth of plankton-feeding perch, especially slow secondsummer growth, can lead to delay or inhibition of the later ontogenetic shift from benthivory to piscivory in perch populations (Heibo et al. 2005, Estlander et al. 2012).

The main reason for the slow growth of perch in Lake Valkea-Kotinen was presumably intraspecific food competition, as there were no competing fish species present. Although the growth was not as slow as in some other humic lakes without pike predation (Rask 1983, Lappalainen et al. 1988), it was far below the growth of perch recorded in the more productive and diverse ecosystems of larger lakes (Sarvala and Helminen 1996, Ruuhijärvi et al. 2010) or in
oligotrophic lakes with low population density (Raitaniemi et al. 1988). The slight positive trend in the growth of larger perch, those of age groups of 4 and 5 years, indicates their different trophic position and better food resources or higher feeding efficiency as compared with the small planktivorous perch. The positive relation of older perch (age 4 and 5) to the first RDA axis suggests a lower dependence of larger fish on pelagic food resources and on light conditions, as they mainly consume littoral macroinvertebrates and small fish (Rask et al. 1998, Estlander et al. 2010). It appears that reduced light conditions have not affected the growth of older perch, although no data are available concerning possible changes in the benthic or littoral production of the lake during the study period. Additionally, as the average size of perch decreased during the monitoring period, there is more intrinsic prey of suitable size for older perch which might partly explain their improved growth.

In conclusion, the results of our 20-year study indicate that the population dynamics of perch is mainly regulated by intraspecific processes, whereas the changes in growth were more closely related to the trends in environmental factors. The decrease in the early growth of perch was most clearly related to the increase in water colour and subsequent decrease in general productivity of the lake, including decrease in food resource production for small perch and subsequent intraspecific food competition. These changes outweighed the expected positive effect of increasing temperature for young but not for the older perch.

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## References

Allen K.R. 1935. The food and migration of perch (Perca fluviatilis) in Windermere. J. Anim. Ecol. 4: 264-273.
Alm G. 1952. Year class fluctuations and span of life of perch. Rep.Inst. Freshw. Res. Drottningholm 33: 17-38. Arvola L., Rask M., Ruuhijärvi J., Tulonen T., Vuorenmaa J.,

Ruoho-Airola T. \& Tulonen J. 2010. Long-term patterns in pH and colour in small acidic boreal lakes of varying hydrological and landscape settings. Biogeochemistry 101: 269-279.
*Arvola L., Salonen K., Keskitalo J., Tulonen T., Järvinen M. \& Huotari J. 2014. Plankton metabolism and sedimentation in a small boreal lake - a long-term perspective. Boreal Env. Res. 19 (suppl. A): 83-96.
Ask J., Karlsson J., Persson L., Ask P., Bystrom P. \& Jansson M. 2009. Whole-lake estimates of carbon flux through algae and bacteria in benthic and pelagic habitats of Clearwater lakes. Ecology 90: 1923-1932.
Bergman E. 1988. Foraging abilities and niche breadths of two percids, Perca fluviatilis and Gymnocephalus cernua, under different environmental conditions. J. Anim. Ecol. 57: 443-453.
Brett M.T., Kainz M.J., Taipale S.J. \& Seshan H. 2009. Phytoplankton, not allochthonous carbon, sustains herbivorous zooplankton production. Proceedings of the National Academy of Science 106: 21197-21201.
CEN 2005. Water quality - sampling fish with multi-mesh gillnets. European Standard EN 14757:2005.
Estlander S., Nurminen L., Olin M., Vinni M., Immonen S., Rask M., Ruuhijärvi J., Horppila J. \& Lehtonen H. 2010. Diet shift and food selection of (Perca fluviatilis) and roach (Rutilus rutilus) in humic lakes of varying water colour. J. Fish Biol. 77: 241-256.
Estlander S., Horppila J., Olin M., Vinni M., Lehtonen H., Rask M. \& Nurminen L. 2012. Troubled by the humics - effects of water colour and interspecific competition on the feeding efficiency of planktivorous perch. Boreal Env. Res. 17: 305-312.
Forsius M., Saloranta T., Arvola L., Salo S., Verta M., AlaOpas P., Rask M. \& Vuorenmaa J. 2010. Physical and chemical consequences of artificially deepened thermocline in a small humic lake - a paired whole-lake climate change experiment. Hydrol. Earth Syst. Sci. 14: 2629-2642.
Futter M., Forsius M., Holmberg M. \& Starr M. 2009. A long-term simulation of the effects of acidic deposition and climate change on surface water dissolved organic carbon concentrations in a boreal catchment. Hydrology Research 40: 291-305.
Guma'a S.A. 1978. The food and feeding habits of young perch, Perca fluviatilis, in Windermere. Freshwat. Biol. 8: 177-187.
Heibo E., Magnhagen C. \& Vollestad L.A. 2005. Latitudinal variation in life-history traits in Eurasian perch. Ecology 86: 3377-3386.
Helfman G.S. 1979. Twilight activities of yellow perch Perca flavescens. J. Fish. Res. Bd. Canada 36: 173-179.
Henriksen A., Lien L., Rosseland B.O., Traaen T.S. \& Sevaldrud I.S. 1989. Lake acidification in Norway: present and predicted fish status. Ambio 18: 314-321.
Hipel K.W. \& McLeod A.I. 2005. Time series modelling of water resources and environmental systems. Avalable at http://www.stats.uwo.ca/faculty/aim/1994Book/.
Horppila J., Ruuhijärvi J., Rask M., Karppinen C., Nyberg K. \& Olin M. 2000. Seasonal changes in the food composition and relative abundance of perch and roach - a
comparison between littoral and pelagial zones of a large lake. J. Fish Biol. 56: 51-72.
Horppila J., Olin M., Vinni M., Estlander S., Nurminen L., Rask M., Ruuhijärvi J. \& Lehtonen H. 2010. Perch production in forest lakes: the contribution of abiotic and biotic factors. Ecology of Freshwater Fish 19: 257-266.
Jeppesen E., Mehner T., Winfield I.J., Kangur K., Sarvala J., Gerdeaux D., Rask M., Malmquist H.J., Holmgren K., Volta P., Romo S., Eckmann R., Sandström A., Blanco S., Kangur A., Ragnarsson Stabo H., Meerhoff M., Ventelä A.-M., Søndergaard M. \& Lauridsen T.L. 2012. Impacts of climate warming on lake fish assemblages: evidence from 24 European long-term data series. Hydrobiologia 694: 1-39.
Jones R.I., Carter C.E., Kelly A., Ward S., Kelly D.J. \& Grey J. 2008. Widespread contribution of methane-cycle bacteria to the diets of lake profundal chironomid larvae. Ecology 89: 857-864.
*Jylhä K., Laapas M., Ruosteenoja K., Arvola L., Drebs A., Kersalo J., Saku S., Gregow H., Hannula H.-R. \& Pirinen P. 2014. Climate variability and trends in the Valkea-Kotinen region, southern Finland: comparisons between the past, current and projected climates. Boreal Env. Res. 19 (suppl. A): 4-30.
Kankaala P., Taipale S., Grey J., Sonninen E., Arvola L. \& Jones R. 2006. Experimental $\delta^{13} \mathrm{C}$ evidence for a contribution of methane to pelagic food webs in lakes. Limnol. Oceanogr. 51: 2821-2827.
Karlsson J., Byström P., Ask J., Ask P., Persson L. \& Jansson M. 2009. Light limitation of nutrient-poor lake ecosystems. Nature 460: 506-509.
Kempe O. 1962. The growth of roach (Leuciscus rutilus L.) in some Swedish lakes. Rep. Inst.Freshw. Res. Drottningholm 44: 43-104.
Krebs C.J. 1989. Ecological methodology. Harper \& Row, New York.
Lappalainen A., Rask M. \& Vuorinen P. 1988. Acidification affects the perch, Perca fluviatilis L., populations in small lakes of southern Finland. Env. Biol. Fish. 21: 231-239.
Lappalainen J. \& Lehtonen H. 1997. Temperature habitats for freshwater fishes in a warming climate. Boreal Env. Res. 2: 69-84.
*Lehtovaara A., Arvola L., Keskitalo J., Olin M., Rask M., Salonen K., Sarvala J., Tulonen T. \& Vuorenmaa J. 2014. Responses of zooplankton to long-term environmental changes in a small boreal lake. Boreal Env. Res. 19 (suppl. A): 97-111.
Monteith D.T., Stoddard J.L., Evans C.D., de Wit H.A., Forsius M., Høgåsen T., Wilander A., Skjelkvåle B.L., Jeffries D.S., Vuorenmaa J., Keller B., Kopácek J. \& Vesely J. 2007. Dissolved organic carbon trends resulting from changes in atmospheric deposition chemistry. Nature 450: 537-540.
Nyberg K., Raitaniemi J., Rask M., Mannio J. \& Vuorenmaa J. 1995. What can perch population data tell us about the acidification history of a lake? Water Air Soil Pollut. 85: 395-400.
Ohlberger J., Edeline E., Vollestad L.A., Stenseth N.C. \& Claessen D. 2011. Temperature-driven regime shifts in
the dynamics of size-structured populations. American Naturalist 177: 211-223.
Olin M., Rask M., Ruuhijärvi J., Kurkilahti M., Ala-Opas P. \& Ylönen O. 2002. Fish community structure in mesoand eutrophic lakes of southern Finland: the relative abundances of percids and cyprinids along a trophic gradient. J. Fish Biol. 60: 593-612.
Olin M., Vinni M., Lehtonen H., Rask M., Ruuhijärvi J., Saulamo K. \& Ala-Opas P. 2010. Environmental factors regulate the effects of roach Rutilus rutilus and pike Esox lucius on perch Perca fluviatilis populations in small boreal forest lakes. J. Fish Biol. 76: 1277-1293.
Persson L. 1983. Effects of intra and interspecific competition on dynamics and size structure in a roach (Rutilus rutilus) and a perch (Perca fluviatilis) population. Oikos 41: 126-132.
Persson L. 1994 Natural shifts in the structure of fish communities: mechanisms and constraints on perturbation sustenance. In: Cowx I.G. (ed.), Rehabilitation of freshwater fisheries, Fishing News Books, Hartnoll, Cornwall, pp. 421-434.
Persson L., Andersson J., Wahlström E. \& Eklöv P. 1996. Size-specific interactions in lake systems: predator gape limitation and prey growth rate and mortality. Ecology 77: 900-911.
Persson L., Byström P. \& Wahlström E. 2000. Cannibalism and competition in Eurasian perch: population dynamics of an ontogenetic omnivore. Ecology 81: 1058-1071.
Posch M., Aherne J., Forsius M. \& Rask M. 2012. Past, present and future exceedance of critical loads of acidity for surface waters in Finland. Environmental Science \& Technology 46: 4507-4514.
Raitaniemi J., Rask M. \& Vuorinen P.J. 1988. The growth of perch, Perca fluviatilis L., in small Finnish lakes at different stages of acidification. Ann. Zool. Fennici 25: 209-219.
Rask M. 1983. Differences in growth of perch (Perca fluviatilis L.) in two small forest lakes. Hydrobiologia 101: 139-144.
Rask M. 1986. The diet and diel feeding activity of perch, Perca fluviatilis L., in a small lake in southern Finland. Ann. Zool. Fennici 23: 49-56.
Rask M., Holopainen A.-L., Karusalmi A., Niinioja R., Tammi J., Arvola L., Keskitalo, J. Blomqvist I., Heinimaa S., Karppinen C., Salonen K. \& Sarvala J. 1998. An introduction to the limnology of Finnish Integrated Monitoring lakes. Boreal Env. Res. 3: 263-274.
Rask M., Olin M. \& Ruuhijärvi J. 2010. Fish-based assessment of ecological status of Finnish lakes loaded by diffuse nutrient pollution from agriculture. Fisheries Management and Ecology 17: 126-133.
*Ruoho-Airola T., Hatakka T., Kyllönen K., Makkonen U. \& Porvari P. 2014. Temporal trends in the bulk deposition and atmospheric concentration of acidifying compounds and trace elements in the Finnish Integrated Monitoring
catchment Valkea-Kotinen during 1988-2011. Boreal Env. Res. 19 (suppl. A): 31-46.
Ruuhijärvi J., Rask M., Vesala S., Westermark A., Olin M., Keskitalo J. \& Lehtovaara A. 2010. Recovery of the fish community and changes in the lower trophic levels in a eutrophic lake after a winterkill of fish. Hydrobiologia 646: 145-158.
Salonen K., Arvola L., Tulonen T., Hammar T., Metsälä T.-R. \& Kankaala P. 1992. Planktonic food chains of a highly humic lake. I. A mesocosm experiment during the spring primary production maximum. Hydrobiologia 229: 125-142.
Saloranta T., Forsius M., Järvinen M. \& Arvola L. 2009. Impacts of projected climate change on thermodynamics of a shallow and deep lake in Finland: model simulations and Bayesian uncertainty analysis. Hydrology Research 40: 234-248.
Sarvala J. \& Helminen H. 1996. Year-class fluctuations of perch (Perca fluviatilis) in Lake Pyhäjärvi, Southwest Finland. Ann. Zool. Fennici 33: 389-396.
Siefert R.E. 1972. First food of larveal yellow perch, white sucker, bluegill, emerald shiner and rainbow smelt. Trans. Am. Fish. Soc. 101: 219-225.
Tammi J., Appelberg M., Hesthagen T., Beier U., Lappalainen A. \& Rask M. 2003. Fish status survey of Nordic lakes: effects of acidification, eutrophication and stocking activity on present fish species composition. Ambio 32: 98-105.
Tammi J., Rask M., Vuorenmaa J., Lappalainen A. \& Vesala S. 2004. Population responses of perch (Perca fluviatilis) and roach (Rutilus rutilus) to recovery from acidification in small Finnish lakes. Hydrobiologia 528: 107-122.
ter Braak C.J.F. \& Šmilauer P. 2002. CANOCO Reference Manual and CanoDraw for Windows User's Guide: Software for Canonical Community Ordination (version 4.5). Microcomputer Power, Ithaca, NY.

Thorpe J. 1977. Synopsis of biological data on the perch Perca fluviatilis Linnaeus, 1758 and Perca flavescens Mitchill, 1814. FAO Fish. Synop. 113: 1-138.
Tonn W.M., Magnuson J.J., Rask M. \& Toivonen J. 1990. Intercontinental comparison of small-lake fish assemblages: the balance between local and regional processes. American Naturalist 136: 345-375.
Ukonmaanaho L., Starr M., Hirvi J.-P., Kokko A., Lahermo P., Mannio J., Paukola T., Ruoho-Airola T. \& Tanskanen H. 1998. Heavy metal concentrations in various aqueous and biotic media in Finnish Integrated Monitoring catchments. Boreal Env. Res. 3: 235-249.
*Vuorenmaa J., Salonen K., Arvola L., Mannio J., Rask M. \& Horppila P. 2014. Water quality of a small headwater lake reflects long-term variations in deposition, climate and in-lake processes. Boreal Env. Res. 19 (suppl. A): 47-65.
Zingel P., Paaver T., Karus K., Agasild H. \& Nöges T. 2012. Ciliates as the crucial food source of larval fish in a shallow eutrophic lake. Limnol. Oceanogr. 57: 1049-1056.

