

Title	Characterization of primary organic aerosol from domestic wood, peat, and coal burning in Ireland						
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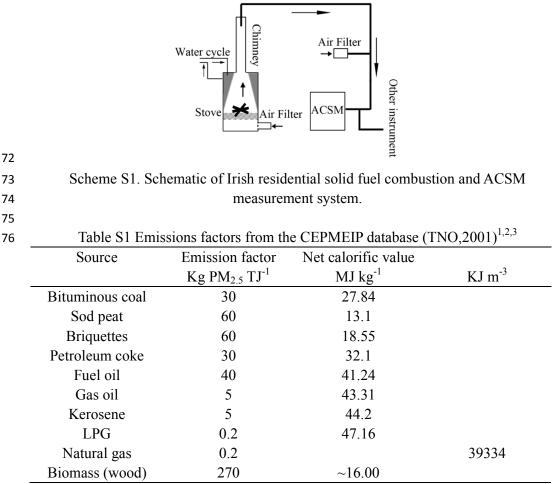
Coláiste na hOllscoile Corcaigh

1	Supplementary Information							
2	Characterization of primary organic aerosol from domestic wood, peat, and coal							
3	burning in Ireland							
4								
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6	Francesco Canonaco ^d , André Stephan Henry Prévôt ^d , Ru-Jin Huang ^{a,b,d,*} , Colin							
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21								
22	The following materials are included:							
23								
24	• Emission factors and caloric values for oil, peat, coal, and wood, Table S1							
25	• Correlation coefficient (R^2) between ACSM profiles of different sources and							
26	PMF factors							
27	• The households by the type of central heating (oil, peat, coal, and wood) from							
28	Central Statistics Office, 2011, Figure S1							
29	• Relative fraction of ACSM measured species, Figure S2							
30	• Mass spectra of each type of fuel under different states, Figure S3-5							
31	• Relative difference of dry wood and smoky coal MS profile compared to peat							
32	at each m/z, Figure S6							
33	• Time series and mass spectra of PMF solutions, Figure S7-9							
34	• Relative contribution of the resolved factors and correlation between OOA and							
35	sulfate with different <i>a</i> values (0-0.2), Figure S10							
36	• Back trajectory during the measurement period in Galway, Ireland, Figure S11							

37 Summary: 13 pages, 1 scheme, 2 tables, and 11 figures

Figerprinting Setup and ACSM data analysis: A boiler stove is used for both home 38 39 heating and the generation of hot water. The combustion chamber was built into a wall with the water pipe network inside the wall behind the chamber. The water pipes 40 41 will take up part of the heat generated in the chamber, producing hot water for everyday use and also circulating through the central heating system warming up the 42 house. The open fire chamber was directly connected to a chimney having no 43 emission control. During each type of fuel sampling, new fuel was added to maintain 44 the combustion which is always the case for the real application instead of waiting for 45 its extinction and igniting a new burning. And each type of sampling fuel was 46 continuously burned for at least 1 hour with a total use of fuel >5kg. ACSM measured 47 the emission with ~ 1 min resolution and 1 h ACSM data was averaged to get the 48 49 relative mass contribution and mass spectrum. The NR-PM1 aerosols generated from 50 the combustion of fuels were collected using a sampling line connected to the chimney. The sampling line was made of ordinary $\frac{1}{2}$ inch copper pipe which extended 51 approximately 10 cm inside the chimney flue. An automobile fuel filter was fitted 2 m 52 downstream of the inlet, which was effective in trapping moisture and large 53 particulate matter. This was followed by a gate valve to restrict the flow of smoke and 54 allow dilution with clean air. The gate valve was adjusted to allow a dilution rate in 55 the range of 80-160:1. The total length of copper line between the chimney and the 56 mobile station was around 10 meters, which provided sufficient time for the aerosol to 57 cool down to ambient temperature before ACSM measurement. 58

ACSM spectra analysis were performed using the standard ACSM analysis 59 software (version: ACSM local 1.5.12.0) provided by Aerodyne which is written 60 within Wavemetrics IgorTM. Collection efficiency (CE) in terms of the mass fraction 61 of ammonium nitrate, particle acidity, and water content should be considered to 62 account for sampling losses as suggested by Middlebrook et al (2012). However, in 63 this study, CE-corrected NR-PM1 results in higher mass concentration than 64 simultaneous PM₁₀ measured by TEOM at several evening/night time peaks. Thus, the 65 CE need to be further investigated. Here, we assumed a CE of 1, which provides a 66 lower limit for ACSM-measured mass concentration. A CE of 1 was also used by 67 Canonaco et al. (2013) in which they also found composition dependent CE would 68 underestimate CE resulting in a higher CE-corrected PM₁ than collocated TEOM 69 70 PM₁₀. However, changes in CE won't affect the relative contribution of all species since CE is applied to all measured specie 71



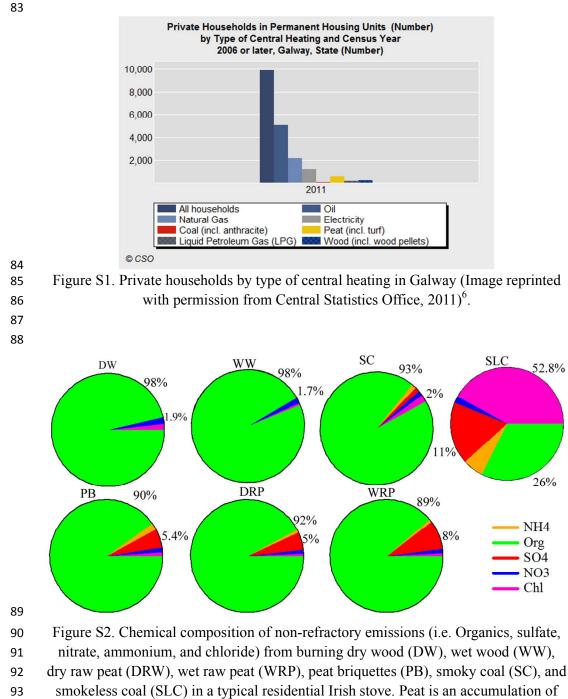
⁷⁷

Table S2. Correlation coefficient (R^2) between ACSM profiles of different sources and

79 PMF factors (dry wood (DW), wet wood (WW), dry raw peat (DP), wet raw peat

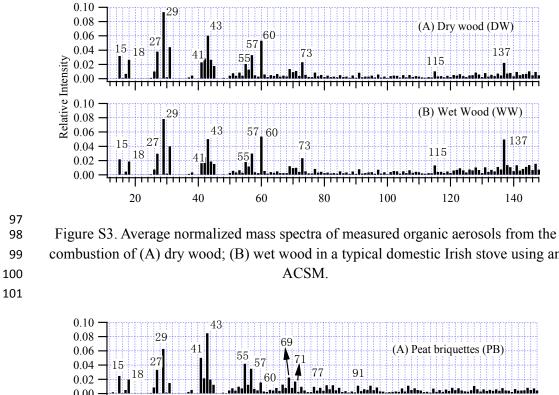
(WP), peat briquettes (PB), bituminous (smoky) coal (SC), and ovoids (smokeless,
 based on anthracite) coal (SLC))

R^2	DW	WW	DP	WP	PB	SC	SLC	HOA^4	BBOA ⁵
DW	1	0.91	0.62	0.62	0.69	0.37	0.69	0.30	0.77
WW	0.91	1	0.47	0.48	0.54	0.32	0.54	0.21	0.57
DP	0.62	0.47	1	0.98	0.96	0.87	0.88	0.8	0.78
WP	0.62	0.48	0.98	1	0.99	0.81	0.89	0.82	0.84
PB	0.69	0.54	0.96	0.99	1	0.77	0.92	0.77	0.88
SC	0.37	0.32	0.87	0.81	0.77	1	0.78	0.76	0.51
SLC	0.69	0.54	0.88	0.89	0.92	0.78	1	0.64	0.83
HOA	0.30	0.21	0.80	0.82	0.77	0.76	0.64	1	0.58
BBOA	0.77	0.57	0.78	0.84	0.88	0.51	0.83	0.58	1

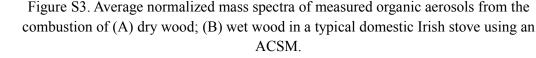


- partially decayed vegetation, it contains more minerals (including sulfur) than fresh
 biomass (e.g. wood).
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S4







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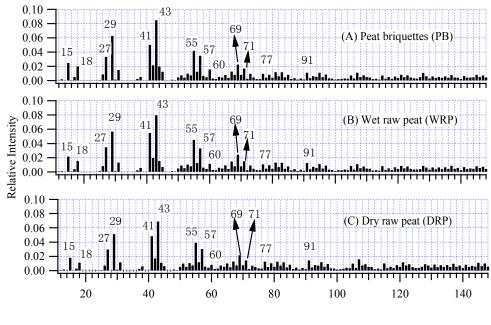
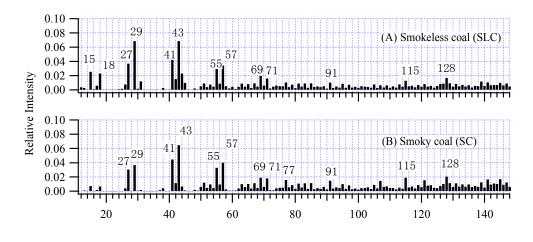


Figure S4. Average normalized mass spectra of measured organic aerosols from the 103 104 combustion of (A) peat briquettes; (B) wet raw pet; (C) dry raw peat in a typical domestic Irish stove using an ACSM. 105

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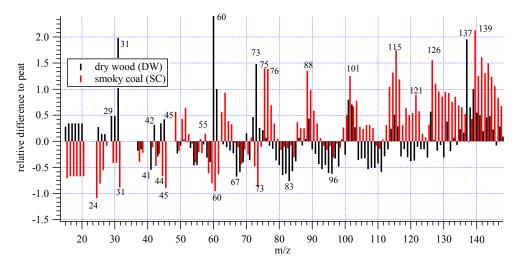




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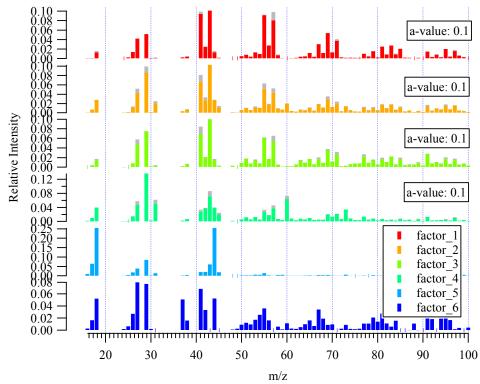
Figure S5. Average normalized mass spectra of measured organic aerosols from the combustion of (A) smokeless coal; (B) smoky coal in a typical domestic Irish stove using an ACSM.



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Figure S6. Relative difference of dry wood and smoky coal MS profile compared to
peat at each m/z.

S6



115 Figure S7. Factor profiles (mass spectra) of the 6-factor solution with four primary 116 factors constrained and two additional left free. Factor 6 is not interpretable by 117 118 comparing with the profiles in AMS database (http://cires1.colorado.edu/jimenez-group/AMSsd/). The α-value method within ME-2 119 was applied. Oil burning (factor 1) profile is from ambient data PMF-derived 120 hydrocarbon-like organic aerosol (HOA) (Crippa et al. 2013)⁴. Peat (factor 2), coal 121 (factor 3), and wood (factor 4) reference profiles are from fingerprinting experiments 122 (Figure 1). Grey bar in the back represents reference profile employed. 123 124

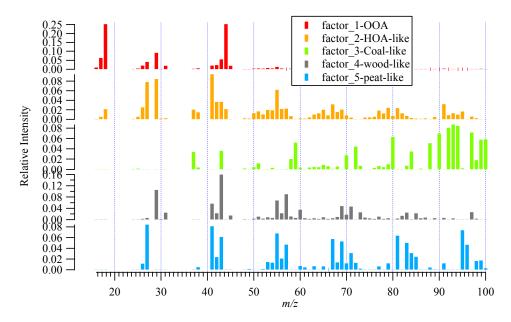
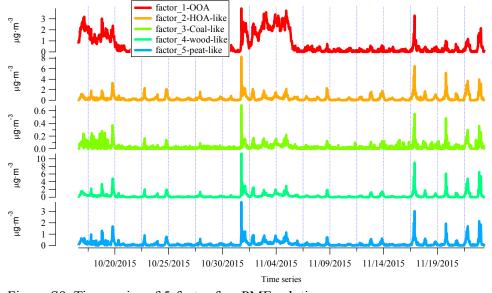
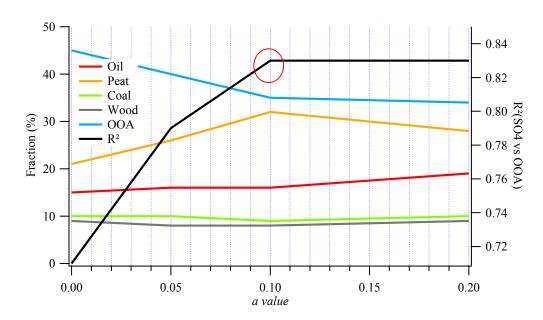




Figure S8. Profiles of 5-factor free PMF solution. Factor 1 is a typical OOA profiles 127 with high m/z 44 signal. Factor 2 shows no signal at m/z 60 and has a higher fraction 128 of signals at lower m/z values, and it is HOA-like. Factor 3 has a higher fraction at 129 higher m/z values, thus it is coal-like. In contrast, both factor 4 and 5 have elevated 130 signals at m/z 60, and the allocation of signals at m/z 29, makes factor 4 wood-like 131 and factor 5 peat like. However, primary factors from free PMF are highly mixed due 132 133 to rotational ambiguity arising from similar emission time. Thus, it is inappropriate to 134 use this solution to estimate the contribution of different sources.



136 Figure S9. Time series of 5-factor free PMF solution.



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Figure S10. The relative contribution of oil, peat, coal, wood, and OOA (left axis) over the whole periods to total OA mass as well as correlation (\mathbb{R}^2) between sulfate and OOA (right axis) as a function of *a* value. An *a* value of 0.1 was selected (red cycle) from which the \mathbb{R}^2 starts to level off.

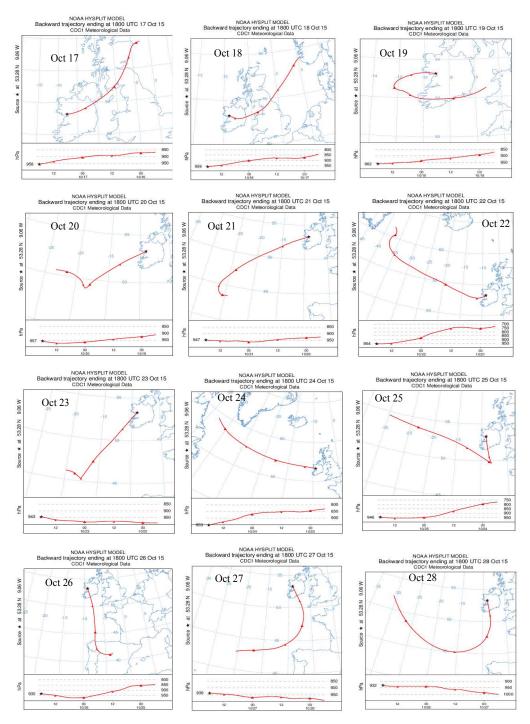
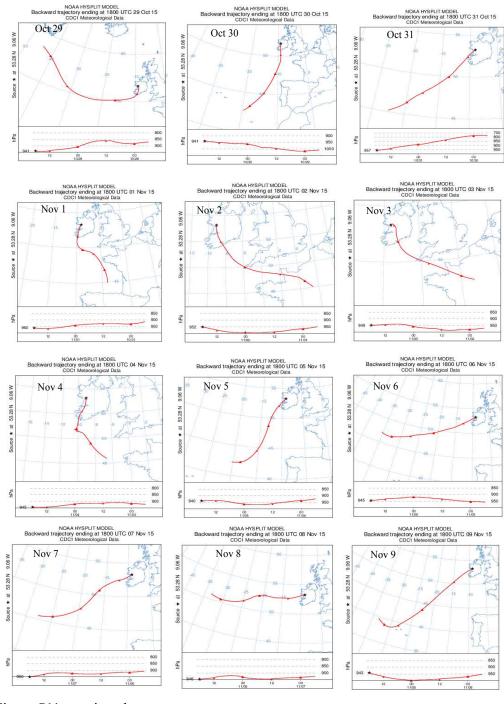
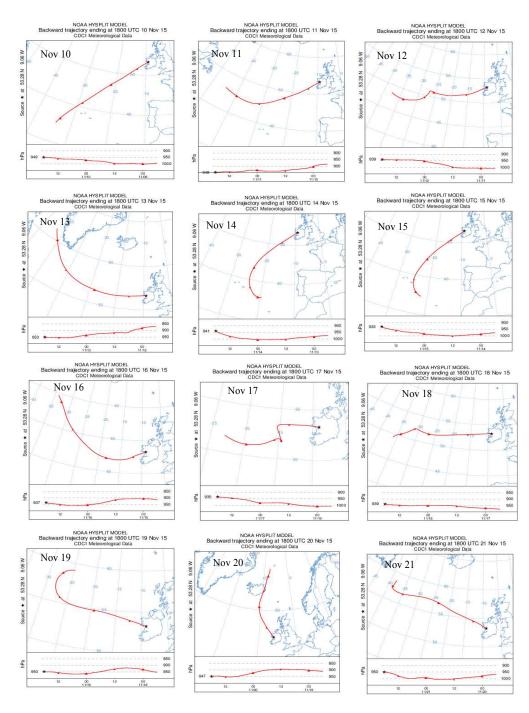




Figure S11. Backward trajectory analysis for 48 h by NOAA Hysplit4 model⁷ ending at 18:00 from October 17 to November 21, 2015 in Galway, Ireland. From Oct 17 to 19 (or S1 in Figure 3) and from Nov 1 to 4 (or S2 in Figure 3), the air masses have a continental origin (from the mainland Europe, the UK, and Ireland itself). From Oct 19 to 31 (or M1 in Figure 3) and Nov 5 to 21 (M2 and M3 in Figure 3), the air masses have a marine origin with short stay in Ireland.









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