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2 Effect of biochar application on mineral and microbial properties of soils growing different
3 plant species

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2 Full-length paper, 7. Fertilizers and soil amendments

3

4

1 **ABSTRACT**

2 Biochar is widely used as a soil amendment to increase crop yields. However, the details of
3 its impact on soil properties have not been fully understood. A pot experiment was
4 conducted using soybean and sorghum under four soil treatment combinations (cattle
5 farmyard manure with or without biochar and rapeseed cake with or without biochar) to
6 elucidate the mechanisms of its beneficial effects on plant growth in terms of the microbial
7 community structure and mineral availability in soils with different types of organic manure
8 application. The application of biochar significantly increased the growth of both species,
9 particularly sorghum with rapeseed cake application by 1.48 times higher than that without
10 biochar. Microbial activity in soil was also enhanced by biochar application in both species
11 with rapeseed cake application, particularly in sorghum. Principal component analysis
12 using Biolog EcoPlate™ data indicated that biochar application changed the microbial
13 community structure in soil, particularly sorghum grown soil. The changes in microbial
14 community structure in sorghum were considered to be at least partly affected by changes
15 in soil pH due to interaction between plant and biochar under organic manure application.
16 Biochar application had little effect on the profile of ammonium-acetate-extractable mineral
17 elements in soil including calcium, potassium, magnesium, sodium, and sulphur with both
18 types of manure application under soybean. Under sorghum, however, biochar with
19 rapeseed cake manure application altered the profile. This alteration is attributable to an
20 increase in the extractable concentration of certain metals in the soil including aluminum,
21 cadmium, and zinc, possibly caused by enhanced organic matter decomposition producing
22 metal-chelating organic compounds. These different changes in the soil properties by

1 biochar application may be directly or indirectly related to the different growth responses of
2 different plant species to biochar application under organic manure application.

3

4 **Keywords:** microbial activity, mineral element, organic manure, sorghum, wood biochar

5

6 **1. Introduction**

7 Driven by population growth, increased human pressure on land has forced the conversion
8 of natural landscapes into agricultural fields while simultaneously depleting the land under
9 agricultural use (Lal 2009). Therefore, there is an urgent need to establish effective
10 agricultural management practices that not only increase food production but also prevent
11 the negative environmental impacts of intensive agriculture. There are various fertilizers
12 and soil amendments that are able to improve soil fertility and crop productivity. Fertilizers
13 are necessary to increase crop production, and are supplied mainly in the form of chemical
14 amendments. However, the continuous and excessive use of chemical fertilizers may result
15 in environmental pollution. In addition, agriculture largely depending on chemical
16 fertilizers is not sustainable in terms of a shortage of their resources and high energy costs
17 for their production (Ryan *et al.* 2012; Woods *et al.* 2010). Organic fertilizers may be
18 alternatively used as chemical fertilizers; such organic fertilizers can also act as soil
19 amendments which improve the physical, chemical, and biological properties of soil.
20 However, organic fertilizers are less effective than chemical fertilizers because nitrogen (N)
21 and phosphorus (P) mainly occur in organic forms in organic fertilizer.

1 Biochar is a product of the thermal degradation of organic material under oxygen-limited
2 conditions. With respect to appearance, it is similar to charcoal produced by natural
3 burning; however, it is distinguished by its use as a soil amendment (Sohi *et al.* 2009;
4 Lehmann and Joseph 2009). Many studies have shown the beneficial effects of biochar on
5 soil chemical properties such as pH (Topoliantz *et al.* 2007; Masulili *et al.* 2010; Yuan *et al.*
6 2011), nutrient availability (Chan *et al.* 2008; Haefele *et al.* 2008), nutrient retention
7 (Glaser *et al.* 2002; Lehmann *et al.* 2003), and cation-exchange capacity (Glaser *et al.*
8 2002; Masulili *et al.* 2010; Yuan *et al.* 2011). Improvements in the growth and yield of
9 plants following biochar application have also been reported in various crop species,
10 including cowpea and rice (Lehmann *et al.* 2003), radish (Chan *et al.* 2008), soybean
11 (Tagoe *et al.* 2007), and maize (Yamato *et al.* 2006).

12 Recently, the use of organic fertilizers in soil to increase crop productivity has received
13 considerable attention. The incorporation of organic fertilizers is a useful approach for
14 maintaining organic matter content in soil and thereby enhance soil biological activity and
15 increase nutrient content, which, in turn, contributes to increasing crop productivity
16 (Dikinya and Mufwanzala 2010; Diacono and Montemurro 2010). However, in order to
17 supply available nutrients to plants, organic fertilizers need to be mineralized by soil
18 microorganisms. Biochars have been shown to have a positive effect on soil fertility and
19 plant growth (as described above); however, little information is available on their effects
20 when combined with manure, particularly in terms of the mineral and microbial properties
21 of soil. Therefore, this study assessed the effects of biochar on the microbial community

1 structure and mineral availability in soils growing different crop species under different
2 organic manure treatments.

3

4 **2. Materials and methods**

5 *2.1. Experimental setup*

6 A pot experiment was conducted using soybean and sorghum under four soil treatment
7 combinations (cattle farmyard manure with/without biochar and rapeseed cake with/without
8 biochar) to examine the effects of wood biochar on the microbial community structure and
9 mineral availability in soils (Table 1). Soils (Gleyic Fluvisol) were collected from the 0–25
10 cm layer at the experimental farm of Hokkaido University. The soil was air-dried and
11 passed through a 2.0 mm mesh screen. Then, 0.8 L of soil and 0.8 L of perlite were mixed
12 and placed in a plastic pot (1.6 L). Since soil used in a preliminary experiment showed a
13 compacted layer and little penetration of plant root, we mixed soil with perlite to improve
14 soil physical conditions. The biochar used in this experiment was purchased from
15 Shimokawa City Forest Organization Carbon Industry and was produced from broad-leaved
16 trees at 400°C; the biochar had a Carbon (C) content of 71.8%. The amount of fine biochar
17 (<0.25 mm) applied to each pot was 28 g (equivalent to a field application rate of 35 t ha⁻¹).
18 Two different types of organic fertilizer were used: cattle farmyard manure (0.8% N, 1.7%
19 P₂O₅, and 1.8% K₂O) and rapeseed cake (5.3% N, 2.0% P₂O₅, and 1.0% K₂O); 31.75 g of
20 cattle farmyard manure (providing 0.254 g N, 0.540 g P₂O₅, and 0.572 g K₂O) or 4.793 g of
21 rapeseed cake (providing 0.254 g N, 0.096 g P₂O₅, and 0.048 g K₂O) was applied to each
22 pot. In treatments with rapeseed cake, 0.878 g of calcium superphosphate and 0.381 g of

1 K₂SO₄ were added to make the application rates of each of N, P₂O₅, and K₂O equal to 100
2 kg ha⁻¹. These organic fertilizers were used in the present study because they are very
3 common in crop cultivation but their chemical properties differ remarkably. Chemical
4 properties of biochar and each organic manure used in this study were shown in the
5 Supplementary material, Table S1. After mixing of the soil, perlite, fertilizer, and biochar,
6 the pots were incubated for 4 weeks in a greenhouse under moderately moist conditions
7 (40–60% of field capacity depending on the soil condition).

8 Seeds of soybean (*Glycine max* (L.) Merr. cv. Toyoharuka) and sorghum (*Sorghum bicolor*
9 (L.) Moench cv. Hybrid Sorgo) were sterilized with 10% (v/v) NaClO solution for 1 min
10 and then rinsed in deionized water. The seeds were sown and germinated in vermiculite.
11 After the first two leaves appeared (post 10–12 days), two of each species were
12 transplanted to each pot. Depending on the soil condition during the experiment, all pots
13 were then watered with deionized water to 40–60% of their field capacity. The pot
14 experiment was performed for 30 days (November 15–December 12, 2015) for soybean
15 and 40 days (November 15–December 23, 2015) for sorghum, according to their growth
16 rates in the vegetative stage. The experiment was conducted in a greenhouse at an almost
17 constant average temperature of 25°C.

18 **2.2. Soil sampling and analysis**

19 Soil samples were collected at the time of plant sampling (30 and 40 days after sowing for
20 soybean and sorghum, respectively). After removing the plants (as described later), the soil
21 in each pot was mixed and remaining roots were removed. Fresh soil was taken for

1 determining the microbial community structure and activity analyses using EcoPlate™
2 (Biolog Inc., CA, USA). The remaining soil was air-dried and sieved for chemical analysis.
3 EcoPlate contains three replicate wells of 31 of different carbon sources and water (no
4 substrate; tetrazolium dye only as a blank) for soil community analysis. For assessing
5 microbial carbon utilization patterns, a 1 g soil sample was thoroughly shaken by hand with
6 10 ml of sterile saline solution (0.85% NaCl) and diluted 1000 times with the same saline
7 solution. A subsample of 150 μL was inoculated directly into each well of the EcoPlate.
8 Three replicate suspensions were prepared for each soil sample. The EcoPlates were then
9 placed in an incubator at 25 °C; purple color was formed when the microbes utilized the
10 carbon source and began to respire. The color development was measured every 24 h for 5–
11 6 days using a microplate reader (Sunrise Remote, TECAN A-5082, Austria) at 595 nm.
12 Changes in the pattern were compared and analyzed using principle components analysis
13 (PCA). The average well color development (AWCD) in each plate, which indicates
14 microbial activity, was calculated as follows:

$$15 \quad \text{AWCD} = [\Sigma(R_i - C)]/31,$$

16 where R_i and C are the optical density (OD) values at 595 nm of the response wells
17 (containing sole carbon sources) and the control well (water), respectively. A high value of
18 AWCD reflects high microbial activity.

19 Soil pH (H_2O) was determined at a soil/water ratio of 1:2.5 using a pH meter (Mettler
20 Toledo, MP220, 2005). For determination of inorganic N (NH_4 -N and NO_3 -N)
21 concentration in the soils, 4 g samples were extracted with 40 mL of 2 M KCl by shaking
22 for 1 h. The soil extracts were passed through filter paper (No. 6, Advantec, Tokyo, Japan).

1 The $\text{NH}_4\text{-N}$ and $\text{NO}_3\text{-N}$ concentrations were determined by colorimetric methods.
2 Available P was extracted with Truog's solution and measured by spectrophotometry (U-
3 5100, HITACHI, Japan) at 710 nm. Excluding inorganic N and available P, the
4 concentrations of mineral elements in the soil were determined by extracting 2 g of soil
5 with 40 mL of 1 M ammonium acetate, shaking for 30 min, and passing through fine filter
6 paper (No. 5C, Advantec, Tokyo, Japan) to exclude fine soil particles. Thereafter, 5 mL of
7 the filtrate was digested with 2 mL of 61% HNO_3 as described later. The concentrations of
8 aluminum (Al), arsenic (As), boron (B), calcium (Ca), cadmium (Cd), cobalt (Co), cesium
9 (Cs), copper (Cu), iron (Fe), potassium (K), lithium (Li), magnesium (Mg), manganese
10 (Mn), molybdenum (Mo), sodium (Na), nickel (Ni), rubidium (Rb), sulphur (S), selenium
11 (Se), strontium (Sr), vanadium (V), and zinc (Zn) in the digested solution were measured
12 using inductively coupled plasma mass spectrometry (ICP-MS) (ELAN, DRC-e, Perkin
13 Elmer, MA, USA).

14 **2.3. Plant sampling and analysis**

15 Plants were harvested at the end of the vegetative growth. Roots of the plants were washed
16 clean with tap water. The plants were then separated, washed with de-ionized water, and
17 dried in an oven at 70°C for 7 days, before being weighed and ground for mineral analysis.
18 The concentrations of mineral elements in the plant samples were determined as described
19 in Watanabe *et al.* (2015). Briefly, plant samples were digested in 2 mL of 61% HNO_3 (EL
20 grade, Kanto Chemical, Tokyo, Japan) at 110°C in a DigiPREP apparatus (SCP Science,
21 QC, Canada) for approximately 2 h until the solution had almost disappeared. After the
22 samples had cooled, 0.5 mL H_2O_2 30% (semiconductor grade, Santoku Chemical, Tokyo,

1 Japan) was added, and the samples were heated at 110°C for a further 20 min. Once
2 digestion was complete, the tubes were cooled and made up to a volume of 10 mL by
3 adding 2% HNO₃. The concentrations of elements were measured using ICP-MS.

4 **2.4. Data analysis**

5 All experimental data were statistically analyzed using Minitab 16 (Minitab, Inc, United
6 States). Analysis of variance (ANOVA) followed by Tukey's test were used to detect
7 significant differences among treatments. To compare the results of the treatments with and
8 without biochar, paired Student's t-tests were applied. PCA was used to profile the
9 microbial communities and minerals in the soils.

10

11 **3. Results**

12 **3.1. Growth and mineral accumulation of plants**

13 The total dry weight of both the plant species grown with rapeseed cake significantly
14 increased by the biochar application, particularly for sorghum (1.21 and 1.48 times higher
15 than that without biochar for soybean and sorghum, respectively) (Fig. 1). A similar trend
16 was also found for cattle farmyard manure; however, this was not statistically significant
17 (Fig. 1). The concentrations of some metal elements in leaf material are shown in Table 2.
18 Overall, irrespective of the plant species and manure type, the biochar application
19 decreased or did not affect the concentration of these elements.

20 **3.2. General chemical properties of soil**

21 Soil pH was higher in soil receiving cattle farmyard manure (Table 3). The biochar
22 application significantly increased soil pH under sorghum with each type of organic manure,

1 but not in soybean. The concentration of inorganic N ($\text{NH}_4\text{-N}$ and $\text{NO}_3\text{-N}$) did not differ
2 significantly between the treatments (Table 3). Available P concentration was higher in
3 soils receiving cattle farmyard manure; however, no biochar effect was observed for either
4 plant grown soil (Table 3). The concentration of ammonium-acetate-extractable S in soil
5 increased under sorghum because of biochar application with rapeseed cake (Table 3).

6 ***3.3. Microbial activity and community structure of soil***

7 The microbial activity in soil was estimated by AWCD. A high value of AWCD reflects
8 higher microbial activity. Figure 2 presents the AWCD values obtained from the EcoPlate.
9 For both soybean and sorghum grown soils with rapeseed cake, biochar application
10 significantly increased AWCD compared to the case without biochar. When PCA was
11 conducted to assess the utilization patterns of the different carbon sources, the total
12 variance explained by the first two components was 39% and 44% for soybean and
13 sorghum, respectively. In the soybean grown soil, the score plot of PCA showed a
14 separation between the B0M2 treatment and the other treatments in PC1 (Fig. 3).
15 Meanwhile, in the score plot of the sorghum grown soil, the soil with biochar application
16 shifted negatively along PC1 in both the cattle farmyard manure and rapeseed cake
17 treatments (Fig. 3).

18 ***3.4. Mineral profile of soil***

19 PCA was also used to examine the treatment effects on the mineral profile of the soils.
20 Figure 4 shows the score plot of the first two components from the ammonium-acetate-
21 extractable concentrations of each mineral element in soils growing soybean or sorghum.
22 The first two components accounted for 67% and 92% of the total variance for the soybean

1 and sorghum soils, respectively. In both species, a clear separation in the score plot was
2 observed between the different types of organic fertilizer applied to the soil (Fig 4).
3 Moreover, the biochar application altered the profile of the extractable mineral elements in
4 the sorghum grown soil with rapeseed cake (Fig. 4), whereas it did not affect that in the
5 soybean grown soil (Fig. 4).

6

7 **4. Discussion**

8 It has been reported that the positive effect of biochar on plant growth may be related to the
9 nutrient-retention capacity of biochar (Glaser *et al.* 2002) and its sorption capacity for toxic
10 metals and some phytotoxic compounds (Hille and den Ouden 2005; Lair *et al.* 2006). In
11 the present study, the biochar application significantly enhanced the growth of both the
12 plant species grown in soils with rapeseed cake, and the same trend was also observed in
13 their growth with cattle farmyard manure (Fig. 1), indicating that the combined application
14 of organic manure with biochar is effective at increasing crop yield. This growth promotion
15 effect was more remarkable in the soil with rapeseed cake for sorghum. Therefore, different
16 plant species as well as different types of organic manure may affect soil-biochar
17 interactions differently. This raises the question regarding the factors causing these
18 differences.

19 In both soybean and sorghum grown soils, AWCD-estimated microbial activity was
20 increased by biochar application with rapeseed cake (Fig. 2). It has been reported that
21 biochar provides a suitable habitat for microorganisms (Pietikainen *et al.* 2000) and
22 produces substances that stimulate the growth of microbes (Kasozi *et al.* 2010). Rapeseed

1 cake may have suitable characteristics for the exertion of these positive effects of biochar
2 on microorganisms. The enhanced microbial activity can be expected to enhance the
3 mineralization of rapeseed cake applied to soils. However, significant differences were not
4 found for both plant species in inorganic N and available P concentrations between soils
5 with and without biochar in the rapeseed cake treatment. By contrast, a significant increase
6 was found in concentrations of extractable S, including SO_4^{2-} and soluble organic S such as
7 S-containing amino acids, because of biochar application under sorghum in the rapeseed
8 cake treatment (Table 3). Moreover, significant positive correlation was found between
9 extractable S concentration in soil and utilization (absorbance) for phenylethyl-amine in the
10 EcoPlate in sorghum with rapeseed cake application ($r = 0.94$, Supplementary material,
11 Figure S1), but not in soybean. These results imply that the biochar enhanced the microbial
12 decomposition of organic matter, containing organic S in this soil, resulting in superior
13 growth of the sorghum. In fact, for the rapeseed cake treatments, S concentration in the
14 leaves of sorghum significantly increased by biochar application (data not shown, Student's
15 *t*-test, $P < 0.05$). Considerable difference in soil extractable S between soybean and
16 sorghum (Table 3) may imply considerable difference in the impact of plant root on
17 microbial community structure between different plant species as described later. Although
18 we cannot know exactly why significant effect of biochar application was observed only in
19 S but not in N and P, leaching and active absorption by root in N and fixation by soil in P
20 may be involved in these different results.

21 It has also been suggested that biochar may change the soil microbial community structure
22 (Lehmann *et al.* 2011). In the present study, PCA of the EcoPlate data demonstrated that

1 biochar application clearly changed the microbial community structure, particularly in
2 sorghum grown soils (Fig. 3). Correlation analysis was conducted to determine the factor(s)
3 responsible for biochar-induced changes in the microbial community structure. In the PCA,
4 the PC1 scores using the EcoPlate showed a strong negative correlation ($r = -0.89$, $n = 16$)
5 with soil pH for sorghum but not for soybean (which showed a weak positive correlation, r
6 $= 0.72$, $n = 16$) (Fig. 5). In fact, biochar application did not significantly affect soil pH
7 under soybean but increased it under sorghum (Table 2). Although biochar is commonly
8 alkaline and has a potential to increase the soil pH, the effect depends on the types of
9 biochar (Yang *et al.* 2016). Moreover, changes in soil pH by biochar application may also
10 be affected by microbial activities such as nitrification which produces H^+ (Yuan *et al.*
11 2011). The positive correlation between soil pH and NH_4/NO_3 ratio observed only in
12 sorghum grown soil (Supplementary material, Figure S2) supports the involvement of
13 nitrification in different responses of soil pH to biochar application between soybean and
14 sorghum. Together, some interactions between the sorghum rhizosphere and biochar may
15 affect soil pH; this may be the primary factor altering the microbial community structure in
16 soils. Under soybean, biochar application had little effect on the profile of ammonium-
17 acetate-extractable mineral elements of the soil for both types of manure application (Fig.
18 4). For sorghum, however, biochar application altered the profile of the extractable
19 elements in the soil applied with rapeseed cake (Fig. 4). This alteration was mainly due to
20 the increase in the extractable concentrations of certain metals in soils due to biochar
21 application (Table 4). The biochar application increased soil pH in sorghum grown soil
22 applied with rapeseed cake (Table 3), which cannot explain the results of the extractable

1 metals in this study because increasing the pH normally decreases the availability of certain
2 metal cations such as Al and Cd (von Uexküll and Mutert 1995; Xian and In Shokohifard
3 1989). In order to increase the solubility of these metal cations even at increased pH
4 condition, some chelating compounds could be needed.

5 In contrast to the effects of biochar on soil, concentrations of these metals in the leaves of
6 sorghum grown in the soil with rapeseed cake did not change, or they tended to show a
7 decrease due to the biochar application, particularly in Al (Table 2). In case of Al, it is well
8 known that organic ligands, such as organic acids can solubilize Al (Huang and Keller
9 1971) but also inactivate it for plant uptake (Ma and Hiradate 2000) in soil. Because
10 biochar application increased microbial activity in the soils applied with rapeseed cake (Fig.
11 2), it possibly enhanced organic matter decomposition in this soil, producing chelating
12 organic compounds such as organic acid that solubilized some metals but also made those
13 metals less available to sorghum roots. In fact, when analyzing the correlation between
14 extractable concentration of each of these metal elements and utilization (absorbance) for
15 each carbon source in the EcoPlate, highly significant correlation ($P < 0.01$) was found in
16 several carbon sources (4-hydroxy benzoic acid and Al/Ba; phenylethyl-amine and Zn; α -
17 D-lactose and Zn) only in sorghum grown soil with rapeseed cake application
18 (Supplementary material, Figure S1). These carbon utilization characteristics of microbial
19 community in sorghum grown soil with rapeseed cake might be related to the production of
20 chelating compounds from soil organic matter to solubilize certain metals in soil.

21 In conclusion, biochar application can be an important agricultural practice for increasing
22 the efficiency of organic manure for crop cultivation. However, its effects differ depending

1 on the plant species and organic manure type. These differences may be attributed to the
2 complicated interactions between the plant rhizosphere, biochar, organic manure, and soil
3 microorganisms. In order to elucidate these interactions, detailed analysis of the dynamics
4 of microorganisms and organic/inorganic substances in the rhizosphere of soils applied
5 with different types of organic manure is needed.

6

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10

11 **Figure legends**

12 **Figure 1.** Dry weight of soybean and sorghum. ■ and □ indicate root and shoot,
13 respectively. The error bars represent the standard error of the mean (n = 4). *
14 indicates a significant difference between treatments with and without biochar
15 in each organic manure treatment (M1 or M2) ($P < 0.05$, Student's *t*-test).
16 Relative value of the B1 treatment to the B0 treatment is indicated on the bar of
17 B1 in each manure treatment. B0: without biochar; B1: with biochar; M1: cattle
18 farmyard manure; M2: rapeseed cake.

19

20 **Figure 2.** AWCD values of soybean and sorghum grown soils. Error bars represent the
21 standard error of the mean (n = 4). * indicates a significant difference between
22 treatments with and without biochar in each organic manure treatment (M1 or

1 M2) ($P < 0.05$, Student's t -test). B0: without biochar; B1: with biochar; M1:
2 cattle farmyard manure; M2: rapeseed cake. Data recorded 144 and 120 h after
3 the incubation started were used for soybean and sorghum, respectively.

4
5 **Figure 3.** Principal component analysis of carbon source utilization activity (EcoPlate) in
6 soybean and sorghum grown soils. B0: without biochar; B1: with biochar; M1:
7 cattle farmyard manure; M2: rapeseed cake. Data recorded at 144 and 120 h
8 after the incubation started were used for soybean and sorghum, respectively.

9
10 **Figure 4.** Principal component analysis of ammonium acetate-extractable mineral
11 elements in soybean and sorghum grown soils. B0: without biochar; B1: with
12 biochar; M1: cattle farmyard manure; M2: rapeseed cake.

13
14 **Figure 5.** Correlation of soil pH with PC1 or PC2 of PCA ($P < 0.01$) in the EcoPlate for
15 soybean and sorghum grown soils.

Table 1. Treatments in this study.

Treatment	Description
B0M1	no biochar + cattle farmyard manure
B1M1	biochar + cattle farmyard manure
B0M2	no biochar + rapeseed cake
B1M2	biochar + rapeseed cake

Table 2. Concentration (mg kg⁻¹ dry weight) of some metal elements in plant leaf.

		Al			Ba			Cd			Co			Sr			Zn		
Soybean	B0M1	55.33	±6.48	c	79.4	±5.9	a	0.261	±0.018	ab	0.175	±0.006	b	91.2	±5.5	a	154.0	±8.4	a
	B1M1	70.71	±3.55	bc	48.7	±5.0	b	0.164	±0.013	c	0.181	±0.006	b	71.2	±2.6	b	124.7	±8.3	b
	B0M2	94.16	±4.29	a	13.3	±0.3	c	0.303	±0.007	a	0.223	±0.011	a	79.4	±1.7	ab	173.9	±3.1	ab
	B1M2	85.61	±4.97	ab	20.2	±0.6	c	0.244	±0.004	b	0.204	±0.007	ab	70.5	±0.7	b	148.9	±9.5	c
Sorghum	B0M1	65.38	±7.90	b	14.8	±0.9	a	2.11	±0.22	bc	0.080	±0.006	b	24.7	±1.1	ab	108.6	±3.9	a
	B1M1	40.71	±2.42	b	14.5	±0.8	a	1.71	±0.09	c	0.055	±0.003	b	22.0	±1.3	b	79.4	±3.9	b
	B0M2	360.54	±41.32	a	8.4	±0.8	b	2.81	±0.13	a	0.308	±0.028	a	28.3	±0.5	a	114.3	±3.2	a
	B1M2	61.92	±4.77	b	7.7	±0.4	b	2.61	±0.07	ab	0.068	±0.004	b	24.9	±0.7	ab	93.9	±1.9	c

Different letters indicate significant difference at $P < 0.05$ in each species (Tukey's multiple range test).

Table 3. pH and concentration (mg kg⁻¹ dry soil) of NH₄-N, NO₃-N, Truog-P, and ammonium acetate-extractable S in soil.

		pH			NH ₄ -N			NO ₃ -N			P			S		
Soybean	B0M1	5.26	±0.04	a	105.3	±20.1	a	14.9	±0.9	a	546	±17	a	91	±5	c
	B1M1	5.41	±0.06	a	85.6	±20.8	a	9.0	±2.0	b	568	±22	a	98	±2	c
	B0M2	4.88	±0.02	b	38.7	±10.3	a	13.3	±1.1	ab	287	±12	b	362	±13	a
	B1M2	4.99	±0.02	b	47.6	±8.2	a	10.4	±1.3	ab	267	±5	b	303	±6	b
Sorghum	B0M1	5.23	±0.03	b	38.1	±5.2	a	21.0	±1.3	ab	595	±24	a	92	±2	b
	B1M1	5.47	±0.03	a	33.1	±1.7	a	12.2	±0.7	b	601	±20	a	109	±7	b
	B0M2	4.81	±0.02	d	22.8	±3.1	a	28.8	±3.9	a	293	±23	b	1506	±418	b
	B1M2	4.98	±0.01	c	34.9	±5.0	a	21.8	±2.4	ab	278	±9	b	9245	±2390	a

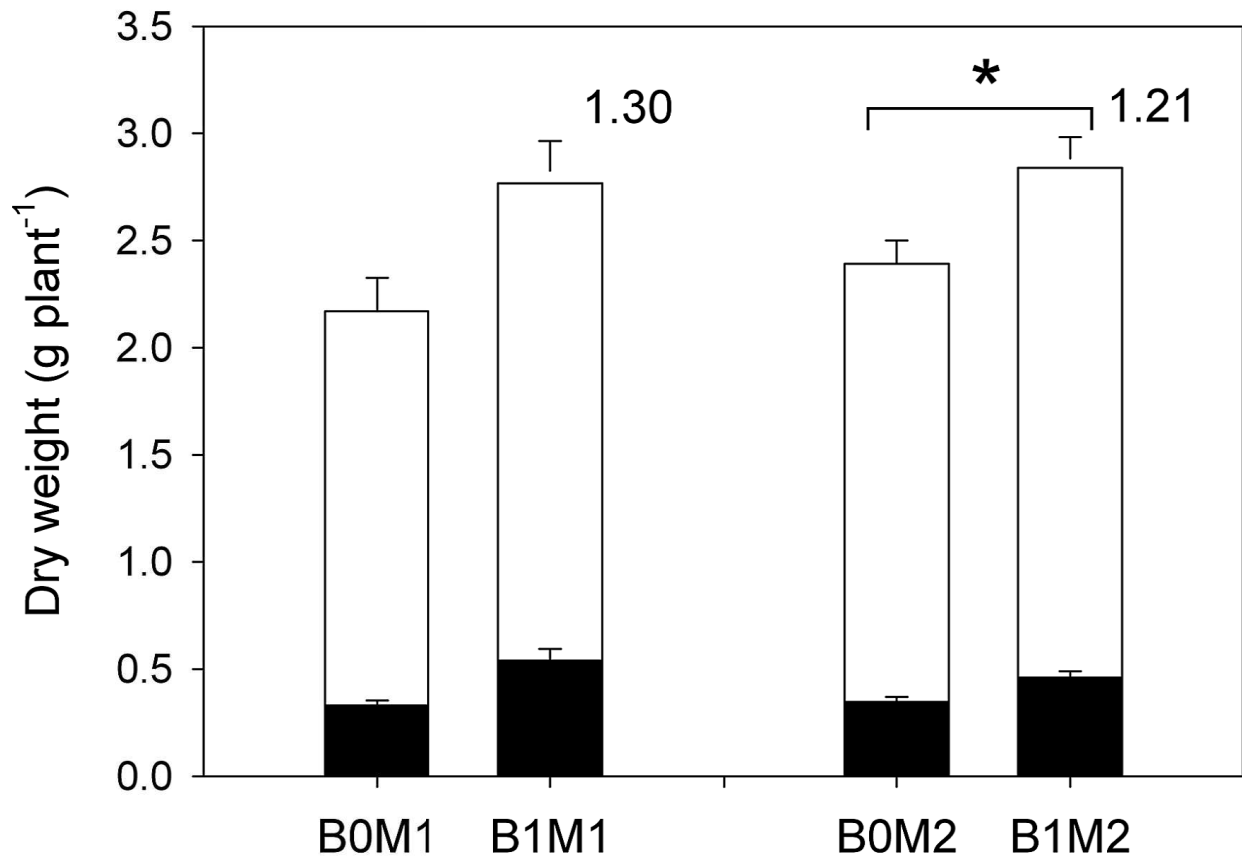
Different letters indicate significant difference at $P < 0.05$ in each species (Tukey's multiple range test).

Table 4. Concentration (mg kg⁻¹ dry soil) of some metal elements in the soil extract.

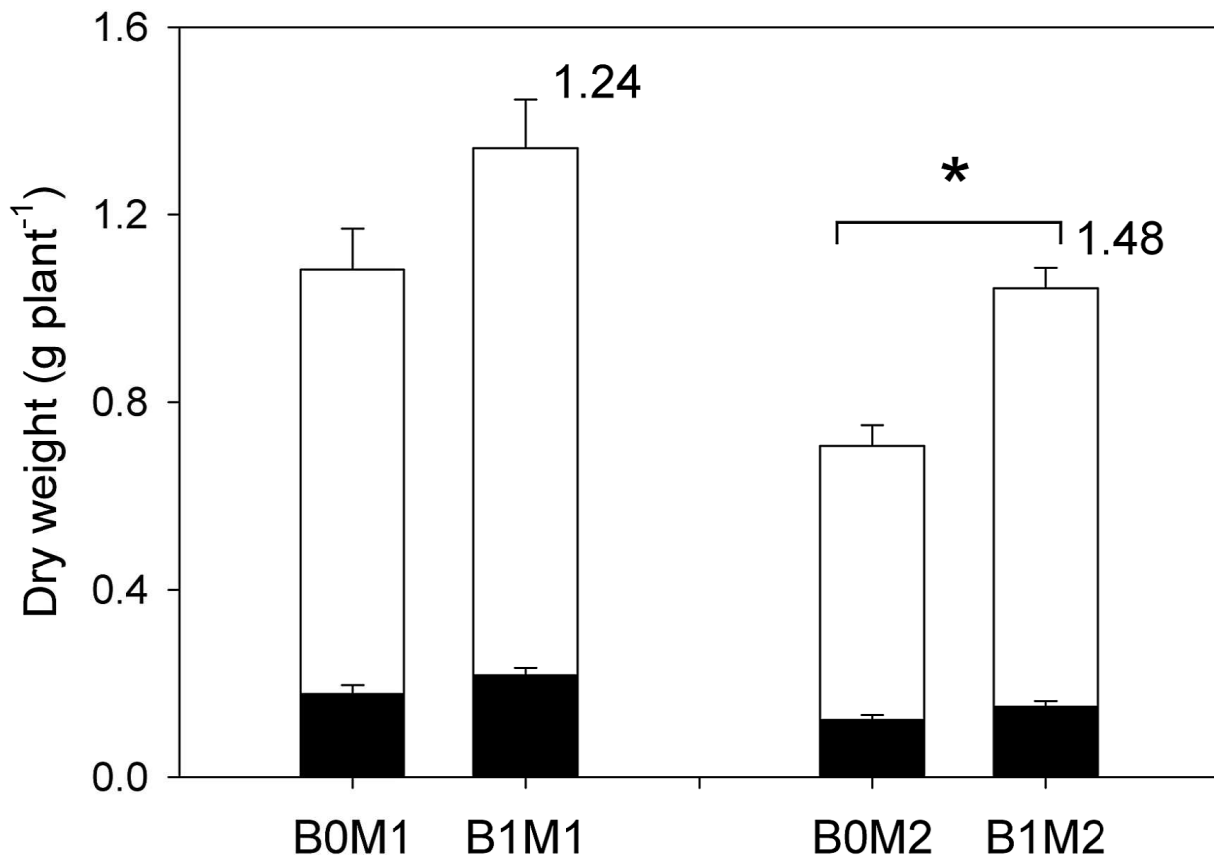
		Al			Ba			Cd			Co			Sr			Zn		
Soybean	B0M1	5.20	±0.59	a	30.0	±1.3	b	0.064	±0.003	b	0.026	±0.002	bc	14.0	±0.6	b	1.19	±0.29	c
	B1M1	4.23	±1.26	a	26.7	±1.8	b	0.060	±0.002	b	0.022	±0.000	c	14.0	±0.7	b	3.95	±0.67	a
	B0M2	3.85	±0.81	a	38.3	±1.9	a	0.100	±0.006	a	0.035	±0.002	ab	19.5	±0.9	a	1.49	±0.44	bc
	B1M2	4.11	±0.37	a	38.4	±1.5	a	0.098	±0.009	a	0.032	±0.003	ab	19.4	±1.4	a	0.63	±0.20	c
Sorghum	B0M1	3.84	±0.47	c	31.6	±0.5	b	0.084	±0.001	c	0.033	±0.001	b	17.2	±0.3	c	1.86	±0.47	c
	B1M1	3.96	±0.59	c	38.4	±3.7	b	0.093	±0.011	c	0.030	±0.003	b	20.2	±1.8	c	2.16	±0.61	c
	B0M2	11.05	±1.20	b	173.9	±41.1	b	0.327	±0.037	b	0.103	±0.012	a	51.0	±5.4	b	11.47	±0.66	b
	B1M2	18.61	±1.80	a	1528.5	±403.8	a	0.543	±0.066	a	0.127	±0.014	a	83.7	±9.9	a	17.51	±1.68	a

Different letters indicate significant difference at $P < 0.05$ in each species (Tukey's multiple range test).

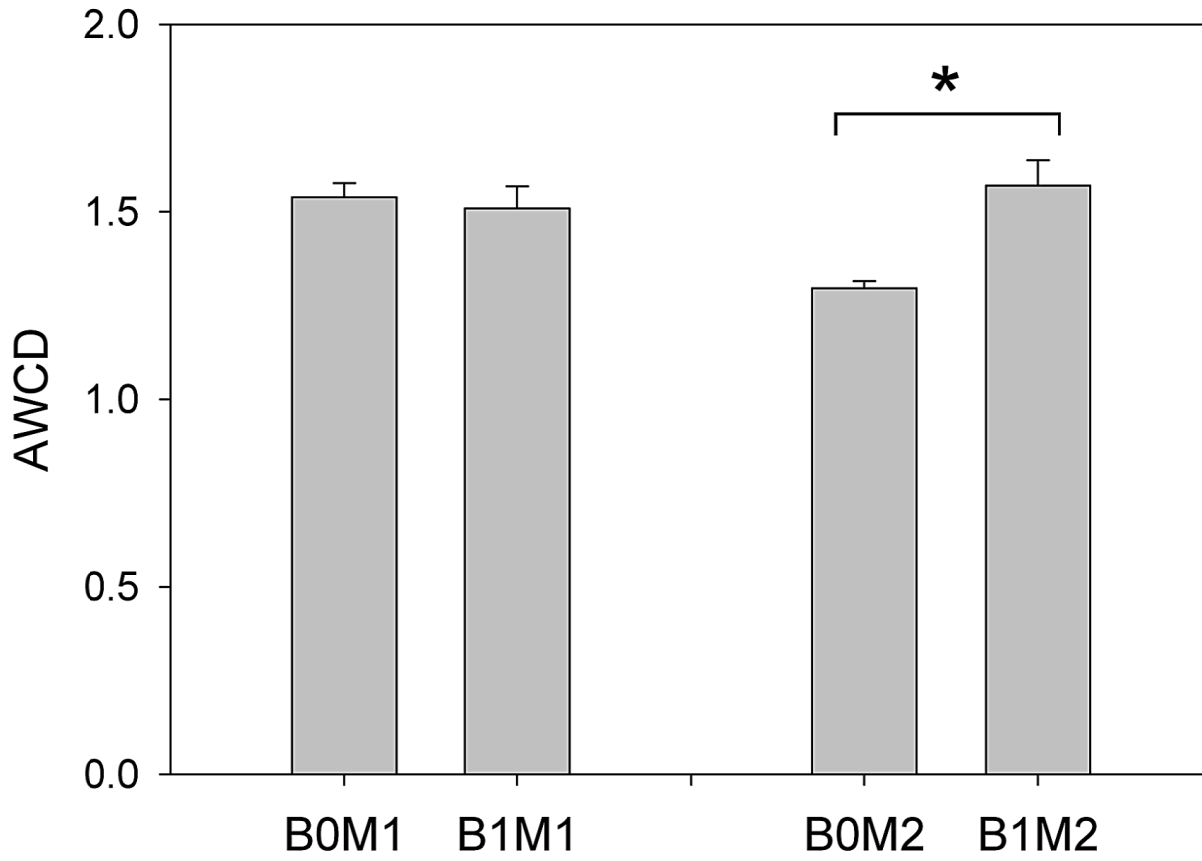
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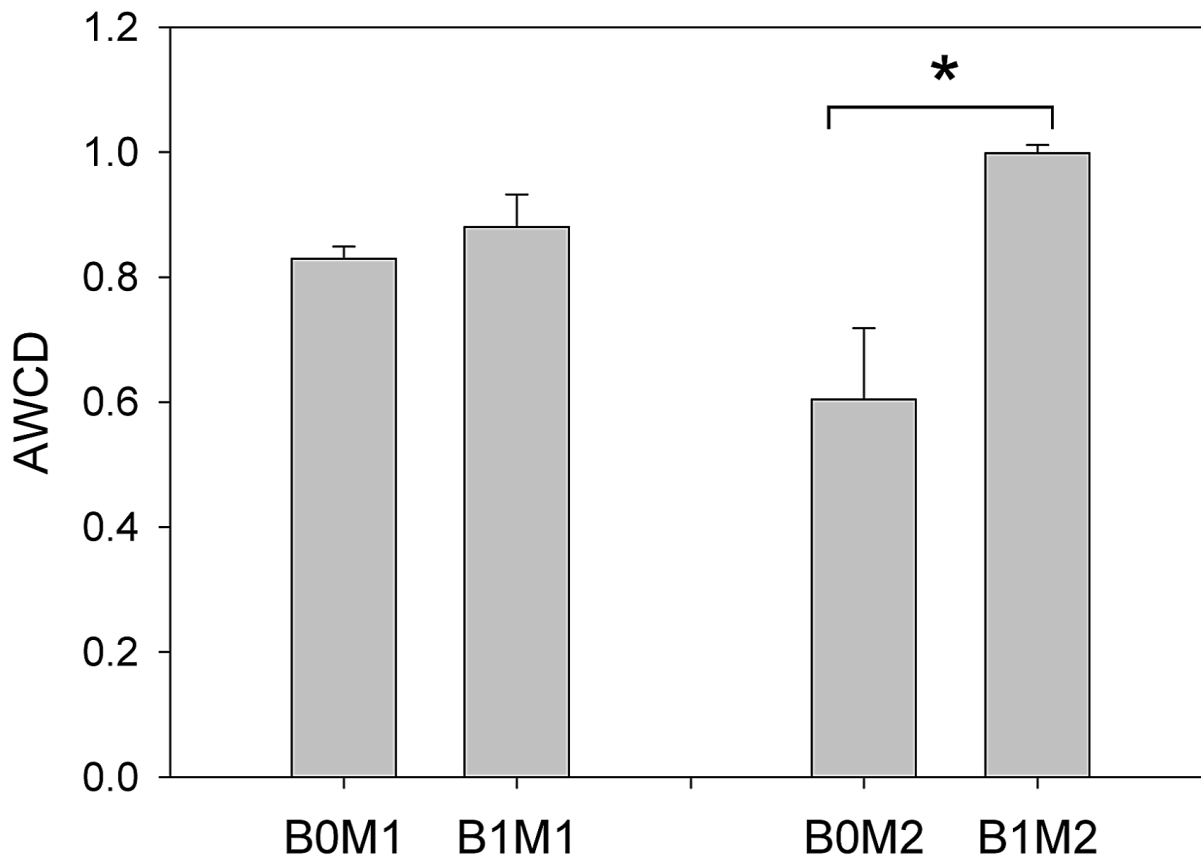
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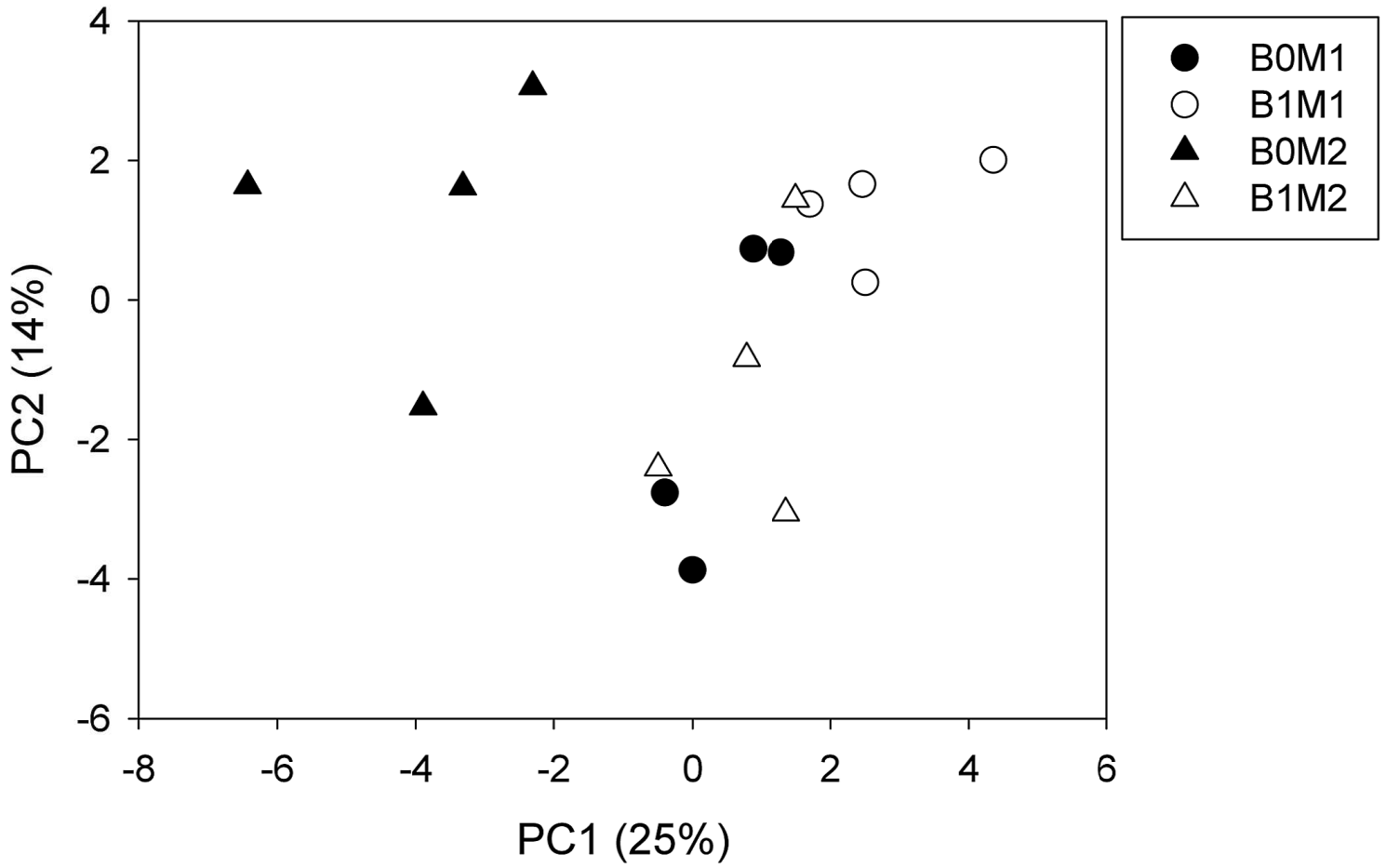
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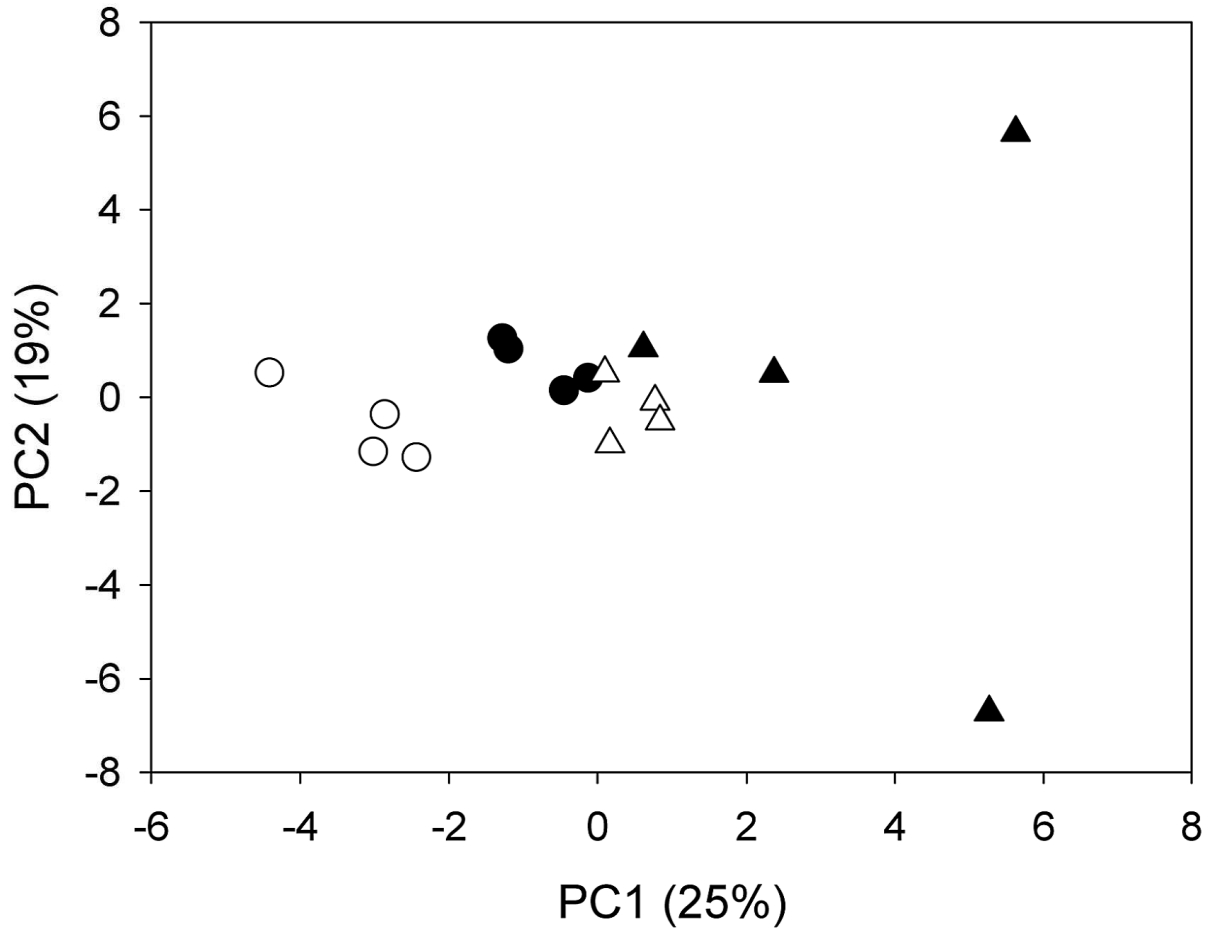
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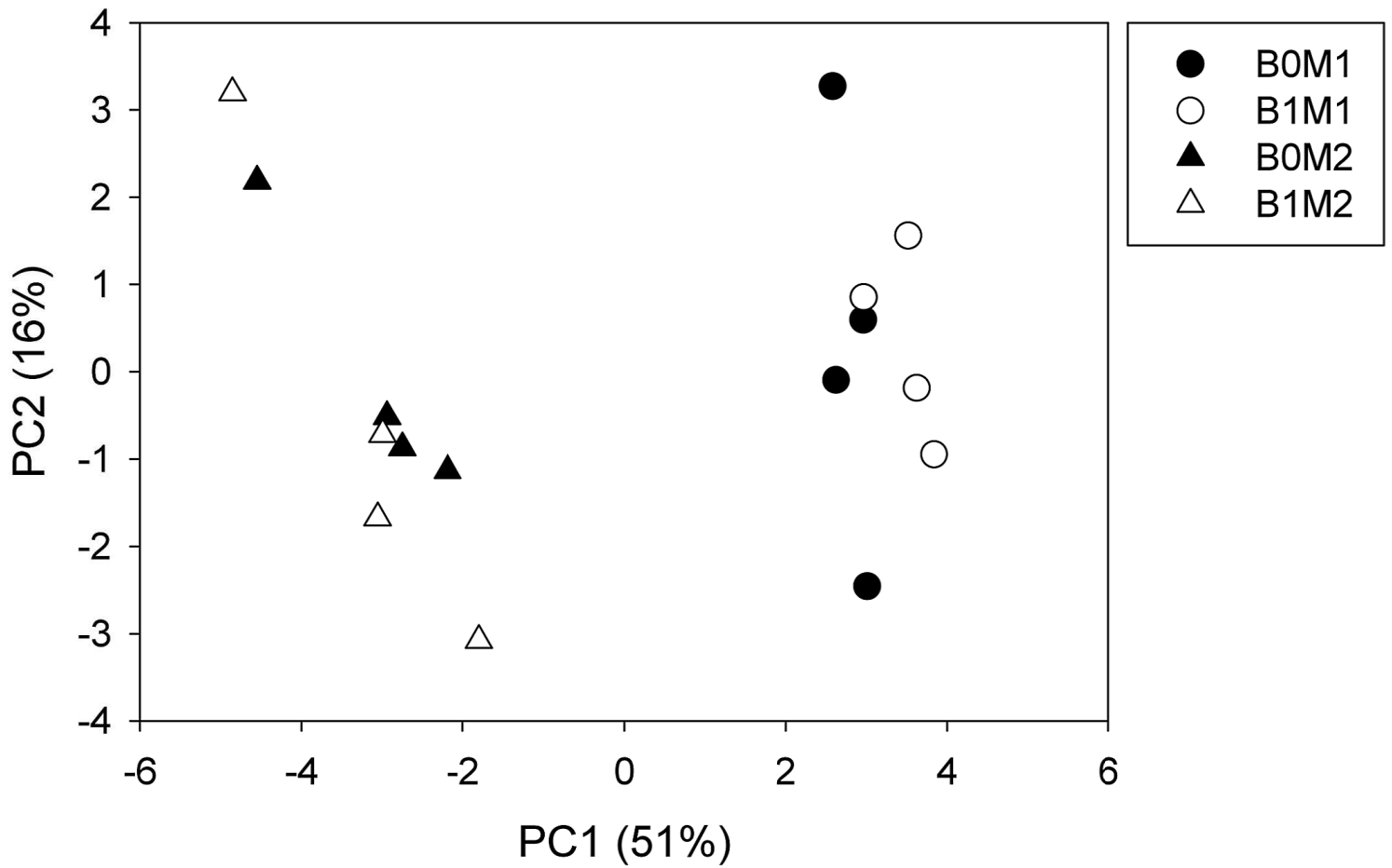
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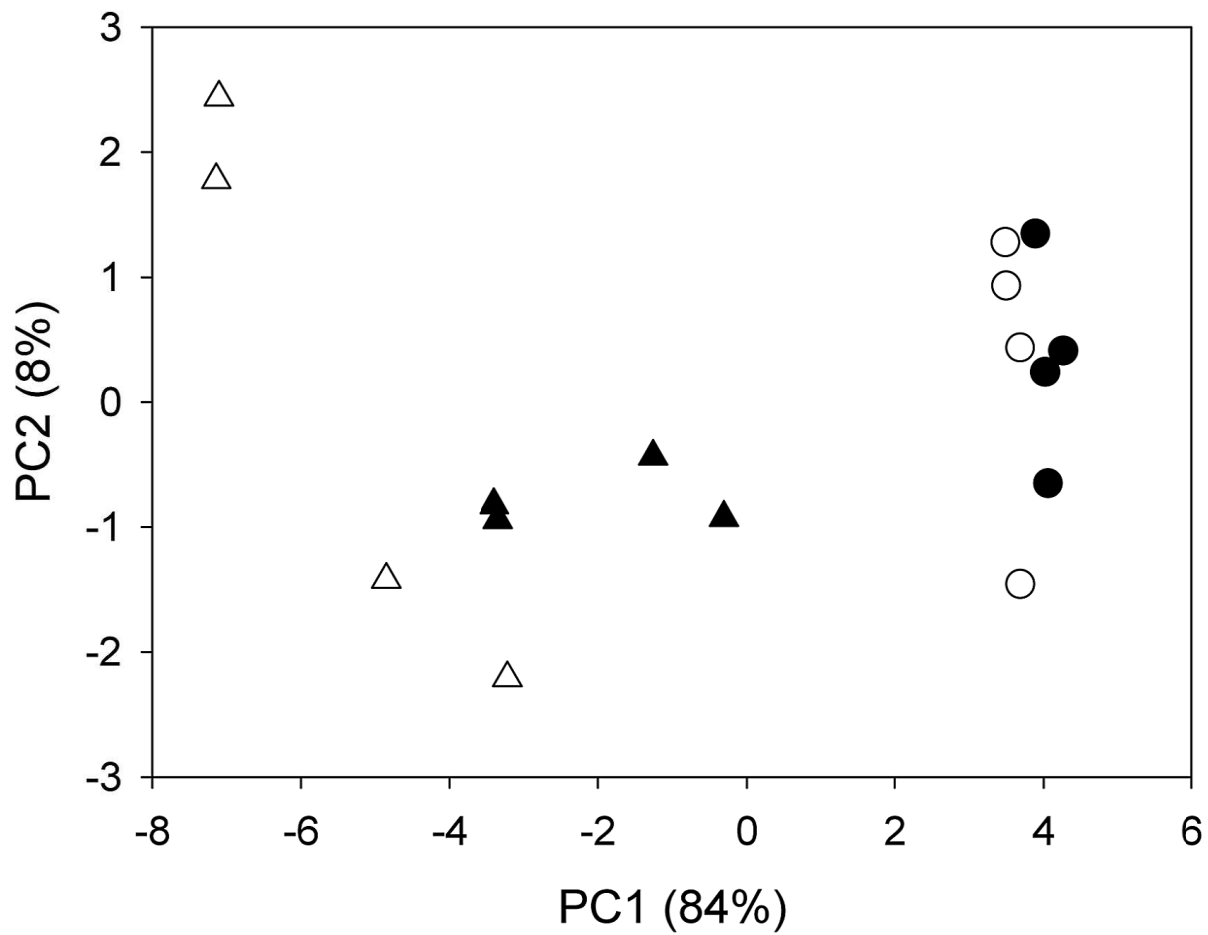
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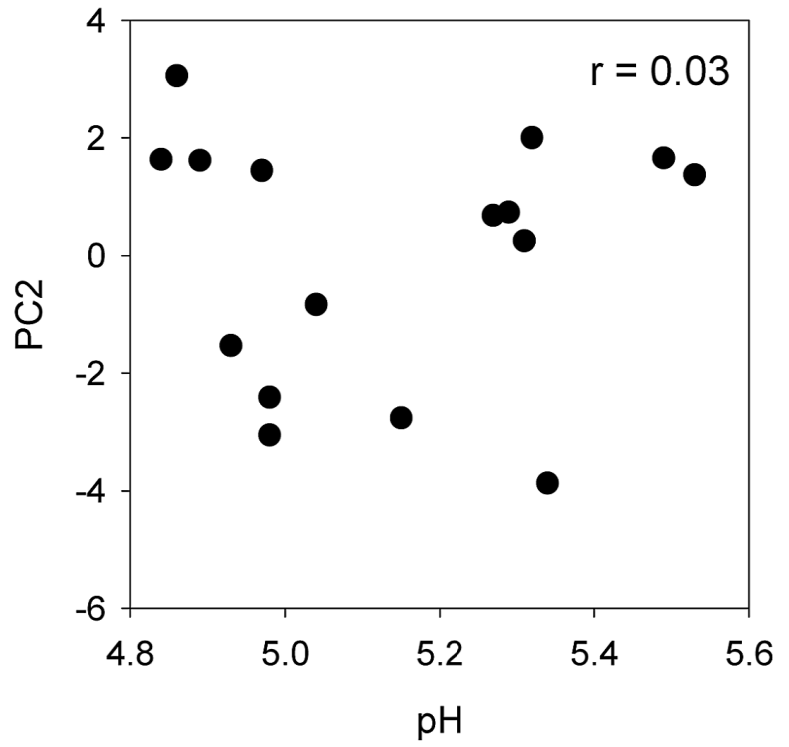
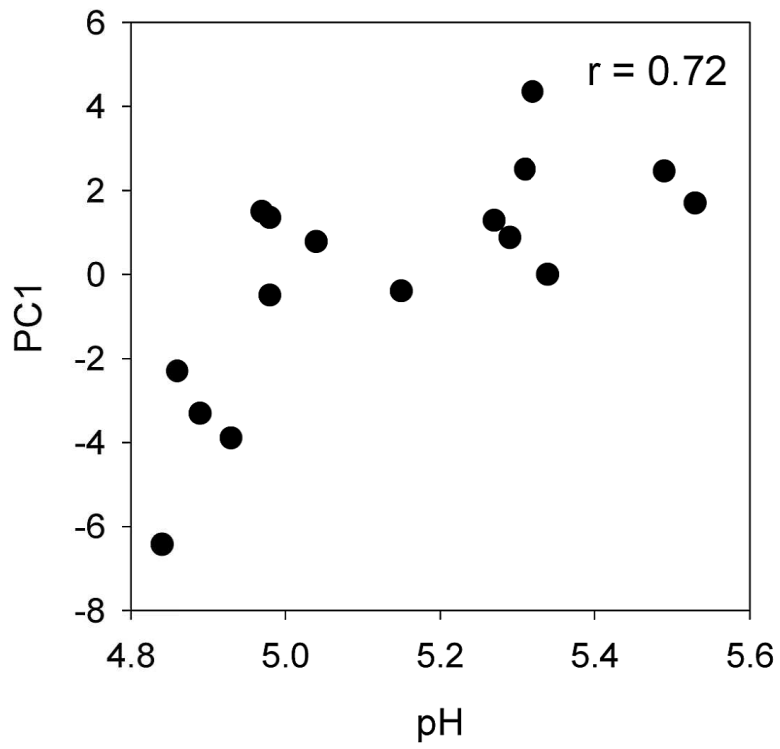
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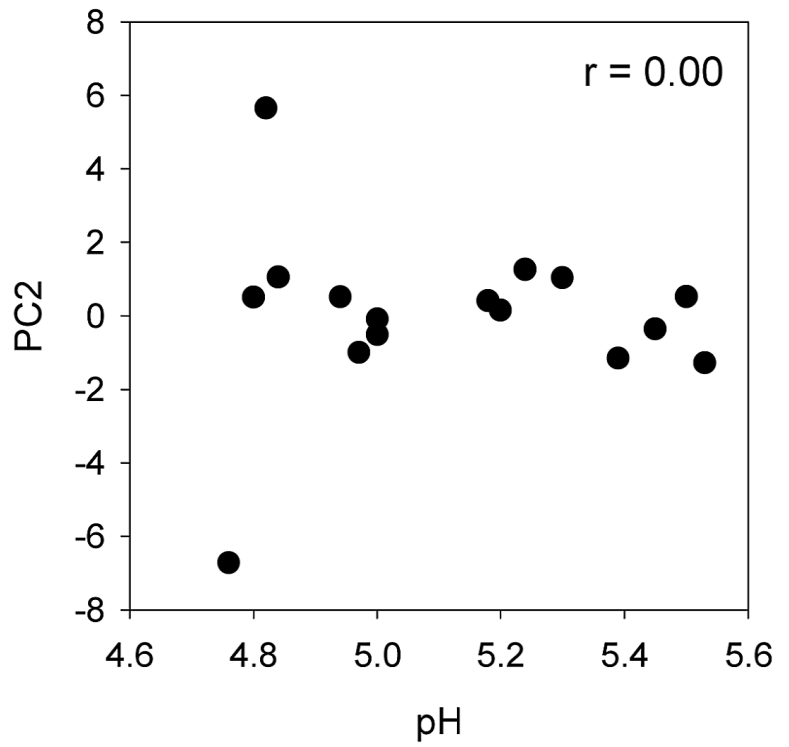
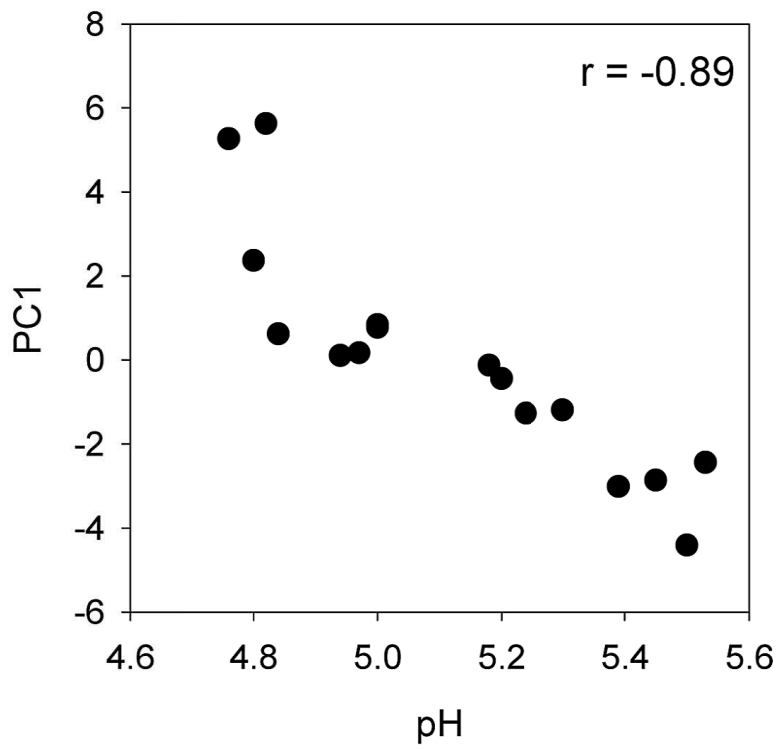
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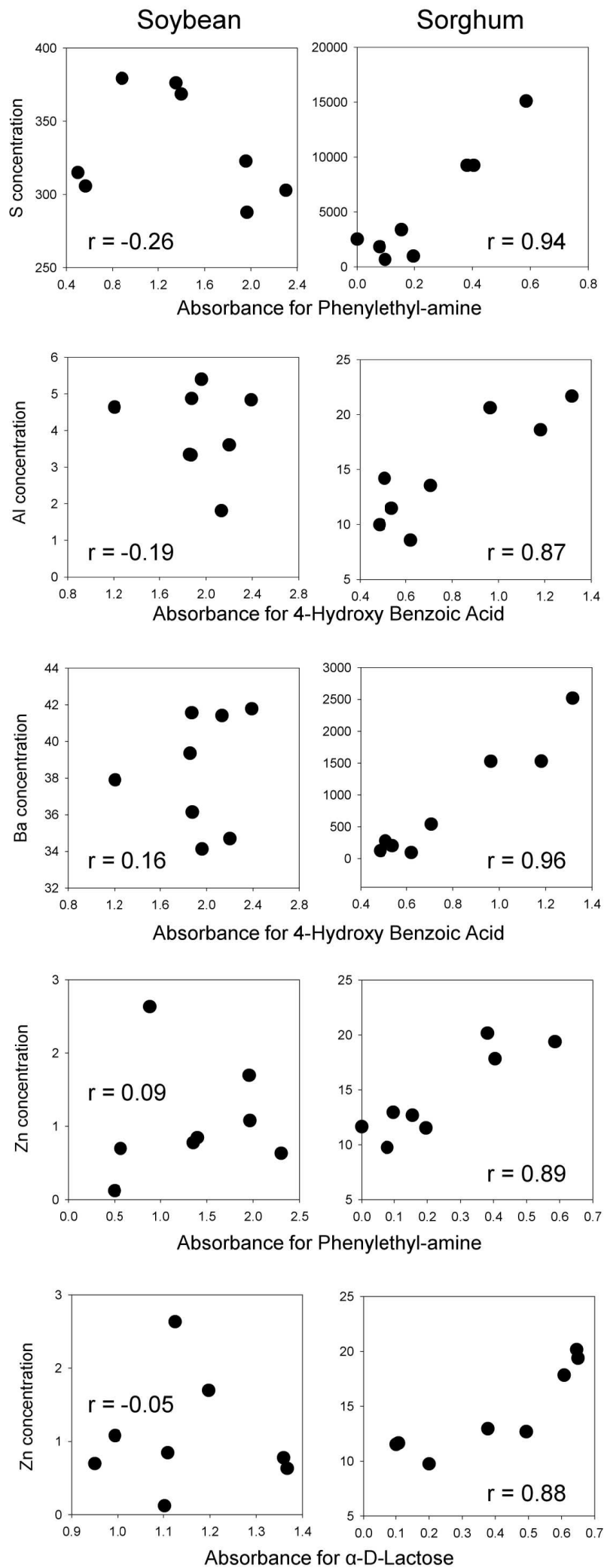


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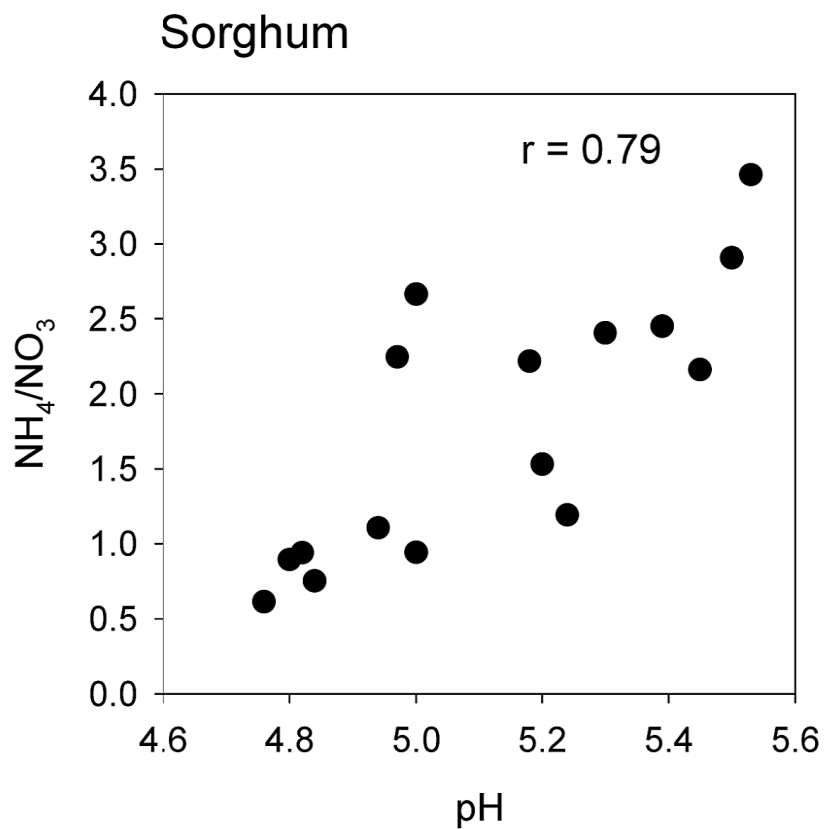
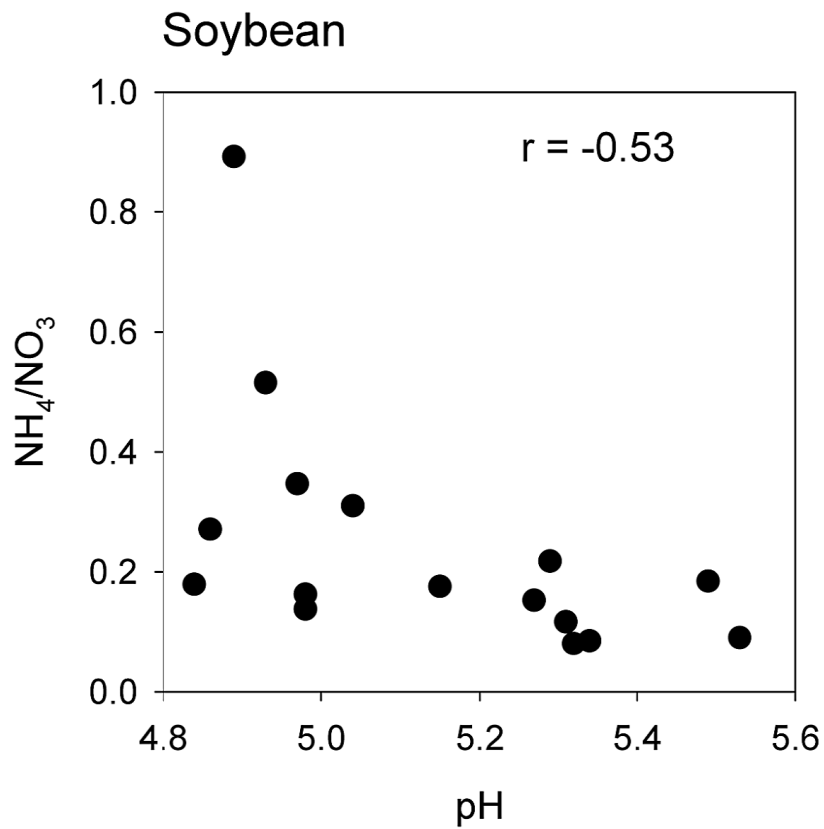


Sorghum





Supplemental Figure S1. Correlation between concentration of extractable element (mg kg^{-1} dry soil) in soil and utilization (absorbance in Biolog Ecoplate) of carbon source in microbial community of soil with rapeseed cake (including both with and without biochar). Only pairs with significant positive correlation ($P < 0.01$) were shown.



Supplemental Figure S2. Correlation between pH and $\text{NH}_4^+/\text{NO}_3^-$ ratio in soils.